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# Enhanced mineral carbonation on surface functionalized MgO as a proxy for mine tailings†

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The escalating demands of industrialization and development underscore the necessity for an efficient and scalable carbon capture and storage (CCS) methodology. Mineral carbonation of MgO presents itself as a promising solution due to its considerable theoretical capacity for CO<sub>2</sub> adsorption. However, the sluggish kinetics of the carbonation process pose a significant challenge. Consequently, a comprehensive understanding of the structural and chemical alterations occurring during carbonation is imperative for material design. In this study, we conduct a thorough structural and chemical investigation of the MgO (sourced from different mine tailings) carbonation process using electron microscopic techniques. Our findings demonstrate that treating MgO with polar solvents enhances its degree of carbonation significantly, offering a promising avenue for improvement. Moreover, we observe a particle size dependency in MgO carbonation and note that the inclusion of additional materials, such as Si-based compounds, further accelerates the carbonation. Density functional theory (DFT) calculations provide insight into surface functionalization as a result of solvent treatment and its mechanistic effect on the origin of the enhanced carbonation of polar solvent-treated MgO, revealing a stronger interaction between CO<sub>2</sub> and the treated MgO (100) surface as compared to the non-polar solvent treated surfaces. These discoveries showcase an alternative approach for enhancing MgO carbonation, thereby offering a potential method for sequestering atmospheric CO<sub>2</sub> more effectively using mine waste rich in MgO.

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## Environmental significance

The study presented in the manuscript describes the structural and chemical changes occurring during direct carbonation of MgO under normal atmosphere which helps in understanding the reaction mechanism and can further pave path for the future development of enhanced Carbon capture and storage (CCS) technologies and contributing to the goal of limiting global warming. The study leverages waste materials (mine tailings) rich in MgO for carbonation, which otherwise would remain unused. This reduces the need for mining new materials, lowering the environmental footprint associated with mining activities. By improving the kinetics of carbonation and utilizing low-cost materials, this approach could be scaled to capture significant amounts of CO<sub>2</sub>, thus contributing to broader carbon capture and storage (CCS) efforts. The product of the carbonation process results in the formation of a stable and non-toxic magnesium carbonate which has potential for industrial application. Further lowers down the need for landfilling of disposal of the mine tailings.

## 1 Introduction

With the rapid development and ever-increasing energy demand in the past decade, the anthropogenic release of CO<sub>2</sub> into the atmosphere has increased its concentration by 20

parts per million (ppm) to reach 421 ppm.<sup>1</sup> This alarming rate of CO<sub>2</sub> release has drawn global attention to the need for effective and scalable carbon capture and storage (CCS) methods.<sup>2–4</sup> Current strategies involve capturing CO<sub>2</sub> emitted from industrial flues (point source capture) before it is released into the atmosphere.<sup>4–9</sup> The most commonly used adsorbents for CO<sub>2</sub> capture from natural and flue gases are liquid amines, such as monoethanolamine, aqueous ammonia, or alkali hydroxide.<sup>10–14</sup> However, liquid sorbents are limited and require high operational energy and cost.<sup>15</sup> In addition to controlling CO<sub>2</sub> emissions, there is also a need to develop methods for directly capturing CO<sub>2</sub> from the atmosphere and storing it.

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In this regard, carbon mineralization has shown significant promise as a chemically and economically effective method of carbon sequestration.<sup>16–18</sup> Carbon mineralization is a natural geological process in which minerals and industrial waste containing alkaline metals (Mg, Ca) interact with CO<sub>2</sub> to form stable carbonates:  $MO + CO_2 \rightarrow MCO_3$ , where M represents the divalent ions Mg<sup>2+</sup> and Ca<sup>2+</sup>.<sup>19,20</sup> Mineral carbonation offers potential advantages over other CCS methods, as the end products of mineralization are solids, environmentally friendly and have industrial applications. Typically, minerals rich in Ca and Mg in the form of oxides or silicates are used for carbon mineralization.<sup>21,22</sup>

Among reactive minerals for the carbonation process, MgO has gained significant attention due to its relatively high theoretical CO<sub>2</sub> capture capacity, abundance, low cost, non-toxicity, and wide operating range.<sup>23,24</sup> Another advantage of carbonation using MgO is its relatively low regeneration temperature compared to other mineral oxides such as CaO, which implicates lower energy requirement during regeneration. The negative  $\Delta G$  for the carbonation of MgO indicates the thermodynamic feasibility of carbonation under ambient conditions. However, the carbonation of MgO is kinetically limited.<sup>25,26</sup> In addition, the low atmospheric CO<sub>2</sub> concentration and the formation of a passivating MgCO<sub>3</sub> layer on the MgO surface result in a slow reaction rate for natural carbonation. The MgCO<sub>3</sub> layer has been proposed to act as a barrier, hindering CO<sub>2</sub> diffusion to the MgO surface for further reaction to occur,<sup>27,28</sup> once surface being carbonated, the lack of an effective primary active site further decrease the carbonation kinetics of MgO. The main challenge associated with carbon mineralization using these minerals is accelerating the kinetics, namely, achieving accelerated carbonation.<sup>29,30</sup> The conversion of MgO to MgCO<sub>3</sub> occurs through a series of chemical reactions, and it has been hypothesized that the thermodynamics and kinetics can be influenced by factors such as hydration (relative humidity (RH)), carbonation duration, and the presence of other chemical components.<sup>24,31</sup> Further several efforts have been made to modify and improve the kinetics of MgO carbonation including varying the particle size, increasing the water content,<sup>16</sup> or surface functionalization by incorporating metal salts such as K<sub>2</sub>CO<sub>3</sub>, KNO<sub>3</sub>, and NaNO<sub>3</sub> or amine.<sup>32–35</sup> Recent work by Bracco *et al.* has demonstrated that increasing relative humidity (RH) promotes the formation of a disordered hydrated reaction front on MgO during carbonation reactions.<sup>36</sup> The morphology of hydration products over MgO have been shown to depend on the rate of dissolution and diffusion. The slow rate of diffusion and dissolution leads to the epitaxial growth of Mg(OH)<sub>2</sub> and faster rates lead to the bulk growth.<sup>37–39</sup> Surface functionalization has also been shown to enhance CO<sub>2</sub> uptake. For instance, Alkadhhem *et al.* reported that amine functionalization of MgO surfaces enables stronger interactions between amine groups and

CO<sub>2</sub>, resulting in improved carbonation adsorption.<sup>35</sup> Similar functionalization strategies have been explored for other materials, such as carbon molecular sieves, with positive outcomes for CO<sub>2</sub> adsorption.<sup>40</sup> Furthermore, combined theoretical and experimental investigations have revealed that organic polar groups can adsorb onto MgO surfaces and undergo partial dissociation,<sup>41–43</sup> contributing to surface functionalization. This phenomenon offers potential for further optimization of carbonation processes. Despite these advancements, a comprehensive nano- and atomic-scale understanding of the structural and chemical transformations occurring during mineral carbonation remains elusive and warrants further investigation.

Industrial mine wastes rich in alkali ions are another inexpensive yet less exploited alternative material for mineral carbonation.<sup>44,45</sup> Despite the existence of a cheap and cost-effective feedstock supply, extensive-scaled application of mine waste for carbonation suffers from low reactivity.<sup>46,47</sup> Therefore, even a slight increase in reactivity of the mine waste could open a new paradigm for industrial-scale mineral carbonation. Thus, it is crucial to understand the mineral carbonation mechanism structurally and chemically under ambient conditions. Fundamental insight into the mechanism may help to accelerate mineral carbonation efforts.

The purpose of this work is to provide a mechanistic understanding of the structural changes that occur during the carbonation of surface functionalized MgO minerals for proxy mine waste using a combination of experimental and computational tools and have been compared with non-functionalized MgO. The use of organic polar molecules as a solvent provides an added advantage to just functionalize the surface without much alteration to the initial structure of MgO due to extensive hydration and provides evidence of surface functionalization effects for the carbonation process. The minerals were obtained directly from the mines and contained various impurities, primarily Si-based, which we found have variable impact on carbonation. While this complicates the analyses presented herein, we focus on generally observed trends.

## 2 Materials and method

### 2.1 Material synthesis

Two MgO samples, hereafter referred to as sample A and sample B, were prepared by calcinating MgCO<sub>3</sub> rich minerals as a proxy for mine tailings obtained from the Calix mine in Australia and the Premier mines of Nevada, USA, respectively. The minerals obtained were calcined at 600 °C for 2 h to regenerate the MgO. TEM studies of carbonation kinetics and mechanism in MgO derived from mine tails were studied by preparing the TEM grid under two different conditions. One set of the TEM grid was prepared by sprinkling the dry powder on a carbon-coated copper grid, hereafter referred to as the “dry sample”, and the other was prepared by dispersing the MgO samples in methanol or isopropyl



alcohol (IPA) using sonication, followed by drop-casting a drop of solution on TEM grids and drying, hereafter referred to as the “treated sample”. TEM grids were kept in an open-air container to undergo the natural carbonation process under ambient humidity and air, and were examined at different intervals under the electron microscope.

## 2.2 Characterization

The phase and crystallinity of the initial samples were characterized using X-ray diffraction (XRD) Rigaku MiniFlex 6G theta-2theta vertical goniometer benchtop powder diffraction system. The morphology, microstructure, and structural evolution during the carbonation process were investigated using transmission electron microscopy (TEM). TEM imaging was performed using a JEOL F200, operating at 200 kV. Scanning transmission electron microscopy (STEM), energy dispersive spectroscopy (EDS), and electron energy loss spectroscopy (EELS) were performed using JEOL NEOARM, operating at 200 kV with a convergence angle of 27 mrad. STEM images were acquired with a 4 cm camera length and a probe current of 160 pA. EDS and EELS were acquired in dual spectrum imaging mode (SI) with a probe current of 160 pA to acquire sufficient signal without damaging the samples. A camera length of 2 cm and 0.1 eV per channel or 0.25 eV per channel dispersion was used to obtain EELS spectra with an enhanced signal-to-noise ratio. EELS spectra were acquired with a dual GIF camera and a K2-IS camera operating in the summit mode, provided by GATAN Inc., using a 5 mm aperture fitted at the Gatan imaging filter (GIF) entrance.

## 2.3 Computational methods

All modeled systems were setup and managed using the atomic simulation environment (ASE).<sup>48</sup> First-principles non-spin polarized density functional theory (DFT) calculations were performed with the Quantum espresso software package,<sup>49,50</sup> using the Bayesian error estimation exchange correlation functional with van der Waals correlation (BEEF-vdW).<sup>51</sup> A plane-wave basis set was used to expand the Kohn–Sham wave functions, with 500 and 5000 eV kinetic energy and density cutoff, respectively. Ultrasoft pseudopotentials<sup>52</sup> were used to approximate core-electrons. The bulk lattice constant of the cubic MgO crystal<sup>53</sup> was found to be 4.282 Å, which is close to the experimentally reported values of 4.21–4.24 Å.<sup>54,55</sup> A  $4 \times 4 \times 1$  ( $4 \times 4 \times 4$ ) Monkhorst–Pack  $k$ -point grid<sup>54</sup> was used for all surface (bulk) calculations. The (100) and (110) surfaces of MgO were modeled using a  $2 \times 2$  repeated cell in the  $x$ - and  $y$ -directions, a 4-layer slab, and a 10 Å vacuum spacing between periodic images in the  $z$ -direction. The atomic positions of the two lower slab layers were constrained to their bulk lattice coordinates, allowing only the atoms in the top two slab layers to relax. The individual methanol, isopropanol (IPA), hexane, water, CO<sub>2</sub>, and H<sub>2</sub> molecules were modeled in a supercell

with 10 Å vacuum in all directions since these energies are needed as reference energies to obtain the thermodynamics (see details below).

