



Cite this: *Environ. Sci.: Adv.*, 2022, 1, 769

Non-stoichiometric $\text{Cu}_x\text{In}_{1-x}\text{S}$ quantum dots for robust photodegradation of gemifloxacin: influencing parameters, intermediates, and insights into the mechanism†

Deeptimayee Prusty, Sriram Mansingh, Kundan Kumar Das, Jyotirmayee Sahu and K. M. Parida *

Non-stoichiometric quantum dots (QDs) have gathered wide attention in advanced oxidation processes for environmental remediation due to their ability to generate highly reactive oxygen active species. Here, a series of promising non-stoichiometric $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs were fabricated *via* the reflux method and examined toward the degradation of gemifloxacin (GMF) under visible light. $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD showed optimum activity toward GMF degradation (95% in 2 h, rate const = $258 \times 10^{-4} \text{ min}^{-1}$) with a TOC (total organic carbon) efficiency of 70% due to the non-stoichiometric arrangement of Cu and In metals, which not only promoted the spatial excitons separation efficiency but also bestowed the reducing power of photo-induced electrons as supported by the required characterization. Further, the compositional ratio and phase pure chalcopyrite $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs indicate the good formation of the composition tunable materials. Additionally, an in-depth study was performed to study the effect of anions and pH on the degradation efficiency. Moreover, the scavenging experiments illustrate that O_2^- and OH (indirect route) radicals play a dominant role in GMF degradation, whereas the cycling test showed long-term durability of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD. Eventually, the degradation route was traced by recognizing the intermediate products *via* LC-MS study. This research provides an innovative platform for efficient photocatalytic pollutant degradation by non-stoichiometric $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs in a sustainable environment.

Received 22nd June 2022
Accepted 8th September 2022

DOI: 10.1039/d2va00143h

rsc.li/esadvances

Environmental significance

The occurrence of antibiotics in the environment resulting in the development of antibiotic resistive microbes has gathered increasing attention from the scientific community because of potential adverse ecological and human impacts. Solar-light driven photocatalysts with high efficiency and durability provide environmental benign and sustainable pathway to address the growing concerns of these hazardous and life threatening micropollutants. Herein, a series of promising non-stoichiometric visible light active $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs were fabricated *via* reflux method and examined towards the degradation of pharmaceutical micro pollutant gemifloxacin (GMF) under LED light. Out of the prepared QDs, $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ portrayed the optimum activity towards GMF degradation due to the non-stoichiometric arrangement of Cu and In metals, which not only promoted the spatial excitons separation efficiency but also bestowed the reducing power of photo-induced electrons. The finding is expected to open a new avenue for the fabrication of non-stoichiometric QDs for water treatment in particular and environmental remediation in general.

Introduction

The rapid growth of civilization and modernization has led to excessive production and use of pharmaceuticals and personal-care products, which adversely affect the water ecosystem and hence have become a global issue requiring urgent attention.^{1,2} Among the water contaminants, the antibiotic gemifloxacin (GMF) is a major concern for the research community because both the human and animal kingdoms have been largely infected by these chronic and toxic water pollutants.³ Thus, several countries are avoiding the excessive and unnecessary use of antibiotics in veterinary, agricultural, meat, and dairy

Centre for Nanoscience and Nanotechnology, Siksha 'O' Anusadhan (Deemed to be University), Bhubaneswar-751030, Odisha, India. E-mail: paridakulamani@yahoo.com; kulamaniparida@soa.ac.in

† Electronic supplementary information (ESI) available: Which unfolds chemical information, characterization techniques, the fabrication of working electrodes, CB and VB potentials, a comparison table, EDX patterns, color-mapping image, XPS survey spectra, the Bode phase, TR-PL plots, TR-PL fitting calculations, the reusability plots, XRD, the absorbance plots before and after use, PZC plot, NBT, TA plots, ESR plots and mineralization percentage plots. See <https://doi.org/10.1039/d2va00143h>



based foodstuffs.⁴⁻⁶ Further, the presence of these contaminants in surface and groundwater causes the growth of antibiotic-resistant micro-organisms (bacteria, parasites, viruses, and fungi) (ARM) termed superbugs.⁷ Broadly, GMF is a non-biodegradable, successful fluoro-quinoline antibacterial oral drug, which is used worldwide for the treatment of severe bacterial amplification of chronic inflammation of bronchial tubes and mild to modest pneumonia in veterinary and human sectors causing acute environmental and health problems (ARM and genotoxicity) with limited adverse effect as per the report issued by WHO in 2014.⁷

Several methods such as electro-Fenton, adsorption, ozonization, biological decontamination, and membrane filtration have been developed and tested toward different antibiotic degradation reactions but are found to be ineffective.⁸ For example, the Fenton process requires excessive addition of H₂O₂, which limits its practical implementation. Similarly, in the adsorption process, the recovery of the adsorbate and separation of pollutants from adsorbate are some of the associated shortcomings of the process. Furthermore, the membrane filtration process is subjected to fouling, where the pollutant gets attached to the membrane surface reducing its efficiency and increasing energy and cost usages. Beside, in the ionization process, the major drawbacks are lack of an ozone residual time, corrosion of pipes and fixtures, and requiring storage tanks for treatment purposes.⁸ Therefore, the current scenario crucially demands a cost-effective, eco-friendly, highly efficient treatment for pollutant removal. In this regard, photo-degradation of GMF is a feasible and green technique to circumvent the harmfulness and mineralize the antibiotic by forming bio-degradable products. Meanwhile, our group has extensively studied photocatalytic degradation of various antibiotics over a wide range of materials such as ZFCN@PPY, B/S doped-TiO₂, ZnFe₂O₄@WO_{3-x}/polypyrrole, WO_{3-x}@ZnFe₂O₄, and Ce₂Zr₂O₇@rGO.¹⁰⁻¹⁴

Currently, semiconductor arbitrated photocatalytic ministrations of harmful pollutants have gathered huge research interest owing to their economical and environment-friendly nature. For an efficient photocatalytic reaction, a semiconductor must have the necessary band edge potential and absorb broad light energy to produce the exciton pairs, which are effectively separated. Recently, immense efforts have been made by researchers in designing competent light-sensitive photocatalysts, based upon dimensional-oriented metal chalcogenides owing to their efficacious optical properties.¹⁵⁻²⁰ From this standpoint, quantum dots (QDs) are regarded as effectual photocatalysts due to their extensive absorption from the visible to near IR regions, ultra-small size (less than 10 nm), large surface binding capacity, and quantum confinement effect.²¹ Additionally, ternary QDs have recently generated significant attention and considered environmentally benign alternatives against conventional binary QDs containing heavier toxic metals such as Hg, Cd, and Pb. In this regard, non-stoichiometric copper indium sulphide QDs, a part of group I-III-VI, are considered as robust photocatalysts showing an optical bowing effect, *i.e.*, the optical properties can be tuned by changing the molar ratio of the respective elements.²² Besides,

Hu *et al.* made an interesting observation in Cu-rich CIS QDs, where the major transitions are from the conduction band to the indium vacancy (V_{In}) or the indium site substituted by copper (Cu_{In}) and from the sulfur vacancy (V_S) to the valence band.²³ However, in Cu-deficient CIS QDs, the key transitions are from the sulfur vacancy to the copper vacancy and from the copper site substituted by indium (In_{Cu}) to the copper vacancy (V_{Cu}). Here, In_{Cu} and V_S performed as donors, and V_{In}, V_{Cu} and Cu_{In} acts as acceptors, suggesting that defect states in the CIS are mostly dependent on the composition ratio of the constituent elements. Hence, by varying the composition, it is possible to tune their optical properties and photocatalytic activity for different reduction reactions.²⁴ Non-stoichiometric Cu_xIn_{1-x}S QDs are dazzling semiconductor photocatalysts for pollutant degradation due to their suitable bandgap energy ($E_g = 1.6-2.4$ eV), which can successfully capture most part of the solar light. The conduction band (CB) potential of Cu_xIn_{1-x}S makes it more appropriate for the generation of superoxide radical ($\cdot\text{O}_2^-$), which acts as a powerful species in the antibiotic degradation process by following a photo-reduction mechanism.²⁵

