



**Environmental
Science**
Water Research & Technology

Purification and removal of the low molecular weight fraction of PolyDADMAC reduces N-nitrosodimethylamine formation during water treatment

Journal:	<i>Environmental Science: Water Research & Technology</i>
Manuscript ID	EW-ART-04-2020-000392.R1
Article Type:	Paper

SCHOLARONE™
Manuscripts

Nitrosamines form in the presence of chloramine disinfectants during drinking water treatment. Drinking water treatment relies upon chemical addition, including cationic polymers (e.g., polyDADMAC), to achieve turbidity and pathogen removal. The present study shows the contributions of cationic polymer to nitrosodimethylamine (NDMA) formation, and how polyDADMAC purification can dramatically reduce NDMA formation. Diaultrafiltration of polyDADMAC to remove low molecular polymer residuals reduced NDMA formation by over greater than 50% without impacting other water treatment goals.

1 **Purification and removal of the low molecular weight fraction of**
2 **PolyDADMAC reduces N-nitrosodimethylamine formation during water**
3 **treatment**

4 **Ariel J. Atkinson,¹ Natalia Fischer¹, Samantha Donovan,² Justin Bartlett,³ Omar Alrehaili,¹**
5 **Shahnawaz Sinha,¹ Sunil Kommineni,³ Pierre Herckes,² Paul Westerhoff^{1*}**

6 ¹School of Sustainable Engineering and the Built Environment, Arizona State University, Tempe,
7 Arizona

8 ²School of Molecular Science, Arizona State University, Tempe, Arizona

9 ³KIT Professionals, Inc., Houston, Texas

10

11 **Abstract**

12 Cationic polymers are critical coagulant aids at drinking water plants, but the same polymers
13 are simultaneously N-nitrosodimethylamine (NDMA) precursors upon chloramine exposure. We
14 hypothesized the low-molecular fraction, which represents a small mass residual in
15 polyDADMAC cationic polymer, reacts with chloramines, may not be well removed during
16 coagulation, and is thus responsible for forming parts-per-trillion (ppt) concentrations of NDMA
17 in finished water. To identify strategies capable of reducing post-coagulation residual polymer
18 associated with NDMA formation, this study fractionated polyDADMAC by molecular weight,
19 characterized reactivity of each fraction, and quantified precursor contributions to NDMA
20 formation potential during bench-scale test and NDMA concentration due to formation during
21 treatment during pilot tests. Diaultrafiltration of the cationic polymer produced a low molecular
22 weight fraction (<10 kDa) and a purified polyDADMAC (>10 kDa). Native organic matter in
23 bench- and pilot- tests contributed 19–38% of NDMA formed throughout treatment, while
24 polymers were responsible for 62–81% of NDMA formed. The <10 kDa fraction of
25 polyDADMAC was more reactive (450–540ngNDMA/mgDOC) and formed >10x NDMA than
26 non-purified polyDADMAC in jar tests. Purified polyDADMAC, with <10 kDa fraction
27 removed, formed 54% less NDMA during pilot tests (and 63% less during jar tests) than non-
28 purified polyDADMAC. There was no adverse effect on coagulation or subsequent filter
29 performance by using purified polyDADMAC, but significantly ($p<0.05$) less NDMA formed in
30 filtered water when using diaultrafiltered cationic polymer. Thus, removing the low molecular
31 weight impurities (<10 kDa) by polymer suppliers would lead to an equally-effective coagulant-
32 aid that substantially lowers NDMA formation during drinking water treatment.

33 **Keywords:** N-nitrosodimethylamine, disinfection by-products, coagulation, precursors

34

Introduction

35 Nitrosamines are potential carcinogens and disinfection by-products in drinking water.¹ N-
36 nitrosodimethylamine (NDMA) is the most frequently detected and highest concentration
37 nitrosamine produced during chloramination of drinking waters.^{2,3} NDMA is part of the United
38 States Environmental Protection Agency (USEPA) contaminant candidate list 4 (CCL4), and
39 regulatory limits exist for drinking water at concentrations of 10 to 40 ng/L in several states and
40 countries (e.g., Canada, Australia, Massachusetts, California). Organic NDMA precursors
41 include natural organic matter, trace organics in wastewater, and cationic polymers used as
42 coagulant aids in water treatment plants.^{2,4} Polydiallyldimethylammonium chloride
43 (polyDADMAC) is the most commonly used cationic polymer, used in approximately 60% of
44 water treatment plants in the US, and has been linked to significant NDMA formation.⁵⁻⁷
45 Polyamines are the second most commonly used cationic polymer class. Polymeric coagulant
46 aids contain long chain polymers, oligomers (i.e., short-chain polymers), polymer precursors
47 (e.g., monomer), and impurities.⁸⁻¹⁰ Low molecular weight organics, including dimethylamine
48 (DMA) and diallyldimethyl ammonium chloride (DADMAC), present in polyDADMAC, have
49 been identified as NDMA precursors but have yields of <1%.^{5,10-13} We hypothesized that
50 commercial polyDADMAC contains residual low molecular weight polymers from the
51 manufacturing process and that these residuals or impurities (1) are irrelevant as coagulant aids
52 (i.e., do not influence coagulation, sedimentation, or filtration of particulates), (2) are poorly
53 removed during water treatment, (3) have higher NDMA reactivity (ngNDMA/mg polymer) than
54 bulk polyDADMAC, and (4) are significant sources of NDMA in chloraminated drinking water.