Herein, we will computationally model the reactivity and carbonation of the (100) and (110) MgO surface facets which are the two predominant surface facets observed experimentally using XRD (see Fig. S4 in ESI†). Hence, we assume that the other less pronounced surface facets, *e.g.* the (111) and (311) surfaces, will have minimal impact on the overall carbonation performance. The treatment of the material with the polar and non-polar solvents was hypothesized to modify the termination of the (100) and (110) surface facets through adsorption of the intact or partially dissociated solvent molecule, which could result in different smaller species adsorbed on the surface. The different species are modeled to adsorb on the bulk-cleaved surfaces and the corresponding Gibbs free energy of adsorption is defined in eqn (2). This assumes that the surfaces are in fast equilibrium with both carbon and oxygen elements, hence the change in solvent chemical potential only affects the chemical potential of hydrogen. The considered dissociated species include individual sub-components of the solvent molecules and combinations of multiple sub-components. Specifically, this analysis included the following single species H, O, OH, CH<sub>x</sub>, CH<sub>x</sub>O, and H<sub>2</sub>O as well as mixed combinations including H + OH, H + CH<sub>x</sub>O, OH + CH<sub>x</sub> at coverage 0.125 ML and 1 ML (see legend of Fig. 8). Here,  $x$  is either 2 or 3, depending on the nature of the solvent reservoir. The monolayer coverage is defined by an (or a pair of) adsorbate per MgO unit. We further assume that the entropic contributions of the crystal structures and the volume changes are negligible. The Gibbs free energy of adsorption is given by

$$\Delta G_{\text{ads}} = \Delta F_{\text{ads}} + P\Delta V - n_{\text{solvent}}\mu_{\text{solvent}} \approx (E_{\text{sys}} - E_{\text{slab}} - G_{\text{solvent}}) - n_{\text{solvent}}\mu_{\text{solvent}} \quad (1)$$

$$= E_{\text{sys}} - E_{\text{slab}} - \sum_{i \neq H} n_i \mu_i - n_H \left( g_{\text{solvent}} + \mu_{\text{solvent}} - \sum_{i \neq H} \tilde{n}_i \mu_i \right), \quad (2)$$

where  $E_{\text{slab}}$  and  $E_{\text{sys}}$  are the DFT calculated energies of the clean and functionalized MgO surfaces, respectively.  $n_i$  and  $\tilde{n}_i$  are the number of individual chemical elements in the adsorbate and in the solvent molecule, respectively. The reference energy of each element  $\mu_i$  is determined from the total electronic energy of a stable gas phase molecule or a reference gas phase reaction.<sup>56</sup> The chemical potentials of the solvent molecules are calculated from their ideal gas Gibbs free energies:

$$\mu_{\text{solvent}}(T, P) = \left( \frac{\partial G_{\text{solvent}}}{\partial n_{\text{solvent}}} \right)_{T, P} \quad (3)$$

$$G_{\text{solvent}}(T, P) = H_{\text{solvent}}(T) - TS_{\text{solvent}}(T, P), \quad (4)$$



where  $H_{\text{solv}}(T)$  and  $S_{\text{solv}}(T, P)$  are the ideal gas enthalpy and entropy, respectively. Calculation details for these thermodynamic terms are given in ESI.†

After the treatment of the material with the solvents, which could result in a functionalized material, the remaining solvent is retrieved from the system and the solid-liquid equilibrium transforms into a solid-air equilibrium. For the new solid-air equilibrium, the dissociatively adsorbed species no longer obey boundary condition with the treatment solvent reservoir, leading to a possibility for these adsorbed species to desorb from the surface. To assess the stability of the adsorbed dissociates species at the new equilibrium, we perform a desorption energy analysis. The desorption energy of a single/pair of the adsorbate species for a given coverage is defined by:

$$\Delta E_{\text{des}} = E_{n-1} - E_n + \sum E_i, \quad (5)$$

where  $E_n$  is the energy of the slab covered with  $n$  (or  $n$  pairs of) species, and  $E_{n-1}$  is the energy of the slab with one less (or one less pair of) species. In this study, we focus solely on presenting an analysis of a single/individual (or a single pair of) adsorbate coverage and a full monolayer (1ML) coverage. Investigating intermediate coverage levels introduces the challenge of sampling short-range interactions among the adsorbate pairs, significantly increasing the computational costs.

Finally, to investigate the carbonation thermodynamics of the solvent-modified (100) and (110) MgO surfaces, we

performed a comprehensive computational screening of CO<sub>2</sub> adsorption on multiple possible adsorption sites and initial configurations of the molecule.<sup>57</sup> These results are compared with the bare untreated surfaces, that is, clean bulk-cleaved surfaces. One specific adsorption map for the bare MgO (100) surface is presented in Fig. 1. The adsorption maps for other differently faceted and terminated surfaces can be found in Fig. S12–S18 in the ESI.† The CO<sub>2</sub> adsorption energy is defined by:

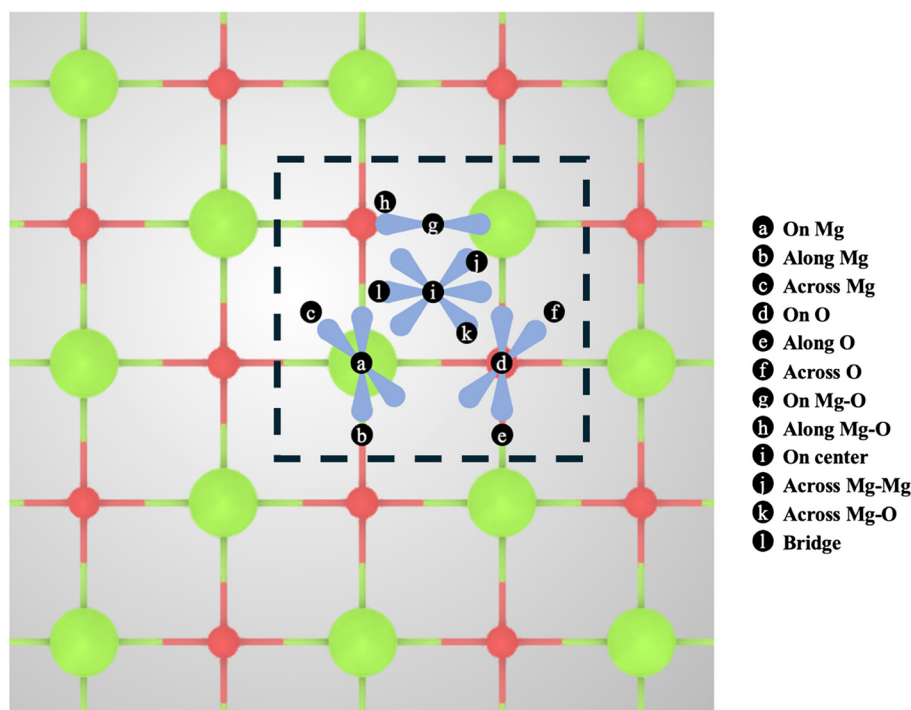
$$\Delta E_{\text{ads}} = E_{\text{CO}_2+\text{slab}} - E_{\text{slab}} - E_{\text{CO}_2}, \quad (6)$$

where  $E_{\text{CO}_2+\text{slab}}$  is the total energy of the system,  $E_{\text{slab}}$  is the energy of the relaxed slab without CO<sub>2</sub>, and  $E_{\text{CO}_2}$  is the energy of a CO<sub>2</sub> gas molecule.

### 3 Results

#### 3.1 Carbonation of MgO nanoparticleless in dry as-produced conditions

Morphological analyses of the dry samples in their original condition (day-0) were imaged (Fig. 2 and S1†), revealing clusters of crystalline MgO nanoparticles. The high-resolution image of MgO shows the predominant (100) surface of MgO. Periodic microscopic observations were conducted to track the carbonation process of the dry samples. After 120 days, no structural changes associated with carbonation were observed in the dry sample A and aggregates similar to those observed on day-0 were still present (Fig. 2(c and d)). Electron energy



**Fig. 1** Adsorption site and orientation map for considered starting configurations for CO<sub>2</sub> adsorption on the clean bulk-cleaved MgO (100) surface (top-down view). Mg and oxygen atoms are represented in green and red, respectively.



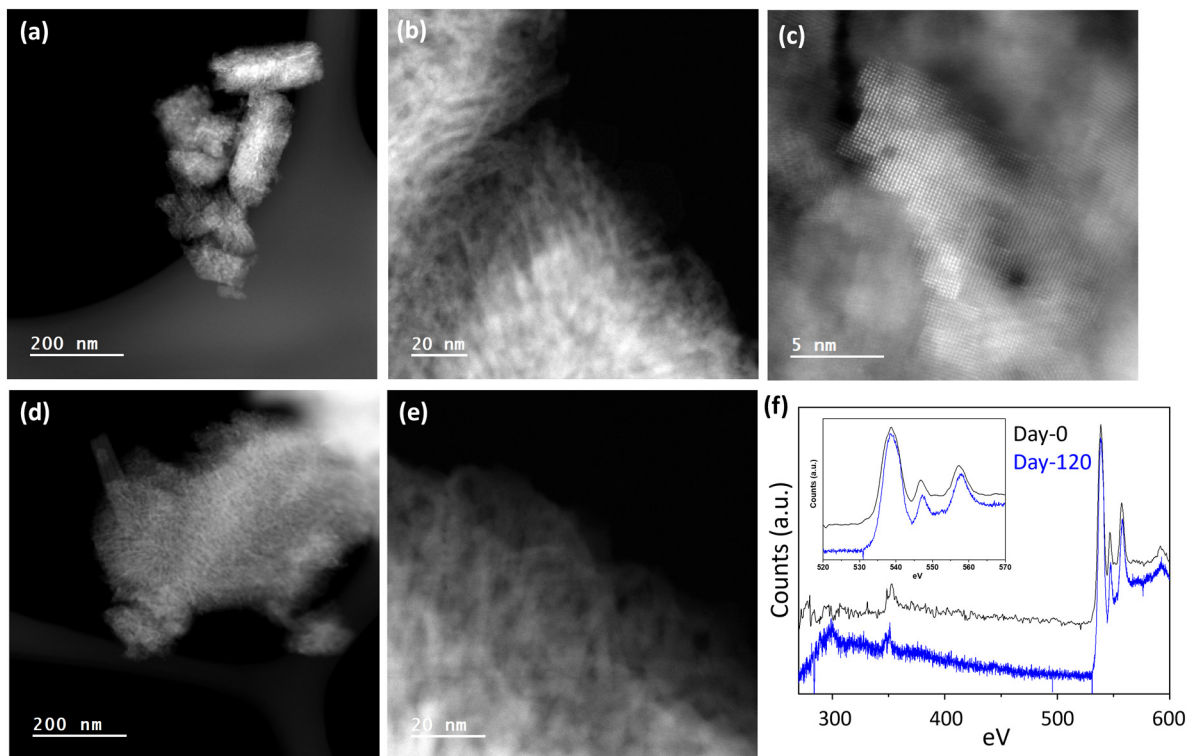


Fig. 2 High-angle annular dark-field scanning transmission electron micrographs of dry-sprinkled MgO sample A (a–c) day-0, (d and e) day 120, and (f) corresponding EELS spectra (inset: magnified O K edge).