This study reports exotic Cu_xIn_{1-x}S photocatalytic systems fabricated successfully by the reflux strategy by varying Cu and In ratios of 0.25 : 0.75, 0.5 : 0.5, and 0.75 : 0.25, respectively. Here, TGA was used as a capping agent, which controlled the size of the particle and averted the formation of binary QDs. Also, tuning of the bandgap energy was observed, while varying the molar ratio of Cu : In, resulting in the variation of the optical properties. Further, the non-stoichiometric 0D QDs showed better charge-carrier separation as well as migration, as confirmed by PL, EIS, and Bode phase analysis, which eventually boosted the photocatalytic degradation efficiency of antibiotics *via* direct superoxide and an indirect hydroxyl radical pathway, as evidenced by the scavenger test, TA, NBT, and EPR analysis. Additionally, the effect of anions and pH variation were well studied. In addition, to the best of our knowledge, Cu_xIn_{1-x}S QD photocatalysts have not been explored so far for the degradation of GMF, so this work will provide new scope for researchers working in this field.

Experimental section

Synthesis of Cu_xIn_{1-x}S QDs

A single pot reflux technique was adopted to synthesize the Cu_xIn_{1-x}S composite by following our previous literature, where favorable amounts of CuCl₂·2H₂O, InCl₃·4H₂O, and Na₂S·9H₂O were picked as precursors of copper, indium, and sulphur, respectively.²⁶ A total of 1 mmol of Cu and In precursors along with 10 mmol of TGA as the capping ligand were taken into two-necked round bottom (RB) flask and dissolved in the required amount of distilled water. In order to maintain the pH = 8.5 of the mixture solution, 1 M NaOH solution was added and subsequently, an aqueous solution of 0.52 mmol Na₂S·9H₂O was inserted into the RB. The entire solution was refluxed for 4 h at 100 °C, which was precipitated by adding ethanol, followed by dehydration at 40 °C under vacuum. Typically, a series of Cu_xIn_{1-x}S QDs were prepared with three different molar ratios of Cu : In *i.e.*, 0.75 : 0.25, 0.50 : 0.50, and 0.25 : 0.75,



respectively, as illustrated in Scheme 1. All the above-used chemicals were of analytical grade and high purity (Merck company), so used as such without any further purification.

Results and discussion

Structure, morphology, and surface composition analysis

The crystallinity and phase purity of the fabricated catalysts were analyzed using PXRD (powder X-ray diffraction), as revealed from the data in Fig. 1(a). It was observed that $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs with varying compositional ratios of Cu and In showed similar XRD patterns with 2θ values at 28.7° (112), 46.95° (220), and 54.9° (312), corresponding to (ICDD card no. 85-1575) the chalcopyrite phase without any impurity peaks, confirming phase purity of the crystals.²⁴ Further, from Fig. 1(b), the fringe pattern of the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ nanocrystal, confirmed the existence of (112) and (220) planes with d_{spacing} values of 0.32 and 0.20 nm, respectively, which showed a good correlation with the XRD data. The angle between the two planes was calculated to be 54° , confirming the chalcopyrite phase in the material, which correlated well with the XRD results. Again, the FESEM image of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ showed particle-like morphology as shown in Fig. 1(c). Additionally, the HRTEM image, as presented in Fig. 1(d), indicated the formation of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD with an average particle size of 3.62 nm, as shown in the inset of Fig. 1(d). Furthermore, the energy dispersive X-ray (EDX) image and color mapping picture of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD showed the presence of all the constitutional elements such as Cu, In, and S without any impurity element, which was well supported by XPS analysis (Fig. S1(a)–(e)†).

XPS (X-ray photoelectron spectroscopy) is an important tool to inspect the elemental compositions with their ratios and chemical states of the corresponding elements in the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD, as observed from the data in Fig. 1. The XPS

survey spectrum of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD shown in the ESI file (Fig. S2†), revealed the presence of all the constituent elements (Cu, In and S) correlating with the EDX analysis, indicating compositional purity. From Fig. 1(e), it is notable that, the binding energies of 952.1 and 931.6 eV corresponded to two deconvoluted peaks of Cu $2p_{1/2}$ and Cu $2p_{3/2}$ spin levels, respectively, validating the existence of Cu^+ in the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ unit with no impurity peaks. Further, the spin–orbit coupling states, *i.e.*, $3d_{5/2}$ and $3d_{3/2}$ of In^{3+} are located at 445.2 and 452.8 eV, respectively, as shown in Fig. 1(f).^{26,27} Additionally, sulphur was manifested by a doublet peak at binding energies of 157.7 eV and 163.1 eV corresponding to S $2p_{3/2}$ and S $2p_{1/2}$, respectively, owing to spin–orbit coupling.²⁸ This revealed the binding of S to Cu and In in $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD, as demonstrated in Fig. 1(g). Besides, the compositional ratio of Cu : In in $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD was revealed by XPS analysis and was found to be 0.74 : 0.26, which is quite reliable with the molar ratios of Cu and In precursors taken during synthesis. Therefore, it is noticeable from the above data that copper and indium are very well balanced during the fabrication time and non-stoichiometric $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD was formed as per the estimation. Similarly, the compositional ratios for other prepared catalysts are tabulated in Table S1.†

Optical absorption and charge separation properties

The optical and charge separation properties were characterized by UV vis-DRS, PL, and EIS analysis. Fig. 2(a) depicts the absorbance of the synthesized materials, which indicates that all copper–indium–sulfide QDs exhibit a wider absorbance in the visible region and extends up to the near IR region due to the inherent property of the QD materials.²⁹ As a result, the materials can utilize almost all portions of solar energy in the visible and as a small fraction of the IR spectrum, leading to a superior optical property. Here, from the figure, it is obvious



Scheme 1 Scheme for the synthesis of $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs by varying Cu : In molar ratios *via* the reflux strategy.



