


 Cite this: *RSC Adv.*, 2021, 11, 4873

Robust immobilization of anionic silver nanoparticles on cellulose filter paper toward a low-cost point-of-use water disinfection system with improved anti-biofouling properties†

 Gongyan Liu,^{abc} Ruiquan Yu,^{abc} Jing Jiang,^a Zhuang Ding,^b Jing Ma^c and Ruifeng Liang^{ib}*^d

Silver nanoparticle (AgNP)-decorated cellulose filter paper (FP), a low-cost point-of-use (POU) water disinfection system, can supply affordable and safe drinking water for people in desperate need, especially in rural areas in developing countries. However, owing to the unstable immobilization of AgNPs, silver can leach into the treated drinking water from the FP and exceed the World Health Organization (WHO) drinking water limit ($<100 \mu\text{g L}^{-1}$), which is a potential threat to both human health and the environment. In this work, in order to robustly immobilize anionic silver nanoparticles (GA@AgNPs), we facilely prepared lipoic acid-modified cellulose filter paper (LA-FP), in which GA@AgNPs were robustly immobilized onto filter paper (GA@AgNPs-LA-FP) by strong chelation *via* the disulfide bond of LA with the surface of the silver nanoparticles. GA@AgNPs-LA-FP exhibited both excellent bacterial anti-adhesion activity and strong bactericidal activity, which can synergistically mitigate biofouling by inhibiting biofilm formation on the paper surface. Moreover, employed as a gravity-driven bactericidal filter, the GA@AgNPs-LA-FP membrane treated 100 mL of river water within 10 min, and the resulting water quality met the WHO drinking water standards, indicating this material's practical application for POU water disinfection.

 Received 27th October 2020
 Accepted 14th January 2021

DOI: 10.1039/d0ra09152a

rsc.li/rsc-advances

1. Introduction

A clean and safe drinking water supply is a cornerstone for human life.^{1–3} However, acute waterborne illnesses caused by pathogens in drinking water are still serious threats to human health, especially in rural areas in developing countries.^{4–6} Currently, over two million deaths a year, mostly of children, are caused by waterborne illnesses such as diarrhea, vomiting, and gastroenteritis.^{7,8} Therefore, point-of-use (POU) technologies

that can reduce pathogens in water sources have been recommended as effective solutions to provide safe drinking water.^{9,10}

Since the early twentieth century, filtration technologies capable of removing colloids, suspended solids, and pathogens from drinking water sources have been widely implemented in POU water treatment systems.^{11,12} The removal mechanism of filtration is size exclusion, so that microorganisms larger than the pore size of the filter can be removed from water sources.^{13,14} Particularly, membrane filters with dense porous structures and large surface areas are considered to be one of the most effective POU water treatment processes used in the past decade.^{15,16} For example, mature reverse osmosis (RO) membrane filtration has been used in households for drinking water purification. Owing to their extremely small pore sizes ($<1 \text{ nm}$), RO membranes provide an absolute barrier for microorganisms, keeping them in the water source and preventing waterborne diseases.¹⁷ However, operating the RO process requires an external driving force such as electrical pumping to achieve the desired flow rate.^{18,19} The requirements of a stable electricity supply and maintenance of RO water treatment systems increase the operation costs, making them impractical for use in developing countries.^{20,21} In addition, biofouling caused by accumulation of microorganisms is still a major challenge for the RO membrane filtration process, which not only decreases the performance of

^aCollege of Biomass Science and Engineering, Sichuan University, Chengdu 610065, China. E-mail: lrjcy@163.com

^bNational Engineering Research Center of Clean Technology in Leather Industry, Sichuan University, Chengdu 610065, China

^cThe Key Laboratory of Leather Chemistry and Engineering of Ministry of Education, Sichuan University, Chengdu 610065, China

^dThe State Key Laboratory of Hydraulic and Mountain River Engineering, Sichuan University, Chengdu 610065, China

† Electronic supplementary information (ESI) available: Schematic illustration of the synthetic route of GA@AgNPs; HRTEM image of GA@AgNPs; zeta potential of GA@AgNPs at different pH values; and particle size histograms of GA@AgNPs. Ag 3d core-level XPS spectra of the GA@AgNPs-LA-FP. ATR-FTIR spectra of pristine FP, LA-FP and GA@AgNPs-LA-FP. The cytotoxicity of GA@AgNPs-LA-FP. CLSM images of biofilm structures developed on FP and GA@AgNPs-LA-FP. Bactericidal stability test of the GA@AgNPs-LA-FP. See DOI: 10.1039/d0ra09152a



the membrane but also shortens its lifespan.^{22–24} Therefore, a low-cost and non-energy-driven membrane filtration system with anti-biofouling activity is considered to be a promising POU water treatment system and highly recommended for people living in rural areas.^{25–27}

Compared with RO membranes possessing dense nanopores, microfiltration (MF) membranes with microporous network structure (*i.e.* cellulose filter paper) have drawn increasing interest as an easy-to-use POU water treatment technology owing to their low cost, high water flux, and non-energy-driven properties.^{28–30} Despite MF membrane-based POU systems being able to remove colloids and suspended solids effectively from drinking water sources, pathogens are hardly blocked at all because of the micropores of MFs.^{31–33} For this reason, the quality of drinking water treated by conventional MF membranes generally fails to meet the drinking water guidelines set by the World Health Organization (WHO), which state that coliforms must not be detected.^{34–36} To address this issue, silver nanoparticles (AgNPs) have been widely used as disinfectants to decorate MF membranes because of their strong and broad-spectrum antibacterial activities.^{37–39} In contrast to conventional MF membranes, these AgNPs-decorated MF membranes can inactivate or kill microorganisms by releasing bactericidal Ag⁺ ions during gravity filtration, resulting in low-cost and efficient POU water disinfection devices to supply safe drinking water.^{40–42} However, the decoration strategies of AgNPs on/in MF membranes mainly depend on noncovalent interactions between membrane substrates and AgNPs, leading to weak immobilization of AgNPs.^{42–44} In most of these cases, the leaching of AgNPs into treated drinking water will be inevitable, which is a potential threat to both humans and the environment.^{45–47} Hence, the robust immobilization of AgNPs onto MF membranes is a primary consideration.

The aim of this work was to immobilize AgNPs on cellulose filter paper (FP) to produce a low-cost POU water disinfection system. In contrast with our previous work,^{40,45} to realize the robust immobilization of AgNPs onto FP, natural lipoic acid-modified cellulose filter paper (LA-FP) was first prepared through the esterification reaction between the carboxyl group (–COOH) of lipoic acid (LA) and primary alcohol groups (–OH) of β -glucose units on the filter paper (shown in Scheme 1).^{41,46} Then, anionic silver nanoparticles (GA@AgNPs) developed in our previous work were anchored onto LA-FP through chelation with the disulfide of LA bond (GA@AgNPs-LA-FP).^{41,42} The silver loading stability was investigated to determine whether the total leaching of Ag⁺ ions in the treated water met the WHO drinking water standard. Taking advantage of the bacterial anti-adhesion and bactericidal activities of immobilized anionic GA@AgNPs, a GA@AgNPs-LA-FP filter with improved antifouling activity exhibited excellent gravity-driven POU disinfection performance for both model bacteria-contaminated water and natural river water. This study proposes a low-cost, non-energy-driven, and anti-biofouling POU disinfection system to provide safe drinking water for people living in rural areas and prevent waterborne disease.