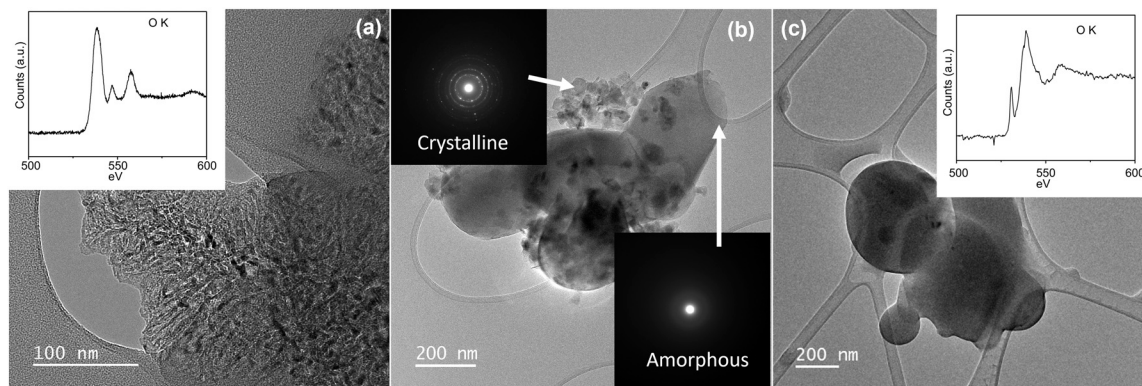
loss spectroscopy of the dry sample A confirmed the initial findings from the STEM images. Initially, the freshly dispersed samples exhibited strong O K edge peaks at 538.8 eV, 547 eV, and 557.5 eV, which are characteristic of MgO, with a minor impurity peak of Ca, and have very small quantity of adventitious carbon, indicating the samples were clean. After 120 days, sample A displayed no structural changes and the strong crystalline peaks of the O K edge for MgO remained unchanged. Adventitious carbon was found on the sample surface on day-120, as evidenced by the presence of a broad amorphous C K edge, contrary to what was observed for the as-prepared condition. A similar carbonation behavior was also observed for dry sample B (Fig. S1†). However, in contrast to the sample A, the dry sample B exhibited evidence of water adsorption after 6 months, as identified through EELS acquisition, showing an additional pre-peak around  $\sim 531.6$  eV near the O K edge. The pre-peak before the main O K edge can be corroborated to the presence of oxygen molecules from the decomposition of  $\text{H}_2\text{O}$  or  $(-\text{OH})$  groups adsorbed on the MgO surface.<sup>58,59</sup> These observations indicate that untreated MgO minerals exhibit slow carbonation kinetics.

### 3.2 Carbonation of MgO nanoparticles dispersed and sonicated in polar solvents: methanol and IPA

It has been suggested that carbonation kinetics can be enhanced by adding surface functionalization, moisture/

humidity, seed nuclei or the presence of secondary minerals.<sup>60</sup> To understand the impact of surface functionalization on carbonation, we prepared MgO nanoparticles functionalized with different functional groups by treating them with different solvents. The crystalline MgO sample A and sample B were dispersed in a polar solvent (methanol or IPA), drop-casted onto a TEM grid, dried and stored in an open container similar to the dry-sprinkled MgO. The carbonation of MgO dispersed in polar solvent and sonicated proceeds sporadically for well-dispersed MgO in contact with ambient air as it converts into amorphous  $\text{MgCO}_3$  (Fig. 3). The amorphous carbonate that is formed has a very smooth and round surface, consistent with previously observed amorphous spherules formed from CaO carbonation.<sup>61</sup> The area occupied by the newly formed  $\text{MgCO}_3$  is approximately the same as the area initially occupied by MgO before undergoing carbonation, which indicates slight or no area expansion during carbonation (Fig. S2†). In addition, the general shape of the original MgO aggregates is maintained throughout the aggregate's carbonation and after an aggregate has fully carbonated with the collapse of the individual MgO nanoparticles to one large grouping of carbonates. The amorphous carbonate features initiate and grow from the surface of the crystalline MgO aggregates. As the carbonate grows, the MgO gets depleted, eventually leading to the complete conversion of the original MgO aggregates. The EELS signature of the O K edge of the aggregates can also be used to characterize the formation of





**Fig. 3** Representative images of MgO carbonation. (a) MgO was prepared by calcinating mined  $\text{MgCO}_3$  and freshly dispersed on a TEM grid (day-0), inset: O K EELS spectra. (b) MgO aggregate dispersed on a TEM grid undergoing carbonation, inset: selected area electron diffraction (SAED) pattern from the carbonated and uncarbonated region. (c) Fully carbonated MgO aggregate ( $\text{MgCO}_3$ ) dispersed on a TEM grid, inset: O K EELS spectra of fully carbonated aggregates.

carbonate phase. The EELS signature for the O K feature of carbonate is different from the one for MgO, as shown in the insets of Fig. 3.

The formed carbonate phase are electron beam sensitive. If amorphous carbonate phase is exposed to the electron beam in vacuum conditions, it rapidly bubbles off the  $\text{CO}_2$  and de-carbonates, regaining a nanocrystalline MgO structure (Fig. S3†). Although the beam induced MgO has similar crystal structure to the pre-carbonated MgO, its morphology differs, and the newly formed MgO has small crystallite sizes ( $<10$  nm). This process is fast and complete conversion is observed in  $<10$  s unless the electron beam dose and dose rate are very low.

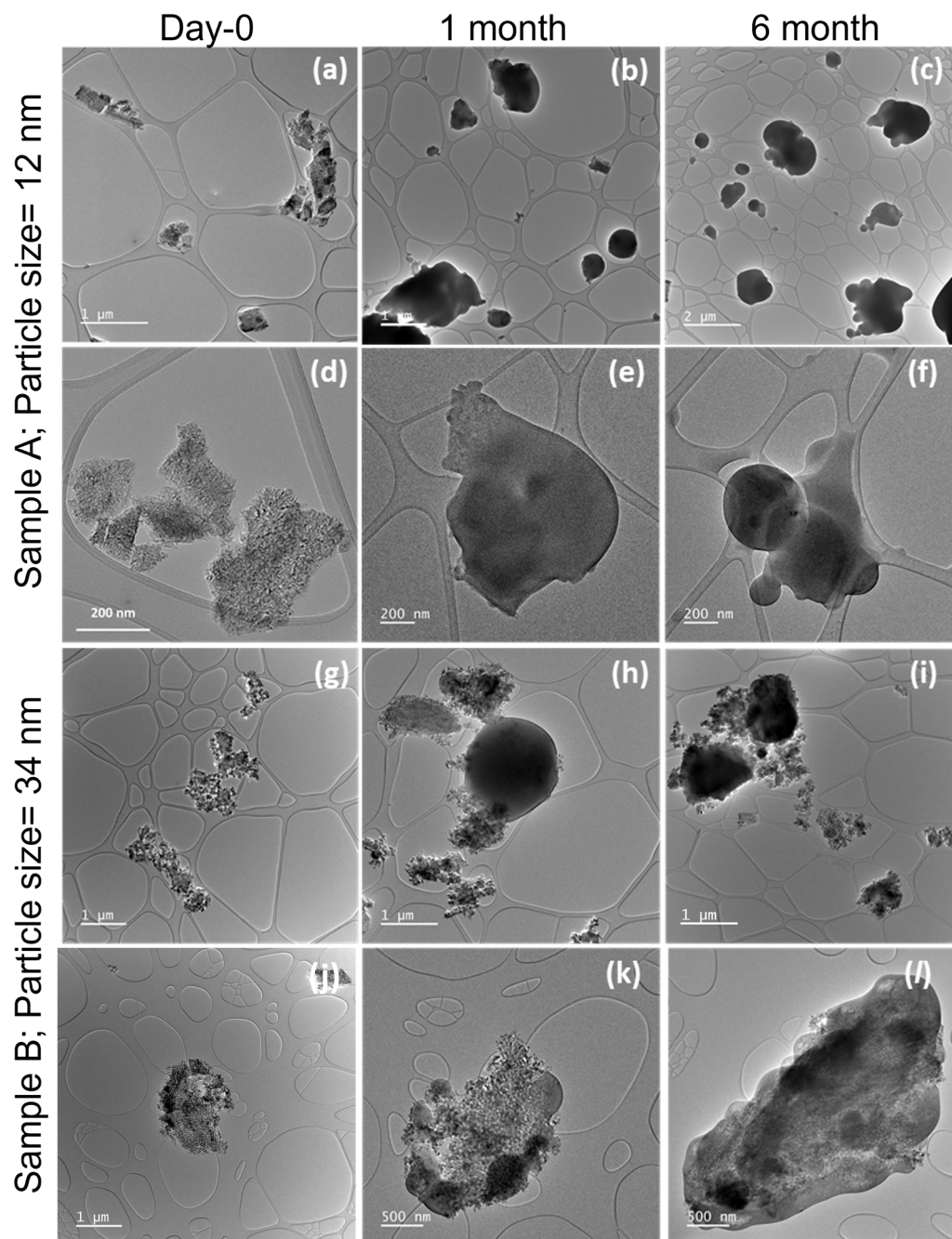
A qualitative difference in carbonation was observed for the polar solvent treated sample A and sample B. The carbonation process is faster for the treated sample A (Fig. 4). The vast majority of the treated sample A was fully converted to the amorphous carbonate even after only one month, while in the case of treated sample B, most aggregates were not fully converted during the same carbonation duration. To dive deeper into difference in the degree of carbonation of treated sample A and sample B, XRD analysis was performed (Fig. S4†). XRD analysis showed that both samples are structurally periclase and exhibit similar levels of preferential orientation, favouring the (100) facets, consistent with high-resolution TEM imaging (Fig. S5†). This is one of the primary facets that was used in our computational models. The broader XRD peaks for sample A as compared to sample B depict the particle size difference between sample A and sample B. Using the Scherrer equation  $\tau = K\lambda/(\cos \theta \beta)$  on the intense 200 reflection of the XRD pattern, an average particle size was calculated, where  $\tau$  is the crystallite size,  $K$  is the shape factor  $\beta$  is the full width at half maxima of the peak and  $\theta$  is the Bragg angle. An average particle size of 12 nm was found for sample A, compared to 34 nm for sample B. The estimated particle size is consistent with the observations from TEM analysis of freshly dispersed sample A and sample B (Fig. S6†). In summary, sample A, with

smaller particles of the two samples, shows more carbonation when treated with the polar solvent. While both MgO samples carbonated sporadically and quickly after dispersed in methanol or IPA (Fig. 3 and 4) and drop-casted onto a TEM grid, neither showed carbonation when left as untreated powder form in contact with air for an extended period (Fig. 2 and S1†).

The carbonation of MgO aggregates does not happen at the same rate for all MgO aggregates. Likewise, it does not occur homogeneously over each aggregate. While certain aggregates show complete conversion after two months some remain completely uncarbonated (Fig. 5(a and b)). The aggregates that remain as MgO and do not carbonate are small. For example, the average aggregate size of uncarbonated MgO for freshly dispersed is  $<1$   $\mu\text{m}$  ( $n = 30$ ). After two months, virtually all large aggregates have carbonated to  $\text{MgCO}_3$  (Fig. 4 and 5(a–f)). The average aggregate size of the remaining uncarbonated material is only 100 nm ( $n = 30$ ), showing that only very small aggregates are left uncarbonated (Fig. 4 and 5(a)). In contrast, the large aggregates exhibit extensive carbonation in all cases. Furthermore, the uncarbonated aggregates typically exhibit significant evidence of adventitious carbon at their surface, which is in contrast to the large MgO aggregates that have undergone carbonation and seldomly have surface adventitious carbon (Fig. 5(c–f)). These findings indicate that larger MgO aggregate sizes are more prone to convert to  $\text{MgCO}_3$  at the tested conditions.

