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Preparation and characterization of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ magnetic visible light photocatalyst for water treatment

Congzhi Fu,  Xijun Liu,* Yuwei Wang, Li Li and Zihao Zhang

In this work, $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ magnetic photocatalyst was successfully prepared by a sol-gel method and a hydrothermal method. The crystalline structure and performance of the resulting catalyst have been characterized by scanning electron microscopy (SEM), transmission electron microscopy (TEM), X-ray diffraction (XRD), X-ray photoemission spectroscopy (XPS), Fourier transform infrared (FT-IR) spectroscopy and ultraviolet-visible light (UV-Vis) spectroscopy. The magnetic photocatalyst consists of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co}$ active particles and rGO carriers. The active particles have a double-shell core-shell structure with a size of about 500 nm and are supported on the rGO lamellae. TiO_2 doping with a small amount of metal Co and rGO can significantly improve the catalytic effect of magnetic photocatalyst, and rGO can also significantly improve the adsorption of pollutants by magnetic photocatalyst. The catalyst exhibited high photocatalytic activity in the degradation of methylene blue (MB) under visible light. 92.41% of this ability was retained after five times of repetitive use under the same conditions. The magnetic photocatalyst is easy to recover, and a recovery rate of 93.88% is still maintained after repeated use for 5 times.

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1. Introduction

Environmental pollution is considered to be the most serious and the most difficult problem that currently plagues human beings, and the problem of water pollution is even more important. Printing and dyeing wastewater is a kind of polluted wastewater with high chroma, high organic pollution content, complex composition, high chemical oxygen demand (COD) and high biological oxygen demand (BOD). It has seriously affected the quality of human life and the natural environment.¹ In fact, the water treatment includes catalytic reduction of pollutants²⁻⁵ and adsorption removal of pollutants.^{6,7} However, the adsorption method cannot completely remove pollutants, and there is a problem of secondary pollution, which requires further follow-up treatment. Studies have shown that photocatalytic oxidation has been considered as the most effective method for treating printing and dyeing wastewater, which has become the focus of research and development by scientists in various countries.⁸ Studies have shown that advanced oxidation processes (AOPs) as the most effective methods for degradation of organic pollutants at present.⁹ TiO_2 photocatalytic degradation is the most appropriate and available method of AOPs for the oxidation of organic compounds.¹⁰ Nano- TiO_2 has excellent photocatalytic activity,¹¹ and the photocatalytic efficiency is

dependent on the crystalline phase, particle size and specific surface area. The sol-gel method is an effective method for preparing nano- TiO_2 .^{12,13}

The application of TiO_2 in water treatment has been limited, due to the large band gap of TiO_2 (about 3.2 eV which can only use 3% of solar energy), recombination of generated electrons and also difficult recycle of TiO_2 particles. The researchers have combined TiO_2 with metals, doped with semiconductors with narrow band gap, and carbon-based nanomaterials with unique structures (such as carbon nanotubes, graphene and graphite oxide) are used for compounding in order to improve the photocatalytic characteristics of TiO_2 catalyst.¹⁴⁻¹⁸ At present, slight doping of TiO_2 with transition metals such as cobalt (Co^{2+}) has been widely investigated to show that the doped particle has a better photocatalytic ability.^{19,20} Not only can be the photocatalytic degradation efficiency of TiO_2 be improved, but also the band gap width of TiO_2 can be reduced so that it also has light response in the visible light region, thus improving the utilization rate of sunlight.²¹⁻²⁴ On the other hand, graphene, new nanostructure of carbon, has proper mobility of electrons, large specific area and special thermal and electrical conduction.^{25,26} In recent years, semiconductor-graphene has been widely investigated owing to the unique properties of graphene.²⁷ Graphene oxide (GO) is obtained by oxidation of graphene and contains a large amount of oxygen-containing functional groups on the surface, which allows the nanoparticles to successfully adhere to the surface.²⁸ The reduced graphene oxide (rGO) is obtained by high temperature reaction

College of Materials Science and Engineering, Qiqihar University, Heilongjiang, Qiqihar 161006, China. E-mail: 1104611198@qq.com; liuxijun2002@163.com; wyw032378@163.com; qqhrll@163.com; 570028946@qq.com



or the action of strong reducing agent.^{29,30} RGO can fill the shortcomings of low conductivity of TiO₂ and improve electrochemical performance when TiO₂ and rGO form composite material.³¹ RGO in composition with TiO₂ could act as an effective contaminant adsorbent and decrease electron-hole pairs recombination rate.³² Previous studies have shown that higher photocatalytic activity of these nanocomposites compared to pure TiO₂.³³

One of the other issues which can limit the application of TiO₂-Co/rGO photocatalyst is the separation of photocatalyst from reaction solution.^{34,35} Magnetic photocatalyst provides a practicable method for separating catalyst and aqueous solution, in a magnetic field. Hence, in this study, we prepared a novel photocatalyst of Fe₃O₄@SiO₂@TiO₂-Co supported by rGO. Provide magnetic properties to the catalyst by designing the Fe₃O₄ core. To prevent any reduction in catalytic effect of TiO₂ because of iron ion destroy the TiO₂ crystal structure, SiO₂ is utilized for encapsulation of magnetite.³⁶ Fe₃O₄@SiO₂@TiO₂-Co/rGO has the characteristics of good dispersibility, strong adsorption and easy recovery. Meanwhile, the photocatalytic performance of the novel photocatalyst was evaluated by degrading methylene blue (MB) under visible light. Fig. 1 shows the schematic of the formation of Fe₃O₄@SiO₂@TiO₂-Co/rGO nanocomposite.

2. Experimental

2.1 Reagents and apparatus

Reagent. Graphene oxide, Shenzhen Tuling Evolution Technology Co., Ltd.; Fe₃O₄, Luoyang Haorun Information Technology Co., Ltd.; tetrabutyl titanate (TBOT), tetraethyl orthosilicate (TEOS), methylene blue (MB), sodium dodecylbenzene sulfonate (SDBS), Co(NO₃)₂·6H₂O, NH₃·H₂O, HNO₃, HCl, C₂H₅OH, Tianjin Komiou Chemical Reagent Co., Ltd. All chemical reagents are of analytical grade, and them were used without further purification.

Apparatus. DZF-6090 vacuum drying oven, Shanghai Jinghong Experimental Equipment Co., Ltd.; YHG-9070A blast drying oven, Shanghai Yaoshi Instrument Equipment Factory; XQ-250E Ultrasonic Instrument, Kunshan Ultrasonic Instrument Co., Ltd.; TGL-16 High Speed Centrifuge, Yingyi Yuhua Instrument Factory, Gongyi City; SKGL-1200 Tube Furnace, Shanghai Precision Instrument Manufacturing Co., Ltd.; CEL-HXF300 Xenon Light Source System, Beijing Zhongjiao Jinyuan Technology Co., Ltd.; UV-7504 Ultraviolet Spectrophotometer, Shanghai Precision Instrument Co., Ltd.