55 To quantify the contribution of the different constituents, commercial polyDADMAC
56 solutions were separated into a “purified” higher molecular weight fraction (>10 kDa) and a

57 lower molecular weight fraction (<10 kDa) using dead-end and recirculation diaultrafiltration
58 and then used in bench- and pilot-scale experiments simulating full-scale water treatment
59 conditions. Bench-scale tests, measured NDMA formation potential (NDMAFP) and were used
60 as proof-of-concept, that the removal of the <10 kDa fraction reduced NDMAFP. Pilot tests,
61 were used to demonstrate at full-scale the reduction of NDMA formed during treatment when a
62 purified (>10 kDa) polymer was used. The objectives of this study were to: (1) compare the
63 NDMA formation potential (NDMAFP) reactivity of the polyDADMAC molecular weight
64 fractions and combined fractions with reactivity and jar tests, (2) quantify the contribution of
65 different molecular weight fractions to NDMA formed during treatment under realistic drinking
66 water conditions with pilot tests, and (3) identify strategies that the polymer manufacturers
67 and/or water treatment plants can use to minimize NDMA formation.

68 **Experimental Methods**

69 **Polymer Sources, Fractionation, and Purification**

70 Commercial cationic polymers were provided by a major chemical vendor (SNF Polychemie).
71 During screening experiments comparing NDMA formation potential (NDMAFP) for Clarifloc
72 C308 and C318, C308 produced roughly 70% less NDMA (Figure S1). Polyamines are the
73 second most common polymeric coagulant aid used in water treatment facilities and are known
74 sources of NDMA precursors. Two sources of polyamine polymers had higher NDMAFP than
75 C308 (Figure S2). Therefore, because of the common use of C308 and the lower NDMAFP from
76 the polymer upon chloramination, most experiments used C308 (20% active polymer).^{6,11}

77 A 10kDa membrane was selected to fractionate the polymer based upon prior research,⁵ and it
78 was utilized in a either a dead-end cell (Amicon) or larger-scale recirculating diaultrafiltration
79 system (Millipore Ultracel and Pellicon 10 kDa membranes). Each polymer fraction was

80 analyzed for total organic carbon (TOC) and total dissolved nitrogen (TN) to ensure mass
81 balance (TOC whole polyDADMAC= TOC >10 kDa + TOC <10 kDa) was within 10% and to
82 inform dosing in subsequent reactivity, jar, and pilot tests. TOC is used to normalize results in
83 reactivity and NDMAFP tests, as a surrogate for the active polymer mass.

84 The dead-end stirred cell operated at 50 psi was used to filter 400 mL of feed until 150 mL
85 remained. The feed was rediluted to 400 mL with ultrapure water and filtered again. The
86 concentrated feed was rediluted to 400 mL, and this solution was considered the bench-scale
87 purified polyDADMAC (>10 kDa). The 500 mL of permeate collected throughout filtration was
88 combined, constituting the bench-scale lower molecular weight fraction (<10 kDa). Separate
89 experiments collected permeate from a 1 kDa ultrafilter membrane (termed <1 kDa fraction), and
90 used in a limited number of experiments. The bench-scale fractions were used in reactivity and
91 jar tests.

92 Prior to diaultrafiltration, Millipore Ultracel and Pellicon 10 kDa membranes (for bench- and
93 pilot- scale, respectively) were cleaned by soaking in ultrapure water for 24 hours (with three
94 water changes) and used to filter ultrapure water (200 mL and 10 L, respectively) that was
95 discarded. C308 was diluted with ultrapure water ($>18 \text{ M}\Omega\cdot\text{cm}^{-1}$) to 10% active polymer
96 concentration and used as the feed. A cross-flow filtration apparatus was used, with pressure
97 being regulated by a valve on the concentrate stream. The concentrate was recirculated to the
98 feed and continuously diluted with ultrapure water to maintain a constant feed volume of 45
99 liters. Filtration continued until 45 liters of permeate was collected (appx. 30 hours of continuous
100 operation). The recirculated concentrate was termed “pilot-purified polyDADMAC (>10 kDa),”
101 and the permeate was considered the “pilot low molecular weight fraction (<10 kDa)”.

102 **Bench-Scale Jar Tests**

103 Bench-scale experiments were conducted in a standard jar test apparatus (PB900, Phipps and
104 Bird, Virginia, USA) with 5 and 30 minutes of rapid mixing and flocculation, respectively,
105 followed by 60 minutes settling time.¹⁴ Jar tests were performed with Colorado River Water
106 (CRW) collected from the Central Arizona Project in Scottsdale (Arizona, USA). CRW has a
107 low watershed NDMA formation potential and is useful to differentiate NDMA associated with
108 polymer addition. Coagulation conditions—20 mg/L alum dosages and 2 mg/L polyDADMAC
109 as active polymer—were selected to optimize particle and dissolved organic carbon removal and
110 closely mimic full-scale water treatment conditions at the point of CRW collection. Three
111 fractions of Clarifloc C308 polyDADMAC were utilized in the testing: (1) as-received, non-
112 fractioned polymer, (2) low molecular weight fraction (<10 kDa), and (3) purified
113 polyDADMAC (>10 kDa fraction). Turbidity in each jar was measured in triplicate. Supernatant
114 samples were filtered (Millipore GF/F filters) prior to water quality analysis and NDMAFP tests.

115 **Pilot-Scale Tests with Purified Cationic Polymers**

116 Briefly, a two-train pilot-scale facility (Houston, TX) that replicates key full-scale facility
117 operations (Figure S3) was operated using either the as-received C308 polyDADMAC or the
118 purified higher molecular weight fraction (>10 kDa) polyDADMAC as the coagulant aid.
119 Mimicking full-scale operations, the raw water consisted of 100% Trinity River water that was
120 chloraminated at the plant intake through addition of sodium hypochlorite at 2.5 to 3 mg/L and
121 liquid ammonium sulfate at 0.5 mg/L. Due to chloramination at the intake, NDMA could form
122 across the treatment train through reaction between chloramines and the native water matrix,
123 polymer added, or residual polymer (after sedimentation). The raw water was pumped into two
124 parallel treatment trains where each treatment train included mixing and flocculation tanks where
125 35 mg/L ferric sulfate was added with either an optimized polymer dose condition (2.6 mg/L

126 active polymer) or an overdosed polymer condition (5.2 mg/L active polymer). Following
127 flocculation, process waters in each train flowed over a weir and into an upflow sedimentation
128 tank. The settled water was lifted via pumping to two sand/anthracite filter columns per train.
129 Raw, settled, and filtered water samples were collected from each train. Samples were analyzed
130 for pH, temperature, turbidity, total chlorine, total ammonia, monochloramine residual, UV
131 absorbance at 254 nm, TOC, and NDMA. Note, NDMA was directly measured, NDMAFP was
132 not performed in pilot tests. With the exception of NDMA, there were no significant differences
133 ($p > 0.05$) in the water quality measurements between the two trains.