The carbonation of MgO can also be influenced by the presence of certain impurities which could act as hydrating agents by adsorbing the moisture and maintaining the hydration of the MgO. The mine-tailed samples have  $\text{SiO}_2$ -based impurities, which when dispersed in solvent, surround the MgO in various regions and can facilitate the carbonation of MgO in those aggregates. Fig. S7† shows the STEM-EDS maps of the solution dispersed sample A showing the presence of  $\text{SiO}_2$ -based impurities layering over the MgO aggregates which were aggregated in dry powder. The



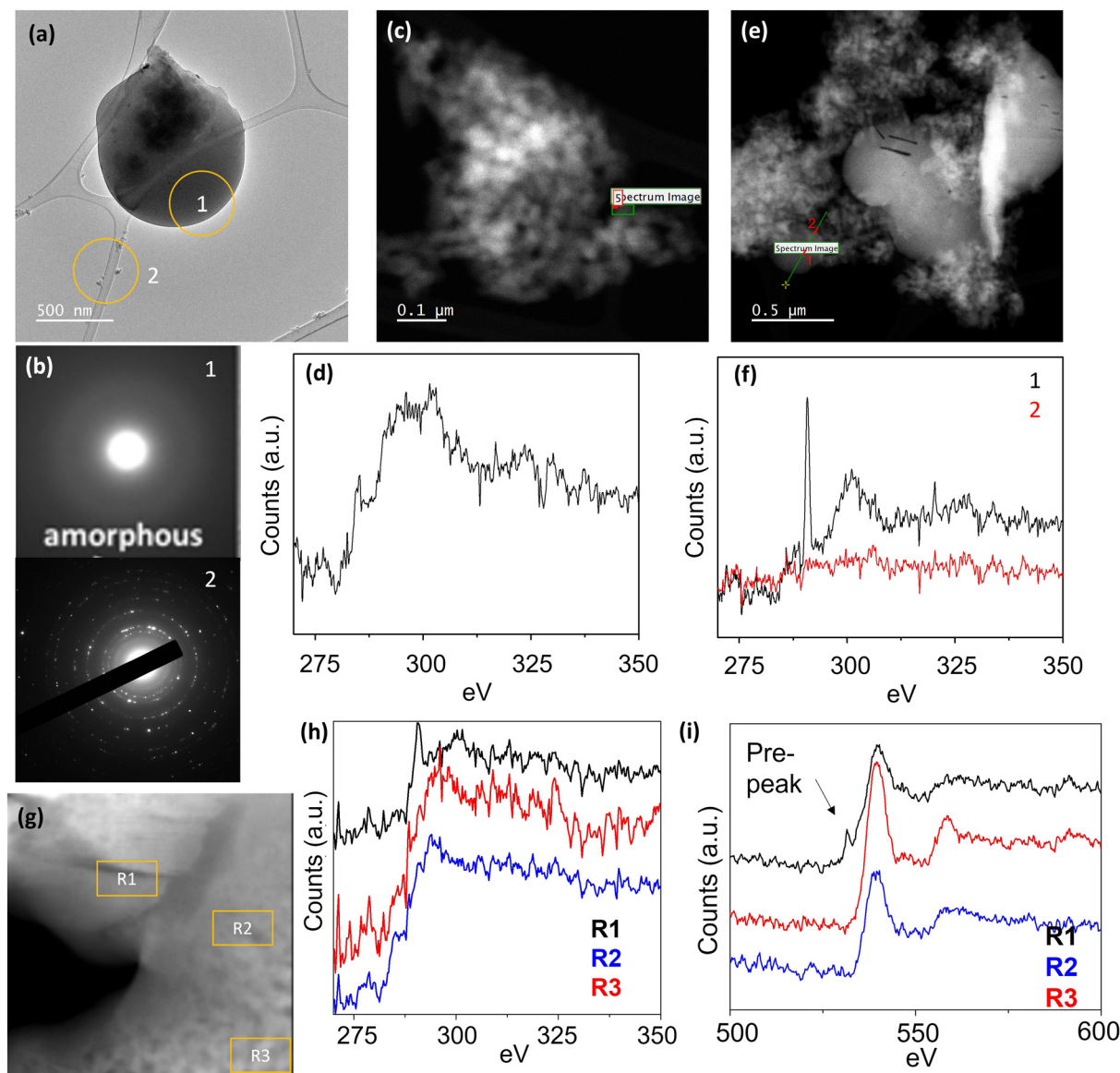


**Fig. 4** Time-evolution of two MgO samples with different average particle size. (a and d) Freshly dispersed sample A; (b and e) sample A after 1 month of dispersion on the TEM grid; (c and f) sample A after 6 months of dispersion on TEM grid; (g and h) freshly dispersed sample B; (i and j) sample B after 1 month of dispersion on TEM grid; (k and l) sample B after 6 months of dispersion on the TEM grid.

aggregates that carbonated more quickly have a significant fraction of  $\text{SiO}_2$ -based impurities in the nearby area and have attracted moisture (Fig. 5(g-i)). The presence of moisture in the carbonating MgO aggregates can be identified from the EELS spectra of carbonated and uncarbonated regions (Fig. 5(g-i)). The carbonated area shows a pre-edge peak in O K edge, corresponding to the gaseous release of  $\text{O}_2$  due to presence of either  $\text{H}_2\text{O}$  or  $(-\text{OH})$ .<sup>58,59</sup> At the same time, the uncarbonated MgO regions have less Si in those aggregates (Fig. S7†). The observed

$\text{SiO}_2$ -modified carbonation indicates that the hygroscopic nature of  $\text{SiO}_2$  can also further affect the carbonation of the MgO aggregates by maintaining enough hydration around the aggregates. MgO is shown to form metal-silicate hydrate upon reaction with  $\text{SiO}_2$ -based mineral in presence of water further reinforcing our observation.<sup>62,63</sup> However, the exact quantification ratio of MgO to  $\text{SiO}_2$  needed to maximize the carbonation was not achieved as the amount of  $\text{SiO}_2$  kept changing due to the surface diffusion of  $\text{SiO}_2$  over the aggregates. The hygroscopic nature of  $\text{SiO}_2$  can be further





**Fig. 5** (a) TEM image of amorphous carbonate and smaller uncarbonated aggregates marked in circles as region 1 and 2, respectively; (b) SAED pattern from the marked region 1 and 2 of (a); (c and d) STEM image of uncarbonated aggregates and corresponding C K EELS spectra showing the adventitious C features (e and f) STEM image of carbonated aggregates and corresponding C K EELS spectra from the marked region as 1 and 2; (g–i) STEM image, C K and O K EELS spectra from carbonated and uncarbonated regions for an MgO aggregates.

supported by observing formation of amorphous regions between the aggregates after an extended period (Fig. S8†). A layering of SiO<sub>2</sub> was observed between the MgO aggregates, which was absent in the freshly prepared samples.

The impact of sonication is indirectly related to the interaction of solvent with the MgO nanoparticles. A drop of IPA was added to the dry-sprinkled MgO nanoparticles and examined over an extended period. The carbonation kinetics were found to be slower as compared to the sonication-induced dispersion. Very few regions of MgO show carbonation, presumably due to insufficient solvent interaction with MgO and hence small quantities of functionalization which could influence the reaction. Most

areas showed uncarbonated MgO with adventitious carbon (Fig. S9†).

The influence of solvent polarity was investigated by dispersing MgO nanoparticles in the non-polar solvent hexane. The hexane-dispersed MgO nanoparticle aggregates showed no structural changes and retained their high crystallinity following dispersion, as confirmed by electron energy loss spectroscopy (EELS) analysis of the O K-edge (Fig. S10(a, b and e)†). No adventitious carbon contamination was detected initially. However, upon exposure to air for approximately one month, the MgO aggregates exhibited significant adsorption of adventitious carbon, as evidenced by the C K-edge EELS spectra (Fig. S10(e)†). Prolonged air exposure of hexane-treated samples led to partial



carbonation, although some uncarbonated MgO remained even after two years of atmospheric contact (Fig. S10(f-h)†). The non-polar nature of hexane, lacking functional groups conducive to surface interactions, does not provide any additional benefit for carbonation. On the contrary, it may hinder the carbonation process by facilitating the adsorption of adventitious carbon, leading to surface passivation. Interestingly, unlike samples treated with polar solvents, hexane-treated MgO surfaces were devoid of silica (SiO<sub>2</sub>) deposition. In contrast, polar solvent-treated samples typically exhibited diffuse SiO<sub>2</sub> based layers surrounding the MgO aggregates, as shown in Fig. S11.† This difference underscores the role of solvent polarity in surface chemistry and structural interactions.

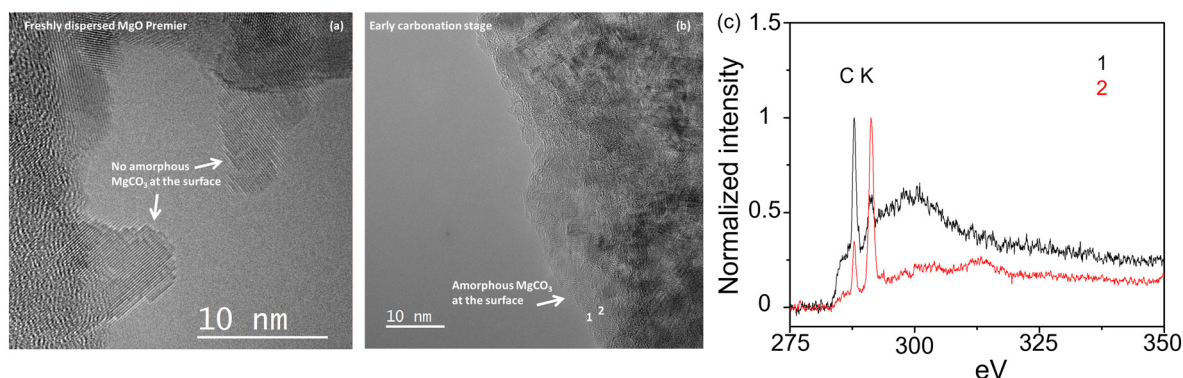
### 3.3 Understanding MgO carbonation at the nanoscale level

To gain a deeper understanding of the MgO carbonation process, we employed electron energy loss spectroscopy (EELS) to analyze the evolution of MgO aggregates into MgCO<sub>3</sub> (Fig. 6). The EELS images revealed a clean surface of MgO with well-defined crystal edges (Fig. 6(a)). To facilitate carbonation, we dispersed the MgO on a TEM grid and introduced air by deliberately exhaling humidified CO<sub>2</sub> onto the grid. High concentration of CO<sub>2</sub> with humidity have been shown to enhanced the carbonation process<sup>64,65</sup> and, in less than 1 hour, an amorphous carbonated layer was observed at the surface of freshly dispersed sample A aggregates (Fig. 6(b)). The freshly dispersed MgO aggregates undergoing carbonation initially did not contain any carbon on their surface. However, during the carbonation process, a clear EELS carbonate signature emerged in the amorphous carbonated materials on the surface of the MgO aggregates, indicating the uptake of CO<sub>2</sub> during the carbonation process (Fig. 6(c)). The carbon K edge spectra further revealed an additional peak at approximately ~287 eV, in addition to the conventional carbonate carbon K edge. This additional peak can be attributed to C–OH groups, implying the formation of bicarbonate species along with the carbonation process.

In Fig. 7, we observe the aggregates undergoing the natural carbonation process, and the point EELS spectra of the carbon K edge display a systematic variation between carbonate and bicarbonate, indicating the possible formation of carbonates *via* a transition zone of bicarbonates. The fully carbonated region of the aggregates shows characteristic peak features of carbonates. However, as we move towards the uncarbonated MgO there is an increase in the intensity of the additional peak corresponding to bicarbonates, suggesting a concentration gradient of these different species within the aggregates. Continuing inward, the carbon peak entirely vanishes in the MgO region of the aggregates, as MgO does not inherently contain carbon before the carbonation process starts. This finding supports the notion that the carbonation process progresses *via* a carbonation front, which starts at the surface of the MgO aggregate and propagates inward into the bulk of the aggregate. Overall, the EELS analysis provides critical insights into the carbonation mechanism of MgO, demonstrating the simultaneous formation of bicarbonate species and the subsequent conversion to carbonates. The identification of a carbonation front, that starts at the surface of the MgO nanoparticle aggregate and propagates inward, further contributes to our understanding of the spatial dynamics of the process.

### 3.4 Solvent-modified MgO surface functionalization

To understand the possibility of surface functionalization depending on the solvent and its effect on the experimentally enhanced carbonation process, we performed DFT calculations. We modeled the stability of functionalized MgO (100) and (110) surfaces as a result of MgO being treated with and in contact with a polar (methanol, IPA) and non-polar (hexane) solvent. Fig. 8(a–c) shows the thermodynamically favored surface termination and coverage of different solvent derived species, which taken together define the functionalization, when MgO (100) is exposed to the solvent as a function of the chemical potential for each solvent. The changes in Gibbs free energies for a full monolayer coverage



**Fig. 6** Carbonation of MgO upon being exposed to exhaled air. (a) Freshly dispersed MgO Calix with clean surface. (b) Freshly dispersed sample A (MgO Calix) upon being exposed to human exhaled air. (c) EELS C-edge for region t 1 (fully carbonated) and region 2 (front of carbonation) in panel (b).