2.2 The preparation of Fe₃O₄@SiO₂

Weigh Fe₃O₄ (0.3 g) into a 100 mL beaker, and dilute HCl (50 mL, 0.1 mol L⁻¹) was added for sonication for 15 min. Then, the Fe₃O₄ solid was magnetically separated and washed three times with deionized water. The magnetically separated Fe₃O₄ solid was put into a 250 mL three-necked flask, and deionized water (18 mL) and absolute ethanol (80 mL) were added. Then, NH₃·H₂O (2 mL) and TEOS (0.6 mL) was slowly added to the solution under stirring which continued for 12 h at room temperature. The Fe₃O₄@SiO₂ NPs were magnetically separated and washed three times with deionized water and absolute ethanol, respectively, and dried at 60 °C under for 12 h.

2.3 The preparation of Fe₃O₄@SiO₂@TiO₂-Co

Co(NO₃)₂ (0.08 g) was weighed into a 100 mL beaker, and then a liquid was prepared by sequentially adding H₂O (1.5 mL), HNO₃ (0.2 mL), and absolute ethanol (20 mL). Weigh Fe₃O₄@SiO₂ (0.2 g) in a 100 mL beaker, add absolute ethanol (20 mL) and glacial acetic acid (0.25 mL), then ultrasonically disperse for 30 min, then add TEOT (5 mL), and mechanically stir for 30 min in a 30 °C water bath to prepare liquid B. The solution A was slowly added dropwise to the solution B, stirred well until a gel was formed, and then aged at 30 °C for 18 h. The obtained gel was dried in an oven at 80 °C for 24 h, and then calcined at 450 °C

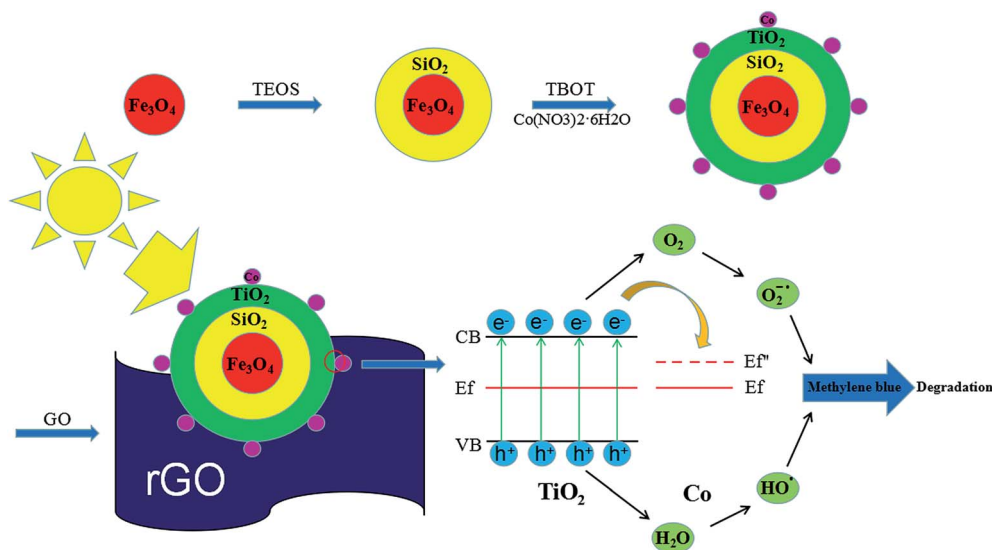


Fig. 1 Schematic diagram of synthesise catalysts process and mechanism of catalyst performance.



for 2 h under a nitrogen atmosphere to obtain Fe₃O₄@SiO₂@TiO₂-Co powder.

2.4 The preparation of Fe₃O₄@SiO₂@TiO₂-Co/rGO

Weigh graphene (0.08 g) oxide into a 100 mL beaker, add absolute ethanol (40 mL) and deionized water (20 mL) in turn, and disperse ultrasonically for 1 h. Then SDBS (0.15 g) and Fe₃O₄@SiO₂@TiO₂-Co (0.2 g) were added, and ultrasonic dispersion was continued for 1 h. The reaction solution was transferred to a 100 mL autoclave and placed in an oven at 120 °C for 3 h. The product was washed three times with absolute ethanol and deionized water, and then dried at 60 °C for 24 h to obtain Fe₃O₄@SiO₂@TiO₂-Co/rGO photocatalyst.

2.5 Characterization

The morphology and microstructure of the sample were observed by a scanning electron microscope (SEM, Hitachi S-3400), with an acceleration voltage of 20 kV. Prior to the analysis, the samples were coated with a thin layer of gold. The microstructure and size of sample have been characterized with transmission electron microscope (TEM, Hitachi H-7650), with an acceleration voltage of 100 kV. Dropping the sample dispersion liquid on a copper mesh coated with a carbon film, and drying for observation. X-ray diffraction diffractometer (XRD, Bruker-AXE D8 Advance) with an incident radiation of Cu K α (50 kV and 50 mA) and scanning range of 10–80° was used on the identification of the crystallographic phase of the sample. Sample chemical composition was confirmed by X-ray photoelectron spectrometry (XPS, Thermo Fisher EscaLab 250Xi). The chemical structure of the sample was analyzed by a Fourier transform infrared spectroscopy (FT-IR, PE Spectrum One), and the sample was mixed with KBr and then pressed for testing. The light response range of the sample was analyzed by an ultraviolet-visible-near-infrared spectrophotometer (UV-Vis-NIR PE Lambda 750).

2.6 Photocatalytic activity tests

Photocatalyst (0.3 g) was dispersed into 100 mL of MB aqueous solution with a concentration of 10 mg L⁻¹ in a beaker, and stirred in the dark for 20 min to reach an adsorption-desorption equilibrium. A 300 W Xe lamp was used as the visible light source (the photon flux of the radiation source is about 80 mW cm⁻²), place the beaker under the light source and adjust the distance between the light source and the liquid surface to 15 cm. During the stirring, 5 mL of suspension was taken every 20 min after commencing irradiation, and then centrifuged and analyzed using a UV-7504 spectrophotometer. Degradation rate (*D*) was used to measure the degree of degradation of MB (λ_{max} = 664 nm):

$$D = \frac{A_0 - A}{A_0} \times 100\% \quad (1)$$

where *A*₀ represents the absorbance of the MB solution before illumination, and *A* represents the absorbance of the MB solution after irradiation.

2.7 Photocatalyst recycling tests

Each photodegradation experiment takes 160 min to complete, and the photocatalyst is separated magnetically after the end of the experiment. Washing with deionized water for three times, drying at 80 °C for 2 h, weighing to calculate the recovery rate, and then carrying out a second photodegradation experiment, thus repeating the experiment for 5 times.