134 **NDMA Reactivity and Formation Potential Test Conditions**

135 To measure NDMAFP reactivity of polymer solutions were conducted using 10 mg/L active
136 polymer with ultrapure water, and TOC was measured. NDMAFP of dilute polymer solutions
137 and jar test effluent was measured in duplicate with 500 mL samples. These solutions were
138 buffered to pH 8 with 10 mM borate buffer before dosing to 18 mg/L as Cl_2 with pre-formed
139 monochloramine. Samples were sealed and left in the dark at room temperature for 72 hours,
140 after which the residual monochloramine was measured (Monochlor-F reagent Hach kit) and was
141 > 3.5 mg/L for all samples. Samples were immediately quenched with ascorbic acid and spiked
142 with 100 ng/L NDMA- d_6 internal standard. Samples were kept in the dark at 4° C until NDMA
143 extraction and analysis could be performed.

144 **Analytical Methods**

145 ***NDMA extraction and analysis.*** EPA method 521 was used for NDMA extraction and analysis.

146 ¹⁵ Briefly, 500 mL sample was passed through a Restek EPA 521 charcoal solid phase extraction
147 (SPE) cartridge after it was conditioned with dichloromethane (DCM), methanol, and high
148 performance liquid chromatography (HPLC) grade water. After the sample was loaded, it was

149 dried with nitrogen gas, and the organic compounds were eluted from the cartridge with DCM.
150 The DCM extracts were blown down with nitrogen gas to 1 mL, dried using anhydrous sodium
151 sulfate cartridges, and stored at -4° C. Extracts were analyzed with Agilent 6890N/5973 inert gas
152 chromatography–mass spectrometry (GC/MS) in positive chemical ionization mode with Agilent
153 DB-1701P (Santa Clara, CA) column, ammonia reagent gas, and He carrier gas. The column
154 oven temperature was 40 °C for 3 minutes, increased by 4 °C per minute to 80 °C, and increased
155 to 120 °C at 20 °C per minute. The column interface temperature was 200 °C. The mass selective
156 detector analyzed mass-to-charge 92 (NDMA + NH₄⁺) and 98 (NDMA-d₆ + NH₄⁺). Calibrations
157 were performed using a series of NDMA standards from 1 to 1000 µg/L, each with a 100 µg/L
158 NDMA-d₆ internal standard.

159 ***¹H-NMR spectroscopy.*** To prepare samples for nuclear magnetic resonance (NMR), polymer
160 solutions were freeze dried to remove all water. Samples were rapidly frozen in a -80 °C freezer
161 then dried using a LABCONO Free Zone freeze dryer, which kept samples at -57 °C and 0.47
162 mbar for 24-48 hours until all water was removed from the sample. All samples were prepared in
163 D₂O with 0.05 wt.% 3-(trimethylsilyl)propionic-2,2,3,3-d₄ acid and sodium salt (TSP) as an
164 internal reference. Proton NMR was performed on a Bruker 400 MHz spectrometer; a single
165 pulse experiment was conducted on all samples.

166 ***Basic water quality analysis.*** Standard methods (USEPA, APHA) were used for pH, TOC, UV-
167 254, bromide, turbidity, and other water quality parameters. Non-purgeable organic carbon
168 (NPOC) and total nitrogen (TN) were measured using a Shimadzu TOC-L analyser with TN unit,
169 after the pH of all samples, blanks, quality control samples, and calibration standards was
170 adjusted to below pH 4.5 with hydrochloric acid. NPOC calibration standards were made with
171 potassium hydrogen phthalate in ultrapure water in the range of 0.5 mg/L to 10 mg/L. TN

172 standards were made with potassium nitrate in ultrapure water in the range of 0.1 mg/L to 5.0
173 mg/L. Blanks and quality control standards were run every 15 samples. NPOC of blanks were
174 always measured as <0.5 mg/L, and quality control standards were within $\pm 15\%$ of expected
175 concentration. Turbidity was measured using a Hach nephelometer, and pH was measured using
176 an Omega bench-top probe and meter.