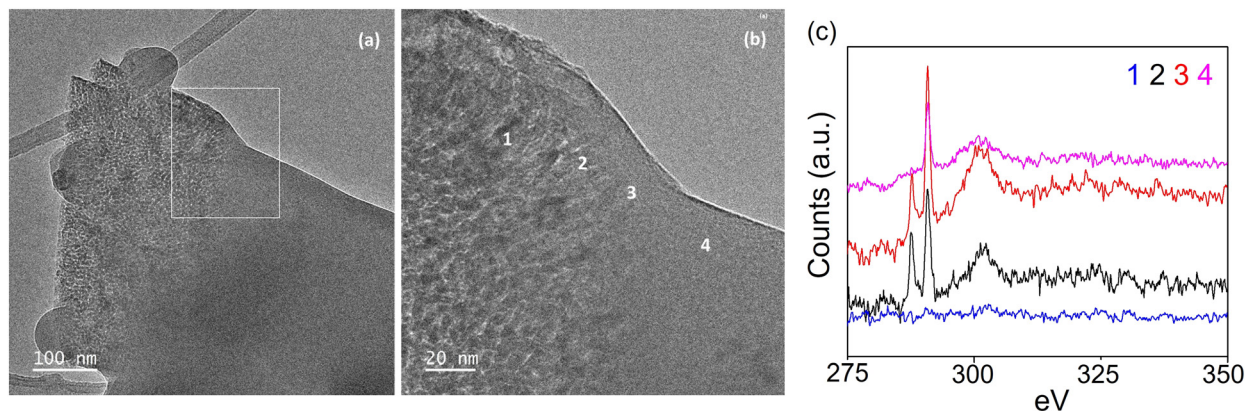


Fig. 7 Carbonation front of an MgO sample. (a) Low-magnification carbonation front of a Calix MgO sample. (b) High-magnification carbonation front of a Calix MgO sample. (c) EELS spectra from the MgO through the carbonation front to the  $\text{MgCO}_3$  (1–4).

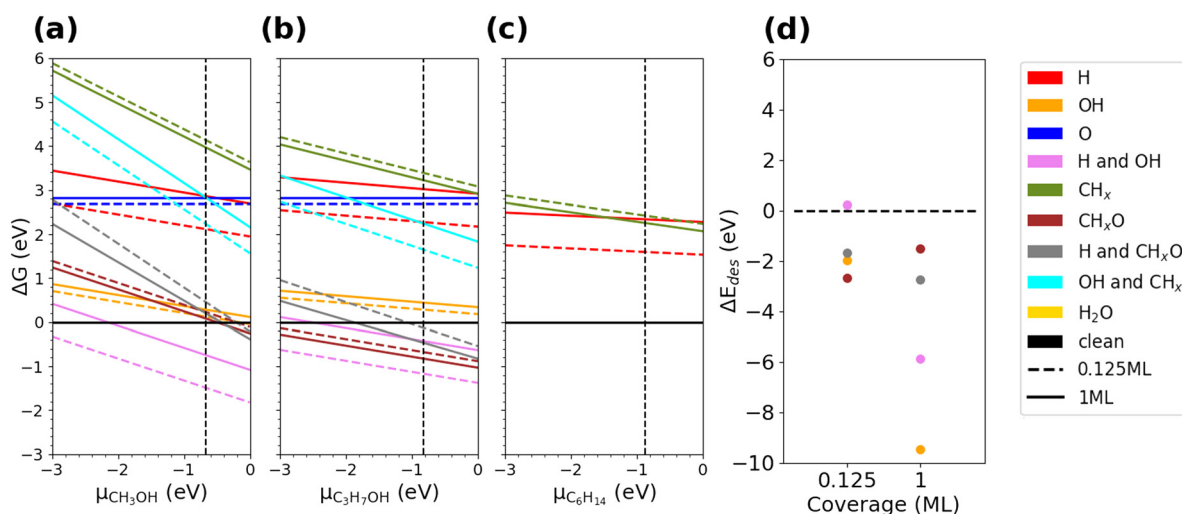


Fig. 8 Stability phase diagrams for an MgO (100) surface exposed to different solvents (a) methanol, (b) isopropanol, and (c) hexane, resulting in adsorption of different solvent-derived species as indicated by the legend. (d) Desorption energetics of the methanol/IPA-treated MgO (100) surface in equilibrium with vacuum.

(solid lines in Fig. 8) were normalized by the number of adsorbate pairs so that the average energy of adsorbing each pair was obtained. The chemical potentials corresponding to the saturated vapor are shown by the dashed black line. Near the saturation conditions, the termination components vary in response to different treatment solvents. For a methanol-treated MgO (100) surface, a low coverage of  $-\text{H}$  and  $-\text{OH}$  pairings is the most stable dominating functionalization. For an IPA-treated MgO (100) surface, the termination components are diverse, consisting of a mixture of  $-\text{H}$ ,  $-\text{OH}$ , and  $-\text{CH}_x\text{O}$  species. On the other hand, the hexane-treated MgO (100) surface prefers to be un-functionalized as none of the solvent derived species stabilize it.

Fig. 8(d) shows the desorption thermodynamics of the adsorbed solvent-derived species from the MgO (100) surface once the remainder of the solvent is removed post treatment. The MgO (100) surface readily loses the adsorbed species when they are not in contact with the solvent reservoir (only exception is in the scenario of sparsely distributed (0.125 ML)

$\text{H}-\text{OH}$  pairings ( $\Delta E_{\text{des}}$  of +0.231 eV)). It is worth noting that 0.125 ML  $-\text{OH}$  species (without H) alone are unstable (−1.983 eV), highlighting the pivotal role of H species as stabilizers. From Fig. 8(a and b), we conclude that the methanol-treated MgO (100) surface maintains its predominant termination when exposed to the air, while the IPA-treated MgO (100) surface transforms from a multi-component system to a surface covered predominantly with  $\text{H}-\text{OH}$  paired species due to their weaker propensity for desorption, suggesting the nature of irreversible adsorption of  $\text{H}-\text{OH}$  pairings on the MgO (100) surfaces.

Similarly, decorating the (110) surfaces with  $\text{H}-\text{OH}$  pairings originating from a methanol or IPA treatment is also readily thermodynamically favored (see Fig. S19(c and d)†). However, the  $\text{H}-\text{CH}_x\text{O}$  pairings are competitively adsorbing, and this trend is more significant for the IPA-treated case. The desorption analysis shows both these adsorbate pairings are stable on the MgO (110) surfaces ( $\Delta E_{\text{des}}$  of +1.951 eV for  $\text{H}-\text{OH}$  and +0.541 eV for  $\text{H}-\text{CH}_x\text{O}$ ). The irreversible binding



of the  $\text{H}-\text{CH}_x\text{O}$  species indicates the  $\text{MgO}$  (110) surfaces are likely susceptible to adventitious carbon. It is also worth noting that the hexane-treated  $\text{MgO}$  (110) surface accumulates  $\text{CH}_x$  groups, implying the potential of  $\text{MgO}$  (110) surfaces for attracting adventitious carbon that could hinder carbonation.

### 3.5 Mechanistic carbonation analysis on functionalized $\text{MgO}$ surfaces

$\text{CO}_2$  adsorption was studied on bare untreated  $\text{MgO}$  (100) surface and functionalized methanol/IPA-treated  $\text{MgO}$  (100) surfaces (terminated with 0.125 ML of  $\text{H}-\text{OH}$  pairings). In addition, these conditions were compared to the scenario where explicit water molecules are present, as to simulate the moisture component in the air. The complete list of considered adsorption configurations and energies are given in Tables S1–S8.† Fig. 9 summarizes the lowest adsorption energies specific to each terminated surface. On the bare  $\text{MgO}$  (100) surface, we only found  $\text{CO}_2$  interacting weakly with the surface in a physisorbed state ( $\Delta E_{\text{ads}}$  of  $-0.292$  eV). On the methanol/IPA-treated  $\text{MgO}$  (100) surface, there were three possible stable  $\text{CO}_2$  adsorption states:  $\text{CO}_2$  adsorbs with a linear geometry ( $-0.609$  eV) (Fig. S20(a and b)†); it could be adsorbed in a bent geometry forming a  $\text{C}-\text{O}$  bond with a lattice oxygen and transforming into a carbonate-like species ( $-0.852$  eV) (Fig. 10(c and d)); or it could form a  $\text{C}-\text{OH}$  bond with a surface  $\text{OH}$  group and become a bicarbonate-like species ( $-0.714$  eV) (Fig. 10(e and f)). To verify if  $\text{CO}_2$  predominantly interacts with the surface and the adsorbed species and not with other  $\text{CO}_2$  molecules as present through the mirrored images due to the periodic boundary conditions, we performed the same calculation on an expanded ( $4 \times 4$ ) supercell. We find similar adsorption results with a value of  $-0.894$  eV compared to  $-0.852$  eV for carbonate-like species and  $-0.744$  eV compared to  $-0.714$  eV for bicarbonate-like species formation. This confirms that the dominating interactions are primarily a consequence of adsorbate–surface interaction and not adsorbate–adsorbate interactions. We next investigated if the improved  $\text{CO}_2$

adsorption thermodynamics can be attributed to the proximity of the adsorbing  $\text{CO}_2$  molecule to the  $\text{H}-\text{OH}$  functional groups. To do this, we adsorbed the  $\text{CO}_2$  as far from  $\text{OH}$  as possible in the  $4 \times 4$  supercell. The resulting  $\Delta E_{\text{ads}}$  is  $-0.290$  eV, which is very close to the adsorption value on a bare untreated surface ( $-0.292$  eV). The redistribution (gain and loss) of electrons of each atom caused by the carbonation processes is shown in Fig. 10 (isosurface level was set to  $0.003 \text{ e } \text{\AA}^{-3}$  for all plots). The electron charge density analysis confirms the capability of  $\text{CO}_2$  chemisorbing on the methanol/IPA-treated (100) surface but not on the bare surface. Two viable pathways for carbonation on a methanol/IPA-treated  $\text{MgO}$  (100) surface (Fig. 10): (1)  $\text{H}$  first recombines with  $\text{OH}$  to form a water molecule, and then interacts with  $\text{O}$  in  $\text{CO}_2$  and stabilizes its structure in a bent configuration, and (2)  $\text{OH}$  stays dissociated from  $\text{H}$ , and directly forms a  $\text{C}-\text{OH}$  bond. In both pathways,  $\text{OH}$  groups are indispensable, but the importance of isolated  $\text{H}$  is unclear. Therefore, we performed additional calculations to isolate  $\text{OH}$  from its  $\text{H}$  counterpart and performed an adsorption analysis on the surface covered with  $\text{OH}$  only. The calculated  $\Delta E_{\text{ads}}$  values are  $-0.469$  eV for linear  $\text{CO}_2$ ,  $-0.447$  eV for carbonate-like species, and  $-0.810$  eV for bicarbonate-like species (see Fig. S23†). Notably, forming carbonate-like species becomes less favored when the surface lacks recombined water. Instead, the isolated  $\text{OH}$  group allows a surface  $\text{Mg}$  to rumple and then facilitates stabilizing  $\text{CO}_2$ , however, this effect has a smaller influence on the adsorption than water. On the other hand,  $\text{OH}$  likely has more freedom to interact with  $\text{C}$  without electron transfer with  $\text{H}$ , hence  $\Delta E_{\text{ads}}$  is more negative. We have also modeled  $\text{CO}_2$  adsorption on the surface with full (1ML)  $\text{H}-\text{OH}$  coverage for comparative purposes, even though this starting surface is not thermodynamically favored. The  $\Delta E_{\text{ads}}$  is  $-10.394$  eV (or  $-1.299$  eV per  $\text{H}-\text{OH}$  pair), but no evidence for chemisorption is found (see Fig. S20(c and d)†). Instead, the very negative energy is due to surface relaxation and not adsorption.