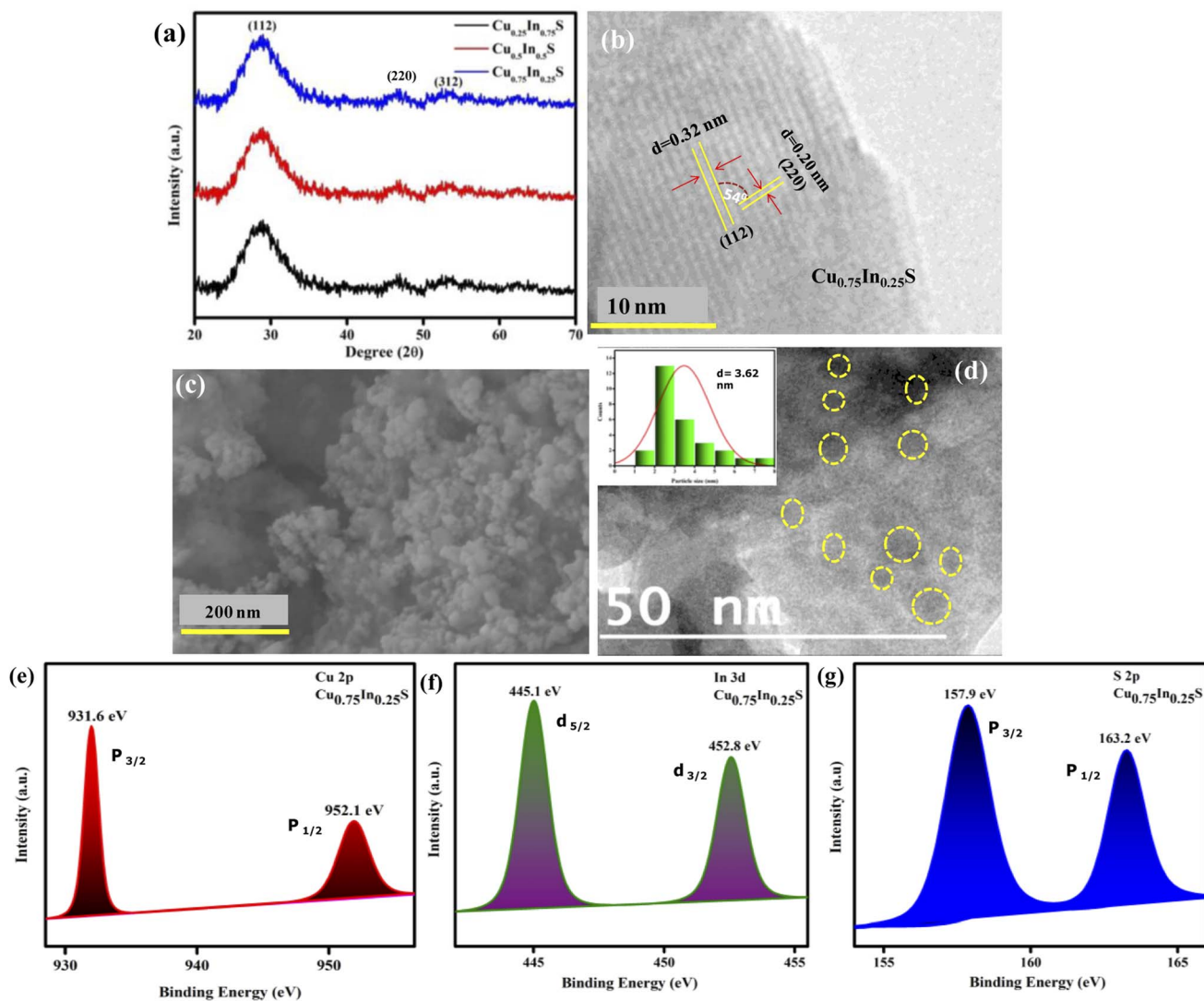


Fig. 1 Structural and compositional characterization: (a) XRD pattern of $\text{Cu}_x\text{In}_{1-x}\text{S}$ QDs with 3 different molar ratios 0.75 : 0.25, 0.50 : 0.50, and 0.25 : 0.75 of Cu : In, respectively (b) fringe pattern, (c) FESEM and (d) HRTEM images of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD; (e) XPS analysis spectra of Cu 2p, (f) In 3d, and (g) S 2p elements in $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD.

that the absorbance shown by $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ is greater than that by the other materials, showing its better photon absorption capability than $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ and $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ QDs. Adding more, Fig. 2(b) represents the Tauc plot of the as-prepared materials, which shows bandgap energies (BE) of 1.84, 2.03, and 2.24 eV for $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$, $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$, and $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QDs, respectively, by using the Kubelka–Munk equation as given below (eqn (1)):¹²

$$(\alpha h\nu)^n = A(h\nu - \text{BE}) \quad (1)$$

where, h = Planck's constant = 6.626×10^{-34} J s, α = absorbance coefficient, ν = frequency of photon, A = proportionality constant and the value of n suggests type of electronic transition, *i.e.*, for indirect transition, $n = 1/2$ and for direct transition, $n = 2$. Additionally, Fig. 2(c) depicts the Mott–Schottky (MS) plots of as-prepared materials with positive MS slopes indicating n-type nature, and flat band potentials (F_p) at the electrode–electrolyte interface were estimated to be -0.94 ,

-1.04 and -1.10 V vs. Ag/AgCl (pH 6.8) for $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$, $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$, and $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QDs, respectively.^{13,26} Further, the higher negative shift in F_p of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ with respect to other samples suggests greater availability of electrons or higher donor density, resulting in improved photocatalytic activity. The increased electron concentration causes the origin of a built-in electric field near the electrode–electrolyte junction leading to effective charge isolation and migration that ultimately boost the degradation rate of GMF over $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$.³⁰ In addition, the conduction band (CB) potentials of the above materials were recalculated on the NHE scale and were found to be -0.65 , -0.75 , and -1.01 V vs. NHE (pH 7) for $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$, $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$, and $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD, respectively. Hence, using eqn (2), valence band (VB) potentials of the as-synthesized samples were calculated and the obtained results are listed in Table S2.†

$$\text{BE} = \text{VB} - \text{CB} \quad (2)$$



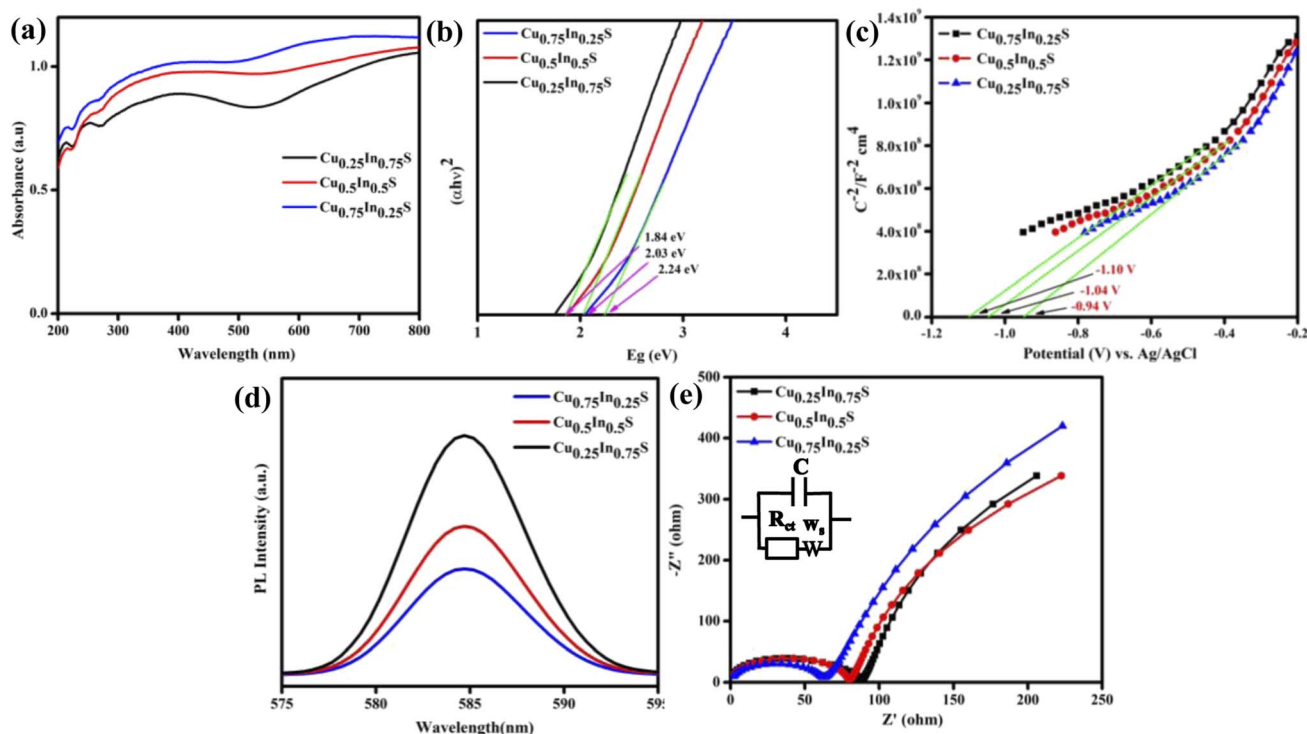


Fig. 2 Optoelectronic study of $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ (black line), $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ (red line), and $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ (blue line) QDs: (a) UV-Vis DRS spectra, (b) Tauc plot, (c) the $M-S$ plot over a frequency of 1000 Hz in 0.1 M Na_2SO_4 , (d) PL spectra, and (e) the EIS plot.