177 **Statistical Analysis**

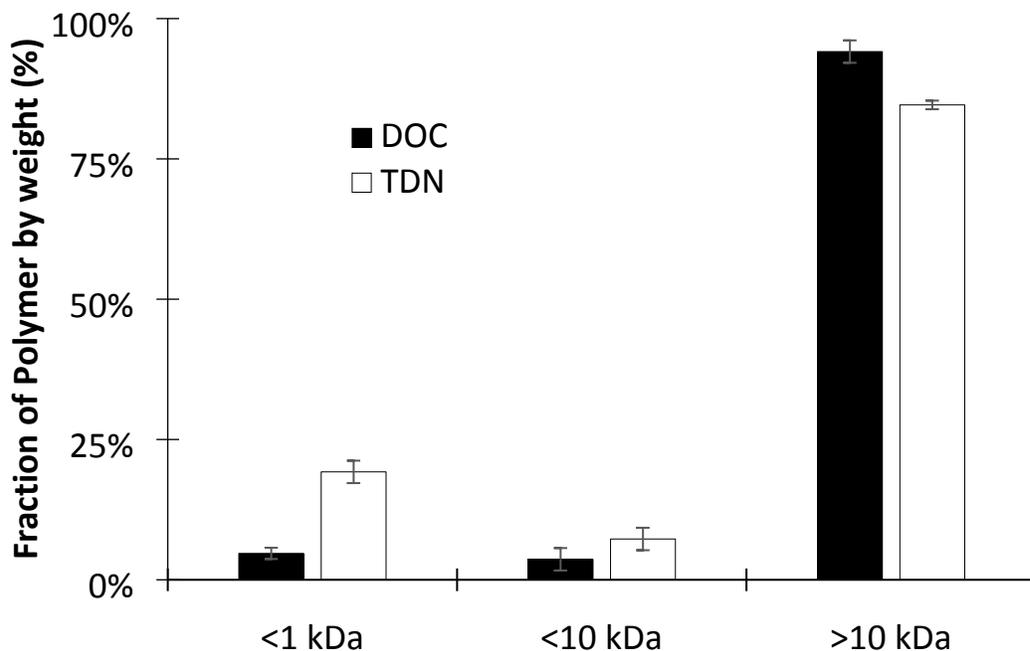
178 Statistical significance tests were performed to compare NDMA concentrations formed and
179 water quality under different test conditions. In general, a two-tail unpaired t-test assuming
180 unequal variances was used ($\mu_1 \neq \mu_2$). Differences were considered to be significant for $p < 0.05$.

181 **Results and Discussion**

182 **Fractionation, Characterization, and NDMAFP Reactivity of polyDADMAC Fractions**

183 Figure 1 shows DOC and TDN of the polyDADMAC after separation into three molecular
184 weight fractions. PolyDADMAC consists mostly of high molecular weight (>10 kDa) organics,
185 with only 5–10% of the constituents in the <10 kDa fraction. In separate ultrafiltration tests,
186 polymer material in the lowest molecular weight size fraction (< 1 kDa) was the most enriched in
187 nitrogen (higher TDN/DOC ratios) compared against the higher MW fractions; material in the
188 <1kDa fraction represents a subset of material in the separately performed <10kDa fraction.
189 Therefore, because there were only minor mass differences for the <10 kDa and <1 kDa fractions
190 and because the 10kDa ultrafilter had a much higher water permeate flux than the tighter
191 membrane (1 kDa), all further experiments used only the 10 kDa ultrafilter, thus limiting the
192 volume of purified polymer needed to meet the pilot-scale treatment polymer feed requirements.

193



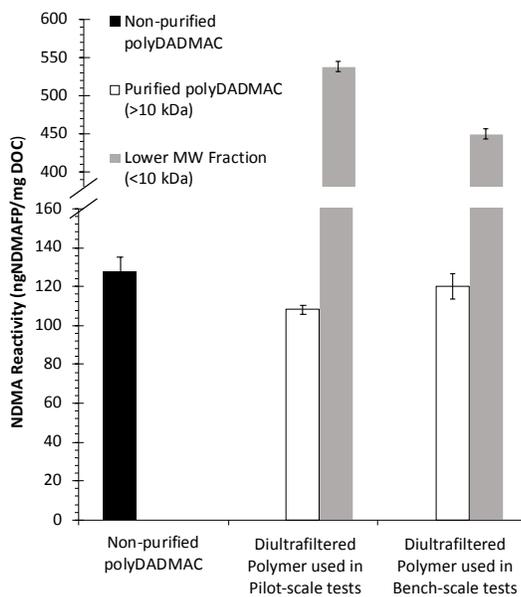
194

195 **Figure 1** The organic (DOC) and nitrogenous (TDN) mass proportions of polyDADMAC by
 196 molecular size, with <1 kDa and <10 kDa fractions representing lower molecular weight
 197 constituents and >10 kDa representing higher molecular weight constituents.

198

199 Figure 2 shows the NDMAFP reactivity of the different polyDADMAC fractions. Non-
 200 purified polyDADMAC exhibited only slightly higher ($p>0.05$) reactivity (130
 201 ngNDMAFP/mgDOC) than the >10 kDa fraction (115 to 120 ngNDMAFP/mgDOC) obtained
 202 from the dead-end and the recirculating diaultrafiltration processes. Note that the dead-end and
 203 recirculating diaultrafiltration processes had comparable reactivity, which is important because
 204 the recirculating system was needed to produce a large enough volume of purified polymer for
 205 the pilot tests. However, the low molecular weight (<10 kDa) fraction is four to five times more
 206 reactive (450 to 540 ngNDMAFP/mgDOC) than the non-purified polyDADMAC. In separate
 207 experiments with a polyamine cationic polymer, similar reductions in NDMA were observed
 208 comparing non-purified polymer solutions (720 ngNDMAFP/mg polymer) against 10 kDa

209 purified polyamine solutions (380 ngNDMAFP/mg polymer) (Figure S2). This demonstrates the
210 NDMA precursors are present in the <10 kDa fractions of different cationic polymers. While the
211 <10 kDa fraction accounts for <10% of the DOC in the non-purified polyDADMAC (Figure 1),
212 mass balance calculations suggest that the <10kDa fraction accounts for roughly 20 to 40
213 ngNDMA/mgDOC or 30-40% of NDMAFP in the non-purified polyDADMAC solution.