When a system is exposed to air, water can participate as a reactant and is particularly critical for lowering the carbonation energy barriers.<sup>66</sup> To examine the role of water,

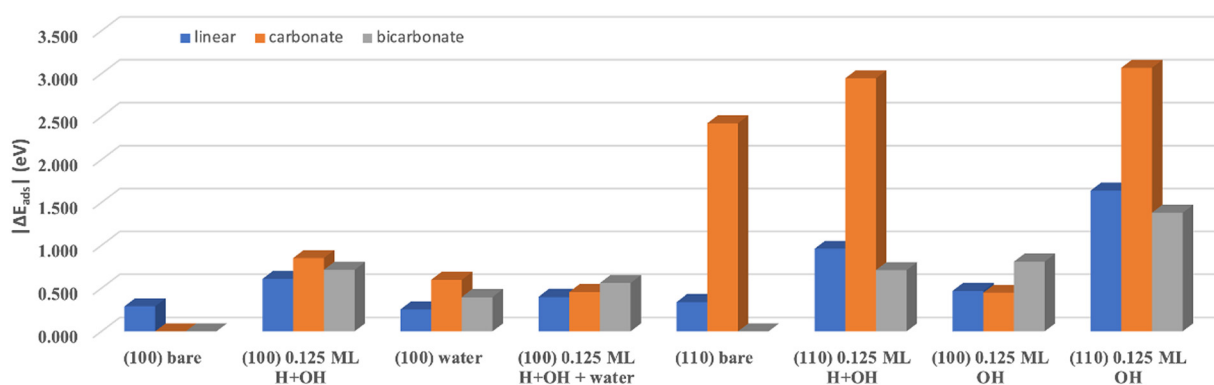
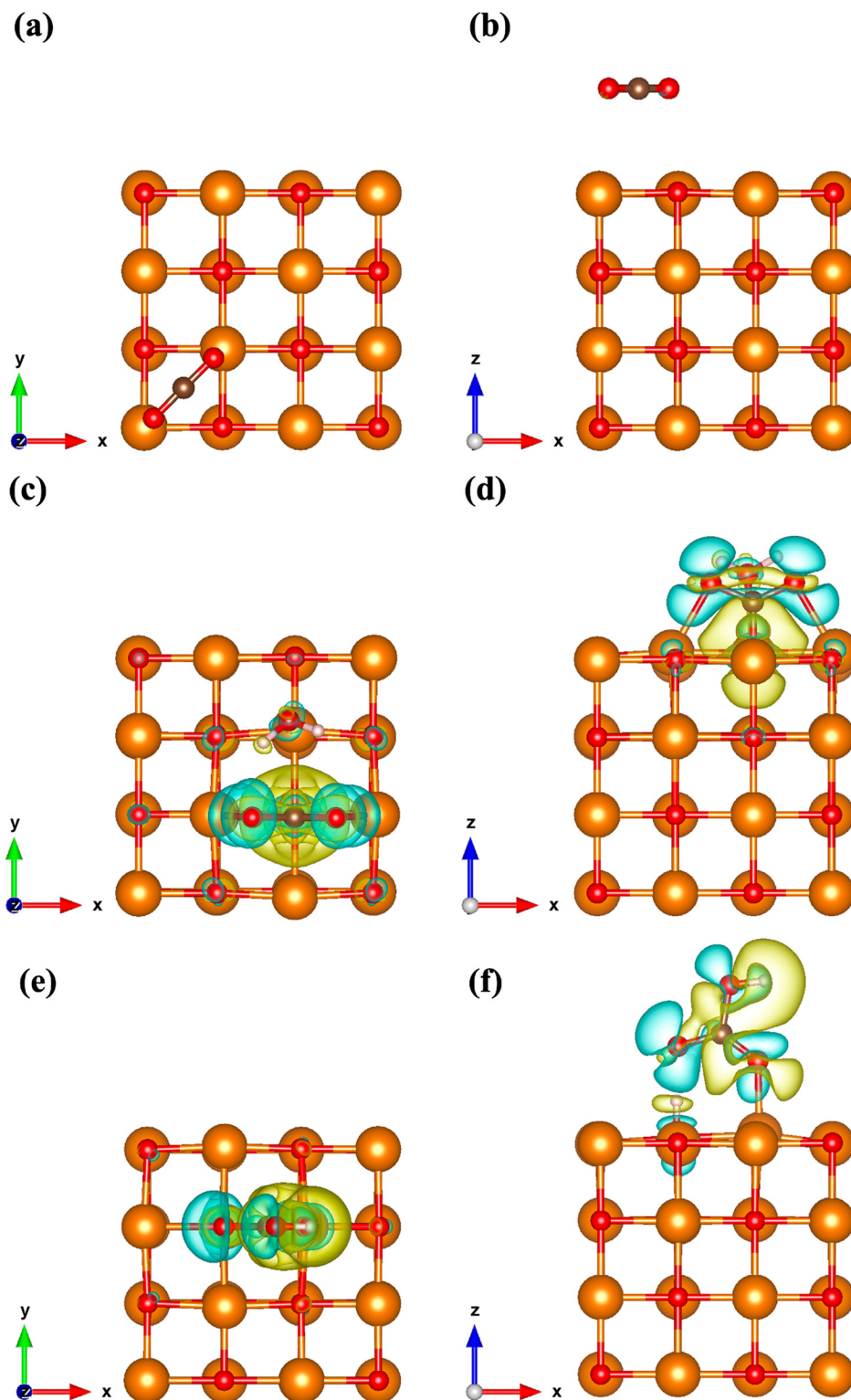


Fig. 9 The lowest  $\text{CO}_2$  adsorption energy on bare and functionalized (100) and (010) surfaces with the  $\text{CO}_2$  in different configurations: linear, carbonate and bicarbonate.





**Fig. 10** Visualized charge difference upon  $\text{CO}_2$  interacting with an  $\text{MgO}$  (100) bare surface (a) top-down view, (b) side view, methanol/IPA-treated surface with carbonate-like species formation (c) top-down view, (d) side view, methanol/IPA-treated surface with bicarbonate-like species formation (e) top-down view, (f) side view.

we assume a 1 : 1  $\text{H}_2\text{O}:\text{CO}_2$  ratio. The bare  $\text{MgO}$  (100) surface was found to have weak interaction with a water molecule ( $\Delta E_{\text{ads}}$  of  $-0.486$  eV), while the water interaction with the

methanol/IPA-treated (100) surface was much stronger ( $-1.024$  eV). This could be due to the hydrogen bonding between water and surface OH groups. With these adsorbed



water configurations on the bare MgO (100) surfaces, adsorption of CO<sub>2</sub> was modeled. The resulting observed CO<sub>2</sub> configurations were both carbonate-like ( $\Delta E_{\text{ads}}$  of  $-0.598$  eV) and bicarbonate-like species ( $-0.398$  eV), in addition to linearly adsorbed CO<sub>2</sub> ( $-0.257$  eV) (see Fig. S21†). For carbonation on the methanol/IPA-treated MgO (100) surfaces with explicit water, we find that the adsorbed linear CO<sub>2</sub> is more favorable ( $-0.400$  eV), but the thermodynamics also shift toward favoring the formation of bicarbonate-like species ( $-0.560$  eV) over carbonate-like species ( $-0.455$  eV) (see Fig. S22†). Prior to the carbonate-like species formation, H and OH recombination takes place first and subsequently triggers the carbonation. Although the presence of H–OH pairings seemingly has no significant effect in lowering  $\Delta E_{\text{ads}}$  as long as molecular water exists, the relatively unfavorable thermodynamics ( $-0.486$  eV) for adsorbing water directly on a bare MgO (100) surface implies diminished feasibility in initiating the carbonation process. More importantly, bicarbonate-like species  $\Delta E_{\text{ads}}$  is much less negative when a water molecule is pre-adsorbed in contrast to the surfaces without adsorbed water molecules but with H–OH pairings. This is because there is an energy cost for breaking the H–OH bond in water to subsequently form a C–OH bond with CO<sub>2</sub>. This highlights the role of an OH surface species that outperforms a molecular water for carbonation. Note that bulk water does not favor a spontaneous dissociation over the bare MgO (100) surfaces (see Fig. S19(a)†). In other words, treating the MgO (100) surface with water wouldn't result in the same favorable conditions (H–OH pairs) with enhanced carbonation as polar organic solvents can achieve.

The adsorption energy calculations were repeated on the MgO (110) surfaces. On the bare MgO (110) surface, the  $\Delta E_{\text{ads}}$  was found to be much more negative ( $-2.424$  eV) than in any other cases. The superior reactivity can be attributed to the presence of surface four-coordinated oxygen (O<sub>4c</sub>) site, which readily forms a C–O bond with CO<sub>2</sub> and turns into a carbonate-like species. The superior thermodynamics likely results in an excessively reactive surface, making it susceptible to passivation by unreactive carbon deposition.<sup>28,36,67</sup> The O<sub>4c</sub> sites are likely occupied, hindering further carbonation on the bare (110) surface. Finally, we modeled the interaction between organic groups and the O<sub>4c</sub> sites by studying the adsorption behaviors of H–CH<sub>x</sub>O groups. When the (110) surface is covered with 0.125 ML H–CH<sub>x</sub>O, the  $\Delta E_{\text{ads}}$  becomes less negative with respect to the distance between the CO<sub>2</sub> adsorption site and the position of CH<sub>x</sub>O. When CO<sub>2</sub> adsorbs on a lattice O that is 3.195 Å away from CH<sub>x</sub>O,  $\Delta E_{\text{ads}}$  is  $-1.673$  eV. When the distance is 4.854 Å, the  $\Delta E_{\text{ads}}$  value becomes  $-2.134$  eV, and at the distance of 5.121 Å we find  $\Delta E_{\text{ads}}$  is  $-2.443$  eV. Hence, we conclude that the presence of CH<sub>x</sub>O screens the surface toward CO<sub>2</sub> adsorption, but the effect drops significantly as the proximity increases. When the surface is fully covered with H–CH<sub>x</sub>O, there is only physisorption ( $-0.223$  eV). This suggests a monotonic decrease in the CO<sub>2</sub> adsorption thermodynamics with respect to the coverage of H–CH<sub>x</sub>O pairings. Conversely,

the presence of H–OH pairings could enhance the interaction between CO<sub>2</sub> and surface species, similar to (100) surfaces. The  $\Delta E_{\text{ads}}$  of adsorbing linear CO<sub>2</sub> and forming carbonate-like species improves from  $-0.341$  eV to  $-0.960$  eV and from  $-2.424$  eV to  $-2.945$  eV, respectively. Regardless, we expect a smaller tendency towards carbonation as the  $\Delta E_{\text{ads}}$  values indicate due to the aforementioned passivation effect. Because of the passivated (110) surfaces, no carbonation was experimentally reported. With methanol/IPA treatment, we have insufficient experimental evidence to show if carbonation has occurred on (110) surfaces. If so, it is likely through the formation of bicarbonate-like species ( $-0.711$  eV), which is less affected by the passivation effect.

