



Cite this: *RSC Adv.*, 2019, 9, 25951

Characterisation of hexagonal birnessite with a new and rapid synthesis method—comparison with traditional synthesis†

Zhangjie Qin,  Xinmin Chen, Nanqi Ouyang, Shuai Lan, Guanjie Jiang, Junxia Zhang and Qin Zhang *

Birnessite is one of the most important manganese oxides that can control the geochemical behaviors of pollutants or can be applied to form industrial products. Many studies have been conducted on the synthesis of hexagonal birnessite because different synthesis methods can affect the structural, morphological, and physicochemical properties of hexagonal birnessite. However, there are still some defects in these synthesis methods. Therefore, a new synthesis method that is rapid, simple, and low-cost was proposed in this study involving the reduction of KMnO_4 by H_2O_2 in a H_2SO_4 solution without controlling the pH, temperature and pressure. Using a series of XRD, chemical composition, AOS, SSA, SEM, FTIR, and TGA analyses, Bir- H_2O_2 was found to have lower crystallinity than Bir-HCl. However, the AOS and SSA of Bir- H_2O_2 were 3.87 and $103 \text{ m}^2 \text{ g}^{-1}$ higher than those of Bir-HCl, *i.e.*, 3.70 and $22 \text{ m}^2 \text{ g}^{-1}$, respectively. Moreover, both Bir- H_2O_2 and Bir-HCl had similar particle morphology and thermal stability; in addition, the maximum adsorption content of Pb^{2+} on Bir- H_2O_2 ($\sim 3006 \text{ mmol kg}^{-1}$) was $\sim 30\%$ greater than that on Bir-HCl ($\sim 2285 \text{ mmol kg}^{-1}$) at pH 5.5; this indicated that the adsorption of Pb^{2+} on Bir- H_2O_2 was better and belonged to a pseudo-second-order model. All the abovementioned results indicate that Bir- H_2O_2 synthesized herein using the proposed synthesis method can have large application value.

Received 4th May 2019
Accepted 21st July 2019

DOI: 10.1039/c9ra03332g

rsc.li/rsc-advances

1. Introduction

Manganese (Mn) is the second most abundant transition metal next to iron on earth and resembles iron in several aspects.¹ Manganese oxides (including oxides, hydroxides, and oxyhydroxides) are widely distributed in the natural environment, controlling the biogeochemical cycles of organic and inorganic compounds.^{2–5} In the recent several decades, Mn oxides as important transition metal oxides have been applied in many fields, such as in ion-exchange, oxidation-reduction, catalysis, and energy conversion and storage, because of their high adsorption capacities;^{6,7} moreover, since the synthesis of Mn oxide materials is cheap and easy when compared with that of other noble-metal catalysts, significant attention has been paid to improve the performance of Mn oxide materials by changing their morphologies, dimensionality, size and so on.⁸

Based on the crystal structure of Mn oxides, they can be divided into two types: tunnel and layer structures.^{1,9} The layer

structure Mn oxides have higher cation-exchange capacity than the tunnel structure Mn oxides,^{10–12} especially birnessite; birnessite consists of edge-sharing MnO_6 that forms a layer structure with hydrated cations in the interlayer, which compensate the layer charge deficit *via* the substitution of Mn^{4+} by either low valent Mn or octahedral layer vacancies.^{13–16} It is commonly found as an alteration product in Mn-rich ore deposits. Furthermore, it can play a significant role in the oxidation-reduction and cation-exchange reactions in soil and water chemistry. Moreover, because of these capabilities, birnessite has been applied in several fields, *e.g.*, ion exchange, batteries and heterogeneous catalysis.^{17–21} However, all known natural birnessite samples are fine-grained and relatively poorly crystalline. Therefore, a large number of birnessites are synthesized in the lab or industry. The different performances of birnessites are closely related to the specific surface area (SSA) and average oxidation degree (AOS) of manganese, which can be affected by the synthetic conditions.

Currently, the direct redox reactions of Mn^{7+} or Mn^{2+} salts to prepare birnessite under hydrothermal conditions or at room temperature are preferred. Usually, the Mn^{2+} salts are oxidized by KMnO_4 or O_2 under alkaline conditions to prepare triclinic Na-birnessite or poorly crystalline $\delta\text{-MnO}_2$ with the morphology of hexagonal flakes or plate crystals;^{22–24} however, this method requires control over the selected temperature, gas bubbling

Key Laboratory of Poyang Lake Basin Agricultural Resource and Ecology of Jiangxi Province, College of Land Resource and Environment, Jiangxi Agricultural University, Nanchang 330045, China. E-mail: chincheung@live.com; Tel: +86 79183813024

† Electronic supplementary information (ESI) available. See DOI: 10.1039/c9ra03332g



conditions, and/or aging time (>4 h) and can easily generate other impurity oxides. Moreover, the reaction of KMnO_4 with hydrochloric acid is often used to synthesize hexagonal birnessite,^{25,26} which needs heating and/or reflux conditions; in addition, using KMnO_4 as the starting reagent, sol-gel processes that require long reaction times and high-temperature calcination have been proposed.²⁷ The application of the abovementioned synthesis methods is usually limited due to the requirements of long reaction times, high reaction temperatures, gas bubbling conditions, calcination, and/or complex synthesis procedures.