Again Fig. 2(d) shows the PL spectra of the as-synthesized samples, which indicates the charge carrier separation and migration efficiency of the materials. The more intense PL peak suggests the least charge carrier separation, whereas, the less intense peak indicates more exciton separation efficiency. From the figure, it is clear that the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD shows greater charge carrier separation efficiency than the other two materials leading to a better photocatalytic activity, which is quite relevant to EIS and TPC analysis. Fig. 2(e) shows the electrochemical impedance spectroscopy (EIS) plot of the synthesized materials in zero applied bias and the fitted circuit diagram is shown in the inset of that figure, where C , W_s , and R_{ct} are double-layer capacitance, Warburg impedance and charge transfer resistance, respectively.³¹ Herewith, the straight line portion entails the Warburg resistance, whereas the semicircle part is about the R_{ct} value, specifying the conductance efficiency of the photocatalysts. The larger the semicircle radius more will be its interface resistance resulting in lesser conductance with smaller charge carrier separation efficiency and *vice versa*.³² Here, in the figure, the smallest semicircle radius of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ indicates the smallest R_{ct} value leads to better charge carrier separation among all other samples, which enhances the photocatalytic activity of the material. Furthermore, the lifetime of the injected photo-electrons of the prepared QDs was estimated from the Bode phase plot (Fig. S3(a)†) following eqn (3). Basically, the migration time of photogenerated electrons from the sample to electrolyte was correlated with the shift in the peak location toward the lower frequency zone, indicating accelerated transport of electrons.³³

$$\text{The life span of electron } (\tau_n) = \frac{1}{2\pi f_{\max}} \quad (3)$$

In addition, a lower value of f_{\max} (frequency), as observed for $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD, corresponds to a greater life time of the injected electrons (41 μs), resulting in a significant increment in the catalytic activity compared to the other two $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$, and $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ QDs having τ_n values of 36 and 38 μs , respectively. This result is also relevant to the PL and TPC analysis, where the charge carrier separation follows the order, $\text{Cu}_{0.25}\text{In}_{0.75}\text{S} < \text{Cu}_{0.5}\text{In}_{0.5}\text{S} < \text{Cu}_{0.75}\text{In}_{0.25}\text{S}$. To double-check the larger exciton life time for $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD, time-resolved-photoluminescence (TR-PL) analysis was carried out. The TR-PL data showed that all synthesized materials could be well fitted to the tri-exponential function by using eqn (S1) and (S2) from the ESI file.† The more the decay time, the greater the exciton separation, thereby, decreasing the recombination rate. From Fig. S3(b),† it is obvious that $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ shows an overhead exciton life-time of 1.43 ns, followed by $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ (1.15 ns) and $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ (1.03 ns), which is well correlated with the Bode phase data, resulting in superior photocatalytic activity.

Photocurrent response

Furthermore, the linear sweep voltammetry (LSV) analysis, as illustrated in Fig. 3(a), suggests the photocurrent density of the synthesized material. It is noteworthy that, the more the photocurrent density more will be its photocatalytic activity. In this context, the current density of the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD is more than



that of $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ and $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ materials under visible light irradiation, leading to better photocatalytic activity. $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$, $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ and $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ materials showed anodic currents of 1.48, 1.22, and 1.09 mA cm^{-2} , respectively, which again supports their n-type character. Additionally, transient photo-current (TPC) measurement indicates the charge carrier generation, separation efficiency, and durability of the excitons by showing photocurrent response under ON-OFF light illumination.³⁴ In dark and light, there was a sudden decrease and increase in photo-current density, suggesting a quick light response of the photo-anodes. The more the photocurrent density more will be the exciton generation and separation efficiency, thereby increasing the photodegradation of antibiotics. From Fig. 3(b), it is clear that $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD shows the highest photo-current density under interrupted light irradiation at -0.3 V (vs. Ag/AgCl) biasing potential, as compared to the other materials, leading to better exciton separation. Moreover, chronoamperometry analysis (CA) was performed to test the photo-stability of the as-synthesized materials. It is noticeable that all the synthesized materials are stable under light illumination for a period of 900 s with

a constant photocurrent value. From Fig. 3(c), it was observed that, $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ showed a higher current density of $24.5 \mu\text{A cm}^{-2}$ at 0.2 V applied voltage under light illumination than $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ ($15.7 \mu\text{A cm}^{-2}$) and $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ ($18.2 \mu\text{A cm}^{-2}$), showed better photo-stability among all materials.

Evaluation of photocatalytic performance

Investigation of the photodegradation performance of the designed catalysts over pharmaceutical pollutant gemifloxacin (GMF). The photodegradation ability of the as-prepared catalysts was examined toward quinoline-type antibiotic GMF (10 ppm) as the model pollutant. The quantitative degradation of GMF over different catalysts was well confirmed from the rapid decrease in its characteristic optical absorption peak intensity observed at about wavelength 270 nm and 343 nm, respectively (Fig. 4(a)). Further, the adsorption of GMF over the synthesized samples was very negligible, hence not an effective value for pollutant removal. As expected the photodegradation efficacy ($D\%$) of GMF by $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$, $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$, and $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ samples were 95%, 86%, and 59%, respectively, in 2 h, as