## 4 Discussions

The findings of our study show that surface functionalization of MgO nanoparticles plays a crucial role in facilitating the carbonation process of proxy mine tailings rich in MgO. We observed that the as-produced dry MgO powder samples showed no signs of carbonation when exposed to air, whereas the carbonation process accelerated significantly once the MgO nanoparticles were dispersed in a polar solvent, sonicated, dried, and then exposed to air. Despite the thermodynamic feasibility of carbonation, the slow carbonation rate observed in the dry sprinkled MgO samples can be attributed to the low CO<sub>2</sub> concentration in the atmosphere and inadequate water content.<sup>24,68</sup> When MgO contacts with methanol/IPA it spontaneously forms surface reactive H–OH groups, which can greatly alter the reaction with CO<sub>2</sub> to form MgCO<sub>3</sub>.<sup>69–71</sup> On the other hand, dispersing the minerals in a non-polar solvent like hexane led to the accumulation of adventitious carbon, hindering the access of CO<sub>2</sub> and H<sub>2</sub>O to the surface of MgO. Furthermore, our Rietveld analysis revealed that both samples exhibited a similar extent of preferential orientation or texture, favoring (100) facets in both cases, which is consistent with our TEM analysis. The particle size difference between sample A and sample B can corroborate the difference in the kinetics of the MgO samples. The carbonation process of MgO is surface-driven, with smaller crystal sizes showing more pronounced effects. However, we couldn't experimentally identify a specific initiation site, such as a crystal corner or edge, that could propagate the reaction. Morphological observations also supported the surface-driven process, as early stages of carbonation consistently formed discrete surface features on the MgO aggregates. Unfortunately, imaging the carbonation reaction front was very challenging due to rapid beam-induced changes in the MgCO<sub>3</sub> area. Nonetheless, our results suggest that the reaction front is not homogeneous, pointing towards mass-diffusion limitations in the conversion process.<sup>72</sup> This is supported by the fact that the shape of an aggregate is preserved during and after the conversion to MgCO<sub>3</sub>, indicating that the reaction occurs entirely in the solid state without dissolution and re-precipitation. Throughout our investigations, we detected the presence of



elements like Si, Ca, Fe, and F. Ca was well mixed within the MgO aggregates, while Si-based impurity aggregates spread out when dispersed in a polar solvent. The hydrophilic nature of  $\text{SiO}_x$  may further promote the carbonation of MgO aggregates by maintaining a hydrating environment around the MgO particles.<sup>73</sup> The initial solvation step is crucial in enhancing carbonation kinetics, as the surface polarity increases with the adsorption of the  $-\text{OH}$  group, attracting moisture from the ambient atmosphere and promoting carbonation.<sup>24</sup> This corroborates the enhanced carbonation achieved for MgO treated with methanol/IPA, which, as discussed above, induces the formation of surface Mg–OH groups.<sup>71</sup> EELS analysis of the carbonating aggregates revealed that the transformation from MgO to  $\text{MgCO}_3$  involved the formation of mixed carbon-containing phases rather than a direct conversion. Additionally, during the transition, a peak corresponding to bicarbonate ion was detected, indicating that bicarbonates are an intermediate species in the conversion of MgO to  $\text{MgCO}_3$ . Further our DFT calculations show that functionalization is possible but dependent on the solvent being polar vs. non-polar. Importantly, the DFT findings also confirm our experimental observation of enhanced carbonation over solvent-treated *i.e.* functionalized MgO. Specifically, partially the H–OH covered MgO (100) surface shows an increased propensity towards water and  $\text{CO}_2$  moiety and subsequently undergoes carbonation process. The presence of H and OH on the surface together with explicit  $\text{H}_2\text{O}$  molecules stemming from the atmosphere results in the formation of bicarbonate along with the carbonate, which is in agreement with the experimental observation. Overall, our study demonstrates the significance of surface functionalization, particle size, and other impurities in influencing the carbonation process of MgO. These findings have implications for understanding and optimizing the carbonation potential of mine tailings rich in MgO for environmental applications.

## 5 Conclusion

We conducted a comprehensive study on the natural carbonation process of MgO derived from Mg-ion-rich mine tailings. While the carbonation of dry MgO at ambient atmospheric conditions was found to kinetically slow although thermodynamically feasible, it can be significantly improved by dispersing in a polar, OH-containing solvent. Notably, we observed that the carbonation process of MgO is size-dependent, with smaller particle size aggregates of MgO carbonate more rapidly than larger ones. Moreover, we discovered that the presence of hygroscopic minerals, such as  $\text{SiO}_x$ , further enhances the carbonation process by maintaining a hydrated atmosphere surrounding the MgO aggregates. In our investigation, we utilized electron energy loss spectroscopy (EELS) to study the carbonation mechanism and the intermediates involved. We found that the carbonation of MgO does not occur directly; instead, it forms mixed phases of carbon-containing substances. Particularly,

the naturally carbonating aggregates revealed that the carbonation takes place through the formation of an intermediate bicarbonate phase. DFT simulations show existence of functionalized surfaces stemming from polar solvent treatment and further strengthen our hypothesis of enhanced carbonation of these modified MgO surfaces. A favorable decomposition of polar solvent over MgO (100) to partially cover the surface with H and OH. Adsorbed isolated H and OH are shown to be crucial for further adsorption of both  $\text{H}_2\text{O}$  and  $\text{CO}_2$  to undergo the carbonation process. These findings shed new light on the carbonation behavior of MgO and provide valuable insights into the role of particle size and hygroscopic minerals in enhancing this process. Our study contributes to a better understanding of the natural carbonation potential of MgO, which may have significant implications for various applications in carbon sequestration and environmental remediation.

## Data availability

Data supporting this study have been included as a part of ESI† file.

## Conflicts of interest

There are no conflicts to declare.

## Acknowledgements

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## Notes and references

- 1 T. P. X. Lan and K. W. Thoning, *Trends in globally-averaged  $\text{CO}_2$  determined from NOAA Global Monitoring Laboratory measurements, version 2024-01*, DOI: [10.15138/9N0H-ZH07](https://doi.org/10.15138/9N0H-ZH07).
- 2 J. Pires, F. Martins, M. Alvim-Ferraz and M. Simões, Recent developments on carbon capture and storage: An overview, *Chem. Eng. Res. Des.*, 2011, **89**(9), 1446–1460.
- 3 R. S. Haszeldine, Carbon capture and storage: how green can black be?, *Science*, 2009, **325**(5948), 1647–1652.
- 4 N. Bahman, M. Al-Khalifa, S. Al Baharna, Z. Abdulmohsen and E. Khan, Review of carbon capture and storage technologies in selected industries: potentials and challenges, *Rev. Environ. Sci. Biotechnol.*, 2023, 1–20.
- 5 Y. Zhu and H. Frey, Integrated gasification combined cycle (IGCC) power plant design and technology, in *Advanced*



- power plant materials, design and technology, Elsevier, 2010, pp. 54–88.
- 6 D. Cebrecuan and I. Ionel, *Biomass co-firing with carbon capture*, 2022.
  - 7 C. Song, Q. Liu, S. Deng, H. Li and Y. Kitamura, Cryogenic-based CO<sub>2</sub> capture technologies: State-of-the-art developments and current challenges, *Renewable Sustainable Energy Rev.*, 2019, **101**, 265–278.
  - 8 U. W. Siagian, A. Raksajati, N. F. Himma, K. Khoiruddin and I. Wenten, Membrane-based carbon capture technologies: Membrane gas separation vs. membrane contactor, *J. Nat. Gas Sci. Eng.*, 2019, **67**, 172–195.
  - 9 D. Fu and J. Xie, Absorption capacity and viscosity for CO<sub>2</sub> capture process using [N1111][Gly] promoted K<sub>2</sub>CO<sub>3</sub> aqueous solution, *J. Chem. Thermodyn.*, 2016, **102**, 310–315.
  - 10 A. Gautam and M. K. Mondal, Review of recent trends and various techniques for CO<sub>2</sub> capture: Special emphasis on biphasic amine solvents, *Fuel*, 2023, **334**, 126616.
  - 11 M. Wilson, P. Tontiwachwuthikul, A. Chakma, R. Idem, A. Veawab and A. Aroonwilas, *et al.*, Test results from a CO<sub>2</sub> extraction pilot plant at boundary dam coal-fired power station, in *Greenhouse Gas Control Technologies-6th International Conference*, Elsevier, 2003, pp. 31–36.
  - 12 D. P. Hagewiesche, S. S. Ashour, H. A. Al-Ghawas and O. C. Sandall, Absorption of carbon dioxide into aqueous blends of monoethanolamine and N-methyldiethanolamine, *Chem. Eng. Sci.*, 1995, **50**(7), 1071–1079.
  - 13 G. T. Rochelle, Amine scrubbing for CO<sub>2</sub> capture, *Science*, 2009, **325**(5948), 1652–1654.
  - 14 Y. Mao, X. Yang and T. V. Gerven, Amine-assisted simultaneous CO<sub>2</sub> absorption and mineral carbonation: effect of different categories of amines, *Environ. Sci. Technol.*, 2023, **57**(29), 10816–10827.
  - 15 M. Ramezan, T. J. Skone, N. Nsakala, G. Liljedahl, L. Gearhart and R. Hestermann, *et al.*, *Carbon dioxide capture from existing coal-fired power plants*, National Energy Technology Laboratory, DOE/NETL Report, 2007, vol. 401, p. 110907.
  - 16 G. Gadikota, Carbon mineralization pathways for carbon capture, storage and utilization, *Commun. Chem.*, 2021, **4**(1), 23.
  - 17 I. M. Power, A. L. Harrison, G. M. Dipple, S. Wilson, P. B. Kelemen and M. Hitch, *et al.*, Carbon mineralization: from natural analogues to engineered systems, *Rev. Mineral. Geochem.*, 2013, **77**(1), 305–360.
  - 18 J. M. Matter, M. Stute, S. Ó. Snæbjörnsdóttir, E. H. Oelkers, S. R. Gislason and E. S. Aradóttir, *et al.*, Rapid carbon mineralization for permanent disposal of anthropogenic carbon dioxide emissions, *Science*, 2016, **352**(6291), 1312–1314.
  - 19 W. J. J. Huigen and R. N. J. Comans, *Carbon dioxide sequestration by mineral carbonation: Literature Review*, Energy Research Centre of the Netherlands, 2003, <https://publications.ecn.nl/E/2003/ECN-C-03-016>.
  - 20 K. S. Lackner, Carbonate chemistry for sequestering fossil carbon, *Annu. Rev. Energy*, 2002, **27**(1), 193–232.
  - 21 A. M. Kierzkowska, R. Pacciani and C. R. Müller, CaO-based CO<sub>2</sub> sorbents: from fundamentals to the development of new, highly effective materials, *ChemSusChem*, 2013, **6**(7), 1130–1148.
  - 22 A. Ababou, M. Ajbary, M. Taleb and A. Kherbeche, Direct mineral carbonation of new materials for CO<sub>2</sub> sequestration, *J. Mater. Environ. Sci.*, 2017, **8**(9), 3106–3111.
  - 23 Y. Hu, Y. Guo, J. Sun, H. Li and W. Liu, Progress in MgO sorbents for cyclic CO<sub>2</sub> capture: A comprehensive review, *J. Mater. Chem. A*, 2019, **7**(35), 20103–20120.
  - 24 N. McQueen, P. Kelemen, G. Dipple, P. Renforth and J. Wilcox, Ambient weathering of magnesium oxide for CO<sub>2</sub> removal from air, *Nat. Commun.*, 2020, **11**(1), 3299.
  - 25 S. Gregg and J. Ramsay, Adsorption of carbon dioxide by magnesia studied by use of infrared and isotherm measurements, *J. Chem. Soc. A*, 1970, 2784–2787.
  - 26 R. Philipp and K. Fujimoto, FTIR spectroscopic study of carbon dioxide adsorption/desorption on magnesia/calcium oxide catalysts, *J. Phys. Chem.*, 1992, **96**(22), 9035–9038.
  - 27 T. Papalas, I. Polychronidis, A. N. Antzaras and A. A. Lemonidou, Enhancing the intermediate-temperature CO<sub>2</sub> capture efficiency of mineral MgO via molten alkali nitrates and CaCO<sub>3</sub>: Characterization and sorption mechanism, *J. CO<sub>2</sub> Util.*, 2021, **50**, 101605.
  - 28 J. Weber, V. Starchenko, K. Yuan, L. M. Anovitz, A. V. Ievlev and R. R. Unocic, *et al.*, Armoring of MgO by a Passivation Layer Impedes Direct Air Capture of CO<sub>2</sub>, *Environ. Sci. Technol.*, 2023, **57**(40), 14929–14937.
  - 29 W. Seifritz, CO<sub>2</sub> disposal by means of silicates, *Nature*, 1990, **345**(6275), 486.
  - 30 P. C. Chiang, S. Y. Pan, P. C. Chiang and S. Y. Pan, Principles of accelerated carbonation reaction, *Carbon dioxide mineralization and utilization*, 2017, pp. 71–96.
  - 31 S. Ó. Snæbjörnsdóttir, B. Sigfússon, C. Marieni, D. Goldberg, S. R. Gislason and E. H. Oelkers, Carbon dioxide storage through mineral carbonation, *Nat. Rev. Earth Environ.*, 2020, **1**(2), 90–102.
  - 32 M. Rekhtina, M. Krödel, Y. H. Wu, A. Kierzkowska, F. Donat and P. M. Abdala, *et al.*, Deciphering the structural dynamics in molten salt-promoted MgO-based CO<sub>2</sub> sorbents and their role in the CO<sub>2</sub> uptake, *Sci. Adv.*, 2023, **9**(26), eadg5690.
  - 33 Y. Xu, F. Donat, C. Luo, J. Chen, A. Kierzkowska and M. A. Naeem, *et al.*, Investigation of K<sub>2</sub>CO<sub>3</sub>-modified CaO sorbents for CO<sub>2</sub> capture using in-situ X-ray diffraction, *Chem. Eng. J.*, 2023, **453**, 139913.
  - 34 T. Harada, F. Simeon, E. Z. Hamad and T. A. Hatton, Alkali metal nitrate-promoted high-capacity MgO adsorbents for regenerable CO<sub>2</sub> capture at moderate temperatures, *Chem. Mater.*, 2015, **27**(6), 1943–1949.
  - 35 A. M. Alkadhém, M. A. Elgzoly and S. A. Onaizi, Novel amine-functionalized magnesium oxide adsorbents for CO<sub>2</sub> capture at ambient conditions, *J. Environ. Chem. Eng.*, 2020, **8**(4), 103968.
  - 36 J. N. Bracco, G. Camacho Meneses, O. Colón, K. Yuan, J. E. Stubbs and P. J. Eng, *et al.* Reaction Layer Formation on MgO in the Presence of Humidity, *ACS Appl. Mater. Interfaces*, 2023, **16**(1), 712–712.