Consequently, it is necessary and appealing to find a new birnessite synthesis method that is a relatively simple synthesis procedure, cost-efficient, and environmentally friendly; the reduction of KMnO_4 by H_2O_2 to prepare $\alpha\text{-MnO}_2$ is an eco-friendly method, which simplifies the synthetic route and effectively improves the AOS and SSA.^{28–30} The only drawback of this method is long synthesis time. Therefore, in this study, a rapid one-step synthesis method was proposed using the reduction of KMnO_4 by H_2O_2 in a H_2SO_4 solution based on the abovementioned approach, which could rapidly produce birnessite with higher AOS and SSA than the traditional mild hydrothermal synthesis using KMnO_4 with hydrochloric acid. This new synthesis method has high potential for industrial application.

2. Materials and methods

2.1 Synthesis method of birnessite

2.1.1 Bir-HCl. Birnessite was prepared according to the McKenzie³¹ study: typically, 45 mL of 6 mol L^{-1} HCl was added dropwise to a boiling solution of 300 mL of 0.667 mol L^{-1} KMnO_4 for 30 min and then aged at 60 °C overnight. The product was washed with DDW, dried, powdered and named Bir-HCl.

2.1.2 Bir- H_2O_2 . A mixture of 100 mL of 12.7 mmol L^{-1} KMnO_4 and 0.35 mL 98% H_2SO_4 was stirred using a magnetic stirrer, and KMnO_4 and H_2O_2 at the molar ratios of 1 : 1.25, 1 : 1.5, and 1 : 1.75 were added to the abovementioned mixture solution at the rate of 5 mL min^{-1} . During the reaction, the temperature, pH, and pressure were not controlled. Then, the product was washed with DDW, dried, powdered and named Bir- H_2O_2 . XRD results of abovementioned three samples showed that the crystallinities of all the samples were similarly low (Fig. S1†). Therefore, we chose the product obtained with the molar ratio of 1 : 1.5 as a representative sample (named Bir- H_2O_2) to be compared with Bir-HCl.

2.2 Sample characterization

The obtained samples were characterized by powder XRD measurements performed *via* the Bruker D8 Advance diffractometer equipped with the LynxEye detector using Ni-filtered $\text{Cu K}\alpha$ radiation ($\lambda = 0.15418$ nm). The diffractometer was operated at the tube voltage of 40 kV and the current of 40 mA with a 12 s counting time per 0.02° 2θ step. The chemical composition of the samples was determined using atomic

absorption spectrometry (AAS, PerkinElmer AA900) and flame spectrometry (Sherwood Model 410) after dissolving 0.1000 g of sample in 25 mL of a mixed $\text{NH}_2\text{OH}\cdot\text{HCl}$ (0.25 mol L^{-1}) and H_2SO_4 (1 mol L^{-1}) solution. The Mn AOS was obtained by a titration method.³² A mass of 0.2000 g birnessite was completely reduced to Mn^{2+} in 5 mL of 0.5000 mol L^{-1} $\text{H}_2\text{C}_2\text{O}_4$ and 10 mL of 1 mol L^{-1} H_2SO_4 . Excess $\text{C}_2\text{O}_4^{2-}$ was determined by back-titration using a KMnO_4 standard solution at 75 °C. The specific surface area (SSA) was examined by nitrogen adsorption at liquid nitrogen temperatures using Micromeritics ASAP 2020. The samples were degassed at 110 °C for 3 h under vacuum prior to the adsorption measurement. Crystallite morphologies of the samples were probed using the FEI Nova NanoSEM 450 field emission-scanning electron microscope (FESEM) after being coated with a gold evaporated film. Thermogravimetric analysis (TGA) was carried out using the NETZSCH TG 209 thermal analysis system under a N_2 atmosphere at the heating rate of 10 °C min^{-1} . Fourier transform infrared (FTIR) spectra were acquired using the Bruker VERTEX 70 FTIR spectrometer (64 scans, 4 cm^{-1} resolution, and wavenumber range 4000–400 cm^{-1}).