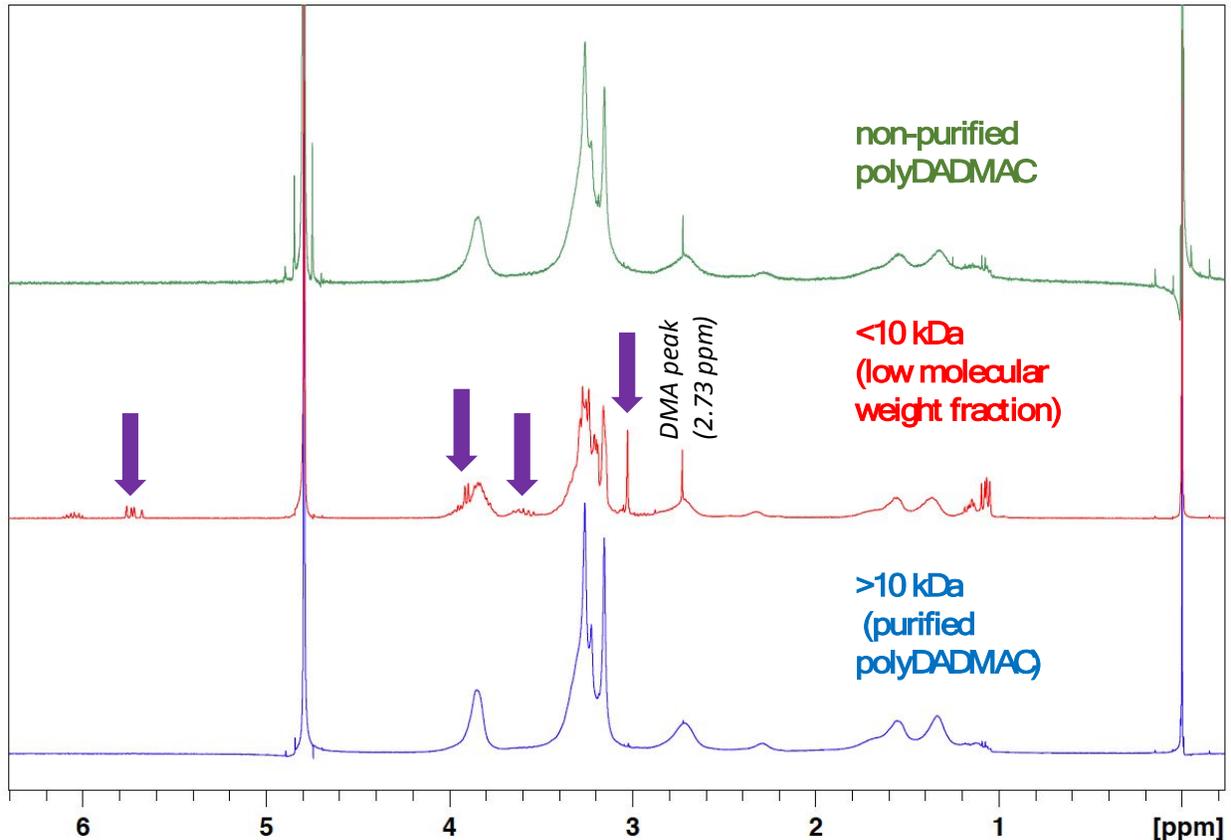


214
215 **Figure 2** NDMAFP reactivity of the non-purified C308 polyDADMAC compared to the lower
216 molecular weight fraction (<10 kDa) and purified (>10 kDa) polyDADMAC.

217
218 To identify why the NDMA FP varies in the different polyDADMAC fractions, ¹H-NMR was
219 performed. Figure 3 shows the ¹H-NMR spectra for purified and non-purified polyDADMAC. A
220 dimethylamine (DMA) peak at 2.73 ppm exists in the non-purified polyDADMAC, but not in the
221 purified >10 kDa polyDADMAC. The DMA peak and peaks at 1 and 3 ppm are higher in the <
222 10kDa than non-purified polymer. The peaks at 1 and 3 ppm are of unknown identity, but they
223 likely include more homogenous compounds than represented by broad NMR peaks and/or
224 represent molecules with higher density of functionalized end-groups. DMA is important

225 because it is thought to be a NDMA precursor.^{9,12,13} Manufacturers report DMA ranging from
226 0.005 to 0.025 wt% (50–250 mg/L) in commercial polyDADMAC. ¹H-NMR spectra indicated
227 that DMA was more prevalent (0.03–0.17%, 300–1700 mg/L) in the <10kDa fraction than the
228 non-purified polymer. Reported NDMA yields from DMA range between 0.2% and 1.5%;^{10,16}
229 which translates to a NDMA reactivity of 0.2– 6.2 ng NDMA per mg-C for the non-purified
230 polymer and 0.9–42 ng NDMA per mg-C for the <10 kDa fraction. Roughly 4% of the NDMA
231 could originate from DMA in the non-purified polymer, and 8% in the < 10 kDa fraction. As
232 supported by other studies,¹⁰ DMA does not *fully* explain the significantly higher reactivity of the
233 <10 kDa fraction of polyDADMAC. Furthermore, there are fewer peaks in the >10kDa fraction.
234 The monomer DADMAC has also been reported by the manufacturer at concentrations up to
235 0.12% or 1200 mg/L in the polyDADMAC solutions, and peaks labeled with arrows in the <10
236 kDa fraction are indicative of the carbon double bonds in DADMAC. DADMAC is reported to
237 have a much lower formation potential than DMA¹⁰ and thus does not contribute substantially to
238 NDMA formation. Finally, in the <10 kDa fraction there appear to be more polymer endgroups
239 detected in the 1.0 to 1.2 ppm range. This makes sense as a greater number of end-groups,
240 relative to the total number of molecules in solution, are expected for shorter-chain length
241 organic molecules present in the <10 kDa fraction. Compared with data shown in Figure 2, while
242 DMA could account for a portion of the NDMA reactivity, the results suggest that other un-
243 identified polymer residuals also contribute to NDMA formation.

244



245

246 **Figure 3** ^1H NMR spectra of non-purified C308 polyDADMAC (top in green), the low
247 molecular weight <10 kDa fraction (middle in red), and the purified polyDADMAC >10 kDa
248 fraction (bottom in blue). The location of the DMA peak at 2.73 ppm is shaded in each spectrum.