3. Results & discussion

3.1 X-ray diffraction patterns

Fig. 2 shows the XRD patterns of the produced photocatalyst. It is shown in Fig. 2a that the six peaks at 2 values of 30.0°, 35.3°, 42.9°, 53.5°, 57.0° and 62.4° are corresponding to (220), (311), (400), (422), (511) and (440) planes of the cubic phase of Fe₃O₄ (JCPDS no. 21-1272), respectively. The diffraction peak of Fe₃O₄@SiO₂ has no significant change compared with Fe₃O₄, which is mainly because of the coating layer SiO₂ belongs to an amorphous body and no new diffraction peak appears. It can be seen from Fig. 2b that the characteristic diffraction peak of Fe₃O₄ still exists, and the diffraction peaks of TiO₂ exhibiting characteristic peaks at $2\theta = 25.3^\circ$ (101), 37.7° (004), 47.9° (200), 54.1° (105), 55.2° (211), 62.7° (204), 68.7° (116) and 74.9° (215) corresponding to the anatase phase of TiO₂ (JCPDS no. 21-1272). This shows that the TiO₂ particles are successfully coated on the surface of Fe₃O₄@SiO₂ particles, and the diffraction peak of Fe₃O₄@SiO₂ inside the particles is seriously weakened due to the large amount of coating. In addition, the analysis of Fe₃O₄@SiO₂@TiO₂-Co/rGO spectra shows that the doping of metal Co and the loading of rGO have not cause the displacement of the characteristic diffraction peak of TiO₂, indicating that the metal Co and rGO have no effect on the crystal phase structure of TiO₂. At the same time, the diffraction peak of Co is not observed in the XRD pattern of the Fe₃O₄@SiO₂@TiO₂-Co/rGO. This may be due to the fact that the doping amount of Co is too small to be detected by XRD. The main diffraction peak of rGO appears at 25.3°, which overlaps exactly with the TiO₂ diffraction peak at the same angle.

3.2 SEM images analysis

The morphologies of the as-prepared Fe₃O₄, Fe₃O₄@SiO₂, Fe₃O₄@SiO₂@TiO₂-Co and Fe₃O₄@SiO₂@TiO₂-Co/rGO were evaluated by SEM (Fig. 3). In Fig. 3a, the Fe₃O₄ NPs are uniform with average particle size of approximately 400 nm and tetragonal body morphology. Comparing Fig. 3a and b, it is not difficult to find that the surface of Fe₃O₄@SiO₂ particles becomes much smoother and the edges and corners gradually become smoother. The particle size increased slightly, and the average thickness increased by about 20–30 nm, indicating that the surface of Fe₃O₄ was successfully coated with a layer of SiO₂. It can be seen from Fig. 3c and d that the surface of Fe₃O₄@SiO₂@TiO₂-Co has become very rough, indicating that the surface of Fe₃O₄@SiO₂ has successfully coated a layer of TiO₂-doped Co particles and successfully loaded on the rGO lamella, and the magnetic particles were dispersed evenly. It can be seen



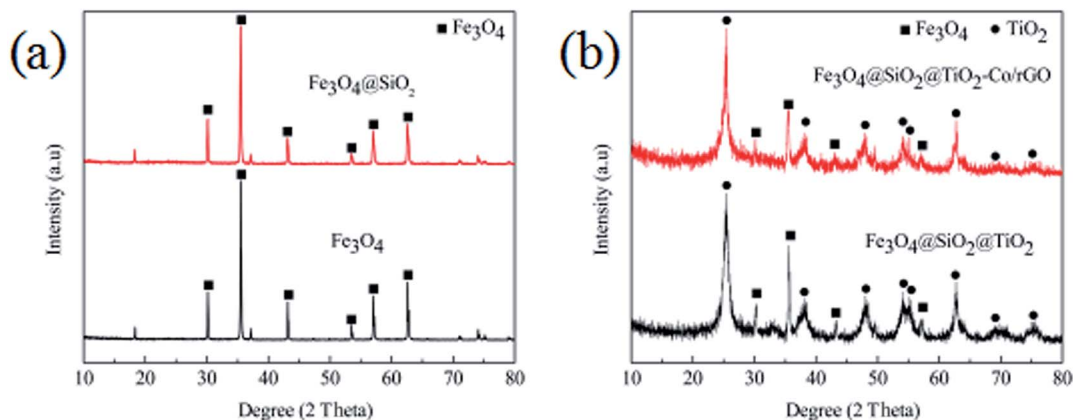


Fig. 2 (a) XRD patterns of Fe₃O₄ and Fe₃O₄@SiO₂, (b) XRD patterns of Fe₃O₄@SiO₂@TiO₂ and Fe₃O₄@SiO₂@TiO₂-Co/rGO.

from the figure that the Fe₃O₄@SiO₂@TiO₂-Co/rGO photocatalyst has been successfully prepared.

3.3 TEM images analysis

In order to observe the morphology of the magnetic powder more clearly, we conducted a transmission electron microscopy analysis of the sample in 3.3. It can be seen from Fig. 4a that the tetragonal form of the Fe₃O₄ magnetic powder is clearer. Referring to Fig. 4a and b, it can be clearly seen that Fe₃O₄@SiO₂ is coated with a layer of SiO₂ (light color part) with a thickness of about 20–30 nm to form a so-called core-shell structure. It can be seen from Fig. 4c and d that the TiO₂-doped Co particles are successfully coated on the surface of

Fe₃O₄@SiO₂ to form a double-shell core-shell structure. And successfully loaded onto the surface of rGO forms the quaternary Fe₃O₄@SiO₂@TiO₂-Co/rGO magnetic photocatalyst.

3.4 XPS analysis

To further confirm the chemical composition and purity of the as synthesized Fe₃O₄@SiO₂@TiO₂-Co/rGO, XPS analysis was carried out and its peak is shown in Fig. 5a. The XPS data show that the composition of surface elements is Ti (14.5 at%), Si (7.26 at%), Co (1.26 at%), Fe (0.51 at%), C (29.03 at%) and O (47.44 at%), which indicates that TiO₂, SiO₂, and Co are presented on the surface, and the C element is provided by rGO. This means that Co is successfully doped, because the amount