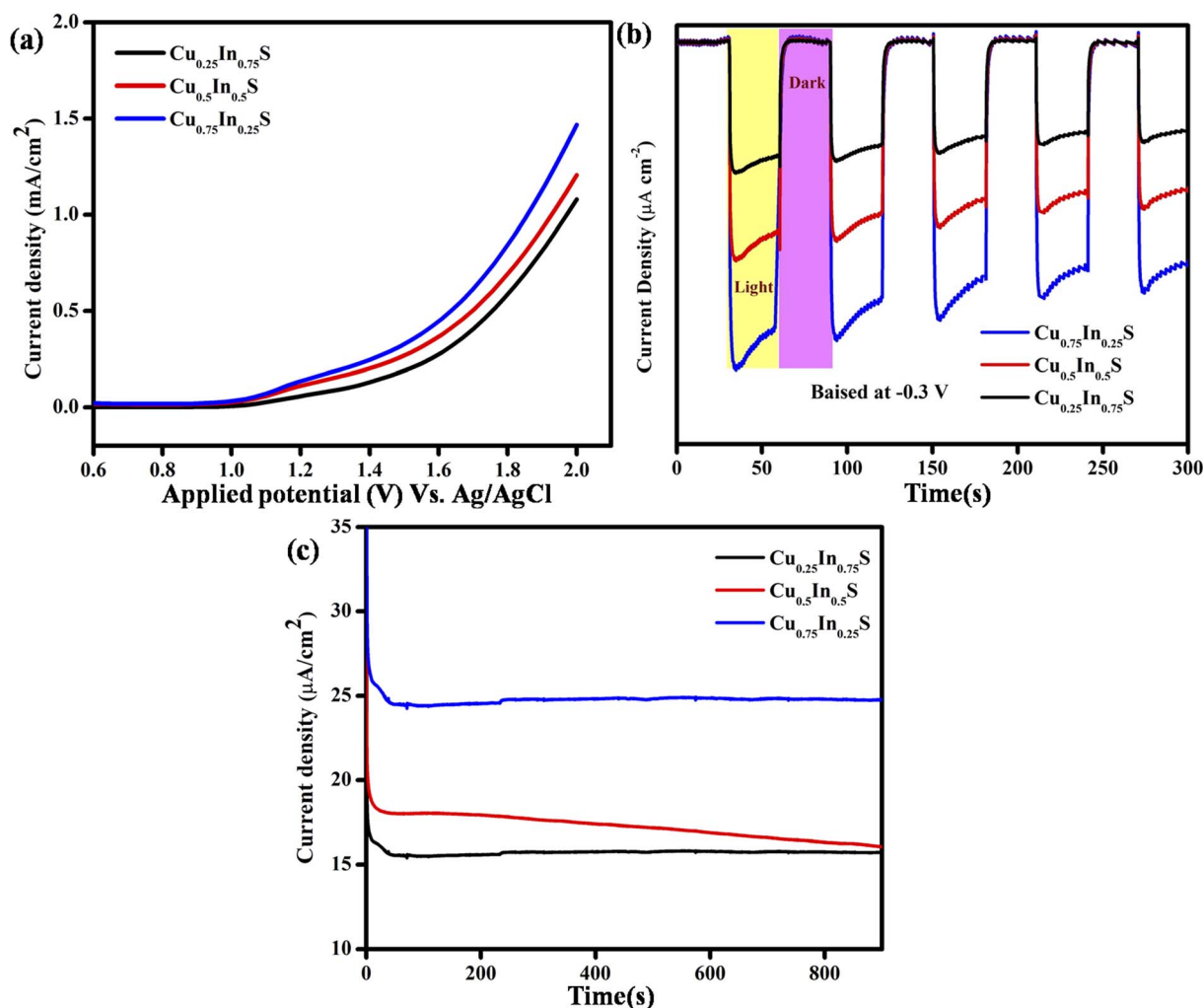


Fig. 3 Photocurrent study of the synthesized materials with various Cu/In ratios: (a) LSV plot, (b) TPC response at -0.3 V biasing potential, and (c) CA plot at 0.2 V applied potential.



illustrated in Fig. 4(b). Additionally, Fig. 4(c) depicts the kinetic model toward the degradation of GMF over different photocatalytic materials, which fits well with the pseudo-first-order kinetic *via* the following eqn (4):^{10,13}

$$\ln\left(\frac{C_0}{C}\right) = k_{\text{app}}t \quad (4)$$

Furthermore, the half-life period ($t_{1/2}$) was calculated employing the below eqn (5):

$$t_{1/2} = 0.693/k_{\text{app}} \quad (5)$$

Here, k_{app} is the rate constant (min^{-1}) and R^2 -co-relation coefficient factor for GMF degradation and the detailed obtained values are included in Table 1. From the above data it is clear that the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ photocatalyst shows the best activity with the highest rate constant and low $t_{1/2}$ value; hence, further studies were performed by taking this sample as the primary material. Also, a comparison table (Table S3†) is given for other catalysts toward GMF degradation, which indicates that our

catalyst showed better photocatalytic activity under 20 W LED light irradiation. Additionally, the photostability of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ QD was also studied to evaluate the reusability of the catalyst under the performed degradation condition. In this case, the GMF degradation experiment was performed five successive times with the same catalyst, and the obtained degradation efficiency after each run was recorded, as depicted in Fig. S4(a).† From the plot, a very little drop in efficiency was observed even after 5 consecutive cycles, which implied excellent durability of the as-prepared QD under operational conditions. However, the visualized reduction in the removal rate corresponded to the loss of the catalyst amount during the recovery process. Besides, XRD and optical absorbance of the material, after and before use (shown in Fig. S4(b) and (c)†) indicated no change in characteristics, which again confirmed the extraordinary stability of the catalyst.

Effect of environmental factors on GMF degradation

Effect of pH. To obtain a detailed insight into the photocatalytic degradation of Gemifloxacin (GMF), some additional

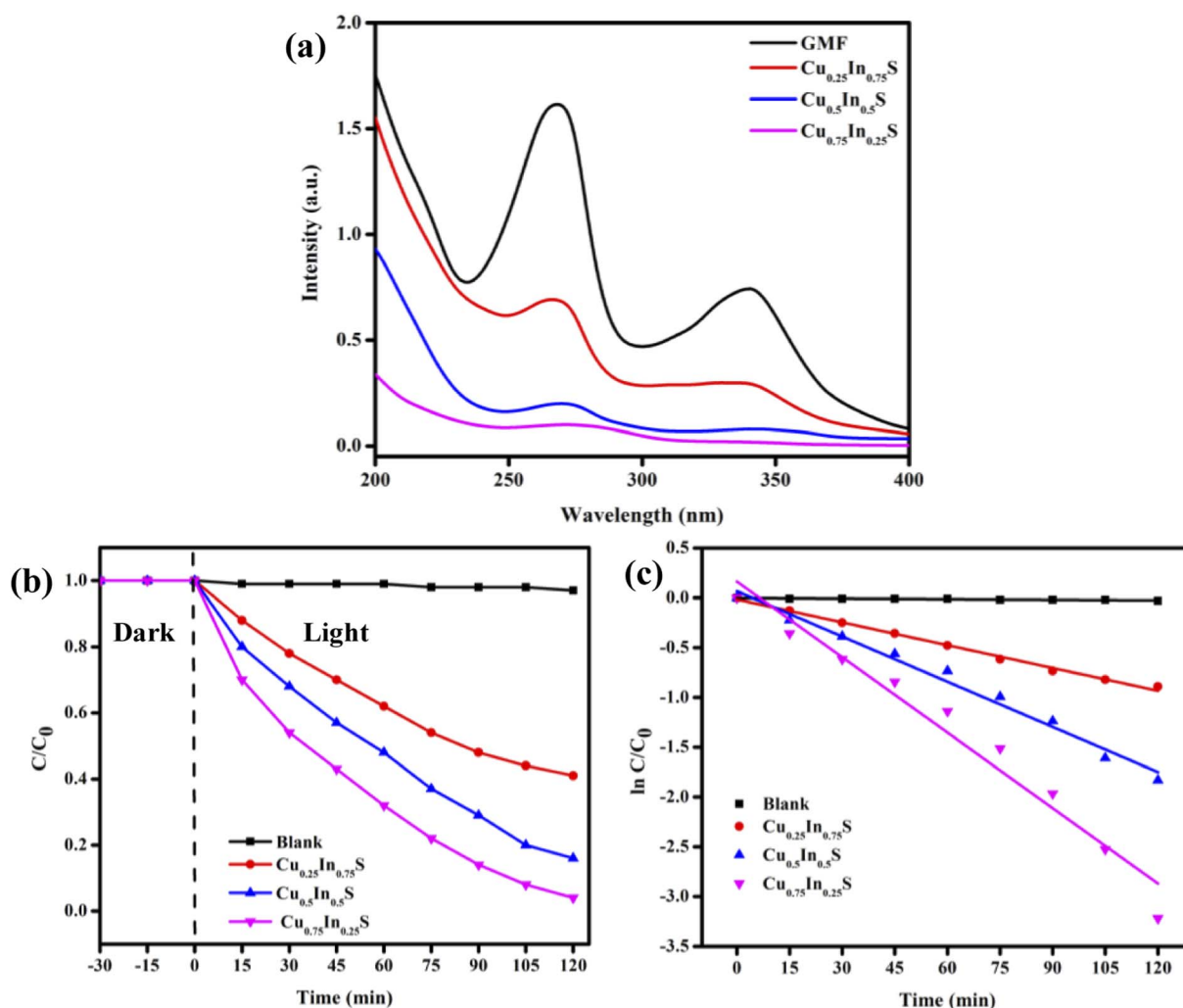


Fig. 4 GMF (10 ppm) degradation in 2 h over synthesized photocatalysts (a) absorbance plot, (b) effect of catalysts on degradation, and (c) fitted kinetic curve for the pseudo-first-order reaction.