249 Purple arrows in the low molecular weight <10 kDa fraction indicate peaks attributed to
250 DADMAC. Peaks between 1.0 and 1.2 ppm are associated with polymer endgroups.

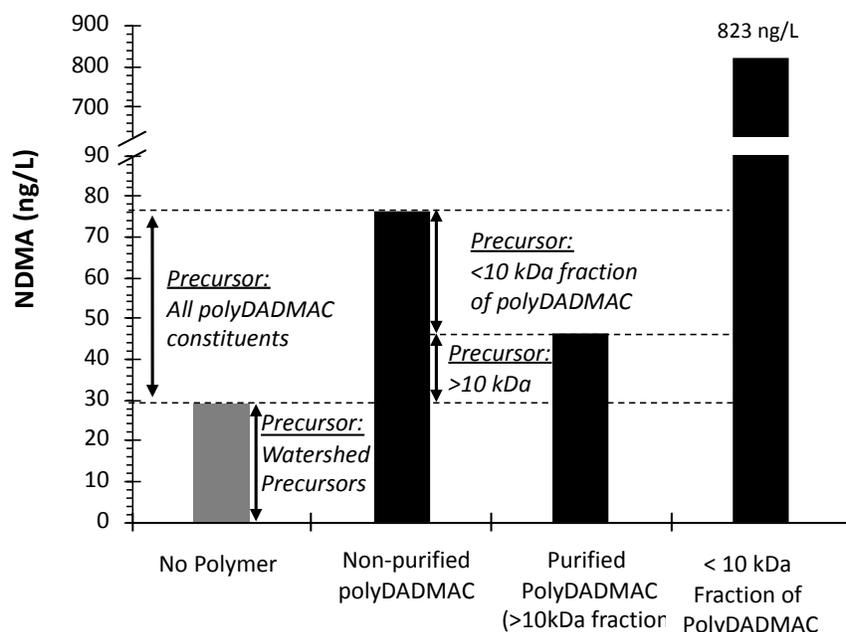
251

252 Contribution of PolyDADMAC Fractions to NDMA Formation During Jar tests

253 Jar tests without polymer and with the different polyDADMAC fractions were performed to
254 evaluate their contribution to NDMA formation in natural waters. Residual turbidity in treated
255 supernatant (i.e., after coagulation, flocculation, and sedimentation) was statistically ($p>0.05$) the

256 same for treatments using non-purified, purified (>10kDa) solution, and <10 kDa fraction
257 polyDADMAC (Figure S4) and was statistically better (>30% lower turbidity, $p<0.05$) compared
258 against alum coagulation alone. These results demonstrate that all the polymer fractions
259 enhanced particle removal.

260 Figure 4 shows NDMAFP in settled supernatant after jar testing. For alum coagulation alone,
261 the native water matrix precursors produced 29 ngNDMAFP/L. Coagulation using alum plus
262 non-purified polyDADMAC approximately doubled (1.6x) the NDMA formation, demonstrating
263 that polyDADMAC significantly ($p<0.05$) contributes to NDMA formation. Coagulation using
264 alum plus <10 kDa fraction of polyDADMAC formed 823 ngNDMAFP/L. This demonstrated
265 the potential for the lower molecular weight fraction of polyDADMAC to contribute NDMA
266 precursors during water treatment. Finally, coagulation was performed using alum coagulation
267 plus a purified polyDADMAC, where the low MW fraction (<10 kDa) was removed via
268 diaultrafiltration. The NDMA formation was 30 ng/L (62%) lower ($p<0.05$) than the non-purified
269 polyDADMAC (Figure 4). This work demonstrated the feasibility for purified polyDADMAC
270 (i.e., to remove <10 kDa fraction of polymer residuals) to achieve turbidity removal goals, and
271 significantly reduce polymer-derived NDMA precursors present after coagulation.
272



273

274 **Figure 4** NDMA formed due to background watershed organics plus polymer precursors in
 275 Central Arizona Project water after coagulation (ferric chloride alone (No polymer) or with
 276 different polyDADMAC fractions), flocculation, sedimentation, filtration, and chloramination
 277 (18 mgCl₂/L).

278

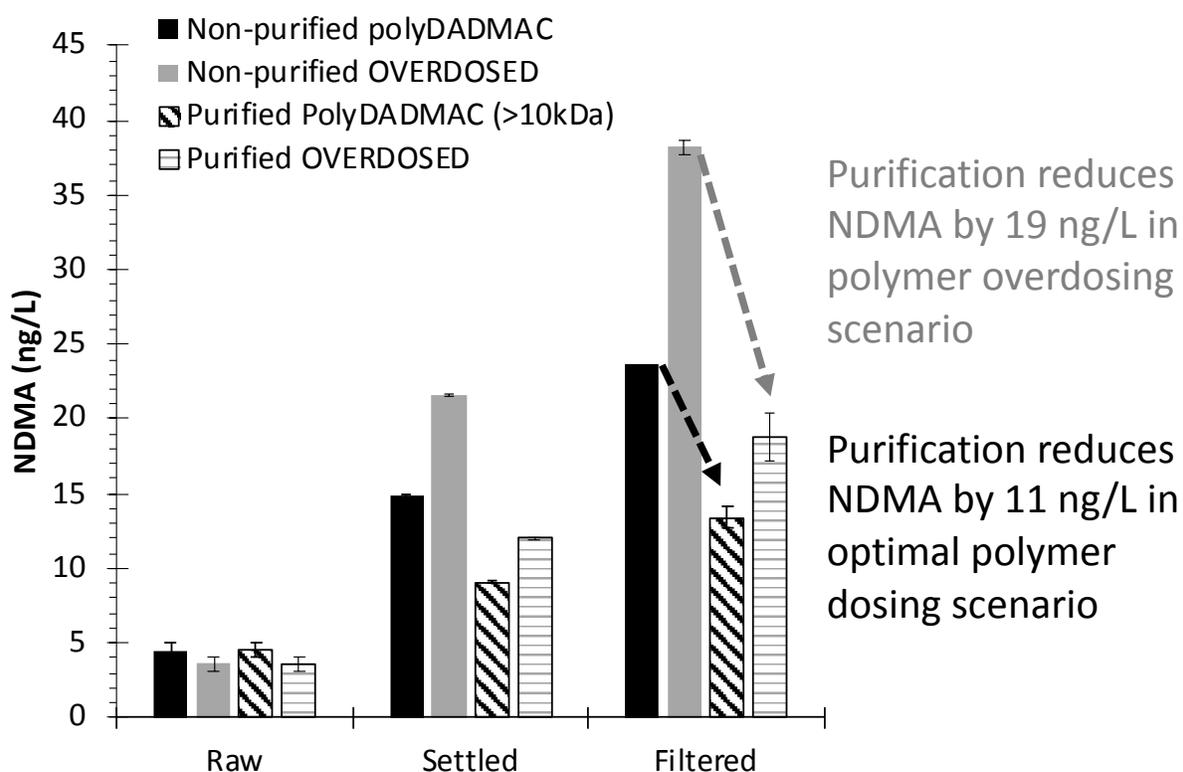
279 **Pilot Tests Confirmed Lower NDMA Formation using Diaultrafiltered PolyDADMAC**