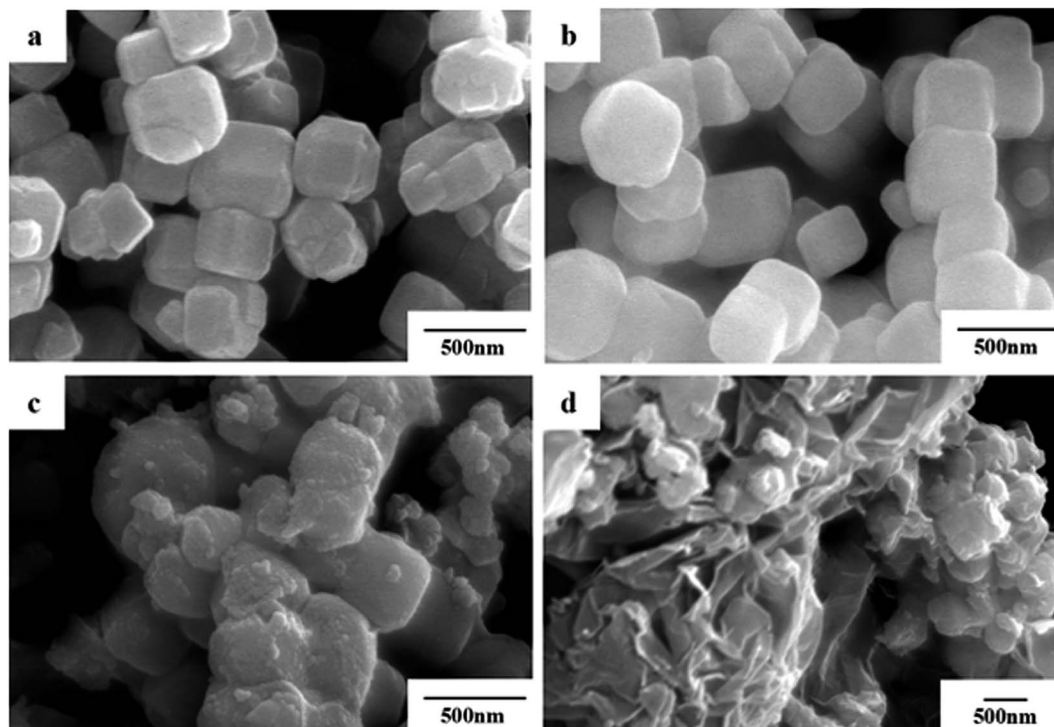


Fig. 3 SEM images of (a) Fe₃O₄, (b) Fe₃O₄@SiO₂, (c) Fe₃O₄@SiO₂@TiO₂-Co and (d) Fe₃O₄@SiO₂@TiO₂-Co/rGO.



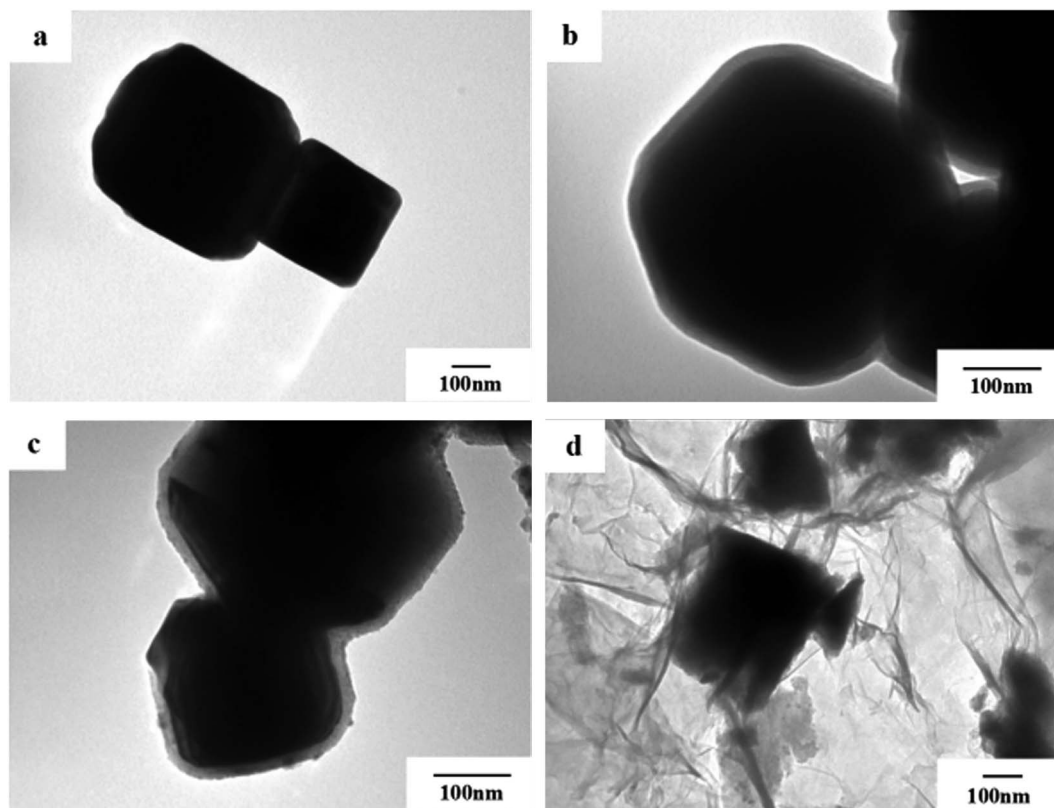


Fig. 4 TEM images of (a) Fe_3O_4 , (b) $\text{Fe}_3\text{O}_4@\text{SiO}_2$, (c) $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co}$ and (d) $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$.

of metal Co added is small, so the content of Co element is low. Compared with the intense Ti_{2p} and Si_{2p} peaks, the peak of Fe_{2p3} is almost undetectable, which confirms the Fe_3O_4 core is almost fully coated by SiO_2 and TiO_2 , resulting in low Fe content. This is consistent with SEM and TEM results and confirms that the structure of the synthesized sample is consistent with the assumptions.

It can be seen from Fig. 5b that the binding energies of the $\text{Ti}_{2p3/2}$ and $\text{Ti}_{2p1/2}$ peaks of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ are respectively at 458.2 eV and 463.8 eV. The binding energies of the $\text{Ti}_{2p3/2}$ and $\text{Ti}_{2p1/2}$ peaks of pure TiO_2 were respectively at 458.8 eV and 464.5 eV. Compared with the pure TiO_2 , the peaks of TiO_2 doped with metal Co are shifted to the direction of low

binding energy, and the interval between the two peaks decreases from 5.7 eV to 5.6 eV. This indicates that the doping of TiO_2 by the element Co causes the chemical environment of Ti to change, resulting in a chemical shift. The above results may indicate that part of Ti^{4+} is replaced by Co^{2+} to form Ti-O-Co structure.³⁷

3.5 FT-IR spectra analysis

The FT-IR spectrum is shown in Fig. 6. The absorption peaks at 1628 cm^{-1} and 3406 cm^{-1} are related to the bending and stretching vibration of the OH group, respectively. In $\text{Fe}_3\text{O}_4@\text{SiO}_2$, the absorption peaks at 567 cm^{-1} is correlated to the asymmetric stretching bonds of Fe-O, 1092 cm^{-1} and 795 cm^{-1}

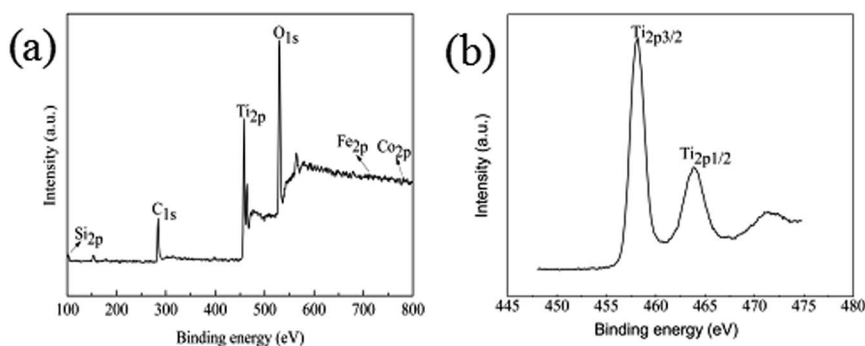


Fig. 5 (a) XPS spectrum of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$, (b) XPS survey spectrum of Ti_{2p} .