2. Experimental

2.1. Preparation of silver nanoparticles immobilized cellulose filter paper (GA@AgNPs-LA-FP)

First, gallic acid-protected silver nanoparticles (GA@AgNPs) were prepared according to our previous work.^{42,45} Then, cellulose filter paper was modified with lipoic acid. Briefly, 0.5 g of cellulose filter paper and 0.12 g of lipoic acid (0.58 mmol) were added into 20 mL dimethyl sulfoxide (DMSO) containing 4-dimethylaminopyridine (DMAP; 0.075 g, 0.61 mmol) and *N,N'*-dicyclohexylcarbodiimide (DCC; 0.075 g, 0.36 mmol). After 24 h reaction under N₂ protection at 35 °C (90 rpm), the cellulose paper was washed with ethanol and dried *in vacuo* to obtain lipoic acid-modified cellulose paper (LA-FP). Further, the LA-FP samples were then immersed in the GA@AgNPs solution, stirred for 10 min at room temperature and washed with deionized water to obtain GA@AgNPs immobilized cellulose paper (GA@AgNPs-LA-FP) with a brown color.

2.2. The silver loading content and the silver loading stability of GA@AgNPs-LA-FP

An inductively coupled plasma atomic emission spectrometer (model JY38S, France) was used to quantify the silver loading content of GA@AgNPs-LA-FP. Details can be found in the ESI.†

2.3. Antibacterial performance of GA@AgNPs-LA-FP

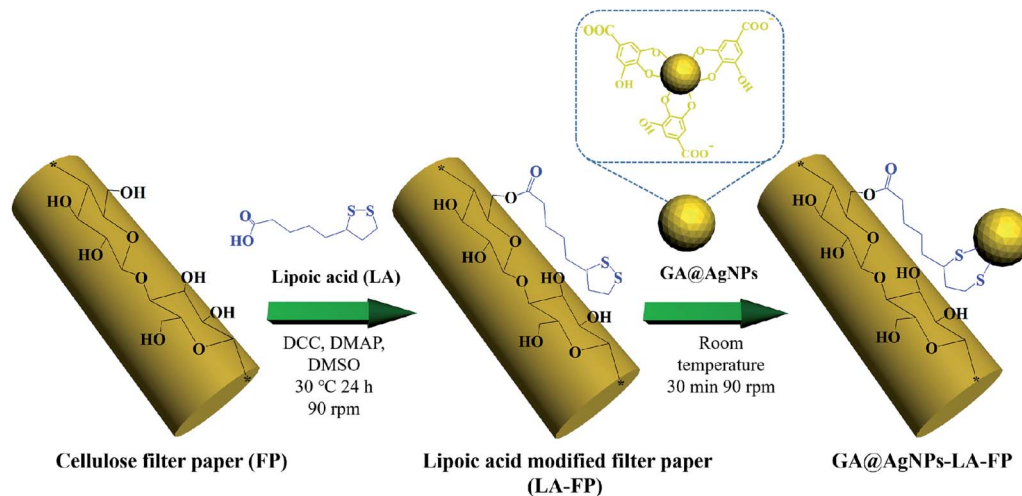
The bacterial anti-adhesion and bactericidal abilities of GA@AgNPs-LA-FP were tested using cellulose filter paper as control. Details can be found in the ESI.†

Then, a 0.1 g circular filter paper was rinsed with 0.04 mL of bacterial solution (containing *Escherichia coli* and *Staphylococcus aureus*; 10⁸ CFU mL^{−1}), placed in a Petri dish, and incubated at 37 °C for 24 h. After that, the filter paper specimen was washed with phosphate-buffered saline (PBS) three times. Last, bacterial cells were fixed on the filter paper surface with 3% glutaraldehyde solution for 4 h at 4 °C and treated with different concentrations of ethanol, followed by observation of scanning electron microscopy (SEM).⁴⁵

2.4. Antifouling property experiments and biofilm characterization

The filtration biofilm measurements were carried out according to previous work.⁴² First, circles with area 12.56 cm² were cut from FP and GA@AgNPs-LA-FP and fixed on a filter support. PBS solution (5 mM) with a pH of 7.4 was used as the medium for bacterial suspension during the filtration. *S. aureus* suspension was diluted in the above medium and cultivated to an OD₆₀₀ of 0.1 to reach a bacterial concentration of 10⁸ CFU L^{−1}. The mixture was used as the feedwater for the filtration test, and the dynamic biofouling experiments were performed using 2 L volumes of this feedwater. Before the permeation test, the pristine FP and GA@AgNPs-LA-FP were used to filter the PBS medium solution for 10 min to obtain initial flux measurements. Then, the permeation tests for GA@AgNPs-LA-FP and FP were performed at room temperature, and the weight increment





Scheme 1 Schematic illustration of the robust immobilization of GA@AgNPs on lipoic acid-modified filter paper surface.

of filtrate was continuously recorded with time by using an electronic balance (BSM-120.4, Zhuojing, China).²¹ At the end of the biofouling experiment, 1 cm² of filter paper was cut from the center of the biofouled paper and the biofilm was characterized by SEM.

2.5. Water disinfection performance of GA@AgNPs-LA-FP

2.5.1. Disinfection of model contaminated water. Bacterial strains of *E. coli* and *S. aureus* were used as model organisms for water contamination.⁴⁰ A 1 mL portion of bacteria suspension with a bacterial concentration of 10⁸ CFU mL⁻¹ was diluted to 100 mL with 99 mL sterile water, for use as a model of contaminated water. Then, this 100 mL of contaminated water was passed through GA@AgNPs-LA-FP under gravity. After filtration, 0.02 mL of the filtrate was collected and dripped on a nutrient agar (NA) plate and cultivated for 24 h to form bacterial colonies. Bacteria concentrations in contaminated water samples and the filtrate of the filter paper were calculated in the same way. In addition, the filtrate and contaminated water samples were stained using the LIVE/DEAD BacLight™ bacterial viability kit and observed by confocal laser scattering microscopy (CLSM).

2.5.2. Disinfection of natural river water. Natural contaminated river water samples are complicated: they contain inorganic substances and organic carbon sources, and are turbid, which could have negative impact on a POU water disinfection system. So, the practical disinfection capability of our system was tested by using natural river water samples from the Jinjiang River near Sichuan University in Chengdu, China. A circle of FP or GA@AgNPs-LA-FP with a diameter of 4 cm was placed in the filter, and then 1 L of natural river water was passed through the filter paper. During the gravity-driven filtration, the height of the water above GA@AgNPs-LA-FP was maintained at 10 cm to achieve constant hydraulic pressure. Coliform bacteria are an important indicator of the probable presence of enteric pathogens in drinking water resources, so the coliforms in the natural

river water samples were detected by coliform test plates before and after filtration.⁴¹