280 To evaluate purified polyDADMAC (>10 kDa) to reduce NDMA formation under realistic
 281 continuous-flow drinking water treatment plant conditions, pilot tests that mimicked the full-
 282 scale treatment processes were performed. Figure S5 shows the DOC (5 mg/L) is lower in the
 283 settled waters than the raw water (6.2 mg/L) and that DOC was the same whether purified or
 284 non-purified polyDADMAC was added. This result demonstrates purified polyDADMAC was as
 285 effective as the non-purified polyDADMAC as a coagulant aid. Furthermore, DOC after
 286 filtration was lower when using the purified polymer compared against non-purified
 287 polyDADMAC.

288 Figure 5 shows NDMA concentrations directly measured across the pilot treatment trains (not
289 NDMAFP). Because there was a chloramine residual in raw water (see methods), NDMA
290 formed during coagulation and other processes. NDMA in the raw water was 4.5 ng/L. In the
291 presence of non-purified polyDADMAC, NDMA in settled and filtered waters was 15 and 24
292 ng/L, respectively. This illustrates a high contribution of NDMA from polyDADMAC. However,
293 using the diaultrafiltered polyDADMAC (i.e., <10 kDa fraction removed) the NDMA in settled
294 and filtered waters was 9 and 13 ng/L, respectively. Overall, the diaultrafiltered polyDADMAC
295 reduced NDMA contributed by the polymer by 53%.

296 To further explore the relationship between NDMA formation and cationic polymer coagulant
297 aids, pilot tests were run to compare optimal versus over-doses (2x optimal dose) of
298 polyDADMAC. Cationic polymer overdosing can occur at a treatment facility if dose is not
299 properly optimized or when the water flowrate varies but polymer feed pumps are not flow-
300 adjusted. All conditions achieved comparable turbidity and DOC removals (not shown). Raw
301 water NDMA averaged 4.5 ng/L throughout the tests. In agreement with trends for optimal
302 polymer dosing, Figure 5 also indicates that polymer overdosing always led to higher NDMA
303 across the pilot plant. Using purified (>10 kDa diaultrafiltered) polyDADMAC significantly
304 ($p < 0.05$) lowered NDMA compared against non-purified polymer. Polymer overdosing using the
305 purified polyDADMAC only increased NDMA by 40% compared against a 61% increase in
306 NDMA when non-purified polyDADMAC was overdosed. Optimal polymer dosing reduced
307 NDMA concentrations in filtered water by 10 ng/L, and 19 ng/L during the polymer overdosing
308 scenario. These are dramatic reductions in NDMA levels relative to existing standards. Clearly,
309 utilities should avoid cationic polymer overdosing as a strategy to reduce NDMA in finished
310 water. Additionally, polymer purification has substantial opportunities to limit NDMA formation

311 in finished water. Further optimization of the purification process can likely achieve even greater
312 reductions in low molecular weight NDMA precursors. While these results confirm the
313 dependence of NDMA on polymer addition, they also demonstrate the efficacy of using
314 diaultrafilter-purified polyDADMAC to reduce NDMA formation under actual operation
315 conditions.
316



317
318 **Figure 5** NDMA in the raw, settled, and filtered waters from pilot tests when optimized and
319 overdosed concentrations of non-purified polyDADMAC and purified polyDADMAC (>10 kDa)
320 were used.

321

322 **Other Methods of Purifying Cationic Polymer Stock Solutions**

323 Diaultrafiltration was selected for the bench- and pilot-scale testing because of its ability to
324 remove polymer residuals based upon molecular weight. However, other purification strategies
325 could be feasible. For example, we considered purging semivolatile compounds from
326 polyDADMAC polymer. While there was no detectable ammonia in the polymer solution, other
327 semivolatile compounds exist, namely DMA. DMA has a Henry's constant of 1.8×10^{-5} atm-
328 m^3/mol . Because the pK_a for DMA is 10.7, we adjusted the pH of the polyDADMAC solution to
329 10.0 and then evaluated NDMA formation without purging versus after purging for several hours
330 using a porous stone diffuser and an inert gas (N_2). Without N_2 purging, the polyDADMAC
331 solution produced 101 ± 2 ngNDMAFP/mg polymer (168 ± 3 ng NDMAFP/mg DOC). Purging
332 reduced formation to 56 ± 3 ngNDMAFP/mg of polymer (92 ± 5 ng NDMAFP/mg DOC), or a 45%
333 reduction in NDMA ($p < 0.05$). These results imply that volatile oligotrophic materials (e.g.,
334 DMA and others), likely of low molecular weight (given their volatility), in the non-purified
335 polyDADMAC solution represent a major portion of NDMA precursors. Thus, there may be
336 other ways to purify cationic polymer solutions to reduce their contribution towards NDMA
337 formation at water treatment facilities that use chloramine disinfectants.

