



Cite this: DOI: 10.1039/d4em00402g

Combining sequential extractions with bulk and micro X-ray spectroscopy to elucidate iron and phosphorus speciation in sediments of an iron-treated peat lake†‡

Melanie A. Münch,^a Andreas Voegelin,^b Luis Carlos Colucho Hurtarte,^{c,d} Jörg Göttlicher^e and Thilo Behrends^a

In shallow lakes, mobilization of legacy phosphorus (P) from the sediments can be the main cause for persisting eutrophication after reduction of external P input. In-lake remediation measures can be applied to reduce internal P loading and to achieve ecosystem recovery. The eutrophic shallow peat lake Terra Nova (The Netherlands) was treated with iron (Fe) to enhance P retention in the sediment. This treatment, however, intensified seasonal internal P loading. An earlier study suggested that Fe addition led to increased P binding by easily-reducible Fe(III) associated with organic matter (OM), which readily releases P when bottom waters turn hypoxic. In this complementary study, bulk and micro Fe K-edge and P K-edge X-ray absorption spectroscopy and micro-focused X-ray fluorescence spectroscopy were applied to characterize the P hosting Fe(III) pool. Combined with sequential extraction data, the synchrotron X-ray analyses revealed that a continuum of co-precipitates of Fe(III) with calcium, phosphate, manganese and organic carbon within the OM matrix constitutes the reducible Fe(III) pool. The complementary analyses also shed new light on the interpretation of sequential extraction results, demonstrating that pyrite was not quantitatively extracted by nitric acid (HNO₃) and that most of the Fe(II) extracted by hydrochloric acid (HCl) originated from phyllosilicate minerals. Formation of an amorphous inorganic–organic co-precipitate upon Fe addition constitutes an effective P sink in the studied peaty sediments. However, the high intrinsic reactivity of this nanoscale co-precipitate and its fine distribution in the OM matrix makes it very susceptible to reductive dissolution, leading to P remobilization under reducing conditions.

Received 5th July 2024
Accepted 11th November 2024

DOI: 10.1039/d4em00402g

rsc.li/espi

Environmental significance

Impairment of surface water quality due to eutrophication is a global issue, causing ecosystem degradation and increasing greenhouse gas emissions. In many lakes, reduction of internal phosphorus loading is needed for ecosystem recovery. Iron addition offers a cost-effective remediation strategy, but little is known about its long-term effects in organic-rich systems. Our study shows that addition of Fe salts to peat lakes can induce the formation of nanoscale inorganic–organic co-precipitates embedded within the organic matter matrix that are prone to reductive dissolution. These co-precipitates can effectively bind phosphorus, but in a form highly susceptible to remobilization under reducing conditions. In organic-rich lakes, iron remediation can thus have adversary effects and potentially increase temporary internal P loading during bottom water hypoxia.

1 Introduction

Phosphorus (P) is an essential and often limiting or co-limiting macro-nutrient in freshwater ecosystems.^{2,3} Increased P input

due to human activities such as fertilizer use or wastewater disposal is one of the main causes for the eutrophication and deterioration of freshwater ecosystems.⁴ In industrialized countries, decades of anthropogenic P input have caused P to

^aUtrecht University, Princetonlaan 8A, 3584CB Utrecht, The Netherlands. E-mail: m.a.munch@uu.nl

^bEawag, Swiss Federal Institute of Aquatic Science and Technology, Überlandstrasse 133, CH-8600 Dübendorf, Switzerland

^cEuropean Synchrotron Radiation Facility, Beamline ID21, Grenoble 38100, France

^dDiamond Light Source, Harwell Science and Innovation Campus, Didcot OX11 0DE, UK

^eKarlsruhe Institute of Technology, Institute for Photon Science and Synchrotron Radiation, Hermann-von-Helmholtz Platz 1, D-76344 Eggenstein-Leopoldshafen, Germany

† Preceding study: M. A. Münch, R. V. Kaam, K. As, S. Peiffer, G. T. Heerdt, C. P. Slomp and T. Behrends, Impact of iron addition on phosphorus dynamics in sediments of a shallow peat lake 10 years after treatment, *Water Res.*, 2023, **248**, 120844, <https://doi.org/10.1016/j.watres.2023.120844>.

‡ Electronic supplementary information (ESI) available. See DOI: <https://doi.org/10.1039/d4em00402g>



accumulate in soils and freshwater sediments.^{5–9} This so-called “legacy P” is a major water management challenge, since it can sustain high surface water P concentrations by internal P loading, even when external P sources have been significantly reduced, thus delaying restoration efforts for decades.^{10–12}