2.3 Pb^{2+} adsorption experiment

The Pb^{2+} adsorption experiments were conducted at a constant supporting electrolyte concentration (NaNO_3 , $I_c = 0.1$ mol L^{-1}). An aliquot of 0–10 mL of 15 mmol L^{-1} $\text{Pb}(\text{NO}_3)_2$ ($\text{pH} = 5.00$ (4.50 ± 0.05)) was pipetted into a series of 50 mL polyethylene tubes; then, the NaNO_3 solution was added to fill the volume up to 10 mL. An aliquot of 5 mL of 1.67 g L^{-1} birnessite suspension, which was pre-equilibrated to the pH level of 4.5 (3.5, 5, and 5.5), was added to each tube, followed by shaking at 250 rpm for 24 h at 25 °C. The pH of the reaction system was maintained at 5.00 (4.50 ± 0.05) using a pH-stat technique. At the end of the reaction, the mixtures were centrifuged, and the supernatants were obtained for Pb^{2+} analyses by atomic absorption spectrometry (AAS, PerkinElmer AA900).³³

3. Results and discussion

The powder XRD patterns of the two synthesized samples show (Fig. 1) that the diffraction peaks agree well with those of turbostratic birnessite reported by Dirst *et al.* (1997). The Bir-HCl sample exhibits five obvious characteristic peaks, *i.e.* (001), (002), (11, 20), (22, 40), and (31, 02), whereas the Bir- H_2O_2 sample has only the first four broad characteristic peaks at similar peak positions. Moreover, all the peak intensities of the Bir-HCl sample are stronger than those of the Bir- H_2O_2 sample. This indicates that the particle crystallinity and/or size of Bir-HCl may be greater than that of Bir- H_2O_2 because of their different synthesis routes and conditions. The d value ratios of the (11, 20) to (22, 40) peaks in the XRD patterns of the two samples are close to 1.73, illustrating their hexagonal symmetry. This symmetry is also confirmed by the symmetrical shape of the (31, 02) band for Bir-HCl.³⁴

For the as-synthesized birnessite, the chemical composition, Mn AOS, and SSA are shown in Table 1. The contents of Mn and



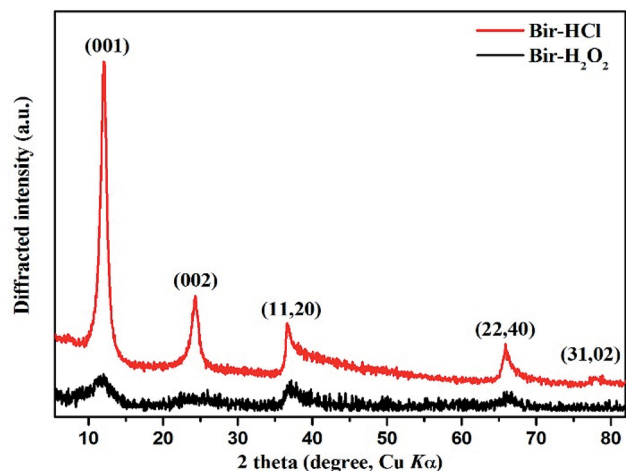


Fig. 1 Powder XRD patterns of the Bir-HCl and Bir-H₂O₂ samples.

Table 1 Mn AOS, SSA, and chemical composition of birnessite

Sample	Element (wt%)		Chemical components	Mn AOS	SSA (m ² g ⁻¹)
	Mn	K			
Bir-H ₂ O ₂	46.49	2.44	K _{0.07} MnO _{1.97} (H ₂ O) _{0.60}	3.87	103
Bir-HCl	3.37	5.81	K _{0.13} MnO _{1.96} (H ₂ O) _{0.50}	3.70	22

K are higher in Bir-HCl than those in Bir-H₂O₂, whereas the content of H₂O is lower in Bir-HCl than that in Bir-H₂O₂. Moreover, Bir-H₂O₂ has higher SSA, which is consistent with the results of XRD indicating its low crystallinity.³⁵ For this type of Bir-HCl, when the Mn AOS is higher, the SSA is always lower.³⁶ However, Bir-H₂O₂ has both bigger SSA and higher AOS, which is more favourable for the adsorption and oxidation of heavy metals and organic pollutants.