338 **Summary and Conclusions**

339 Two major findings were demonstrated in bench-scale jar tests and confirmed in continuous
340 flow pilot-tests. First, cationic polymers contribute to NDMA formation at water treatment plants
341 that utilize chloramine disinfectants. Two source waters with relatively low watershed-precursor
342 NDMA levels were studied, and optimal or overdosing of cationic polymer more than doubled
343 amount of NDMA precursors present in coagulated-settled supernatant waters. This
344 demonstrated that a small portion of non-removable polymer residual was likely contributing to
345 NDMA formation. Mimicking full-scale conditions, pilot testing using raw water containing a

346 chloramine residual showed that NDMA formed during coagulation, where chloramines and
347 polymers were present. Additionally, NDMA formation continued even after sedimentation
348 where both non-removable polymer residuals and chloramines continued to react. These results
349 provide evidence that a portion of polyDADMAC (and polyamine) cationic polymers are
350 reactive in forming NDMA and are not removed during coagulation.

351 Second, for the first time we show that purifying commercial cationic polymers
352 (polyDADMAC or polyamine) to remove low molecular weight residuals can substantially
353 reduce NDMA precursors in the polymer solutions. Research focused on 10 kDa
354 dialtrafiltration to purify polyDADMAC at the bench- and pilot-scale, but the study also
355 illustrated that even alkaline pH adjustment and purging with an inert gas can be used to purify
356 cationic polymers of NDMA precursors. Dialtrafiltration purification of the cationic polymer
357 did not change its ability to aid in turbidity or DOC removal, but it significantly reduced NDMA
358 formation ($p < 0.05$) in settled supernatant water and during pilot-testing using raw water
359 containing chloramine residuals. This work demonstrates the potential for a custom-purified
360 cationic polymer solution, which could be readily manufactured or potentially prepared on-site at
361 water treatment plants, to significantly reduce NDMA formation from cationic polymer-derived
362 precursors.

363 **Conflicts of Interest**

364 There are no conflicts to declare.

365 **Acknowledgements**

366 This research was partially supported by the Water Research Foundation (Project #4622)
367 under the management of Djanette Khari. The authors thank Profs. David Hanigan and Dong An
368 for intellectual input and Laurel Passantino for technical editing.

369

References

- 370 1 H. R. Mian, G. J. Hu, K. Hewage, M. J. Rodriguez and R. Sadiq, *Water Res.*, 2018, **147**,
371 112–131.
- 372 2 S. W. Krasner, W. A. Mitch, D. L. Mccurry, D. Hanigan and P. Westerhoff, *Water Res.*,
373 2013, **47**, 4433–4450.
- 374 3 S. W. Krasner, P. Westerhoff, W. A. Mitch, D. Hanigan, D. L. McCurry and U. von
375 Gunten, *Environ. Sci. Water Res. Technol.*, 2018, **4**, 1966–1978.
- 376 4 P. Sgroi, M., Vagliasindi, F.G.A., Snyder, S.A. and Roccaro, *Chemosphere*, 2018, **191**,
377 685–703.
- 378 5 D. An, Y. Chen, B. Gu, P. Westerhoff, D. Hanigan, P. Herckes, N. Fischer, S. Donovan, J.
379 Philippe and A. Atkinson, *Water Res.*, 2019, **150**, 466–472.
- 380 6 D. A. Cornwell and R. A. Brown, *J. Am. Water Work. Assoc.*, 2017, **109**, E252–E264.
- 381 7 D. Hanigan, J. Zhang, P. Herckes, E. Zhu, S. Krasner and P. Westerhoff, *J. Am. Water*
382 *Work. Assoc.*, 2015, **107**, 90–91.
- 383 8 L. Padhye, Y. Luzinova, M. Cho, B. Mizaikoff, J. H. Kim and C. H. Huang, *Environ. Sci.*
384 *Technol.*, 2011, **45**, 4353–4359.
- 385 9 S. H. Park, L. P. Padhye, P. Wang, M. Cho, J.-H. Kim and C.-H. Huang, *J. Hazard.*
386 *Mater.*, 2015, **282**, 133–140.
- 387 10 S. Park, S. Wei, B. Mizaikoff, E. Taylor, C. Favero, C. Huang, S. Park, S. Wei, B.
388 Mizaikoff and A. E. Taylor, *Environ. Sci. Technol.*, 2009, **43**, 1360–1366.
- 389 11 D. Cornwell, S. Krasner, W. Mitch and J. Pignatello, *Investigating Coagulant Aid*

- 390 *Alternatives to PolyDADMAC Polymers*, Denver, CO, 2015.
- 391 12 K. D. Kohut and S. A. Andrews, *Water Qual. Res. J. Canada*, 2003, **38**, 719–735.
- 392 13 T. Zeng, R. J. Li and W. A. Mitch, *Environ. Sci. Technol.*, 2016, **50**, 4778–4787.
- 393 14 American Water Works Association, *Operational Control of Coagulation and Filtration*
394 *Processes*, 2011.
- 395 15 D. J. Munch and D. . Hautman, *Methods for the Determination of Organic Compounds in*
396 *Drinking Water*, Cincinnati, Ohio, 1995.
- 397 16 J. Choi and R. L. Valentine, *Water Res.*, 2002, **36**, 817–824.

399