Methods to reduce internal P loading include iron (Fe) amendment.^{13,14} This method is based on the classical model of Fe and P diagenesis: the affinity of phosphate to Fe(III)(oxyhydr)oxides, which bind phosphate by adsorption or coprecipitation, closely couples the cycling of P and Fe in environmental systems.^{15–17} Fe amendment thus can enhance the retention of P in lake sediments by binding it to the solid phase and recycling it as a passenger of the Fe redox cycle.^{18,19} The natural coupling of the geochemical cycles of P and Fe makes Fe amendment less invasive than treatments with other chemicals such as *e.g.* aluminium (Al)-compounds or lanthanum (La)-modified bentonite. However, due to the redox sensitivity of Fe, the success of the treatment is highly dependent on the geochemistry of the sediment. In their reviews of case studies, Bakker *et al.*¹³ and Smolders *et al.*¹⁴ concluded that the effect and success of Fe amendment was dependent on the extent of sulfate reduction and the presence of organic matter (OM) in the sediments. Sulfate reduction produces sulfide which precipitates with Fe²⁺ forming Fe-sulfides (FeS_x) and ultimately pyrite (FeS₂), stabilizing Fe(II) and burying it into deeper sediment layers.^{20,21} OM influences the interaction of P with Fe by competition for adsorption sites on the Fe(III)(oxyhydr)oxide surface, hindrance of crystallization of Fe(III) minerals,²² the formation of monomeric and oligomeric Fe-OM complexes^{23–27} and the association with and stabilization of Fe(III)-(hydr)oxide colloids.^{28–30} The OM-associated Fe can bind P with Fe acting as the bridging ion between P and OM, a phenomenon which was also reported involving other metal cations such as Al, calcium (Ca), copper (Cu), zinc (Zn), magnesium (Mg) and manganese (Mn).^{31–33} The resulting P-metal-OM associations account for a substantial part of P in natural environments, as reported for natural waters^{24,34} and postulated for soils and sediments.^{1,29,31,32,35,36}

The effects of sulfate reduction and OM on the coupling of Fe and P cycling in lake sediments is reflected by the speciation of the two elements in the sediments. Hence, knowledge on the solid phase P and Fe speciation is pivotal for the design of Fe-based remediation measures in freshwater systems. Sequential extraction procedures (SEPs), *i.e.*, extraction over multiple sequential chemical extraction steps, are widely used to assess the binding forms of solid-phase Fe and P. Widely used and often adapted SEPs for P (P-SEP) from sediments were developed by Ruttenberg for marine sediments³⁷ and the one by Psenner *et al.*³⁸ for lake sediments and suspended matter. Popular SEPs for Fe (Fe-SEP) from sediments include the procedures developed by Claff *et al.*³⁹ for acid sulfate soils, with special focus on iron sulfide minerals, and by Poulton and Canfield⁴⁰ for modern and ancient sediments.

SEPs are based on consecutive extraction steps of sediment samples at increasing desorptive and/or proton-, ligand-, or reduction-promoted extraction strength, targeting solid phase Fe and P pools of increasing recalcitrance. This approach yields

operationally defined Fe and P pools, which can be hypothetically assigned to distinct Fe and P phases based on the extractability of pure minerals.^{37–41} However, the interpretation of SEPs can be challenging, especially also in very organic-rich sediments like those of peat lakes. These sediments may contain significant amounts of P–Fe–OM phases, which are not specifically targeted in SEPs. Furthermore, the extractability of autochthonous Fe and P phases could be different in sediments of peat lakes where they form in an organic matrix.

To elucidate the nature of reactive Fe and P phases in sediments or soils, SEPs have been combined with other methods such as ³¹P-NMR, XRD, Mössbauer or XAS.^{36,42–45} To date, however, studies that systematically relate chemical fractionation to spectroscopic speciation results for P and Fe in lake sediments are scarce, in particular for organic-rich sediments. A recent study of a boreal peat lake³⁶ combined different P-SEP approaches, with an extraction step for humic-acid-bound P⁴⁶ added to the SEDEX procedure,³⁷ and used the approach of Poulton and Canfield⁴⁰ to fractionate Fe. From the combination of these SEP with ³¹P-NMR, ⁵⁷Fe-Mössbauer and Fe K-edge and P K-edge XAS, they concluded that in spite of the anoxic conditions, a large part of the sedimentary Fe was present as Fe(III) in a continuum of ferrihydrite-like phases with a distribution of crystallinity and variable amounts of associated P and OM. They also showed that spectroscopic analyses can provide complementary information on the interpretation of SEP results. For example, SEP results indicated the presence of carbonate-associated Fe, reducible but more crystalline Fe(III)(oxyhydr)oxides and magnetite. In contrast ferrihydrite-like phases and minor amounts of nano-goethite and Fe in phyllosilicates were identified by Mössbauer spectroscopy but no magnetite or Fe-carbonates.

This study focusses on the Fe and P speciation in sediments from the shallow peat lake Terra Nova (NL), which was treated with FeCl₃ in 2010 to reduce internal P loading.¹ The Fe-amendment efficiently lowered the surface water P concentrations for 1.5 years before seasonal P release started again with increasing intensity, resulting in the highest P concentrations since the start of monitoring and renewed algae blooms. Our investigation of the changes in sediment geochemistry due to the Fe addition suggested that the pronounced seasonal release of P was due to the high redox turnover of Fe associated with OM.¹ The interpretation in this previous paper was based on a combination of P-SEP (adapted from O’Connell *et al.*³⁶ and Slomp *et al.*⁴⁷), Fe-SEP (adapted from Poulton and Canfield,⁴⁰ Raiswell *et al.*⁴¹ and Claff *et al.*⁴⁸), porewater analysis and sediment incubation experiments. The P-SEP and Fe-SEP results suggested the presence of a reactive Fe-OM pool, which interacts with P. The surface sediment, in the depth interval affected by the Fe addition, was characterized by elevated contents of sodium bicarbonate extract P (P-NaHCO₃) and ascorbic acid extractable Fe (Fe-Asc) with respect to the deeper sediment. Sodium bicarbonate is conceived to extract P associated to humic substances over metal cation bridges.^{36,46} Ascorbic acid, a weak reducing agent, targets easily reducible Fe such as in ferrihydrite.⁴¹ Although the two extractions are considered to dissolve different Fe phases and associated P, we hypothesize



that the extracted Fe and P originate from the same highly reactive Fe pool. However, the nature of this postulated, highly reactive Fe pool and the bound P remained elusive.