The typical morphologies of birnessite are shown in Fig. 2. It can be observed that three-dimensional hierarchical microspheres composed of two-dimensional disk-shaped pellets are

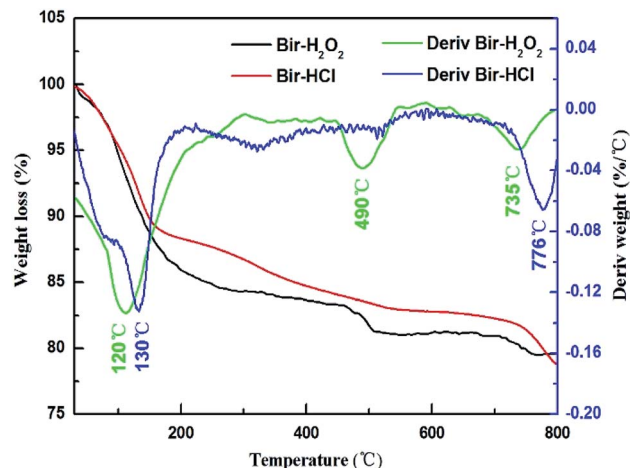


Fig. 3 Thermogravimetric analysis profiles of Bir-H₂O₂ and Bir-HCl.

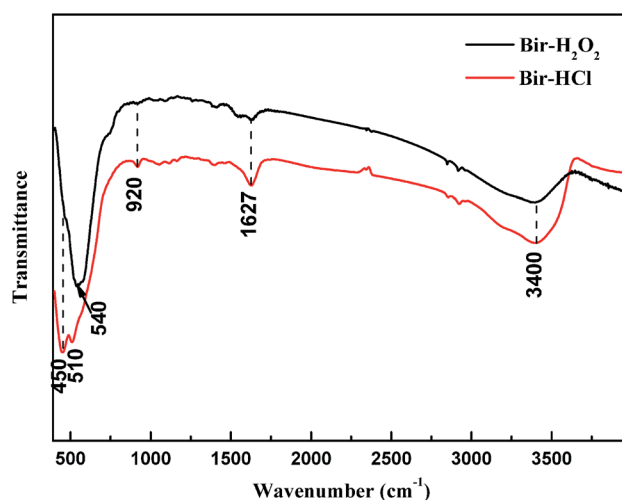


Fig. 4 FTIR spectra of the synthesized birnessite: Bir-H₂O₂ and Bir-HCl.

formed using the two synthetic routes. The diameters of the microspheres are ~500 nm for Bir-H₂O₂ and ~700 nm (or even greater) for Bir-HCl, and the diameters of the disk-shaped

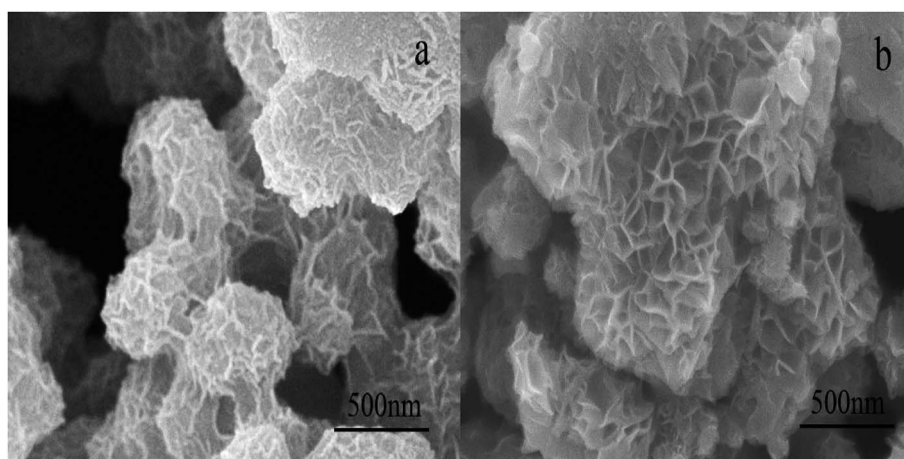


Fig. 2 SEM images of Bir-H₂O₂ (a) and Bir-HCl (b).