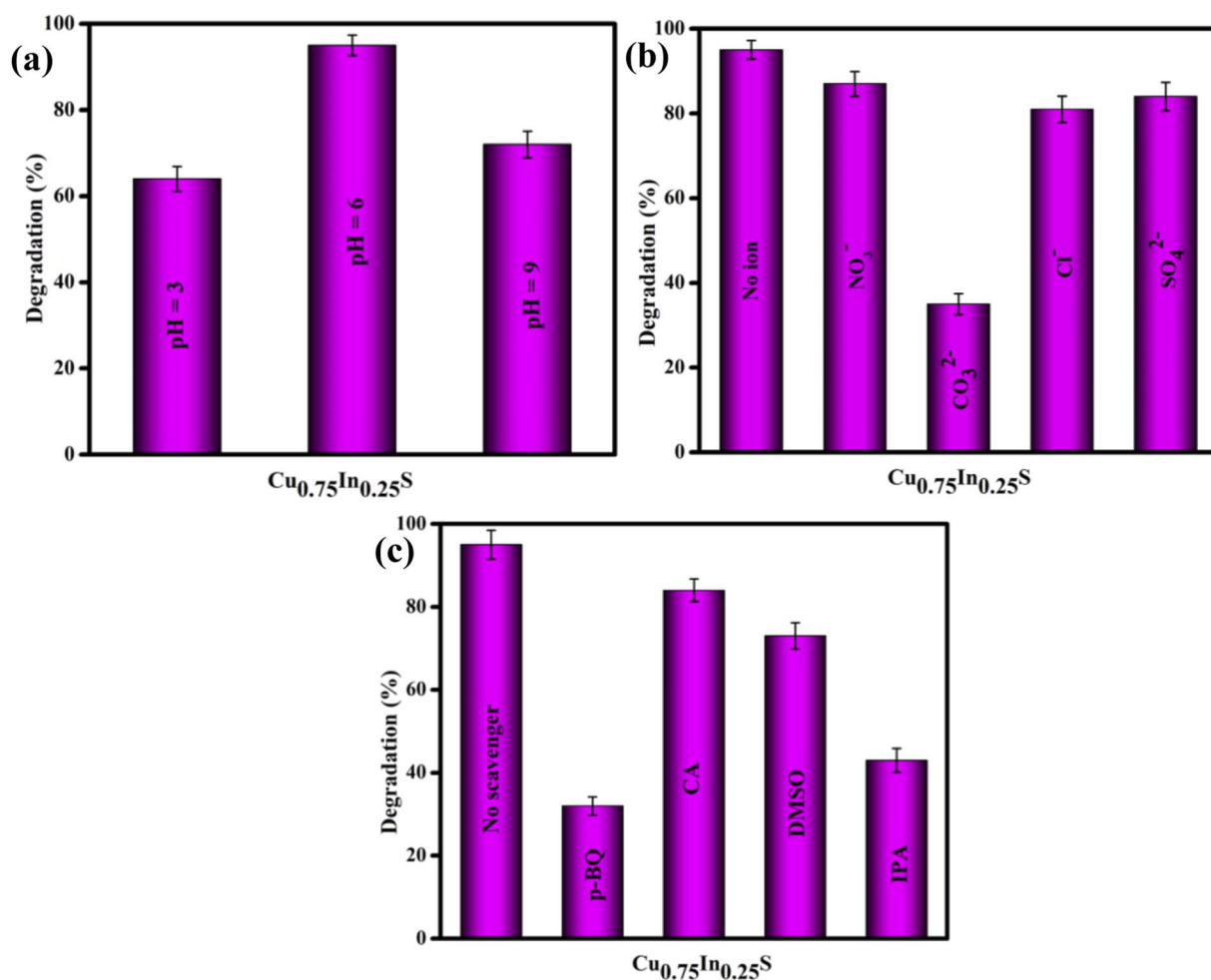


Table 1 Detail values of the reduction percentage, R^2 , k_{app} and $t_{1/2}$ for all synthesized samples after GMF degradation

Photocatalyst	Degradation efficiency (%)	(R^2)	(k_{app}) (10^{-4} min^{-1})	$t_{1/2}$ (min)
Blank	3	0.93	5	0.1386
$\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$	59	0.99	76	0.0091
$\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$	86	0.98	151	0.0045
$\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$	95	0.96	258	0.0026

investigations were conducted with the best ($\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$) photocatalyst and the obtained results are shown in Fig. 5. The solution pH plays a pivotal role in the degradation process. In brief, the photo-degradation experiment was carried out in acidic, neutral, and basic environments, as shown in Fig. 5(a). In this work, the photocatalytic decomposition of GMF was studied under different pH values and the obtained results were found to be 64%, 95%, and 72% at pH 3.0, 6.0, and 9.0, respectively. It can be observed that a low removal efficiency was witnessed under strong acidic (pH = 3) and basic (pH = 9) conditions, while the maximum degradation/removal efficiency was achieved at pH 6 after 120 min of 20 W LED irradiation. However, two pK_a values were reported for GMF, where it exists in cationic (GMF^+) below pK_a 5.53 and anionic (GMF^-) above

pK_a 9.23.³⁵ Fundamentally, the catalyst surface is positively charged, if the pH of the solution is lower than that of point of zero charge (PZC) value and negatively charged, if pH is above the PZC value. Similarly, from the pK_a value perspective, the pollutant behaves positively charged if the $pK_a <$ solution pH and negatively charged at the higher pH value ($pK_a >$ solution pH).¹³ The PZC value of $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ was found to be 5.8 (as shown in Fig. S5†), which means that above this value the catalyst is negatively charged and below that value, it is positively charged. As discussed above, GMF has two pK_a values, $pK_{a1} = 5.53$ and $pK_{a2} = 9.23$. The pH study displayed the best activity at pH = 6, where the catalyst surface becomes negatively charged (pH > 5.8) and GMF appears positively charged (pH < pK_{a1}), so both the adsorbent (photocatalyst) and adsorbate

Fig. 5 Degradation efficiency of GMF over 10 ppm $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ photocatalyst on different (a) pH solutions, (b) anions, and (c) scavengers.

(GMF pollutant) are strongly attached. Moreover, at pH 3 and 9; Cu_{0.75}In_{0.25}S acquires positive (pH < 5.8) and negative (pH > 5.8) charges, while GMF carry positive (pH < pK_{a1}) and negative (pH > pK_{a2}), respectively, causing repulsion between the photocatalyst and pollutant and hence reducing the overall degradation activity.

Effect of anions. We know that the water ecosystem consists of several inorganic salts and impurities, therefore, it is significantly important to inspect the impacts/influence of several ions, specifically anions on the photocatalytic degradation of the GMF pollutant. For the anion study, 0.05 M of NaCl, NaNO₃, Na₂SO₄, and Na₂CO₃ were added separately to the aqueous solution of GMF and the degradation experiments were performed following the same procedure under the same reaction conditions, as stated earlier. The results of this investigation are shown in Fig. 5(b). As per the presented results, SO₄²⁻, NO₃⁻, and Cl⁻ ions were found to have a minimum influence on GMF degradation, however, the photodegradation process was hampered and restricted significantly after the introduction of CO₃²⁻ ions. The CO₃²⁻ ions increase the pH of the reaction medium because of the hydrolyzation and can capture the free radicals or active species ([•]OH and [•]O₂⁻), primarily responsible for photocatalytic degradation of GMF.³⁶

Active species trapping (NBT, TA, and ESR) study. The assessment of trapping experiments was performed to trace the chief active radical/species responsible for the degradation of GMF by the Cu_{0.75}In_{0.25}S photocatalyst. In this direction, citric acid (CA), *para*-benzoquinone (*p*-BQ), isopropyl alcohol (IPA), and dimethylsulfoxide (DMSO) were utilized as trapping agents for holes (h⁺), superoxide radicals ([•]O₂⁻), hydroxyl radicals ([•]OH), and electrons (e⁻), respectively. As highlighted in Fig. 5(c), the maximum degradation of GMF was observed in the absence of scavengers.^{37,38} However, a significant loss in degradation rate, or simply it could be stated that the photocatalytic degradation of GMF was drastically hindered (only 32%) in the presence of *p*-BQ and IPA (43%). This indicates that superoxide and hydroxyl radicals play a predominant role in GMF degradation. Furthermore, a slight inhibition in photocatalytic degradation was displayed after the addition of CA and DMSO, suggesting that the holes and electrons play some contribution toward GMF degradation.³⁹ However, the valence band position of the catalyst is not suitable to produce hydroxyl radicals, therefore, the hydroxyl radicals are generated through an indirect route, *i.e.*, superoxide radicals transfer to hydrogen peroxide, which then decomposes to form [•]OH radicals.⁹ The generation of superoxide and hydroxyl species over the catalyst surface was further confirmed from NBT (nitro blue tetrazolium chloride) and TA (terephthalic acid) experiments, following our previously reported articles.⁴⁰ A peak at 259 nm in UV-Vis spectrum of NBT and PL-emission spectrum at about 423 nm of TA progressively decreased with Cu_{0.75}In_{0.25}S, suggesting that more superoxide and hydroxyl radicals were generated on Cu_{0.75}In_{0.25}S, which ultimately helped in the degradation of pollutants compared to other materials, as shown in the Fig. S6(a) and (b).[†] Further, to ensure the [•]O₂⁻ radicals production ability of the Cu_{0.75}In_{0.25}S system, electron spin resonance (ESR) study was carried out using DMPO as a spin-

trap agent at different time intervals, as shown in Fig. S7.[†] As seen from the figure, four featured signals of DMPO-[•]O₂⁻ were observed, which clearly indicated the formation of [•]O₂⁻ species and the peak intensity gradually increased with an increase in time from 0 to 15 min.