In the present study, we therefore aim at collecting complementary information on the speciation of the corresponding Fe and P pools and their occurrence in the lake sediments. For this, we performed bulk and micro X-ray absorption spectroscopy and X-ray fluorescence and correlated the results to sediment geochemistry data, especially of Fe-SEP and P-SEP, collected in our earlier study.

2 Methods

2.1 Study site

Terra Nova is a shallow, polymictic peat lake of the Loosdrecht lake system in The Netherlands. The local water authority categorizes it as eutrophic based on the EU Water Framework Directive.⁴⁹ It formed due to peat extraction in the early modern age, spanning an area of 0.85 km² with a maximum depth of 2 m and a mean depth of 1 m. Starting in 2010 the lake was treated with 40% FeCl₃ solution over a period of 1.5 years, with a total average Fe dose of 0.5 mol Fe m⁻² as a remediation measure to reduce internal P loading.¹

2.2 Sediment sampling and analysis

Undisturbed sediment cores (length 60 cm, \varnothing 6 cm, min. 10 cm overlying water) were collected at three positions in lake Terra Nova in June 2020 using a gravity corer (UWITEC, Mondsee, Austria). The cores were stored at bottom water temperature (15 °C) under preservation of bottom water oxygen conditions and sliced in an N₂ purged glovebox within three days (sampling resolution: 1 cm for top 10 cm, 2 cm for rest of the cores). The solid phase was separated from the porewater by centrifugation (10 minutes, 3000 rpm), freeze-dried anoxically (ventilation with N₂), homogenized in an N₂ purged glovebox using an agate mortar and pestle, and stored in the glovebox at room temperature.¹

In this study we used depth-resolved solid-phase total elemental contents, sequential extraction derived Fe and P solid-phase speciation and porewater concentrations from three sediment cores, each from its respective sampling site (ESI A_‡). Details on the chemical analysis of the sediment porewater and solid phase and the detailed data sets are reported in the preceding study.¹ Total Fe, P, Ca, Al, S, and Mn contents of the sediment solid phase were determined by HF digestion; TOC (total organic carbon) and TON (total organic nitrogen) were determined using an NCS analyzer as reported by Helmond *et al.*⁵⁰

Fe and the P speciation in the sediment solid phase was determined by dedicated sequential extraction procedures (SEP), conducted on samples across the entire depth of three sediment cores (one per station A to C, ESI A_‡). Here, concise information regarding the SEPs is listed. Details on the used procedures and the corresponding detailed data sets are reported in Münch *et al.*¹ The sequential Fe extraction (Fe-SEP) targeting different, operationally defined, Fe pools in

sediments included ascorbic acid extractable Fe (Fe-Asc; 0.057 M ascorbic acid, carbonate buffered, pH 7.5; easily reducible Fe in ferrihydrite⁴¹), HCl-soluble Fe(II) (Fe(II)-HCl; 1 M HCl; FeS and FeCO₃ (ref. 48)), HCl-soluble Fe(III) (Fe(III)-HCl; 1 M HCl; amorphous Fe(oxyhydr)oxides⁴⁸), dithionite-reducible Fe (Fe-CDB; acetic acid-buffered citrate-dithionite solution, pH 4.8; crystalline Fe (oxyhydr)oxides such as lepidocrocite, goethite and hematite⁴⁸), oxalate-soluble Fe (Fe-Ox; 0.37 M oxalate, pH 3.2; magnetite^{51,52}), HNO₃-soluble Fe (Fe-HNO₃, conc. HNO₃ at room temperature for 2 hours; pyrite⁴⁸). All residual Fe (not-extr. Fe) was regarded as non-reactive Fe, mainly Fe bound in clays.⁴⁰ Our earlier study suggested, that Fe-Asc also included mono- or oligomeric Fe associated with OM.¹ As the extraction only yielded negligible amounts of Fe(III)-HCl and Fe-Ox, this data is not shown here. The sequential P extraction (P-SEP) provided an operational distinction of MgCl₂-soluble P (P-MgCl₂; 1 M MgCl₂, pH 8; adsorbed P⁴⁷), NaHCO₃-extractable P (P-NaHCO₃; 1 M NaHCO₃, pH 8; P bound to OM over metal cation bridges such as Fe, Al, Ca^{46,53}) citrate-dithionite-extractable P (P-CDB; carbonate-buffered citrate-dithionite solution, pH 7.6; P bound to Fe (oxyhydr)oxides and CaCO₃ as well as Fe(II) phosphates⁴⁷), HCl-soluble P (P-HCl; 1 M HCl, room temperature, 24 hours; apatite⁴⁷), and P extractable by HCl after ashing (P-ashed/HCl; 550 °C for 2 hours, 1 M HCl at room temperature for 24 hours; organic P⁴⁷). As the extraction only yielded negligible amounts of P-MgCl₂, this data is not shown here. All solid-phase P was extracted within the 10% error margin.