3. Results and discussion

3.1. Preparation and characterization of GA@AgNPs-LA-FP

As depicted in Scheme 1, in order to robustly immobilize silver nanoparticles onto cellulose filter paper (FP), lipoic acid was coupled onto cellulose fiber *via* the esterification reaction between the carboxyl group (–COOH) of lipoic acid and primary alcohol groups (–OH) of β -glucose units in the cellulose filter paper.^{44,46} Then, as-prepared GA@AgNPs (seen in Fig. S1†) were stably anchored onto LA-modified filter paper (LA-FP) through strong chelation of the disulfide bond and the surface of silver nanoparticles, forming GA@AgNPs immobilized cellulose filter paper (GA@AgNPs-LA-FP). Compared with pristine FP, which is white in color, the appearance of GA@AgNPs-LA-FP was dark brown, indicating the deposition of AgNPs on the paper surface (photographs in Fig. 1a and b). Moreover, the energy dispersive X-ray spectroscopy (EDS) Ag elemental mapping image of GA@AgNPs-LA-FP in Fig. 1b also demonstrates the successful and uniform deposition of GA@AgNPs. In addition, the SEM images in Fig. 1a and b show that the introduction of GA@AgNPs onto the paper surface did not affect the network structure of the cellulose filter paper, which suggests that GA@AgNPs-LA-FP can maintain the original performance of pristine filter paper. Further, plenty of tiny white nanoparticles with good dispersion can be observed in the high-magnification SEM image of GA@AgNPs-LA-FP (Fig. 1d). The large silver signal in the EDS spectrum confirmed that these particles were GA@AgNPs. By comparison with the integrated fibrous structure of the fresh filter paper shown in Fig. 1c, the cellulose fibers of GA@AgNPs-LA-FP can be seen to remain undamaged after AgNPs immobilization.

The robust immobilization process of GA@AgNPs on LA-FP was also investigated by X-ray photoelectron spectroscopy (XPS) analysis. As shown in Fig. 2a, only C 1s and O 1s signals



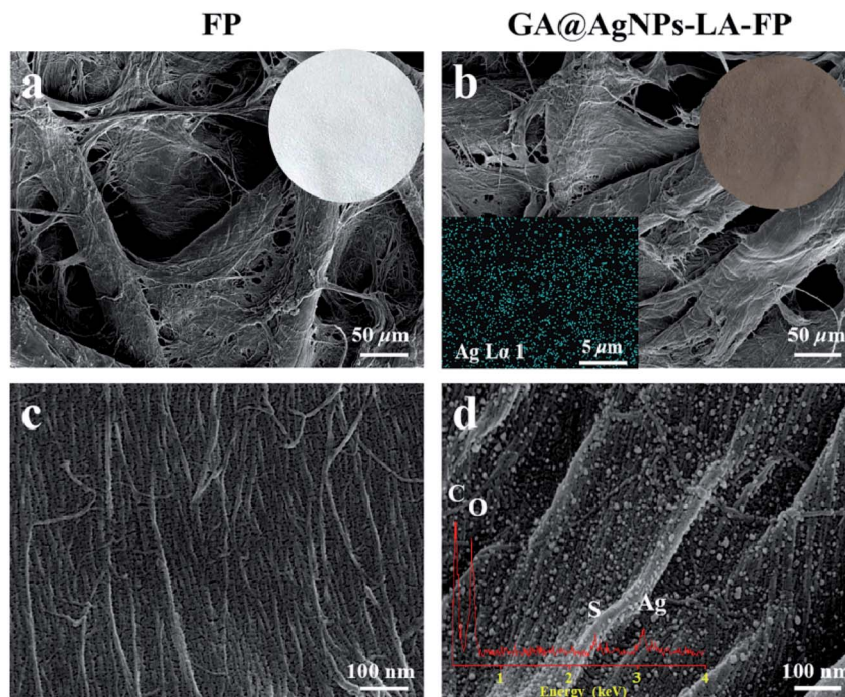


Fig. 1 SEM images of surface morphologies at different magnifications of FP (a and c) and GA@AgNPs-LA-FP samples (b and d). The inset image in (a) is a photograph of a pristine FP sample. The inset images in (b) are a photograph and EDS-mapping of GA@AgNPs-LA-FP. The inset image in (d) is the EDX spectrum of a GA@AgNPs-LA-FP sample.

were observed from the surface of pristine FP. After grafting lipoic acid, the S 2p and S 2s signals assigned to the dithiolane group of lipoic acid were detected for LA-FP (Fig. 2b). Then, the GA@AgNPs deposition was confirmed by the appearance of an Ag 3d doublet near 372.9 eV, Ag 3p at 578.7 and 609.3 eV, and Ag 3s at 724.9 eV, as shown in Fig. 2c. Moreover, in the Ag 3d core-level spectrum (Fig. S2†), the 6 eV splitting of the 3d doublet due to spin-orbit coupling revealed the existence of the Ag⁰ state on the cellulose filter paper. Further, FP, LA-FP, and GA@AgNPs-LA-FP were characterized by high-resolution C 1s XPS to probe

the surface functionalization of the paper samples. The deconvolution of C 1s XPS spectra of FP in Fig. 2d contains three peaks at 284.61 eV, 285.65 eV, and 287.22 eV, which were assigned to C-C, C-O, and C=O bonds in the β-glucose units of cellulose fibers, respectively. Since cellulose is a natural polymer, its surface may be oxidized, which explains the carbonyl (C=O) peak at 287.22 eV in the FP sample. In contrast, a new peak at 288.97 eV appeared in the deconvolution of C 1s XPS spectra of LA-FP (Fig. 2e), which was ascribed to O-C=O groups, suggesting esterification reaction between the carboxyl

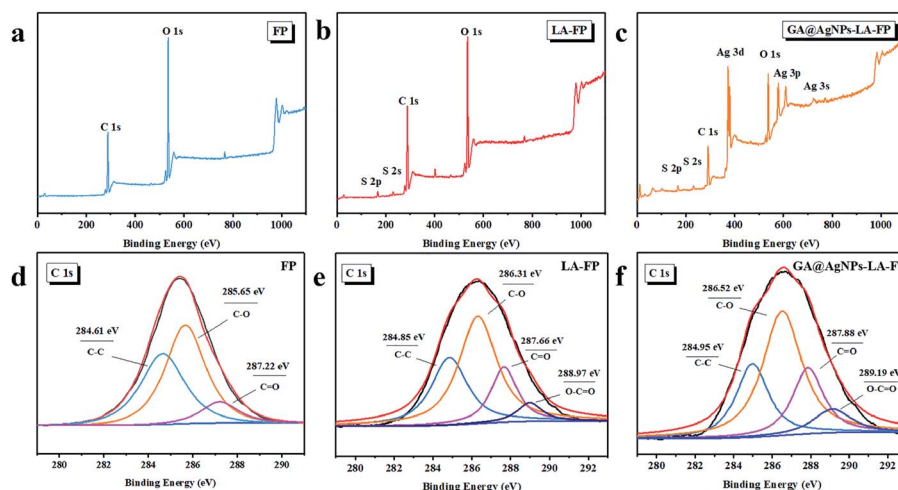


Fig. 2 XPS survey scans of (a) FP, (b) LA-FP and (c) GA@AgNPs-LA-FP; C 1s core-level XPS spectra of (d) FP, (e) LA-FP and (f) GA@AgNPs-LA-FP.



group of LA and primary alcohol groups of cellulose paper. This result indicates the successful conjugation of LA molecules onto FP. In Fig. 2f, showing the GA@AgNPs-LA-FP sample, the C 1s spectrum was shifted to a higher energy near 286.54 eV owing to the deposition of GA@AgNPs, without the emergence of a new deconvoluted peak; however, the O–C=O bond percentage for GA@AgNPs-LA-FP based on the analysis of the XPS C 1s spectrum was 10.2%, which is higher than the O–C=O bond percentage of 6.7% for the LA-FP sample (shown in Table S1†). This is because the immobilized GA@AgNPs have a high carboxyl group density at their interfaces. Furthermore, the chemical structures of FP, LA-FP, and GA@AgNPs-LA-FP were examined by attenuated total reflection Fourier transform infrared spectroscopy (ATR-FTIR). As shown in Fig. S3,† abundant hydroxyl groups (3326 cm^{-1}) on FP, LA-FP, and GA@AgNPs-LA-FP were observed. The presence of a new peak at 1635 cm^{-1} was attributed to ester groups on LA-FP, demonstrating the esterification reaction between LA and FP. The new peaks at 1662 cm^{-1} of GA@AgNPs-LA-FP were attributed to carboxylic groups of GA@AgNPs and ester groups of LA-FP. The above results demonstrate the successful and robust immobilization of GA@AgNPs on the surface of lipoic acid-modified filter paper.