Mineralization/conversion pathways and possible mechanisms of GMF degradation

To determine the intermediate species formed during the photo-mineralization/degradation of GMF, LC-MS analysis was performed and the obtained results are displayed in Scheme 2. Based on the identified compounds, two possible mechanistic pathways are projected toward the degradation of GMF as illustrated below.^{41,42}

Pathway-1. First, GMF is degraded to 2,3-dihydro-1,8-naphthyridin-4-one and then, to 3-[(3-amino-3-oxopropyl)-methylamino] propanamide and 2-nitropyridine. Finally, both the acid and nitro-compound get transferred to H₂O, CO₂, and intermediate species.

Pathway-2. GMF decomposes to 2-amino-1-(piperazin-1-yl) ethan-1-one, which is then converted to prop-2-en-1-ol, which is then ultimately transferred to H₂O, CO₂, and fragmented species.

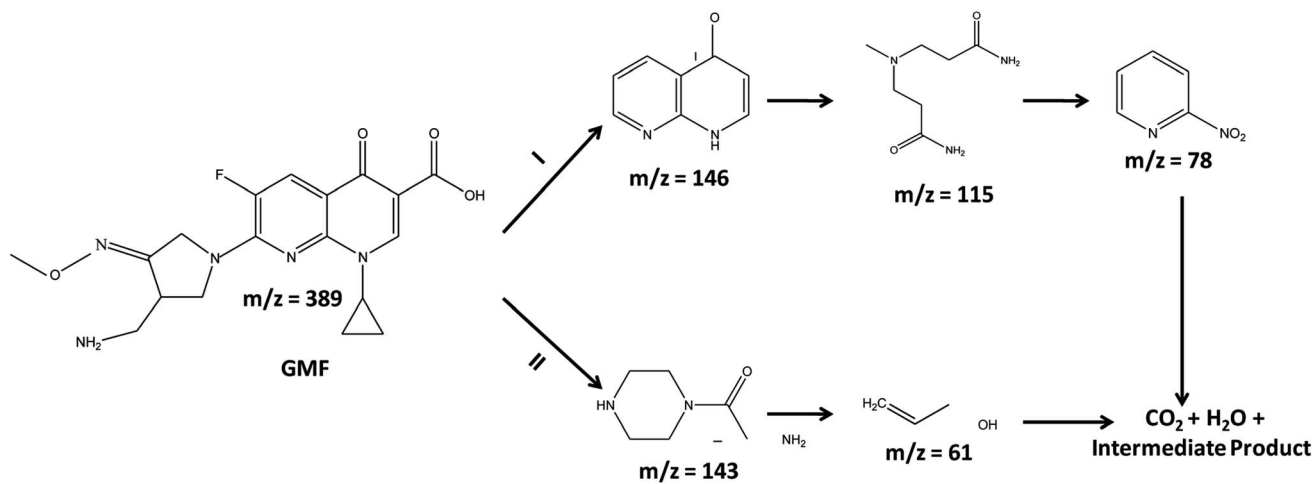
Generally, organic pollutants on mineralization or degradation are converted into carbon dioxide, water, and some intermediate products. However, nitrogen-containing aliphatic organic pollutants decompose into fragments such as amides and carboxylic acids, which further degrade to form the mineralized products. As is known, this mineralization ability of the material plays a vital role in scrutinizing the catalytic potential of the photocatalyst. In this investigation, the decomposition of antibiotic GMF over non-stoichiometric Cu_{0.75}In_{0.25}S QD was tested *via* total organic carbon (TOC) analysis. Further, the extent of mineralization in terms of TOC was calculated using the equation mentioned below (6):

$$\% \text{ of mineralization} = (T_0 - T_t) / T_0 \quad (6)$$

Here, T_0 and T_t stand for the TOC of the model pollutant, *i.e.*, GMF before and after light irradiation, respectively. Fig. S8[†] depicts the percentage of mineralization at the different time gaps. It was found that the removal efficiency was around 70% at the end of 120 min and the remaining TOC may be attributed to the presence of ring-open intermediates and small molecular products, which were detected and confirmed from the LCMS study.

Hence, from the above-mentioned results, a favorable photodegradation mechanism of GMF over the best photocatalyst Cu_{0.75}In_{0.25}S QD was illustrated in Scheme 3. The CB and VB potentials of Cu_{0.75}In_{0.25}S were found to be -1.01 V and 1.22 V vs. NHE, respectively, from *M-S* and UV analysis. As light energy falls on the photocatalyst, electron-hole pairs are generated, and these photo-induced electrons get excited to the CB of the semiconductor; simultaneously, holes are formed at the valence band. Moreover, the photo-induced electrons at the CB of Cu_{0.75}In_{0.25}S are competent for forming [•]O₂⁻ radicals due to a stronger reduction potential than the O₂/[•]O₂⁻ potential

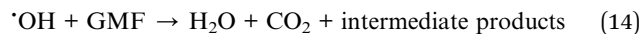
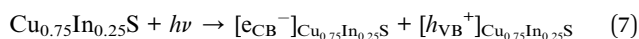




Scheme 2 Anticipated photodegradation pathways for the removal of gemifloxacin.

(−0.33 V vs. NHE). Additionally, the VB of Cu_{0.75}In_{0.25}S is 1.22 V vs. NHE, which is not suitable for the H₂O/·OH generation (+1.99 V vs. NHE), yet the degradation rate decreases with IPA as a scavenger, which signifies the important role of hydroxyl radicals. Therefore, ·OH radicals are produced indirectly from superoxide radicals and the detail ·OH radical formation and degradation mechanism with superoxide and hydroxyl radicals are formulated below. Hence, the superoxide and hydroxyl formation ability of Cu_{0.75}In_{0.25}S was well confirmed through NBT, TA, and EPR analysis.^{43,44}

The step-wise degradation mechanism is equated as below (eqn (7)–(14)):



Scheme 3 Feasible mechanism for the degradation of GMF over Cu_{0.75}In_{0.25}S QDs via ·O₂⁻ and ·OH radical under 20 W visible-LED lamp.



Conclusion

In summary, a novel composition tunable $\text{Cu}_x\text{In}_{1-x}\text{S}$ QD catalyst was designed by a reflux synthetic strategy, and its photocatalytic degradation activity was studied towards GMF under different experimental conditions. The photo-degradation activity of the $\text{Cu}_{0.75}\text{In}_{0.25}\text{S}$ catalyst was remarkably enhanced and it was superior to that of $\text{Cu}_{0.25}\text{In}_{0.75}\text{S}$ and $\text{Cu}_{0.5}\text{In}_{0.5}\text{S}$ QDs because of the improved photon absorption-ability, better anti-recombination, and effective charge carrier diffusion from the bulk to the surface. Also, the catalytic activity under various pH values, scavengers, and the anionic environment was well studied. This study proposed the degradation of GMF in the superoxide generation pathway directly and hydroxyl radical pathway indirectly, which was confirmed by active species trapping experiment, NBT test, TA test, and ESR analysis. Our findings proved that composition tunable ternary $\text{Cu}_x\text{In}_{1-x}\text{S}$ QD leads to a high rate of electron migration at the CB potential, offering a new platform for the photocatalytic degradation of quinoline-type antibiotic GMF in a greener and cleaner environment.