2.3 X-ray absorption spectroscopy (XAS) and micro X-ray fluorescence (μ -XRF)

Bulk Fe K-edge XAS measurements were performed at the SUL-X beamline at the KIT Light Source (Karlsruhe Institute of Technology, Germany), at room temperature in transmissions mode. A Si(111) double-crystal monochromator was used for X-ray energy monochromatization. Bulk P K-edge XAS, μ -XRF and spatially resolved micro-focused XAS (μ -XAS) at the P and Fe K-edges were performed at the ID21 beamline at the European Synchrotron Radiation Facility (ESRF; Grenoble, France; storage ring energy: 6.03 GeV; Beamline design: Salomé *et al.*, 2013;⁵⁴ Beamline configuration: Werner *et al.*, 2017 (ref. 55)), at room temperature in fluorescence mode. A Si(111) double-crystal monochromator was used for X-ray energy selection and a compact Kirkpatrick-Baez mirror system (KB), based on elliptically shaped, fixed-focus Ni coated mirrors, was used for X-ray focussing.

Photon energy was calibrated based on the first maximum of the first derivative of the K-edges of reference compounds (metallic Fe: 7112 eV; tricalcium phosphate: 2153 eV), respectively.

2.3.1 Bulk Fe K-edge XAS measurements. Sediment samples from fifteen depth intervals (1–36 cm) of one sediment core from station B (ESI A_‡, Münch *et al.*¹) were analyzed by Fe K-edge X-ray absorption spectroscopy (XAS), including X-ray absorption near edge structure (XANES) and extended X-ray absorption fine structure (EXAFS) spectroscopy. For analysis, ~35 mg of freeze-dried and ground sediment was thoroughly



mixed with ~65 mg cellulose and pressed into 7 mm diameter pellets to achieve an Fe content of ~1 mg per pellet, equivalent to an adsorption edge step of ~0.5–1. After preliminary tests aimed at minimizing radiation exposure and damage, between 3 to 5 scans of 15 min each were collected at different locations on each pellet, examined for spectral variations, and merged to obtain an averaged sample spectrum.

2.3.2 Bulk P K-edge XAS measurements. XANES spectra at the P K-edge were collected from sediment samples retrieved at four depth intervals (2, 8, 38 and 40 cm depth) of one sediment core from station A (ESI A, ‡ Münch *et al.*¹). For analysis, freeze-dried and ground sediment was thoroughly mixed, pressed into 3 mm diameter pellets and mounted on aluminum sample holders using sulfur-free tape (FLUXANA GmbH). The XAS data was recorded in fluorescence mode with an unfocussed beam (350 × 350 μm beam size, no KB mirrors).

2.3.3 Scanning X-ray microscopy (SXM). Spatially resolved μ-XRF and μ-XAS analyses were performed on two sediment samples (2 and 40 cm depth) of one sediment core from station A (ESI A, ‡ Münch *et al.*¹). Freeze-dried and ground samples were mounted on aluminum sample holders using sulfur-free tape (FLUXANA GmbH) on the shadow side and XRF thin window film (Spex SamplePrep 3525 Ultralene® window film, 4 μm thick, Cole-Parmer) on the beam side. The samples were handled and stored in a glovebox under N₂ atmosphere until right before measurement. μ-XRF maps were collected above the Fe (7200 eV) and P (2400 eV) K-edge energies, with a beam dwell time of 0.1 s and a beam size of (vertical × horizontal) ~500 × 1000 nm and ~500 × 1300 nm, respectively. Selected points of interest were measured with spot XAS across the Fe K-edge (7000–7650 eV) or across the P K-edge (2103–2400 eV) with a resolution of 0.5 and 0.25 eV and a dwell time of 0.05 s and 0.1 s, respectively. All spectra were collected in fluorescence mode.

Positions of the μ-XRF maps and μ-XAS point spectra were selected to cover the largest possible variety of Fe and P phases. Positions for lower resolution overview maps (step size 10 μm) were chosen based on differences in material morphology as assessed by optical microscopy on the mounted samples prior to X-ray beam exposure. Positions for high-resolution maps (step sizes 2 μm or 0.5 μm) were chosen based on Fe and P distributions (hotspots, occurrence of Fe and P in higher and lower intensities) and elemental correlations (Fe, P, S, Ca, Si, Al) observed in the low-resolution overview maps. Positions of μ-XAS point spectra (Fe and P K-edges) were selected per high-resolution map based on the same criteria. μ-XAS point spectra were measured right after collection of the respective high-resolution maps to guarantee position precision.