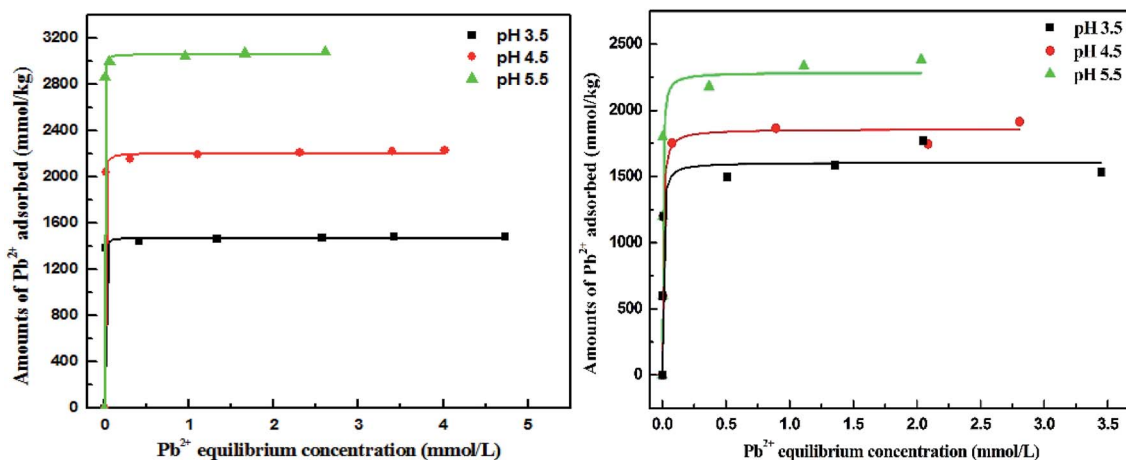


Fig. 5 Isothermal curves of Pb^{2+} uptake by $\text{Bir-H}_2\text{O}_2$ (right) and Bir-HCl (left) at pH 3.5, 4.5, and 5.5.

pellets are ~ 100 – 200 nm for $\text{Bir-H}_2\text{O}_2$ and ~ 150 – 300 nm for Bir-HCl . The particle sizes of both the three-dimensional hierarchical microspheres and the two-dimensional disk-shaped pellets in $\text{Bir-H}_2\text{O}_2$ are smaller than those in the case of Bir-HCl , which are consistent with the results of XRD and SSA.

The TGA data of the two samples are shown in Fig. 3. Compared to the case of Bir-HCl , the weight loss of $\text{Bir-H}_2\text{O}_2$ is faster, as shown in the TGA curve. $\text{Bir-H}_2\text{O}_2$ has three obvious weight loss peaks at 120°C , 490°C , and 735°C , and Bir-HCl has only two obvious weight loss peaks at 130°C and 776°C in the derivative weight curves, corresponding to three (or two) weight loss stages (physisorbed water, structural water, and lattice oxygen).³⁷ Moreover, the temperatures of the weight loss peaks of $\text{Bir-H}_2\text{O}_2$ are slightly lower than those in the case of Bir-HCl . This indicates that the thermal stability of $\text{Bir-H}_2\text{O}_2$ is less than that of Bir-HCl . This is due to the low crystallinity and small SSA for physisorbed water and interlayer water molecules.

Both the synthesized birnessite samples were characterized by FTIR spectroscopy, as shown in Fig. 4. The dominant adsorption peaks at 1627 and 3400 cm^{-1} are assigned to the stretching and bending vibration of crystal water and adsorbed water, respectively.³⁸ The adsorption band at 920 cm^{-1} is attributed to the vibration of Mn-OH located at vacant sites in the MnO_6 layer,³⁹ and the bands at ~ 450 , 510 , and 540 cm^{-1} are owing to the Mn-O lattice vibration of poorly ordered birnessite.^{40–42} This indicates that there are different disorders, such as lattice defects, in the lattices of $\text{Bir-H}_2\text{O}_2$ and Bir-HCl and structural distributions of low valence manganese.

Isothermal adsorption curves for Pb^{2+} on $\text{Bir-H}_2\text{O}_2$ and Bir-HCl are plotted in Fig. 5. The curves of Pb^{2+} adsorption on $\text{Bir-H}_2\text{O}_2$ and Bir-HCl conform to the L-type isotherm,⁴³ and the Pb^{2+} adsorption capacities increase with the increasing pH. For $\text{Bir-H}_2\text{O}_2$, the maximum Pb^{2+} adsorption capacity is 1471 , 2205 , and 3006 mmol kg^{-1} at pH 3.5, 4.5, and 5.5 obtained using the Langmuir fitting, as shown in Table 2, respectively. All values of R^2 in the Langmuir fitting results are 0.99, indicating that the fitting method is appropriate. As the main adsorption sites of Pb^{2+} on birnessite are vacancy sites and edge sites,^{20,44} an

increase in the pH can improve the content of both kinds of adsorption sites. In addition, for Bir-HCl , the maximum Pb^{2+} adsorption capacity is 1606 , 1858 , and 2285 mmol kg^{-1} at pH 3.5, 4.5, and 5.5, respectively (Table 2). Therefore, the maximum Pb^{2+} adsorption capacity for $\text{Bir-H}_2\text{O}_2$ is almost greater than that for Bir-HCl when the pH value is greater than 3.5; for example, the maximum Pb^{2+} adsorption capacity for $\text{Bir-H}_2\text{O}_2$ increases by $\sim 30\%$ than that for Bir-HCl at pH 5.5. Furthermore, the equilibrium constant k for $\text{Bir-H}_2\text{O}_2$ is greater than that for Bir-HCl . The abovementioned results can be caused by the difference in the SSA.