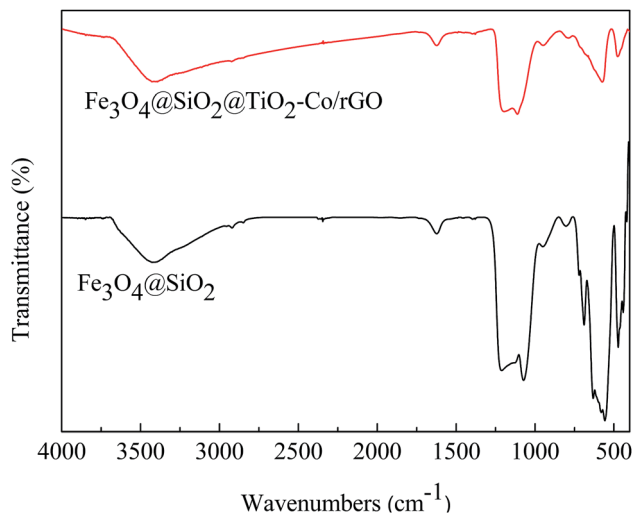


Fig. 6 FT-IR spectra of $\text{Fe}_3\text{O}_4@SiO_2$ and $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$.

corresponds to the asymmetric stretching vibration and symmetric stretching vibration of Si–O–Si. In $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$, the absorption peaks at 1093 cm^{-1} is correlated with the asymmetric vibration of Si–O–Si. Since the outer layer of $\text{Fe}_3\text{O}_4@SiO_2$ is coated with a TiO_2-Co layer, the infrared absorption intensity is lowered. Peaks in the range of $450-700\text{ cm}^{-1}$ were attributed to Ti–O–Ti and Ti–O–Si vibration. Signals in the range of $1400-1600\text{ cm}^{-1}$ were attributed to C=C bonding belongs to graphene. Since the amount of metal Co doping is small, no relevant infrared absorption is detected.

3.6 UV-Vis DRS analysis

DRS spectra provided information about the wavelength region in which the catalyst absorbs light (Fig. 7). It can be seen that pure TiO_2 is absorbed in the ultraviolet region and has almost

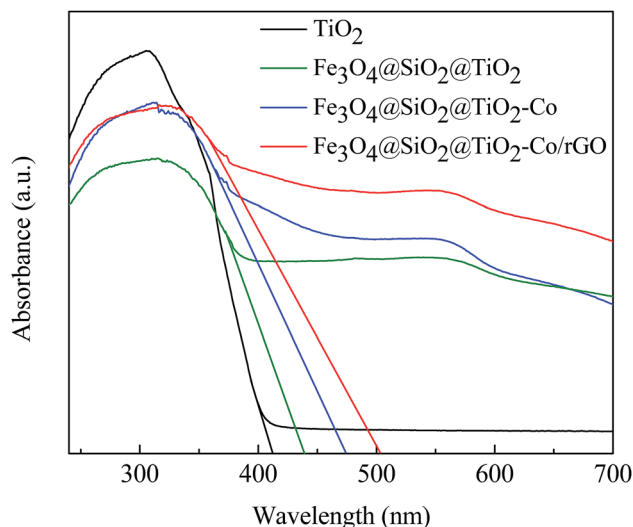


Fig. 7 UV-Vis DRS spectra of pure TiO_2 and $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$ photocatalysts.

no absorption in the visible region. The $\text{Fe}_3\text{O}_4@SiO_2@TiO_2$ magnetic photocatalyst not only absorbs in the ultraviolet region but also has weak absorption in the visible region. However, $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co$ and $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$ magnetic photocatalysts still have strong absorption in the visible region. This is because of the doping of Co reduces the forbidden bandwidth of TiO_2 , and the calculate the forbidden bandwidth of 2.62 and 2.47 eV by the equation $\lambda_g = 1240/E_g$. Similarly, the forbidden bandwidth of pure TiO_2 and $\text{Fe}_3\text{O}_4@SiO_2@TiO_2$ is 3.01 and 2.83 eV. The forbidden bandwidth of $\text{Fe}_3\text{O}_4@SiO_2@TiO_2$ magnetic photocatalyst is 0.18 eV lower than that of pure TiO_2 . The main reason is that SiO_2 and TiO_2 form a composite semiconductor, which suppresses the recombination of photogenerated electrons and holes, and improves the response range of TiO_2 to light. With the introduction of Co, the peak increases correspondingly, indicating that Co doping can promote the absorption of visible light by the photocatalyst. And loading of rGO also leads to the reduction of the forbidden bandwidth of TiO_2 , thus increasing the response range to light.³⁸ The experimental results show that $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$ magnetic photocatalyst can also realize the degradation treatment of organic wastewater under visible light irradiation.

3.7 Adsorption of MB onto catalyst

In order to evaluate the adsorption capacity of the magnetic photocatalyst, different situations of the catalyst were added to 10 mg L^{-1} MB solution stir for 40 min, in the dark environment, and then the residual concentrations of MB were measured using a spectrophotometer. Results as shown in Fig. 8a, the adsorption capacity of $\text{Fe}_3\text{O}_4@SiO_2$ was only 2%, and there was almost no adsorption. The adsorption capacity of the $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co$ increased to 12%, indicating that the specific surface area of the calcined TiO_2 was significantly increased. The adsorption amount of $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$ reaches 31%, which is because of the large specific surface of rGO with excellent adsorption performance. This is also an important factor to improve the photodegradation efficiency of pollutants.

The texture of the as-prepared photocatalyst was characterized by N_2 physisorption experiments, and the corresponding N_2 adsorption–desorption isotherms is shown in Fig. 8b. It can be seen that the $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$ photocatalyst have type IV isotherms (according to IUPAC classification), which suggests that it is mesoporous structures.³⁹ The specific surface area test results show that the specific surface area of the sample is $30.17\text{ m}^2\text{ g}^{-1}$, and the pore size distribution is mainly between 3.4 and 6.5 nm, which is mainly 4.9 nm mesopores.