2.3.4 XAS data processing and analysis. All collected XAS spectra were extracted and evaluated by linear combination fitting (LCF) using the software Athena.⁵⁶

2.3.4.1 Bulk Fe K-edge XAS. For the bulk Fe K-edge XAS spectra, the pre-edge was fitted by a first-order polynomial for background removal (–120 to –20 eV before E_0 ; with E_0 set to 7123 eV) and the post edge was fitted by a third-order polynomial for normalization (50 to 500 eV after E_0). The Autobk algorithm ($R_{\text{bkg}} = 0.85$; k -weight = 2; spline k -range 0–12.4 Å^{–1})

was used to extract the EXAFS spectra, and a Kaiser–Bessel window ($dk = 2$) was used to calculate the Fourier-transform of the k^3 -weighted EXAFS spectra (k -range = 2–10 Å^{–1}). The XANES spectra were analyzed by LCF over an energy-range of –20 to 180 eV and the EXFAS spectra over a k -range of 2 to 10 Å^{–1}. Individual fractions were constrained to values between 0 and 1, whereas the sum of all fractions was not constrained.

Preliminary LCF analyses were performed on the EXAFS and the XANES spectra separately, including an extended set of reference spectra considered relevant for the studied samples. References that were not included in significant fractions in preliminary fits or did not improve the r -factor (normalized sum of square residuals = $\sum(\text{data} - \text{fit})^2 / \sum \text{data}_i^2$) of the fits by at least 10%, were consecutively removed. Finally, six reference spectra that allowed to reproduce all EXAFS and XANES spectra were identified and used for LCF analysis: Fe(III)–HA, CaFeP, Fe(II)aq., smectite, biotite and pyrite (Table 1, ESI B1, and B2 ‡).

2.3.4.2 Bulk P K-edge XAS. Bulk P K-edge XAS spectra were normalized by applying linear functions for the pre-edge (–22 to –5 eV before E_0 , $E_0 = 2152$ eV, post edge 28 to 45 eV above E_0). The corresponding energy ranges were optimized to give consistent normalizations including spectra or references collected in transmission mode or in previous experiments. The spectra were analyzed by LCF over an energy-range of –10 to 30 eV (XANES), constraining individual fractions to values between 0 and 1 and the sum of the fractions to 1.0.

Preliminary LCF analyses were performed using an extended set of reference spectra considered relevant for the studied samples, consecutively removing references that were not used for the fits. Finally, all bulk P-XANES spectra could be described using DNA and Mg inositol hexaphosphate (Mg-IP6) as references (Table 2, ESI C ‡).

2.3.4.3 Fe and P K-edge μ-XAS. For the Fe K-edge point spectra, the pre-edge was fitted by a first order polynomial for background removal (–100 to –10 eV before the edge) and the post edge was fitted by a third order polynomial for normalization (50 to 800 eV, edge step at 7120 eV). Per spot one spectrum was collected, as the first measurements with multiple scans per spot showed indications of beam damage (ESI D1 ‡).

The P K-edge μ-XAS spectra were extracted and normalized as were the bulk P K-edge XAS spectra (Section 2.3.4.2). At each spot, three scans were collected.

A total of 190 and 90 viable μ-XANES spectra were recorded at the Fe and P K-edges, respectively. The spectra collected at the respective K-edges were compared by eye after extraction and normalization. All spectra that were identical within the range of noise (variations within the small high frequency features of the spectra) were merged using I_0 normalized spectra to yield concentration weighed average spectra. Average spectra of point spectra measured on different maps from the same sample were classified to represent an abundant type (Type 1 to n). Average spectra of point spectra all located on the same map (distance larger than beam size) were classified as regions (Region 1 to n). Unique single spectra were classified as spots (Spot 1 to n). Numbering of the classified spectra was done according to abundance with the number 1 of one class being the one comprised of most



Table 1 Reference spectra used for LCF analysis of Fe K-edge XAS sample spectra

Name	Material	Reference
Fe(III)-HA	Pahokee peat humic acid (International Humic Substances Society) loaded with 10 000 mg kg ⁻¹ Fe(III)	23
CaFeP	Amorphous Ca-Fe(III)-phosphate formed by the oxidation of 0.5 mM Fe(II) in synthetic bicarbonate-buffered groundwater at pH 7.0 in presence of 4 mM Ca and 0.75 or 1 mM P, without/with 0.5 mM Si (average of four identical spectra) (CaFeP*)	57
Fe(II)aq.	Aqueous solution of 50 mM FeSO ₄ and 1 mM HCl	This work
Smectite	SWy-2 (Clay Minerals Society), Fe accounting for 15% of cations in octahedral sheet; 7% Fe(II)	58
Biotite	Natural biotite (Fe accounting for 71% of cations in octahedral sheet, 87% Fe(II))	58
Pyrite	Natural pyrite (obtained from the mineral collection of ETH Zurich)	59
Greigite	Greigite synthesized using the method of ref. 60 according to ref. 61. Unpublished spectrum recorded in fluorescence mode at room temperature kindly provided by ed. D. Burton (Southern Cross University, Australia)	This work
Mackinawite	Amorphous FeS synthesized using the method of ref. 62	59
Hematite	Commercial synthetic hematite (Riedel-de-Haën)	This work
Vivianite	Synthetic vivianite, synthesized according to ref. 63, as reported in ref. 64	59