3.2. Silver loading stability of the GA@AgNPs-LA-FP

As filtration membrane, the Ag loading content of GA@AgNPs-LA-FP was 8.4 mg per gram of cellulose paper, measured by inductively coupled plasma atomic emission spectroscopy (ICP-AES). Because of the possible adverse human health effects from silver exposure, the amount of Ag leached from GA@AgNPs-LA-FP into the effluent water was investigated using a flow-through test to ensure the safety of the drinking water (Fig. 3a). The concentration of silver was about $36\text{ }\mu\text{g L}^{-1}$ in the first 0.1 L of filtrate. With the increase of permeate volume to 2 L, the silver concentration gradually decreased to $24\text{ }\mu\text{g L}^{-1}$, which meets the US EPA guideline for drinking water (<100 ppb), demonstrating the safety of GA@AgNPs-LA-FP for water disinfection.¹⁴ To prove the non-toxicity of the released Ag⁺ ions, the solution containing released Ag⁺ ions was further used to culture fibroblast L929 cells to test for *in vitro* cytotoxicity. As shown in Fig. S4,† there was no statistical difference between

the cell viability of L929 cultured in PBS and in the solution containing released Ag⁺ ions. The results demonstrate that the released Ag⁺ ions are non-toxic to normal human cells. In addition, *in vitro* cytotoxicity of free GA@AgNPs was studied with fibroblast L929 cells. The result revealed low cytotoxicity of the prepared GA@AgNPs (shown in Fig. S5†). Understandably, the results of cytotoxicity tests revealed that GA@AgNPs-LA-FP exhibited no cytotoxicity. Such a low level of silver leaching was mainly attributed to the slow release of Ag⁺ ions from GA@AgNPs robustly chelated to cellulose paper by the disulfide bond of LA. To clarify the silver ion release behavior, GA@AgNPs-LA-FP was immersed in 50 mL deionized (DI) water and shaken at 150 rpm, and the leaching rate of silver was quantified with ICP-AES. As shown in Fig. 3b, the release rate of silver steadily decreased from an initial 2.1 μg per day to 1.49 μg per day within 5 days. As a result of this relatively low steady-state silver leaching rate, the release of Ag from the filter paper could be regarded as the leaching of silver ions, rather than leaching of GA@AgNPs from the filter paper. Compared with the total loading content of GA@AgNPs in 50 mg paper ($\sim 420\text{ }\mu\text{g}$), the cumulative leaching of Ag on the 5th day was only 2.15%, confirming the sustained Ag⁺ release from GA@AgNPs-LA-FP. These results indicate that silver nanoparticles were stably loaded onto GA@AgNPs-LA-FP and can potentially provide long-lasting antibacterial activity.

3.3. Antibacterial activities of GA@AgNPs-LA-FP

3.3.1. Bacterial anti-adhesion activity. In our previous study, we proved that GA@AgNPs immobilized or incorporated onto solid surfaces can prevent bacterial adhesion through electrostatic repulsion.^{40,45} Because the high density of anionic carboxyl groups on GA@AgNPs can form a negative interface at neutral pH, it can electrostatically repel bacteria (shown in Fig. 4a). To investigate the effects of the deposition of GA@AgNPs on the surface charge, the surface zeta-potential of GA@AgNPs-LA-FP was analyzed in an aqueous environment at pH 7.0 using pristine filter paper as control. As shown in Fig. 4b, the FP sample presented a negatively-charged surface with a zeta-potential of $17.1 \pm 0.5\text{ mV}$ on account of the many –OH groups on the cellulose fibers. However, the GA@AgNPs-LA-FP showed an enhanced negative charge ($-23.1 \pm 0.6\text{ mV}$), which

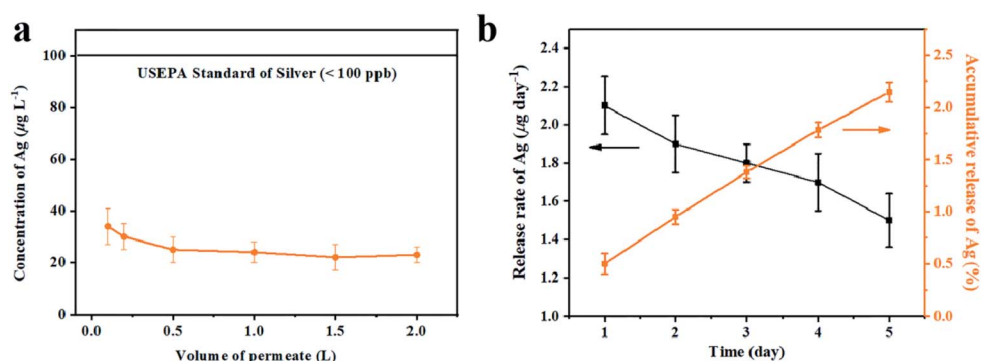


Fig. 3 The release of Ag from GA@AgNPs-LA-FP in the filtration (a) and immersion (b) experiment.



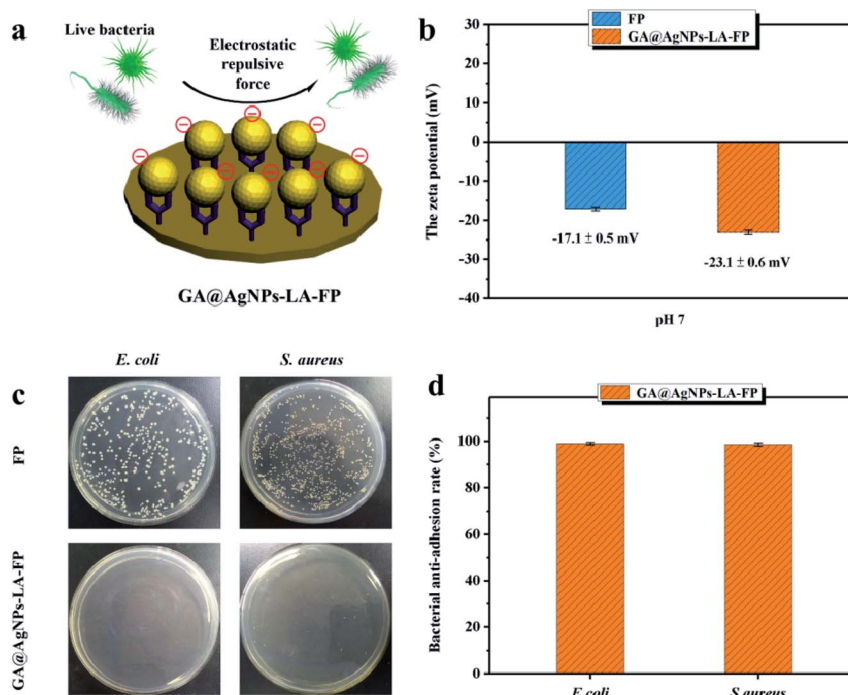


Fig. 4 (a) Schematic of anti-adhesive action of the GA@AgNPs-LA-FP. (b) The surface zeta-potential of the FP and GA@AgNPs-LA-FP at pH 7.0. (c) The corresponding bacterial culture plate photographs, and (d) bacterial anti-adhesion rates of the different samples by plate count method.