3.8 Photocatalytic properties analysis

In order to explore the catalytic efficiency of magnetic photocatalyst, we chose MB as a degradation target for photocatalytic degradation experiments. Fig. 9 shows the photocatalytic degradation curves of MB by different samples under visible irradiation. After dark adsorption for 20 min and light for 160 min, the MB degradation of the blank sample is 1.30% while MB degradation of $\text{Fe}_3\text{O}_4@SiO_2$, TiO_2 , $\text{Fe}_3\text{O}_4@SiO_2@TiO_2$



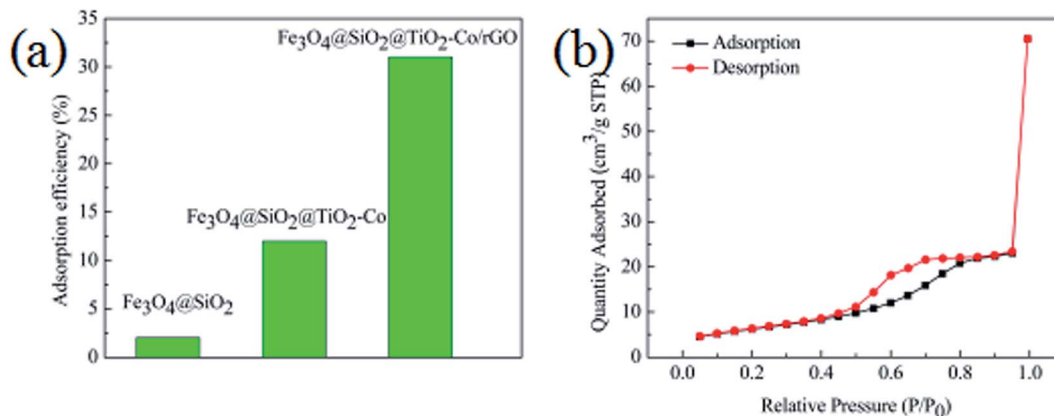


Fig. 8 (a) Adsorption efficiency of MB toward different catalysts in the dark (MB = 10 mg L⁻¹, catalyst loading = 0.3 g L⁻¹, contact time = 40 min), (b) nitrogen adsorption–desorption isotherms of Fe₃O₄@SiO₂@TiO₂-Co/rGO photocatalysts.

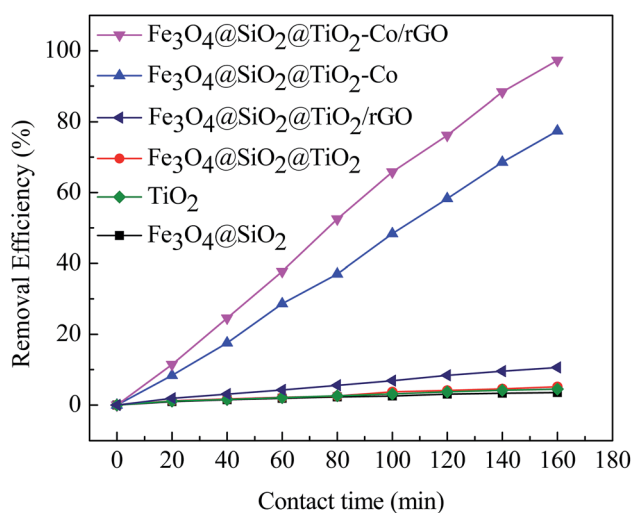


Fig. 9 Plot of percentage degradation of MB by different catalysts under visible irradiation.

and Fe₃O₄@SiO₂@TiO₂/rGO at 5.58%, 14.65%, 17.07% and 35.23%, respectively. It can be confirmed that these four samples have almost no photocatalytic activity under visible light irradiation. There is no photoactive component sensitive to visible light in these cases. When adding Fe₃O₄@SiO₂@TiO₂-Co or Fe₃O₄@SiO₂@TiO₂-Co/rGO as catalyst, an obvious photocatalytic degradation for MB is observed, and the percent degradation of MB dyes is significantly increased. The shows that after doping the metal Co, the response area of TiO₂ to light is broadened, which results in rapid degradation under visible light irradiation. In the presence of for Fe₃O₄@SiO₂@TiO₂-Co, the percent degradation was about 78.63% after 160 min. Under the same conditions, the percent degradation of Fe₃O₄@SiO₂@TiO₂-Co/rGO was 98.87%. This is because rGO has excellent conductivity. Valence electrons can move freely on the planar structure of rGO after introducing rGO as the carrier of magnetic photocatalyst. Finally, electron–holes are generated, which can significantly improve the catalytic efficiency of the photocatalyst.

3.9 Photocatalyst stability and reusability

In order to evaluate the stability and reusability of the Fe₃O₄@SiO₂@TiO₂-Co/rGO magnetic photocatalyst, related experiments were carried out, results as shown in Fig. 10a and b. After completing a photodegradation experiment, the magnetic photocatalyst can be quickly separated from the treated solution by an external magnetic field. It is proved that Fe₃O₄@SiO₂@TiO₂-Co/rGO magnetic catalyst has excellent magnetic properties and can completely realize the recycling of magnetic photocatalyst. The Fe₃O₄@SiO₂@TiO₂-Co/rGO magnetic photocatalyst was reused for five successive cycles, and degradation rate and recovery rate are shown in Fig. 10c. Photodegradation efficiency and recovery rate remained 92.41% and 93.88% after five cycles, respectively. This result approved that we synthesized Fe₃O₄@SiO₂@TiO₂-Co/rGO catalyst is a recyclable and high efficiency magnetic photocatalyst.

3.10 Principle analysis

From the results of the photocatalytic activity test and UV-Vis spectra for photocatalyst, we can conclude that SiO₂, Co and rGO has a significant effect on promoting TiO₂ photocatalytic efficiency. PL spectra can be used to study the charge migration, recombination and transfer during the photocatalytic process, and charge separation efficiency directly affects photocatalytic activity. To explore the effect of SiO₂, Co and rGO on promoting the separation of electron–hole pairs, PL emission spectra of TiO₂, Fe₃O₄@SiO₂@TiO₂, Fe₃O₄@SiO₂@TiO₂-Co and Fe₃O₄@SiO₂@TiO₂-Co/rGO were measured (Fig. 11a). Clearly, compared with the pure sample of TiO₂, the PL strength of Fe₃O₄@SiO₂@TiO₂, Fe₃O₄@SiO₂@TiO₂-Co and Fe₃O₄@SiO₂@TiO₂-Co/rGO samples decreased in turn, indicating SiO₂, Co and rGO can effectively separates the electron–hole pairs photo-generated by TiO₂. To obtain further evidence to support the analysis of PL spectra, the transient photocurrent experiments were also carried out. As shown in Fig. 11b, the TiO₂ pure sample had the weakest photocurrent response due to the quick recombination of photogenerated electron–hole pairs. After SiO₂, Co and rGO are added, the photocurrent density of