To further investigate whether the adsorption mechanism of Pb^{2+} on $\text{Bir-H}_2\text{O}_2$ is different from that on Bir-HCl at the maximum equilibrium constant k , the adsorption behaviors of Pb^{2+} on $\text{Bir-H}_2\text{O}_2$ at different times at pH 4.5 are shown in Fig. 6. The adsorption capacity of $\text{Bir-H}_2\text{O}_2$ towards Pb^{2+} sharply increases within ~ 1 h, and almost no further increase occurs after ~ 7 h (Fig. 6a). The Pb^{2+} adsorption rate is fast in the beginning 1 h owing to out-surface with a large number of adsorption sites, and just a little Pb^{2+} is needed to slowly enter the interlayer adsorption sites in the long time.⁴⁴ The pseudo-first-order and pseudo-second-order kinetic models were employed to interpret the adsorption mechanism. The equations of the two different models are expressed as follows:

The pseudo-first-order equation is

$$\ln(q_e - q_t) = \ln q_e - k_1 t \text{ or } q_t = q_e(1 - \exp(-k_1 t))$$

Table 2 Langmuir parameters for the adsorption of Pb^{2+} by $\text{Bir-H}_2\text{O}_2$ at different pH values and background electrolytes and Bir-HCl at pH 5

		$A_{\text{max}} (\text{mmol kg}^{-1})$	k	R^2
$\text{Bir-H}_2\text{O}_2$	pH = 3.5	1471	239	0.99
	pH = 4.5	2205	724	0.99
	pH = 5.5	3006	391	0.99
Bir-HCl	pH = 3.5	1606	207	0.76
	pH = 4.5	1858	181	0.95
	pH = 5.5	2285	288	0.91



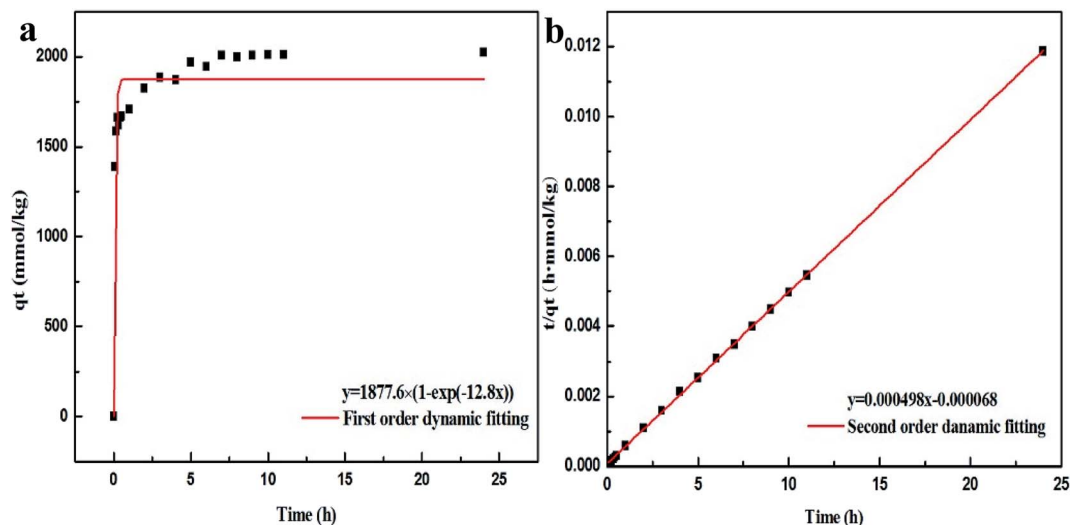


Fig. 6 Pb^{2+} adsorption kinetics on Bir- H_2O_2 : (a) first order dynamic fitting and (b) second order dynamic fitting.

Table 3 Kinetic parameters obtained from kinetic models (pH 4.5, 24 h)

	k (h^{-1} or $\text{kg} (\text{mmol}^{-1} \cdot \text{h}^{-1})$)	q_e (mmol kg^{-1})	R^2
Pseudo-first-order	12.802	1877.6	0.9059
Pseudo-second-order	0.00035	2032.1	0.9998

The pseudo-second-order equation is

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e}$$

where q_e and q_t are the adsorption amount (mmol kg^{-1}) at equilibrium and at time t , respectively. k_1 is the pseudo-first-order rate constant (h^{-1}), and k_2 is the pseudo-second-order equation rate constant ($\text{kg} (\text{mmol}^{-1} \text{h}^{-1})$).