was ascribed to the immobilization of GA@AgNPs with many anionic carboxyl groups. The bacterial anti-adhesion activity of GA@AgNPs-LA-FP with an enhanced negative charge was investigated by incubating paper samples with high concentrations (1×10^8 CFU mL⁻¹) of bacterial suspensions for 4 hours. Afterwards, the bacterial cells attached on the paper surface were separated *via* ultrasonication and cultivated. Fig. 4c shows that bacteria easily adhered to the surface of pristine filter paper. On the contrary, almost no viable bacterial cells were observed on the surface of GA@AgNPs-LA-FP, implying effective inhibition of bacterial adhesion *via* electrostatic repulsion. Moreover, the measurable bacterial anti-adhesion rates for the GA@AgNPs-LA-FP sample against *E. coli* and *S. aureus* were 99.5% and 98.6%, respectively (Fig. 4d). Since bacterial adhesion to a membrane surface is the initial phase before biofilm formation, the excellent bacterial anti-adhesion activity of GA@AgNPs-LA-FP is expected to mitigate biofilm formation.

3.3.2. Bactericidal activity. Although the negatively-charged surface of GA@AgNPs-LA-FP can effectively restrain initial biofilm formation by preventing bacterial adhesion, this is just a passive anti-adhesive strategy instead of an active contact-killing property; however, the robust immobilization of GA@AgNPs not only provided an anionic surface but also acted as a reservoir to maintain a continuous contact-killing effect *via* the persistent release of silver ions into the surrounding environment. To understand the AgNPs-decorated antibacterial surface, the morphology of bacterial cells in contact with FP and GA@AgNPs-LA-FP were investigated by SEM. Compared with the viable bacteria with intact cell membranes seen on the FP

surface, distinct damage was observed on the cell membranes of both bacterial strains on GA@AgNPs-LA-FP (Fig. 5b). This demonstrates the contact-killing ability of immobilized AgNPs. Further, this contact-killing bactericidal activity of GA@AgNPs-LA-FP was tested by the CFU counting method. As presented in Fig. 5b, the numbers of attached live *E. coli* and *S. aureus* on GA@AgNPs-LA-FP decreased by $99.3 \pm 0.4\%$ and $98.8 \pm 0.7\%$, respectively, compared with the FP. These CFU results confirm the outstanding bactericidal properties of GA@AgNPs-LA-FP due to the Ag⁺ ion-mediated killing mechanism. Considering the negatively-charged surface and long-term release profile of immobilized GA@AgNPs, GA@AgNPs-LA-FP is expected to show long-term antibacterial activity to mitigate membrane biofouling during its application as a POU water disinfection device.

3.3.3. Anti-biofouling activity. The biofouling that usually occurs on filter membranes has been a dominant barrier limiting the application of membranes in water treatment, owing to declining rate of flow and shortened working life.^{47,48} In Fig. 6a, a gradual and persistent decrease in the permeate rate of water flow due to biofilm was calculated for both FP and GA@AgNPs-LA-FP.^{51,52} The rate of water flow with FP declined by about 58% after the filtration process. On the contrary, the decrease in the rate of flow for the GA@AgNPs-LA-FP was only 20%, indicating the anti-biofouling due to the synergy between anti-adhesion and antibacterial activities of immobilized GA@AgNPs. After filtration, FP and GA@AgNPs-LA-FP were further analyzed by SEM, which showed that a biofilm containing many *S. aureus* cells formed on the FP surface (shown in Fig. 6b); in contrast, scarcely any viable bacterial cells were seen



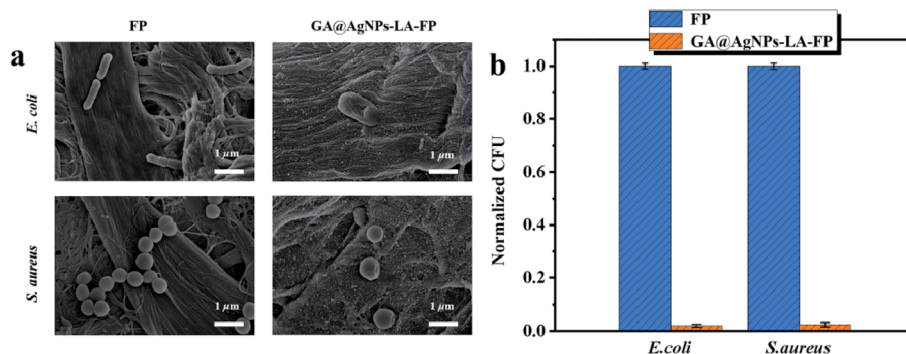


Fig. 5 (a) SEM images of *E. coli* and *S. aureus* cells in contact with the FP and GA@AgNPs-LA-FP. (b) Antibacterial efficiency of the different samples by plate count method. The antibacterial activity was expressed as normalized colony-forming units (CFU) relative to the pristine FP control.

on the surface of GA@AgNPs-LA-FP. The confocal laser scanning microscopy (CLSM) and field emission scanning electron microscopy (FESEM) images in Fig. S6 and S7† clearly demonstrate the biofilms on GA@AgNPs-LA-FP were composed of more dead cells with fewer live cells than biofilms formed on FP. This result provides solid evidence that GA@AgNPs-LA-FP can mitigate biofouling by inhibiting biofilm formation and thus greatly reducing a decline in water flux during operation.

3.4. Point-of-use water disinfection performance of GA@AgNPs-LA-FP during gravity-driven filtration

3.4.1. Disinfection of model contaminated water. To evaluate the water disinfection activity of GA@AgNPs-LA-FP, 100 mL of *E. coli* or *S. aureus* aqueous suspensions with concentrations of 1×10^6 CFU mL⁻¹ were prepared as the model bacteria-contaminated feedwater.^{49,50} As a proof of concept, a water disinfection test was performed by direct filtration under gravity, as shown in Fig. 7a. The average flow rate of bacterial suspensions through the filter paper was about 480 L m⁻² h⁻¹, via controlling the water height above the filter at 5 cm to maintain hydraulic pressure. Before filtration, merged CLSM

images show a large number of live bacteria cells stained with Live/Dead stains in the initial feed solution (Fig. 7b). In contrast, almost all bacteria cells in treated water filtered by GA@AgNPs-LA-FP were dead, indicating that the bacteria in the initial feed solution were effectively contact-killed by the immobilized GA@AgNPs during filtration. To further verify the water disinfection capacity of GA@AgNPs-LA-FP, viable bacteria in the effluent water were cultured on agar medium for 12 h, and bacterial colonies were counted, using bacteria-contaminated feedwater treated using pristine filter paper (FP) as a control. As shown in Fig. 7c, the FP sample with no antibacterial activity exhibited a minor filtration effect for bacteria, with a log reduction lower than 1.0. This low bacteria reduction was attributed to the microporous structure of pristine filter paper (shown in Fig. 1a), which failed to effectively filter bacteria because it only acted as a physical barrier to remove very limited bacteria in the filtration process.