Table 2 Reference spectra used for linear combination fitting of the collected P K-edge XAS spectra

Name	Material	Reference
Apatite	Commercial hydroxyapatite	65
Mg-IP6	Amorphous Mg inositol hexaphosphate (phytate) synthesized using the method of ref. 66 according to ref. 67. Unpublished spectrum kindly provided by Jörg Prietzel (Technical University of Munich, Germany)	This work
Fe-IP6	Amorphous Fe(III) inositol hexaphosphate (phytate) precipitates synthesized following the procedure of: ⁶⁶ 100 ml of 0.2 M HCl was mixed with 30 ml of 0.05 M phytic acid dodecasodium salt (C ₆ H ₆ O ₂₄ P ₆ Na ₁₂) – solution, to which enough Fe(III)Cl ₃ – solution (pH 2) was added to get a P/Fe ratio of 1. The resulting precipitate was filtered and washed three times with boiling deionized water to remove residual adsorbed Fe ions	67
DNA	Commercial deoxyribonucleic acid sodium salt	68

individual spectra. The classified average μ -XANES spectra are listed in Table 3 for Fe and in Table 4 for P, including information on how many spectra were merged and on how many μ -XRF maps those spectra were located. The identity of the individual spectra are listed in ESI Table D2(1).[‡]

Average Fe μ -XANES spectra were analyzed by LCF over an energy-range of –20 to 80 eV (XANES), constraining individual fractions to values between 0 and 1 and leaving the sum of the fractions unconstrained. As a first fit, the same six references were used as for LCF of the bulk Fe XAS spectra (see Section 2.3.4.1). If the fit was not satisfying, additional references were included to find the optimal fit. The quality of a fit was assessed according to its *r*-factor and the visual match between sample and LCF spectra. All fits with an *r*-factor within 20% of the lowest *r*-factor were considered as statistically viable fits. All references used are listed in Table 1 and shown in ESI Fig. B1(1).[‡] The LCF data is listed in ESI Table B4(1).[‡]

Average P μ -XANES spectra were analyzed by LCF in analogy to the bulk P XANES spectra (Section 2.3.4.2). As for LCF of the

bulk P K-edge XANES spectra, preliminary fits with the extended set of reference spectra and consecutive removal of non-used references were performed. Finally, all P μ -XANES spectra could be described using DNA, Mg inositol hexaphosphate (Mg-IP6), Fe inositol hexaphosphate (Fe-IP6) and hydroxyapatite (apatite) as references. All references used are listed in Table 2 and shown in ESI Fig. C1(1).[‡] The LCF data is listed in ESI Table C4(1).[‡]

2.3.5 μ -XRF data processing and analysis. From the μ -XRF data, elemental distribution maps were obtained by deconvolution of the μ -XRF pixel spectra, after incoming flux and detector deadtime correction, using the PyMCA software.⁶⁹ The elemental maps were further processed in Fiji.⁷⁰ The individual element maps were scaled using a histogram-based thresholding algorithm.⁷¹ Subsequently, the scaled element distribution maps were merged with the color channels tool to generate tri-element/tri-color RGB (red, green, blue) maps.

2.3.6 Statistical analyses. To obtain complementary information on the nature of Fe and P pools in the sediment and to further explore the relationship between SEP and XAS analyses, Pearson correlation analysis was performed on the results from solid phase analyses. To this end, rank correlation coefficient matrices of total element contents, SEP-based P fractions and Fe fractions, as well as LCF results from bulk Fe K-edge XAS measurements were produced. For the latter, relative concentrations were multiplied with the respective total element contents so that all variables are in units of μ mol per g dw. The calculations were performed using the *rcorr* function of the *Hmisc* package in R.⁷² For contrasting the sediment layers affected by Fe treatment to those below, correlations were computed separately for total elemental contents, Fe-SEP and P-SEP contents of the top 10 cm of sediment (three sediment cores, 1–10 cm depth, *n* = 30 samples) and of the investigated sediment below 10 cm depth (three sediment cores, 12 – max. 44 cm depth, *n* = 44 samples). Due to the limited number of XAS measurements, a separate correlation matrix was produced for all the samples on which bulk XAS measurements had been performed without depth separation (one sediment core, 1–36 cm depth, *n* = 15 samples).



Table 3 Classification of the merged Fe K-edge μ -XANES spectra. The identities of the individual spectra are listed in ESI Tables D1 and 2

Class	Top: 2 cm depth					Bottom: 40 cm depth				
	Merged spectra		Number of maps	Number of spectra	% of spectra	Merged spectra		Number of maps	Number of spectra	% of spectra
	Name	Abbreviation				Name	Abbreviation			
Abundant type	Type 1	t-T1-Fe	5	41	43	Type 1	b-T1-Fe	4	20	21
	Type 2	t-T2-Fe	4	25	26	Type 2	b-T2-Fe	3	19	20
	Type 3	t-T3-Fe	4	8	8	Type 3	b-T3-Fe	2	7	7
						Type 4	b-T4-Fe	3	6	6
Region	Region 1	t-R1-Fe	1	9	9	Region 1	b-R1-Fe	1	9	9
	Region 2	t-R2-Fe	1	7	7	Region 2	b-R2-Fe	1	8	8
	Region 3	t-R3-Fe	1	2	2	Region 3	b-R3-Fe	1	6	6
	Region 4	t-R4-Fe	1	3	3	Region 4	b-R4-Fe	1	6	6
Spot						Region 5	b-R5-Fe	1	5	5
						Region 6	b-R6-Fe	1	3	3
						Region 7	b-R7-Fe	1	3	3
						Region 8	b-R8-Fe	1	2	2
Total				95	Spot 1	b-S1-Fe	1	1	1	95