Conflicts of interest

The authors declare no competing financial interest.

Acknowledgements

The authors express their profound gratitude toward Siksha 'O' Anusandhan Deemed University for giving all necessary facilities and financial support to conduct this immense research work.

References

- Q. Du, P. Wu, Y. Sun, J. Zhang and H. He, *Chem. Eng. J.*, 2020, **390**, 124614.
- A. Albini and S. Monti, *Chem. Soc. Rev.*, 2003, **32**(4), 238–250.
- A. Kumar, A. Kumar and V. Krishnan, *ACS Catal.*, 2020, **10**, 10253–10315.
- M. Sharma, A. Kumar and V. Krishnan, *Nanotechnology*, 2022, **33**, 275702.
- S. Kumar, A. Dhiman, P. Sudhagar and V. Krishnan, *Appl. Surf. Sci.*, 2018, **447**, 802–815.
- S. Dhingra, M. Sharma, V. Krishnan and C. M. Nagaraja, *J. Colloid Interface Sci.*, 2022, **608**, 1040–1050.
- G. Shankaraiah, P. Saritha, D. Bhagawan, V. Himabindu and S. Vidyavathi, *S. Afr. J. Chem. Eng.*, 2017, **24**, 8–16.
- K. R. Davies, Y. Cherif, G. P. Pazhani, S. Anantharaj, H. Azzi, C. Terashima, A. Fujishima and S. Pitchaimuthu, *J. Photochem. Photobiol., C*, 2021, **48**, 100437.
- K. K. Das, S. Patnaik, S. Mansingh, A. Behera, A. Mohanty, C. Acharya and K. M. Parida, *J. Colloid Interface Sci.*, 2020, **561**, 551–567.
- S. Mansingh, K. K. Das, A. Behera, S. Subudhi, S. Sultana and K. Parida, *Nanoscale Adv.*, 2020, **2**, 2004–2017.
- K. K. Das, D. P. Sahoo, S. Mansingh and K. Parida, *ACS Omega*, 2021, **6**, 30401–30418.
- D. P. Sahoo, K. K. Das, S. Patnaik and K. Parida, *Inorg. Chem. Front.*, 2020, **7**, 3695–3717.
- S. Mansingh, R. Acharya, S. Martha and K. M. Parida, *Phys. Chem. Chem. Phys.*, 2018, **20**, 9872–9885.
- J. Kolny-Olesiak and H. Weller, *ACS Appl. Mater. Interfaces*, 2013, **5**, 12221–12237.
- Z. Mei, G. Wang, S. Yan and J. Wang, *Acta Phys.-Chim. Sin.*, 2021, **37**(6), 2009097.
- J. Bai, W. Chen, L. Hao, R. Shen, P. Zhang, N. Li and X. Li, *Chem. Eng. J.*, 2022, **447**, 137488.
- J. Bai, R. Shen, K. Zhou, Z. Jiang, P. Zhang and X. Li, *Chin. J. Catal.*, 2022, **43**(2), 359–369.
- W. Xue, W. Chang, X. Hu, J. Fan and E. Liu, *Chin. J. Catal.*, 2021, **42**(1), 152–163.
- Z. Xiong, Y. Hou, R. Yuan, Z. Ding, W. J. Ong and S. Wang, *Acta Phys.-Chim. Sin.*, 2022, **38**, 2111021.
- F. Mei, Z. Li, K. Dai, J. Zhang and C. Liang, *Chin. J. Catal.*, 2020, **41**, 41–49.
- D. Prusty, L. Paramanik and K. M. Parida, *Energy Fuels*, 2021, **35**, 4670–4686.
- Z. Long, W. Zhang, J. Tian, G. Chen, Y. Liu and R. Liu, *Inorg. Chem. Front.*, 2021, **8**, 880–897.
- W. Hu, S. Yang and J. Huang, *J. Chem. Phys.*, 2019, **151**, 214705.
- X. B. Fan, S. D. Yu, F. D. Zhang, L. P. Zhang, Y. Tao, C. H. Tung and L. Z. Wu, *ChemSusChem*, 2017, **10**, 4833–4838.
- C. Xia, J. D. Meeldijk, H. C. Gerritsen and C. M. Donega, *Chem. Mater.*, 2017, **29**, 4940–49512017.
- D. Prusty, S. Mansingh, L. Acharya, L. Paramanik and K. M. Parida, *RSC Adv.*, 2022, **12**, 1265–1277.
- Z. Guan, J. Pan, Q. Li, G. Li and J. Yang, *ACS Sustainable Chem. Eng.*, 2019, **7**, 7736–7742.
- Q. Wang, W. Wang, L. Zhong, S. Liu, X. Cao and F. Cui, *Appl. Catal., B*, 2018, **220**, 290–302.
- Y. Chen, S. Li, L. Huang and D. Pan, *Inorg. Chem.*, 2013, **52**, 7819–7821.
- A. Ishikawa, T. Takata, J. N. Kondo, M. Hara, V. Kobayashi and K. Domen, *J. Am. Chem. Soc.*, 2002, **124**, 13547–13553.
- Z. Wei, F. Liang, Y. Liu, W. Luo, J. Wanga, W. Yao and Y. Zhu, *Appl. Catal., B*, 2017, **201**, 600–606.
- A. S. Dezfuli, M. R. Ganjali, H. R. Naderi and P. Norouzi, *RSC Adv.*, 2015, **5**, 46050–46058.
- H. Huang, K. Xiao, Y. He, T. Zhang, F. Dong, X. Du and Y. Zhang, *Appl. Catal., B*, 2016, **199**, 75–86.
- L. Shen, W. Wu, R. Liang, R. Lin and L. Wu, *Nanoscale*, 2013, **5**, 9374–9382.
- F. A. Ibrahim, M. A. Al-Ghobashy, A. El-Rahman, K. Mohamed and I. F. Abo-Elmagd, *Environ. Sci. Pollut. Res.*, 2017, **24**, 23880–23892.
- A. Cai, J. Deng, M. Xu, T. Zhu, S. Zhou, J. Li, G. Wang and X. Li, *J. Chem. Eng.*, 2020, **395**, 125090.
- W. K. Jo and S. Tonda, *J. Hazard. Mater.*, 2019, **368**, 778–787.
- A. Khataee, P. Gholami, B. Kayan, D. Kalderis, L. Dinpazhoh and S. Akay, *Ultrason. Sonochem.*, 2018, **48**, 349–361.



- 39 Z. Chen, X. Li, Q. Xu, Z. Tao, F. Yao, X. Huang, Y. Wu, D. Wang, P. Jiang and Q. Yang, *Chem. Eng. J.*, 2020, **390**, 124454.
- 40 D. Prusty, S. Mansingh and K. M. Parida, *Energy Adv.*, 2022, **1**, 422–437.
- 41 K. Saravanakumar, V. Maheskumar, Y. Yea, Y. Yoon, V. Muthuraj and C. M. Park, *Composites, Part B*, 2022, **234**, 109726.
- 42 P. Gholami, A. Khataee, R. D. C. Soltani, L. Dinpazhoh and A. Bhatnagar, *J. Hazard. Mater.*, 2020, **382**, 121070.
- 43 H. Wang, X. Li, X. Zhao, C. Li, X. Song, P. Zhang and P. Huo, *Chin. J. Catal.*, 2022, **43**, 178–214.
- 44 P. G. Kougias and I. Angelidaki, *Front. Environ. Sci. Eng.*, 2018, **12**, 14.