This result confirms the continuous bactericidal activity of GA@AgNPs-LA-FP in the effluent by the sustained release of silver ions. Moreover, it is noted that the bacterial reduction after filtration through the GA@AgNPs-LA-FP was more than 5

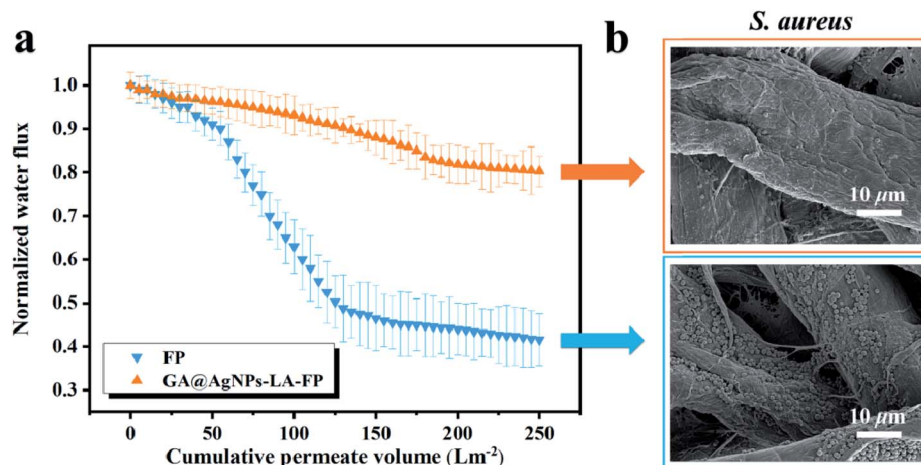


Fig. 6 (a) Normalized water flux of FP and GA@AgNPs-LA-FP samples as a function of the cumulative permeate flux in the dynamic biofouling experiment using *S. aureus*. (b) SEM images of *S. aureus* cells on FP and GA@AgNPs-LA-FP after filtration.

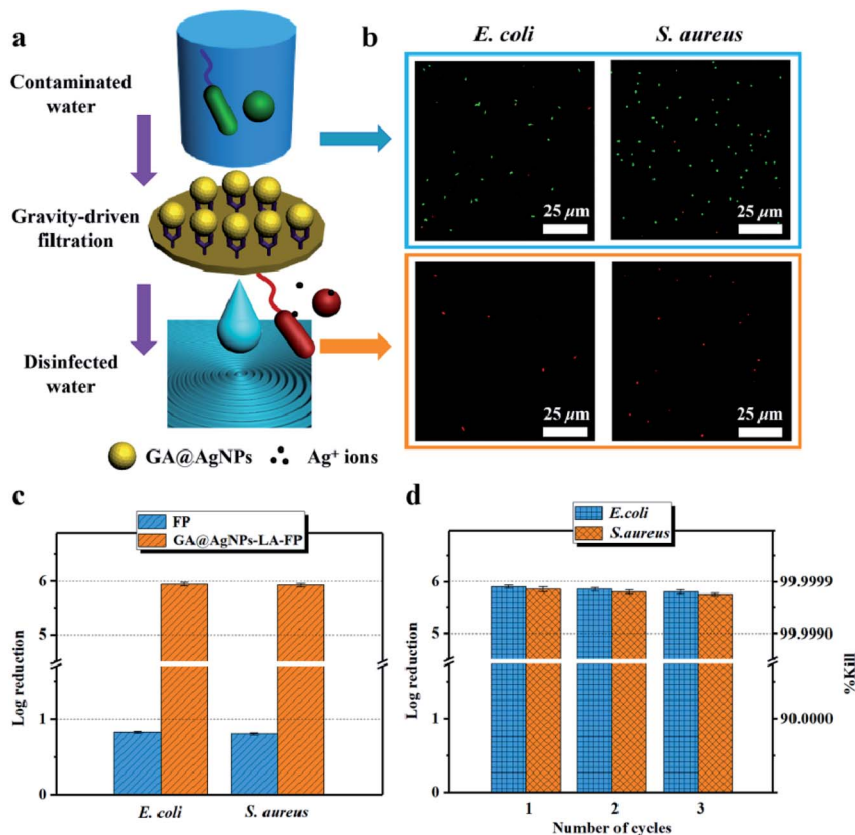


Fig. 7 (a) Schematic POU water disinfection performance of GA@AgNPs-LA-FP. (b) CLSM images of *E. coli* and *S. aureus* cells stained by the Live/Dead kit before and after permeation through GA@AgNPs-LA-FP. (c) Log reduction of *E. coli* and *S. aureus* bacterial count after permeation through FP and GA@AgNPs-LA-FP. (d) Cyclic water disinfection activities of GA@AgNPs-LA-FP against *E. coli* and *S. aureus* solution.

orders of magnitude higher than the control, and meets the US EPA standard for a water purification device. Further, the cyclic water disinfection activity was evaluated by repeatedly filtering 100 mL samples of bacterial solution through GA@AgNPs-LA-FP. After filtration of the first 100 mL sample, the bacterial reduction was measured and set as one cycle. For this test, another three cycles were carried out. As shown in Fig. 7d, even after 3 cycles, the GA@AgNPs-LA-FP still retained satisfactory antibacterial activities, with a log reduction higher than 5.0. This indicates that the active bactericidal effect provided by robust immobilized AgNPs was not abrogated by a covering of dead bacterial cells because of the passive anti-adhesion activity offered by the negatively-charged GA@AgNPs interface; therefore, GA@AgNPs-LA-FP can provide long-term water disinfection performance.

3.4.2. Disinfection of natural river water. GA@AgNPs-LA-FP was also used to disinfect natural river water to investigate its practical disinfection application for drinking water resources. Fig. 8a shows the contaminated river water sample for the POU disinfection test, which was turbid because of its complex composition, including organic microorganisms, inorganic substances, and so on. Within 10 min, 0.1 L of the river water sample was filtered through a circular GA@AgNPs-LA-FP sample. On filtration, the turbid river water became clear (shown in Fig. 8a) because of the porous network structure

of cellulose filter paper, which physically removed suspended solids in the natural water. After that, 0.1 mL samples of both untreated and disinfected water were cultivated on agar plates and standard coliform test plates for 24 h to count their respective colony-forming units. Clearly, for the untreated water sample, lots of colonies grew both on the agar plate and on the standard coliform test plate (seen in Fig. 8b). The total concentrations of viable bacteria and coliforms in contaminated river water, summarized in Table S2,† were 2.1×10^3 CFU mL⁻¹ and 1.5×10^4 MPN/100 mL.³⁷ After filtration, no bacterial colonies on the agar plate or standard coliform test plate were observed for the treated water, implying nearly all bacteria were inactivated by GA@AgNPs-LA-FP. Table S2† shows that the average concentration of viable bacteria in filtered water samples treated with GA@AgNPs-LA-FP was only 11.0 CFU mL⁻¹, which is far below the WHO requirement for drinking water (<100 CFU mL⁻¹). In addition, no coliforms were detected in the treated water. The inactivation rates for GA@AgNPs-LA-FP against total live bacteria and coliforms were 99.5% and 100%, respectively; however, for the water sample treated with FP, the total live bacteria concentration and coliform bacteria were 5.1×10^2 CFU mL⁻¹ and 5.3×10^3 MPN/100 mL. Although some of the bacteria in raw water were physically removed by FP, the quality of the filtered water was not up to the standard