- 37 S. O. Baumann, J. Schneider, A. Sternig, D. Thomele, S. Stankic and T. Berger, *et al.*, Size effects in MgO cube dissolution, *Langmuir*, 2015, **31**(9), 2770–3776.
- 38 J. M. Hanlon, L. B. Diaz, G. Balducci, B. A. Stobbs, M. Bielewski and P. Chung, *et al.*, Rapid surfactant-free synthesis of Mg (OH)<sub>2</sub> nanoplates and pseudomorphic dehydration to MgO, *CrystEngComm*, 2015, **17**(30), 5672–5679.
- 39 P. Liu, P. M. Abdala, G. Goubert, M. G. Willinger and C. Copéret, Ultrathin single crystalline MgO (111) nanosheets, *Angew. Chem., Int. Ed.*, 2021, **60**(6), 3254–3260.
- 40 S. Cho, H. R. Yu, T. H. Choi, M. J. Jung and Y. S. Lee, Surface functionalization and CO<sub>2</sub> uptake on carbon molecular sieves: experimental observation and theoretical study, *Appl. Surf. Sci.*, 2018, **447**, 8–14.
- 41 S. A. Fuente, C. A. Ferretti, N. F. Domancich, V. K. Díez, C. R. Apesteguía and J. I. Di Cosimo, *et al.*, Adsorption of 2-propanol on MgO surface: A combined experimental and theoretical study, *Appl. Surf. Sci.*, 2015, **327**, 268–276.
- 42 R. Kakkar, P. N. Kapoor and K. J. Klabunde, First principles density functional study of the adsorption and dissociation of carbonyl compounds on magnesium oxide nanosurfaces, *J. Phys. Chem. B*, 2006, **110**(51), 25941–25949.
- 43 D. Smith and A. Tench, Reactions of alcohols adsorbed on MgO. Part I. Heterogeneous reactions with hydrogen atoms and trapped electrons at room temperature: a new technique in surface chemistry, *Can. J. Chem.*, 1969, **47**(8), 1381–13819.
- 44 V. Molahid, F. Kusin and M. Soomro, Mineral carbonation for carbon dioxide capture and storage using mining waste as feedstock material, in *IOP Conference Series: Earth and Environmental Science*, IOP Publishing, 2023, vol. 1205, p. 012011.
- 45 A. L. Harrison, I. M. Power and G. M. Dipple, Accelerated carbonation of brucite in mine tailings for carbon sequestration, *Environ. Sci. Technol.*, 2013, **47**(1), 126–134.
- 46 J. Sipilä, S. Teir and R. Zevenhoven, Carbon dioxide sequestration by mineral carbonation Literature review update 2005–2007, *Report Vt.*, 2008, vol. 1, p. 2008.
- 47 D. P. Butt, K. S. Lackner, C. H. Wendt, S. D. Conzone, H. Kung and Y. C. Lu, *et al.*, Kinetics of thermal dehydroxylation and carbonation of magnesium hydroxide, *J. Am. Ceram. Soc.*, 1996, **79**(7), 1892–1898.
- 48 A. H. Larsen, J. J. Mortensen, J. Blomqvist, I. E. Castelli, R. Christensen and M. Dulak, *et al.* The atomic simulation environment—a Python library for working with atoms, *J. Phys.: Condens. Matter*, 2017, **29**(27), 273002.
- 49 P. Giannozzi, S. Baroni, N. Bonini, M. Calandra, R. Car and C. Cavazzoni, *et al.*, ESPRESSO: a modular and open-source software project for quantum simulations of materials, *J. Phys.: Condens. Matter*, 2009, **21**(39), 395502.
- 50 P. Giannozzi, O. Andreussi, T. Brumme, O. Bunau, M. B. Nardelli and M. Calandra, *et al.*, Advanced capabilities for materials modelling with Quantum ESPRESSO, *J. Phys.: Condens. Matter*, 2017, **29**(46), 465901.
- 51 J. Wellendorff, K. T. Lundgaard, A. Møgelhøj, V. Petzold, D. D. Landis and J. K. Nørskov, *et al.*, Density functionals for surface science: Exchange-correlation model development with Bayesian error estimation, *Phys. Rev. B: Condens. Matter Mater. Phys.*, 2012, **85**(23), 235149.
- 52 D. Vanderbilt, Soft self-consistent pseudopotentials in a generalized eigenvalue formalism, *Phys. Rev. B: Condens. Matter Mater. Phys.*, 1990, **41**(11), 7892.
- 53 R. M. Hazen, Effects of temperature and pressure on the cell dimension and X-ray temperature factors of periclase, *Am. Mineral.*, 1976, **61**(3–4), 266–271.
- 54 H. J. Monkhorst and J. D. Pack, Special points for Brillouin-zone integrations, *Phys. Rev. B: Solid State*, 1976, **13**(12), 5188.
- 55 Z. Liu, X. He, Z. Mei, H. Liang, L. Gu and X. Duan, *et al.*, Polarity poly-manipulation based on nanoscale structural transformation on strained 2D MgO, *J. Phys. D: Appl. Phys.*, 2014, **47**(10), 105303.
- 56 R. Michalsky, Y. J. Zhang, A. J. Medford and A. A. Peterson, Departures from the adsorption energy scaling relations for metal carbide catalysts, *J. Phys. Chem. C*, 2014, **118**(24), 13026–13034.
- 57 P. Hu, S. Wang and Y. Zhuo, Research on CO<sub>2</sub> adsorption performances of metal-doped (Ca, Fe and Al) MgO, *Sep. Purif. Technol.*, 2021, **276**, 119323.
- 58 R. Wirth, Water in minerals detectable by electron energy-loss spectroscopy EELS, *Phys. Chem. Miner.*, 1997, **24**(8), 561–568.
- 59 B. Winkler, M. Avalos-Borja, V. Milman, A. Perlov, C. Pickard and J. Yates, Oxygen K-edge electron energy loss spectra of hydrous and anhydrous compounds, *J. Phys.: Condens. Matter*, 2013, **25**(48), 485401.
- 60 A. F. A. Fuhaid and A. Niaz, Carbonation and corrosion problems in reinforced concrete structures, *Buildings*, 2022, **12**(5), 586.
- 61 T. Yang, B. Keller, E. Magyari, K. Hametner and D. Günther, Direct observation of the carbonation process on the surface of calcium hydroxide crystals in hardened cement paste using an atomic force microscope, *J. Mater. Sci.*, 2003, **38**, 1909–1916.
- 62 Y. Jia, J. Zhang, Y. Zou, Q. Guo, M. Li and T. Zhang, *et al.*, Development and Applications of MgO-activated SiO<sub>2</sub> system—achieving a low carbon footprint: A review, *Green Energy and Resources*, 2024, 100072.
- 63 G. B. Singh, C. Sonat, E. Yang and C. Unluer, Performance of MgO and MgO–SiO<sub>2</sub> systems containing seeds under different curing conditions, *Cem. Concr. Compos.*, 2020, **108**, 103543.
- 64 D. P. Prentice, O. AlShareedah, M. Sarkar, J. Arabit, I. Mehdipour and S. Afzal, *et al.*, Process modeling guides operational variables that affect CO<sub>2</sub> utilization during the accelerated carbonation of concrete, *AIChE J.*, 2024, **70**(5), e18387.
- 65 S. Kwon, M. Fan, H. F. Dacosta, A. G. Russell and C. Tsouris, Reaction kinetics of CO<sub>2</sub> carbonation with Mg-rich minerals, *J. Phys. Chem. A*, 2011, **115**(26), 7638–7644.
- 66 M. G. Gardeh, A. A. Kistanov, H. Nguyen, H. Manzano, W. Cao and P. Kinnunen, Exploring mechanisms of hydration



- and carbonation of MgO and Mg (OH)<sub>2</sub> in reactive magnesium oxide-based cements, *J. Phys. Chem. C*, 2022, **126**(14), 6196–6206.
- 67 K. Rausis, A. R. Stubbs, I. M. Power and C. Paulo, Rates of atmospheric CO<sub>2</sub> capture using magnesium oxide powder, *Int. J. Greenhouse Gas Control*, 2022, **119**, 103701.
- 68 P. B. Kelemen, N. McQueen, J. Wilcox, P. Renforth, G. Dipple and A. P. Vankeuren, Engineered carbon mineralization in ultramafic rocks for CO<sub>2</sub> removal from air: Review and new insights, *Chem. Geol.*, 2020, **550**, 119628.
- 69 A. Scott, C. Oze, V. Shah, N. Yang, B. Shanks and C. Cheeseman, *et al.*, Transformation of abundant magnesium silicate minerals for enhanced CO<sub>2</sub> sequestration, *Commun. Earth Environ.*, 2021, **2**(1), 25.
- 70 H. K. Bharadwaj, J. Y. Lee, X. Li, Z. Liu and T. C. Keener, Dissolution kinetics of magnesium hydroxide for CO<sub>2</sub> separation from coal-fired power plants, *J. Hazard. Mater.*, 2013, **250**, 292–297.
- 71 Z. Song and H. Xu, Splitting methanol on ultra-thin MgO (100) films deposited on a Mo substrate, *Phys. Chem. Chem. Phys.*, 2017, **19**(10), 7245–7251.
- 72 I. Bena, M. Droz, K. Martens and Z. Racz, Reaction–diffusion fronts with inhomogeneous initial conditions, *J. Phys.: Condens. Matter*, 2007, **19**(6), 065103.
- 73 C. Gunathilake and M. Jaroniec, Mesoporous calcium oxide–silica and magnesium oxide–silica composites for CO<sub>2</sub> capture at ambient and elevated temperatures, *J. Mater. Chem. A*, 2016, **4**(28), 10914–10924.