The fitting results of the pseudo-first-order and pseudo-second-order models are shown in Fig. 6 and Table 3. It is obvious that the fitting curve using the pseudo-second-order equation is agreed better than using the pseudo-first-order equation (Fig. 6). Moreover, the fitting results of R^2 and q_e in the pseudo-first-order and pseudo-second-order are 0.9059 and 1877.6 mmol kg^{-1} and 0.9998 and 2032.1 mmol kg^{-1} , respectively, which the latter results show $R^2 > 0.99$ and the value of q_e closing to the maximum adsorption (2205 mmol kg^{-1}). This indicates that the adsorption of Pb^{2+} on birnessite follows the second-order model, which suggests that the rate limiting step can be chemisorption involving the exchange of H^+ or Mn^{2+} .¹¹

4. Conclusion

Compared to Bir-HCl obtained using the traditional synthesis method, Bir- H_2O_2 synthesized herein using a new synthesis method has many advantages in terms of material properties and Pb^{2+} adsorption. The lower crystallinity of Bir- H_2O_2 could

lead to high SSA. However, the AOS of Bir- H_2O_2 was greater than that of Bir-HCl. The relationship between the AOS and the SSA in Bir- H_2O_2 differs from that in the case of Bir-HCl. Therefore, Bir- H_2O_2 can have better prospects of application in environmental pollution control. Moreover, Bir- H_2O_2 has a similar particle morphology and thermal stability to Bir-HCl. The maximum equilibrium adsorption capacity of Bir- H_2O_2 for Pb^{2+} was 3006 mmol kg^{-1} at pH 5.5, which was greater by $\sim 30\%$ than that of Bir-HCl. The pseudo-second-order equation could be better fitted to the Pb^{2+} adsorption kinetics for Bir- H_2O_2 at pH 4.5 to explain the adsorption mechanism than the pseudo-first-order equation. All the abovementioned characteristics show that the new synthesis method that is low cost, rapid, and simple has a significant potential for application in the scientific studies, environmental remediation, and formation of industrial products such as hexagonal birnessite.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

The research was financially supported by the National Natural Science Foundations of China (Grants 41201225 and 41867004) and Invention Patent Industrialization Technology Demonstration Project of Jiangxi (20143BBM26104).

References

- 1 B. M. Tebo, J. R. Bargar, B. G. Clement, G. J. Dick, K. J. Murray, D. Parker, R. Verity and S. M. Webb, *Annu. Rev. Earth Planet. Sci.*, 2004, **32**, 287–328.
- 2 R. J. Casale, M. W. LeChevallier and F. W. Pontius, *Manganese control and related issues*, American Water Works Association, 2002.
- 3 A. T. Stone, *Environ. Sci. Technol.*, 1987, **21**, 979–988.