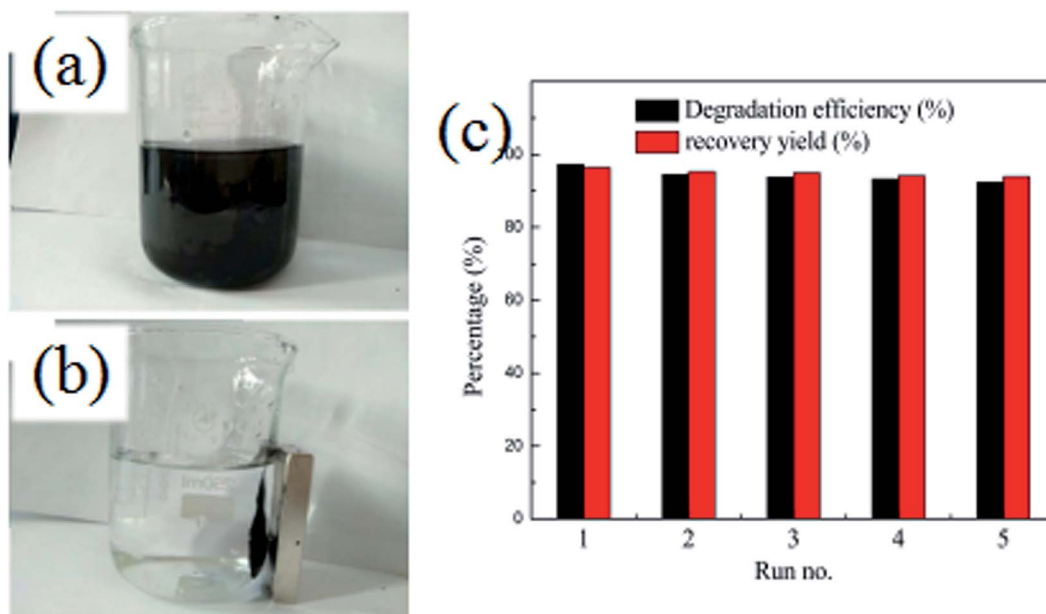


Fig. 10 (a and b) Separation of $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$ by a magnet, (c) recycling experiment for $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$.

photocatalyst is significantly enhanced. The result indicates that SiO_2 , Co and rGO can inhibit the recombination rate of electron and hole, and therefore enhance the charge separation efficiency, which is consistent with that of PL spectra.

EIS analysis can be used to study the interface migration and reaction capability of the charges in different photocatalytic materials. The EIS test was performed in a conventional three-electrode system with a platinum plate as the auxiliary

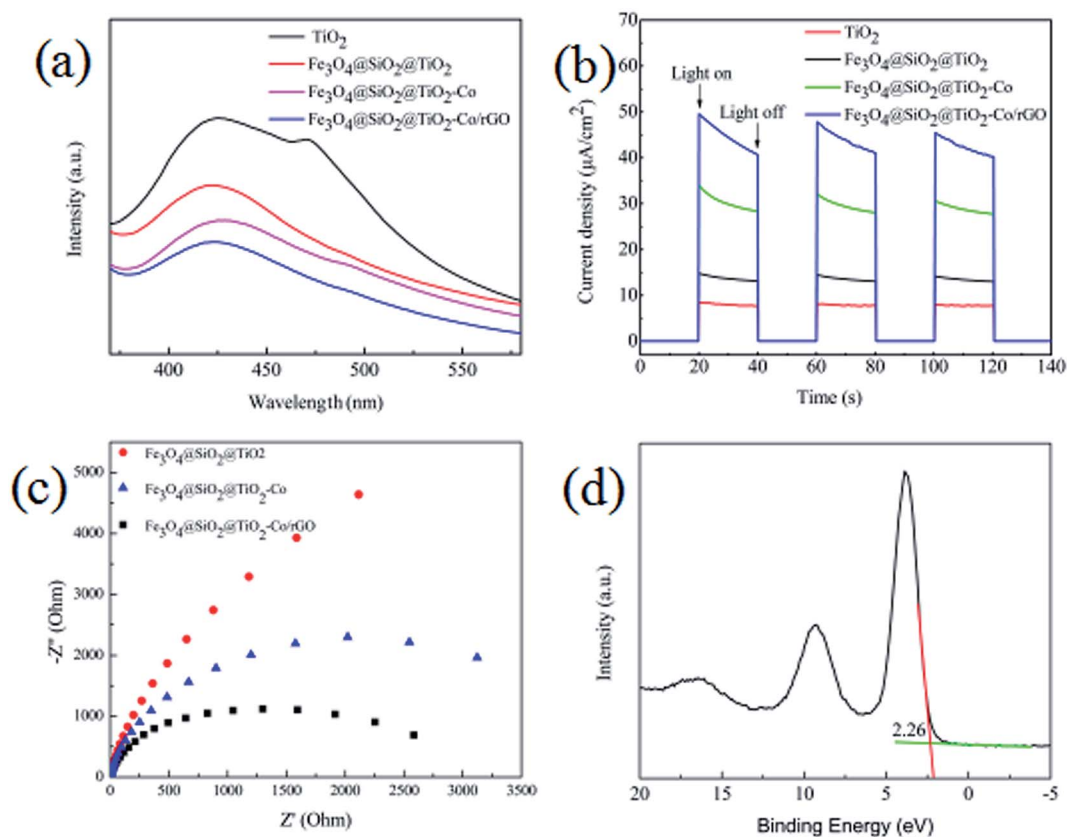


Fig. 11 (a) PL spectra of the as-prepared samples, (b) transient photocurrent response of the as-prepared samples, (c) EIS of the as-prepared samples, (d) XPS valence band spectra of $\text{Fe}_3\text{O}_4@SiO_2@TiO_2-Co/rGO$.



electrode and a saturated calomel electrode as the reference electrode. Fig. 11c shows the typical EIS Nyquist plots of different photocatalyst, the arc radius on the EIS plots of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co}$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ was smaller than that of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2$. It is well known that the smaller the radius is, the lower transfer resistance the interfacial electrons have. It means that the photocatalyst has a better transfer efficiency of photo generated electron-hole pairs. In addition, the arc radius on the EIS plots of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ was smaller than that of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co}$, suggesting that Co doping and rGO loading can made charge transfer easier. This corresponds to the photo-degradation efficiency curve shown in Fig. 9.

In addition, the valence band spectrum of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ can be obtained by XPS energy spectrum, which can explain the electronic structure principle more clearly. It can be seen from Fig. 11d that the valence band top (VB) of the photocatalyst is about 2.26 eV, and in combination with the UV-Vis spectrum result, the conduction band bottom (CB) of the sample can be calculated as -0.21 eV according to the formula $E_{\text{VB}} = E_{\text{CB}} + E_g$.

4. Conclusions

In summary, the recoverable $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ magnetic photocatalyst were successfully prepared by a sol-gel and hydrothermal method. The magnetic photocatalyst consists of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co}$ active particles and rGO carrier. The active particles have a double-shell core-shell structure with a size of about 400 nm and are supported on rGO lamellae. Co as a dopant and loading rGO significantly increases the photocatalytic activity of $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ catalyst. Moreover, rGO can significantly improve the adsorption of contaminants by magnetic photocatalysts and improve the catalytic efficiency. The magnetic photocatalyst shows excellent photocatalytic activity for MB solution under visible light irradiation and no obvious reduction in photocatalytic discoloration efficiency was observed after five cycles. With its low cost, high photocatalytic activity, high chemical stability and easy magnetic separation, $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2\text{-Co/rGO}$ magnetic photocatalyst has broad application prospects in large-scale photocatalytic wastewater treatment.

Conflicts of interest

There are no conflicts to declare.