Table 4 Classification of the merged P K-edge μ -XANES spectra. All P μ -XANES spectra were collected in triplicates. The total number of spectra including triplicates is listed. The identities of the individual spectra are listed in ESI Tables D1 and 2

Class	Top: 2 cm depth					Bottom: 40 cm depth				
	Merged spectra		Number of maps	Number of spectra	% of spectra	Merged spectra		Number of maps	Number of spectra	% of spectra
	Name	Abbreviation				Name	Abbreviation			
Abundant type	Type 1	t-T1-P	4	33	85	Type 1	b-T1-P	3	24	47
						Type 2	b-T2-P	3	18 ^a	35
Region						Region 1	b-R1-P	1	9	18
	Spot	Spot 1	t-S1-P	1	3	8				
		Spot 2	t-S2-P	1	3	8				
Total				39					51	

^a Two scans showed compressed spectra, which were not included in the LCF. Thus, 12 spectra were used for LCF.

3 Results

3.1 Total elemental contents

Fig. 1 shows the averaged depth profiles of elemental contents obtained from sampling the solid phase of three sediment cores (one per station A to C, ESI A†). The enrichment of Fe in the top 5 cm with respect to the deeper sediment, reflected in increased Fe/Al ratios, can be attributed to the Fe treatment.¹ In the same depth range, the contents of Ca and Mn were higher compared to the deeper sediment. Sulfur (S) contents were maximum at 20 cm depth with *ca.* 0.8 mmol per g dw. The Al contents were around 0.95 mmol per g dw in the upper 18 cm, below which they increased to 1.1 mmol per g dw. At the sediment surface, the P content was *ca.* 80 μ mol per g dw and decreased with depth to *ca.* 25 μ mol per g dw. Total organic carbon (TOC) contents were constant with depth at *ca.* 37 wt%, corresponding to a total organic matter content of approximately 75%.⁷³ Total

organic nitrogen (TON) was at *ca.* 3 wt% in the top 10 cm of the core and decreased with depth to 2 wt%.

3.2 Fe and P fractionation according to sequential extraction (SEP) of the sediment solid phase

According to Fe-SEP (Fig. 2A and B) the largest extracted Fe pool throughout the investigated sediment depth was Fe(II) extracted by HCl (Fe(II)-HCl). On average, this step extracted \sim 32% of the total Fe from the top 10 cm of sediment. Below 10 cm, Fe(II)-HCl decreased to a minimum of 19% of total Fe between 18 and 28 cm depth and increased again to \sim 36% towards the lowest probed sediment depth. The second largest Fe pool was ascorbic acid extractable Fe (Fe-Asc) with up to \sim 28% in the top 12 cm and up to \sim 26% of total Fe below 28 cm depth. Between 12 and 28 cm depth, Fe-Asc was the smallest extracted Fe pool, while Fe-HNO₃ was second largest pool with up to \sim 8% of total Fe at 16 cm depth. In the top 6 cm Fe-HNO₃ was the smallest