- 4 A. T. Stone and J. J. Morgan, *Environ. Sci. Technol.*, 1984, **18**, 617–624.
- 5 A. T. Stone and J. J. Morgan, *Environ. Sci. Technol.*, 1984, **18**, 450–456.
- 6 Z. Chen, Z. Jiao, D. Pan, Z. Li, M. Wu, C.-H. Shek, C. L. Wu and J. K. Lai, *Chem. Rev.*, 2012, **112**, 3833–3855.
- 7 D. Portehault, S. Cassaignon, E. Baudrin and J.-P. Jolivet, *Chem. Mater.*, 2007, **19**, 5410–5417.
- 8 L. Wang, Y. Ebina, K. Takada and T. Sasaki, *Chem. Commun.*, 2004, **9**, 1074–1075.
- 9 J. E. Post, *Proc. Natl. Acad. Sci. U. S. A.*, 1999, **96**, 3447–3454.
- 10 K. C. Kang, H. J. Jin, S. S. Kim, H. B. Min and S. W. Rhee, *J. Ind. Eng. Chem.*, 2011, **17**, 565–569.
- 11 H. Yin, W. Tan, L. Zheng, H. Cui, G. Qiu, F. Liu and X. Feng, *Geochim. Cosmochim. Acta*, 2012, **93**, 47–62.
- 12 H. Li, F. Liu, M. Zhu, X. Feng, J. Zhang and H. Yin, *J. Environ. Sci.*, 2015, **34**, 77–85.
- 13 R. Giovanoli, E. Stähli and W. Feitknecht, *Helv. Chim. Acta*, 1970, **53**, 209–220.
- 14 R. Giovanoli, E. Stähli and W. Feitknecht, *Helv. Chim. Acta*, 1970, **53**, 453–464.
- 15 R. G. Burns, V. M. Burns and A. J. Easton, *Philos. Trans. R. Soc. A*, 1977, **286**, 283–301.
- 16 E. D. Glover, *Am. Mineral.*, 1977, **62**, 278–285.
- 17 J. E. Post and D. R. Veblen, *Am. Mineral.*, 1990, **75**, 477–489.
- 18 S. Bach, J. P. Pereiramos and N. Baffier, *Electrochim. Acta*, 1993, **38**, 1695–1698.
- 19 Y. F. Shen, S. L. Suib and C. L. O'Young, *J. Catal.*, 1996, **161**, 115–122.
- 20 Y. Wang, X. Feng, M. Villalobos, W. Tan and F. Liu, *Chem. Geol.*, 2012, **292–293**, 25–34.
- 21 B. A. Manning, S. E. Fendorf, B. Bostick and D. L. Suarez, *Environ. Sci. Technol.*, 2002, **36**, 976.
- 22 J. Luo and S. L. Suib, *J. Phys. Chem. B*, 1997, **101**, 10403–10413.
- 23 J. Luo, Q. Zhang and S. L. Suib, *Inorg. Chem.*, 2000, **39**, 741–747.
- 24 R. Cornell and R. Giovanoli, *Clays Clay Miner.*, 1988, **36**, 249–257.
- 25 M. Tsuji, S. Komarneni, Y. Tamaura and M. Abe, *Mater. Res. Bull.*, 1992, **27**, 741–751.
- 26 F. Leroux, D. Guyomard and Y. Piffard, *Solid State Ionics*, 1995, **80**, 299–306.
- 27 S. Ching, D. J. Petrovay, M. L. Jorgensen and S. L. Suib, *Inorg. Chem.*, 1997, **36**, 883–890.
- 28 J. Villegas, L. J. Garces, S. Gomez, J. P. Durand and S. L. Suib, *Chem. Mater.*, 2005, **17**, 1910–1918.
- 29 Y. Yang, L. Xiao, Y. Zhao and F. Wang, *Int. J. Electrochem. Sci.*, 2008, **3**, 67–73.
- 30 J. Chen, X. Chen, Z. Xu, W. J. Xu, J. J. Li, H. P. Jia and J. Chen, *ChemistrySelect*, 2016, **1**, 4052–4056.
- 31 R. McKenzie, *Mineral. Mag.*, 1971, **38**, 493–502.
- 32 N. Kijima, H. Yasuda, T. Sato and Y. Yoshimura, *J. Solid State Chem.*, 2001, **159**, 94–102.
- 33 W. Zhao, H. Cui, F. Liu, W. Tan and X. Feng, *Clays Clay Miner.*, 2009, **57**, 513–520.
- 34 S. Grangeon, B. Lanson, N. Miyata, Y. Tani and A. Manceau, *Am. Mineral.*, 2010, **95**, 1608–1616.
- 35 H. Yin, H. Li, Y. Wang, M. Ginder-Vogel, G. Qiu, X. Feng, L. Zheng and F. Liu, *Chem. Geol.*, 2014, **381**, 10–20.
- 36 H. Yin, X. Wang, Z. Qin, M. Ginder-Vogel, S. Zhang, S. Jiang, F. Liu, S. Li, J. Zhang and Y. Wang, *J. Environ. Sci.*, 2018, **65**, 282–292.
- 37 S. H. Lee, T. W. Kim, D. H. Park, J. Choy, S. Hwang, N. Jiang, S. Park and Y. Lee, *Chem. Mater.*, 2007, **19**, 5010–5017.
- 38 J. G. Kim, J. B. Dixon, C. C. Chusuei and Y. Deng, *Soil Sci. Soc. Am. J.*, 2002, **66**, 306–315.
- 39 W. Zhao, F. Liu, X. Feng, W. Tan, G. Qiu and X. Chen, *Clay Miner.*, 2012, **47**, 191–204.
- 40 L. Kang, M. Zhang, Z. Liu and K. Ooi, *Spectrochim. Acta, Part A*, 2007, **67**, 864–869.
- 41 R. M. Potter and G. R. Rossman, *Am. Mineral.*, 1979, **64**, 1199–1218.
- 42 C. M. Julien, M. Massot and C. Poinsignon, *Spectrochim. Acta, Part A*, 2004, **60**, 689–700.
- 43 C. H. Giles, T. H. Macewan, S. N. Nakhwa and D. Smith, *J. Chem. Soc.*, 1960, **32**, 3973–3993.
- 44 M. Villalobos, J. Bargar and G. Sposito, *Environ. Sci. Technol.*, 2005, **39**, 569–576.