Notes and references

- J. L. Xie, W. Hong, M. W. Meng, M. K. Tian, C. Y. Kang, Z. M. Zhou, C. S. Chen, Y. H. Tang and G. Y. Luo, *Z. Anorg. Allg. Chem.*, 2018, **23**, 1570–1575.
- G. F. Liao, J. Chen, W. G. Zeng, C. H. Yu, C. F. Yi and Z. S. Xu, *J. Phys. Chem. C*, 2016, **120**, 25935–25944.
- G. F. Liao, Q. Li, W. Z. Zhao, Q. H. Pang, H. Y. Gao and Z. S. Xu, *Appl. Catal., A*, 2018, **549**, 102–111.
- G. F. Liao, W. Z. Zhao, Q. Li, Q. H. Pang and Z. S. Xu, *Chem. Lett.*, 2017, **46**, 1631–1634.
- G. F. Liao, J. S. Fang, Q. Li, S. H. Li, Z. S. Xu and B. Z. Fang, *Nanoscale*, 2019, **11**, 7062–7096.
- X. Q. Qiao, F. C. Hu, F. Y. Tian, D. F. Hou and D. S. Li, *RSC Adv.*, 2016, **6**, 11631–11636.
- G. Y. Zhao, L. J. Liu, J. R. Li and Q. Liu, *J. Alloys Compd.*, 2016, **664**, 169–174.
- K. Huang and Z. Y. Cai, *Z. Anorg. Allg. Chem.*, 2018, **18**, 1072–1077.
- H. Suzuki, S. Araki and H. Yamamoto, *J. Water Process Eng.*, 2015, **7**, 54–60.
- H. Wang, H. L. Wang, W. F. Jiang and Z. Q. Li, *Water Res.*, 2009, **43**, 204–210.
- M. C. Wang, L. Peng, J. H. Wang, C. Q. Li, L. H. Guan and Y. Q. Lin, *J. Nanosci. Nanotechnol.*, 2018, **12**, 7372–7379.
- W. H. Dong, F. Pan, L. L. Xu, M. R. Zheng, C. H. Sow, K. Wu, G. Q. Xu and W. Chen, *Appl. Surf. Sci.*, 2015, **349**, 279–286.
- G. X. Zhang, A. K. Song, Y. M. Duan and S. L. Zheng, *Microporous Mesoporous Mater.*, 2018, **255**, 61–68.
- L. Zhang, Z. Wu, L. W. Chen, L. J. Zhang, X. L. Li, H. F. Xu, H. Y. Wang and G. Zhu, *Solid State Sci.*, 2016, **52**, 42–48.
- W. W. Zhong, W. G. Tu, S. S. Feng and A. J. Xu, *J. Alloys Compd.*, 2019, **772**, 669–674.
- W. W. Zhong, S. j. Shen, S. S. Feng, Z. P. Lin, Z. P. Wang and B. Z. Fang, *CrystEngComm*, 2018, **20**, 7851–7856.
- Y. P. Liu, S. J. Shen, J. T. Zhang, W. W. Zhong and X. H. Huang, *Appl. Surf. Sci.*, 2019, **478**, 762–769.
- W. W. Zhong, Z. P. Lin, S. S. Feng, D. Wang, S. J. Shen, Q. H. Zhang, L. Gu, Z. P. Wang and B. Z. Fang, *Nanoscale*, 2019, **11**, 4407–4413.
- H. Liu, X. N. Dong, X. C. Wang, C. C. Sun, J. Q. Li and Z. F. Zhu, *Chem. Eng. J.*, 2013, **230**, 279–285.
- J. G. Yu, Y. Wang and W. Xiao, *J. Mater. Chem. A*, 2013, **36**, 10727–10735.
- P. Zhou, J. G. Yu and Y. X. Wang, *Appl. Catal., B*, 2013, **142**, 45–53.
- G. F. Liao, J. S. Fang, Q. Li, S. H. Li, Z. S. Xu and B. Z. Fang, *Nanoscale*, 2019, **11**, 7062–7096.
- G. F. Liao, Y. Gong, L. Zhang, H. Y. Gao, G. J. Yang and B. Z. Fang, *Energy Environ. Sci.*, 2019, DOI: 10.1039/c9ee00717b.
- M. Hamadani, A. R. Vanani and A. Majedi, *J. Iran. Chem. Soc.*, 2010, **7**, 52–58.
- S. Mugundan, B. Rajamannan, G. Viruthagiri, N. Shanmugam, R. Gobi and P. Praveen, *Appl. Nanosci.*, 2015, **5**, 449–456.
- K. I. Bolotin, K. J. Sikes and Z. Jiang, *Solid State Commun.*, 2008, **146**, 351–355.
- L. L. Tan, S. P. Chai and A. R. Mohamed, *ChemSusChem*, 2012, **5**, 1868–1882.
- Y. Li, P. P. Zhang, Z. F. Ouyang, M. F. Zhang, Z. J. Lin, J. F. Li, Z. Q. Su and G. Wei, *Adv. Funct. Mater.*, 2016, **26**, 2122–2134.
- L. T. Soo, K. S. Loh, A. B. Mohamad, W. R. W. Daud and W. Y. Wong, *J. Power Sources*, 2016, **324**, 412–420.
- N. I. Zaaba, K. L. Foo, U. Hashim, S. J. Tan, W. W. Liu and C. H. Voon, *Procedia Eng.*, 2017, **184**, 469–477.



- 31 H. Chen, M. Q. Wang, Y. A. Yu, H. Liu, S. Y. Lu, S. Y. Bao and M. W. Xu, *ACS Appl. Mater. Interfaces*, 2017, **9**, 35040–35047.
- 32 H. Y. Shang, M. Ma, F. S. Liu, Z. Miao and A. P. Zhang, *J. Nanosci. Nanotechnol.*, 2019, **6**, 3376–3387.
- 33 P. Ribao, M. J. Rivero and I. Ortiz, *Environ. Sci. Pollut. Res.*, 2017, **24**, 12628–12637.
- 34 H. Khojasteh, M. Salavat-Niasari, M. P. Mazhar and M. Hamadani, *RSC Adv.*, 2016, **81**, 78043–78052.
- 35 Y. Zhou, Y. H. Zhu, X. L. Yang, J. F. Huang, W. Chen, X. M. Lv, C. Y. Li and C. Z. Li, *RSC Adv.*, 2015, **62**, 50454–50461.
- 36 X. Xu, F. Ji, Z. Fan and L. He, *Int. J. Environ. Res. Public Health*, 2011, **8**, 1258.
- 37 Y. D. Liu, F. Xin, F. M. Wang, S. X. Luo and X. H. Yin, *J. Alloys Compd.*, 2010, **498**, 179–184.
- 38 M. Teimouri, S. Waqif-Husain, M. Saber-Tehrani and P. Abroomand-Azar, *Int. J. Environ. Sci. Technol.*, 2016, **13**(12), 2797–2804.
- 39 K. S. W. Sing, R. A. W. Haul, L. Moscou, R. A. Pierotti, J. Rouquerol and T. Siemieniowska, *Pure Appl. Chem.*, 1985, **57**, 603–619.