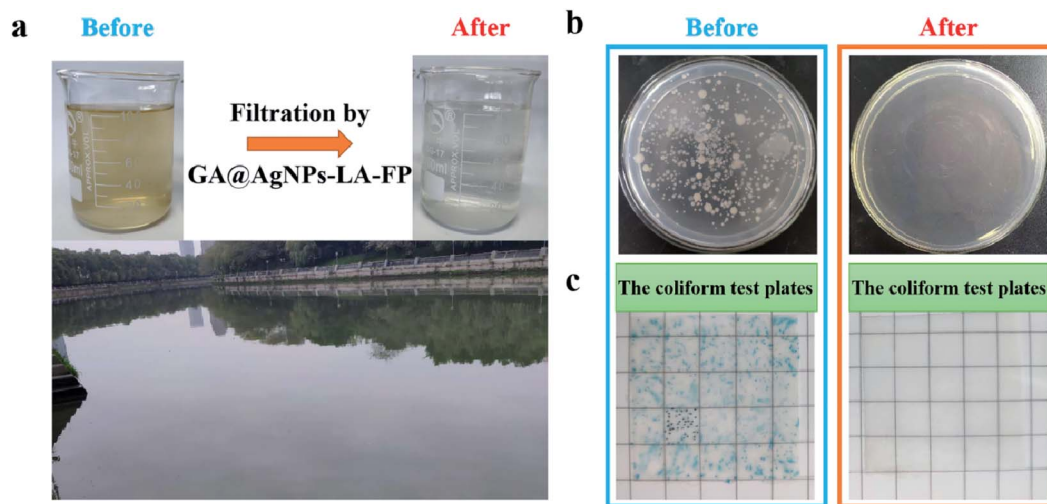


Fig. 8 (a) Photographs of appearance of natural river water before and after treatment with GA@AgNPs-LA-FP. (b) Growth of bacterial colonies on agar plates and standard coliform test plates from natural river water before and after filtration.

for drinking water (total bacteria < 100 CFU mL⁻¹, no coliforms).¹⁴

To test the reusability of GA@AgNPs-LA-FP for effective POU water disinfection, 1 L of contaminated natural river water was filtered by GA@AgNPs-LA-FP in 100 min, and the number of bacteria was counted for each 0.2 L of output. The results indicated that the overall live bacteria concentration was less than 15 CFU mL⁻¹ after treatment, and no coliforms were found (Fig. S8†), which showed that GA@AgNPs-LA-FP had good stability for effective POU water disinfection application. Moreover, using a 10 × 10 cm² GA@AgNPs-LA-FP sample, 2 L of purified clean drinking water was collected within approximately 25 min, which meets the needs of an adult for one day. These results demonstrate the good application prospects for POU water disinfection in remote rural areas or under emergency conditions.

4. Conclusions

A low-cost GA@AgNPs-LA-FP filter was successfully fabricated with robust immobilization of anionic silver nanoparticles onto lipoic acid-modified cellulose filter paper *via* strong chelation between LA disulfide bonds and silver nanoparticles. As a result of the negatively-charged interface and sustained Ag⁺ leaching from immobilized GA@AgNPs, this GA@AgNPs-LA-FP showed both excellent anti-adhesion and strong antibacterial activities, which could synergistically inhibit biofilm formation on the paper surface. In addition, the amount of silver leached from GA@AgNPs-LA-FP was below the WHO drinking water standard for Ag⁺, which verifies that it can be used as a safe water disinfection system. As a novel antibacterial filter, the water disinfection property of GA@AgNPs-LA-FP was evaluated by using it to treat 100 mL of bacterial aqueous suspensions containing concentrations of 1 × 10⁶ CFU mL⁻¹, with a hydraulic retention time of 10 min *via* gravity filtration. The results showed that the bacterial reduction after filtration was more

than 5 orders of magnitude higher than the control, and meets the US EPA standards for a water purification device. Moreover, the potential application of GA@AgNPs-LA-FP for POU water disinfection was also tested using natural river water. The results revealed that almost all live bacteria in the contaminated river water were inactivated when passed through GA@AgNPs-LA-FP *via* gravity filtration within 100 min, demonstrating the paper's viability for practical POU water disinfection.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This work was supported by the support program of Chengdu Science and Technology Bureau (2019-GH02-00001-HZ), opening project of State Key Laboratory of Hydraulics and Mountain River Engineering, Sichuan University (SKHL1816), opening project of laboratory and equipment management, Sichuan University (SCU 203031), and opening project of Key Laboratory of Leather Chemistry and Engineering of Ministry of Education, Sichuan University (20826041C4159).

Notes and references

- 1 M. A. Shannon, P. W. Bohn, M. Elimelech, J. G. Georgiadis, B. J. Marinas and A. M. Mayes, *Nature*, 2008, **452**, 301–310.
- 2 A. Prabhakar, M. Agrawal, N. Mishra, N. Roy, A. Jaiswar, A. Dhawaja and D. Verma, *RSC Adv.*, 2020, **10**, 17479–17485.
- 3 M. S. Mauter, I. Zucker, F. O. Perreault, J. R. Werber, J. H. Kim and M. Elimelech, *Nat. Sustain.*, 2018, **1**, 166–175.
- 4 K. P. Ching and Y. N. How, *npj Clean Water*, 2018, **11**, 1–8.
- 5 H. Zhao, L. Yang, Y. Li, W. Xue, K. Li, Y. Xie, S. Meng and G. Cao, *RSC Adv.*, 2020, **10**, 28267–28276.



- 6 P. J. J. Alvarez, C. K. Chan, M. Elimelech, N. J. Halas and D. Villagran, *Nat. Nanotechnol.*, 2018, **13**, 634–641.
- 7 Z. Zhu, L. Shan, F. Hu, Z. Li, D. Zhong, Y. Yu and J. Zhang, *RSC Adv.*, 2020, **10**, 31295–31304.
- 8 D. H. Seo, S. Pineda, Y. C. Woo, M. Xie, A. T. Murdock, E. Y. M. Ang, Y. Jiao, M. J. Park, S. I. Lim, M. Lawn, F. F. Borghi, Z. J. Han, S. Gray, G. Millar, A. Du, H. K. Shon, T. Y. Ng and K. K. Ostrikov, *Nat. Commun.*, 2018, **9**, 683.
- 9 I. B. Gomes, L. C. Simões and M. Simões, *RSC Adv.*, 2019, **9**, 32184–32196.
- 10 K. L. Kotloff, *Lancet*, 2013, **382**, 209–222.
- 11 S. Ghosh, S. Das and M. E. G. Mosquera, *Polymers*, 2020, **12**, 2993.
- 12 F. Wang, J. Dai, L. Huang, Y. Si, J. Yu and B. Ding, *ACS Nano*, 2020, **14**, 8975–8984.
- 13 Y. Si, J. Y. Li, C. Y. Zhao, Y. Deng, Y. Ma, D. Wang and G. Sun, *ACS Biomater. Sci. Eng.*, 2017, **3**, 854–862.
- 14 M. Fan, L. Gong, J. Sun, D. Wang, F. Bi and Z. Gong, *ACS Appl. Mater. Interfaces*, 2018, **10**, 38239–38245.
- 15 T. F. Clasen, I. G. Roberts, T. Rabie, W. P. Schmidt and S. Cairncross, *Br. Med. J.*, 2006, **334**, 782–785.
- 16 J. Zhu, J. Hou, Y. Zhang, M. Tianb, T. Hee, J. Liu and V. Chen, *J. Membr. Sci.*, 2018, **550**, 173–197.
- 17 S. Ghosh, G. Amariei, M. E. G. Mosquera and R. Rosal, *J. Hazard. Mater.*, 2021, **402**, 123788.
- 18 X. Zhang, Z. Wang, C. Y. Tang, J. Ma, M. Liu, M. Ping, M. Chen and Z. Wu, *J. Membr. Sci.*, 2018, **549**, 165–172.
- 19 Y. Gu, G. Teng, X. Jin, L. Wang, Z. Qiang, W. Ma and C. Zhang, *Ind. Eng. Chem. Res.*, 2020, **59**, 11219–11231.
- 20 S. Ghosh, P. Chakraborty, A. Chakrabarti, M. Ghosh, A. Mandal, P. Saha, A. Mukherje, S. Acharya and M. Ray, *RSC Adv.*, 2016, **6**, 6631.
- 21 W. B. Zhao, M. R. Du, K. K. Liu, R. Zhou, R. N. Ma, Z. Jiao, Q. Zhao and C. X. Shan, *ACS Appl. Mater. Interfaces*, 2020, **12**, 13305–13315.
- 22 S. Ghosh, S. R. Keshri, S. Bera and R. N. Basu, *Int. J. Hydrogen Energy*, 2020, **45**, 6159–6173.
- 23 D. Li, Y. Yan and H. Wang, *Prog. Polym. Sci.*, 2016, **61**, 104–155.
- 24 L. Malaeb and G. M. Ayoub, *Desalination*, 2011, **267**, 1–8.
- 25 G. Kang and Y. Cao, *Water Res.*, 2012, **46**, 584–600.
- 26 D. L. Shaffer, J. R. Werber, H. Jaramillo, S. H. Lin and M. Elimelech, *Desalination*, 2015, **356**, 271–284.
- 27 M. Elimelech and W. A. Phillip, *Science*, 2011, **333**, 712–717.
- 28 S. Bera, S. Ghosh and R. N. Basu, *J. Alloys Compd.*, 2020, **830**, 154527.
- 29 J. M. Daniel, R. D. Daniel, W. B. Christopher, R. P. Donald and D. F. Benny, *Angew. Chem., Int. Ed.*, 2017, **56**, 4662–4711.
- 30 R. Zhang, Y. Liu, M. He, Y. Su, X. Zhao, M. Elimelech and Z. Jiang, *Chem. Soc. Rev.*, 2016, **45**, 5888–5924.
- 31 J. Wang, M. Sui, Z. Ma, H. Li and B. Yuan, *RSC Adv.*, 2019, **9**, 25667–25676.
- 32 A. Zirehpour, A. Rahimpour, A. A. Shamsabadi, M. G. Sharifian and M. Soroush, *Environ. Sci. Technol.*, 2017, **51**, 5511–5522.
- 33 C. S. Ong, P. S. Goh, W. J. Lau, N. Misdan and A. F. Ismail, *Desalination*, 2016, **393**, 2–15.
- 34 T. S. Kim, S. H. Park, D. Parka, J. H. Lee and S. Kanga, *J. Membr. Sci.*, 2019, **576**, 17–25.
- 35 S. Xiong, S. Xu, A. Phommachanh, M. Yi and Y. Wang, *Environ. Sci. Technol.*, 2019, **53**, 3331–3341.
- 36 S. M. Praveena, K. Karuppiah and L. T. L. Than, *Cellulose*, 2018, **25**, 2647–2658.
- 37 X. Zeng, D. T. McCarthy, A. Deletic and X. Zhang, *Adv. Funct. Mater.*, 2015, **25**, 4344–4351.
- 38 T. A. Dankovich and D. G. Gray, *Environ. Sci. Technol.*, 2011, **45**, 1992–1998.
- 39 J. Wang, Y. Wu, Z. Yang, H. Guo, B. Cao and C. Y. Tang, *Sci. Rep.*, 2018, **7**, 1–8.
- 40 G. Liu, J. Jiang, R. Yu, H. Yan and R. Liang, *Ind. Eng. Chem. Res.*, 2020, **59**, 10857–10867.
- 41 Y. L. Li, L. Zhu, Z. Liu, R. Cheng, F. Meng, J. H. Cui, S. J. Ji and Z. Zhong, *Angew. Chem.*, 2009, **121**, 10098–10102.
- 42 G. Liu, H. Gao, K. Li, T. Lan and Z. Zhang, *J. Colloid Interface Sci.*, 2017, **514**, 338–348.
- 43 T. Lan, R. An, Z. Liu, K. Li, J. Xiang and G. Liu, *J. Colloid Interface Sci.*, 2018, **532**, 331–342.
- 44 P. V. Cherepanov, M. A. Rahim, N. Bertleff-Zieschang, M. Abu Sayeed, A. P. O'Mullane, S. E. Moulton and F. Caruso, *ACS Appl. Mater. Interfaces*, 2018, **10**, 5828–5834.
- 45 Q. Xia, L. Yang, K. Hu, K. Li, J. Xiang, G. Liu and Y. Wang, *ACS Appl. Mater. Interfaces*, 2019, **11**, 2352–2363.
- 46 M. S. Islam, N. Akter, M. M. Rahman, C. Shi, M. T. Islam, H. Zeng and M. S. Azam, *ACS Sustainable Chem. Eng.*, 2018, **6**, 9178–9188.
- 47 B. S. Moshe, R. Z. Katherine, G. G. Qi, K. Yan, E. P. Giannelis and M. Elimelech, *Environ. Sci. Technol.*, 2014, **48**, 384–393.
- 48 F. Perreault, H. Jaramillo, M. Xie, U. Mercy, L. D. Nghiem and M. Elimelech, *Environ. Sci. Technol.*, 2016, **50**, 5840–5848.
- 49 L. Qi, Y. Hu, Z. Liu, X. An and E. Bar-Zeev, *Environ. Sci. Technol.*, 2018, **52**, 9684–9693.
- 50 S. M. Praveena, L. S. Han, L. T. L. Than and A. Z. Aris, *J. Exp. Nanosci.*, 2016, **11**, 1307–1319.
- 51 L. Qi, Z. Liu, N. Wang and Y. Hu, *Appl. Surf. Sci.*, 2018, **456**, 95–103.
- 52 S. A. Aani, V. Gome, C. J. Wright and N. Hilal, *Chem. Eng. J.*, 2017, **326**, 721–736.