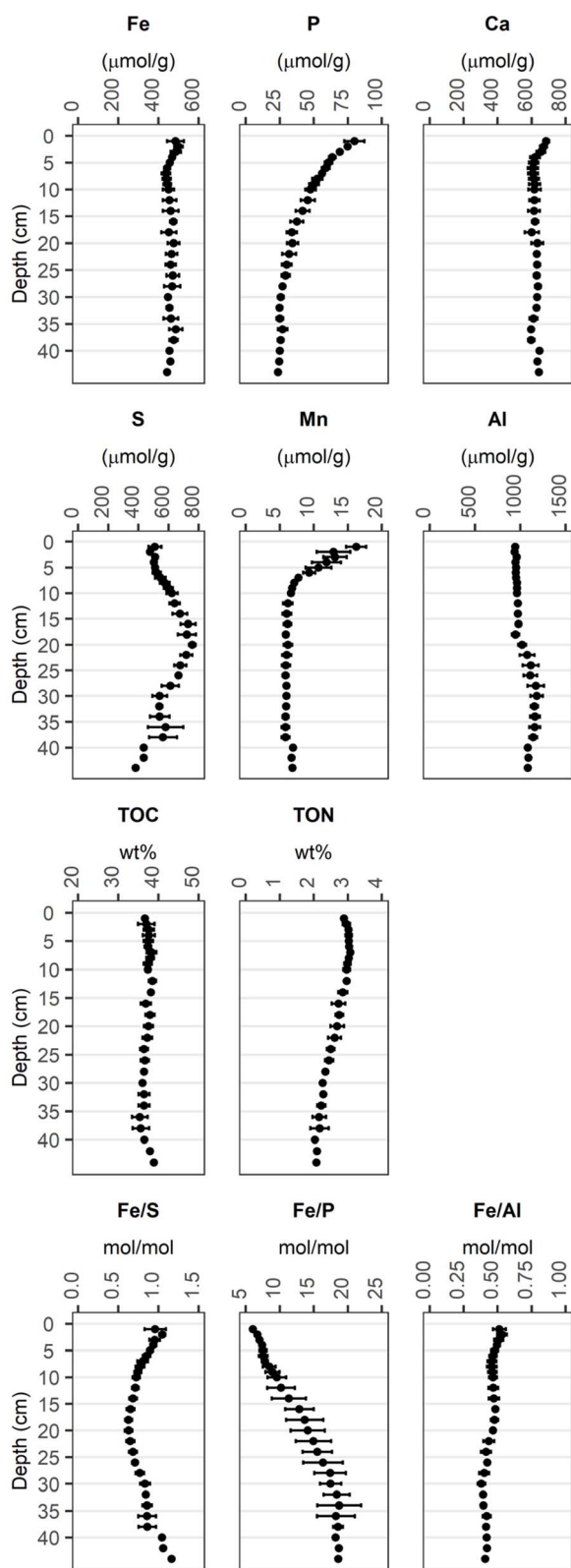


Fig. 1 Average total elemental contents of the sediment solid phase (1–36 cm: $n = 3$, 38 cm: $n = 2$, 40–44 cm: $n = 1$). Error bars denote the standard deviation of the values obtained from 3 sediment cores for the corresponding depth. Individual measurements are reported in Münch *et al.*¹

extracted Fe pool, accounting for a constant $\sim 4\%$ of total Fe, as it did below 40 cm depth. Fe-CDB did not show a clear trend with depth, with slightly higher fractions of $\sim 9\%$ of total Fe in the top 6 cm and oscillating around $\sim 6\%$ of total Fe in the deeper sediment. While in the surface and the deeper sediment more than 50% of Fe was extracted, between 14 and 28 cm depth over 50% and up to $\sim 64\%$ of the present Fe was not extracted by the SEP.

According to the P-SEP (Fig. 2C and D), organic P (P-ashed/HCl) is the dominant P phase over the entire depth accounting for up to $\sim 69\%$ of total P, except in the top 2 cm, where it accounts for only $\sim 40\%$ of total P. In the top 6 cm the two Fe-associated P pools P-NaHCO₃ and P-DCB are enriched with respect to the deeper sediment, accounting for up to $\sim 27\%$ and $\sim 26\%$ of total P, respectively. Apatite (P-HCl) is constant at $\sim 16\%$ of total P over the entire investigated sediment depth, but nevertheless represents the second largest pool below 10 cm depth.

3.3 Bulk XAS analyses

3.3.1 Fe speciation based on Fe K-edge XAS. Six reference spectra were identified that allowed to obtain satisfying fits of all EXAFS and XANES bulk spectra: Fe(III)-HA, CaFeP, Fe(II)aq., smectite, biotite and pyrite (Table 1, ESI B1, and B2†). Whereas smectite, biotite and pyrite represent minerals with defined structure, CaFeP and Fe(III)-HA are synthetic samples from laboratory experiments designed to study Fe precipitate and complex formation in natural systems. The CaFeP reference spectrum was collected on amorphous Ca-Fe(III)-phosphate formed by the oxidation of 0.5 mM Fe(II) in synthetic bicarbonate-buffered groundwater in the presence of 0.4 mM Ca, 0.5 mM Si and 0.75 or 1 mM phosphate (initial P/Fe in solution of 1.5 or 2.0) as reported in Senn *et al.*⁵⁷ The Fe(III)-HA spectrum was measured on oligomeric Fe(III) complexed within humic acid, synthesized by Fe³⁺ hydrolysis in presence of humic acid at near neutral pH.²³ Inclusion of the Fe(II)aq. spectrum in the LCF improved the quality of the fits both in the XANES and EXAFS range, as evidenced in a systematic decrease in the r -factor (averages (and standard deviations; $n = 15$) of -44% (14%) for XANES and -15% (7%) for EXAFS). Small fractions of Fe(II)aq. were mostly compensated by lower fractions of mostly Fe(II)-containing biotite. In the EXAFS LCF, addition of the Fe(II)aq. reference also increased the fraction of CaFeP mostly at the expense of Fe(III)-HA. These effects can be explained by the coordination of Fe in the respective reference materials. Without Fe(II)aq., biotite was the only reference containing O-coordinated Fe²⁺ and its contribution to the fit, hence, was determined mainly by the fraction of O-bound Fe²⁺ in the sample. However, Fe(II) in the octahedral layers of biotite is coordinated to up to six Fe in the second shell, resulting in a pronounced second-shell Fe signal. The inclusion of Fe(II)aq. at the expense of biotite thus implies that the Fe²⁺ coordination in biotite is not a good representation of O-coordinated Fe²⁺ with lower or no second-shell Fe coordination, *e.g.* in clay minerals with lower Fe content or in the form of adsorbed or complexed Fe²⁺. In fits without the Fe(II)aq. reference, the



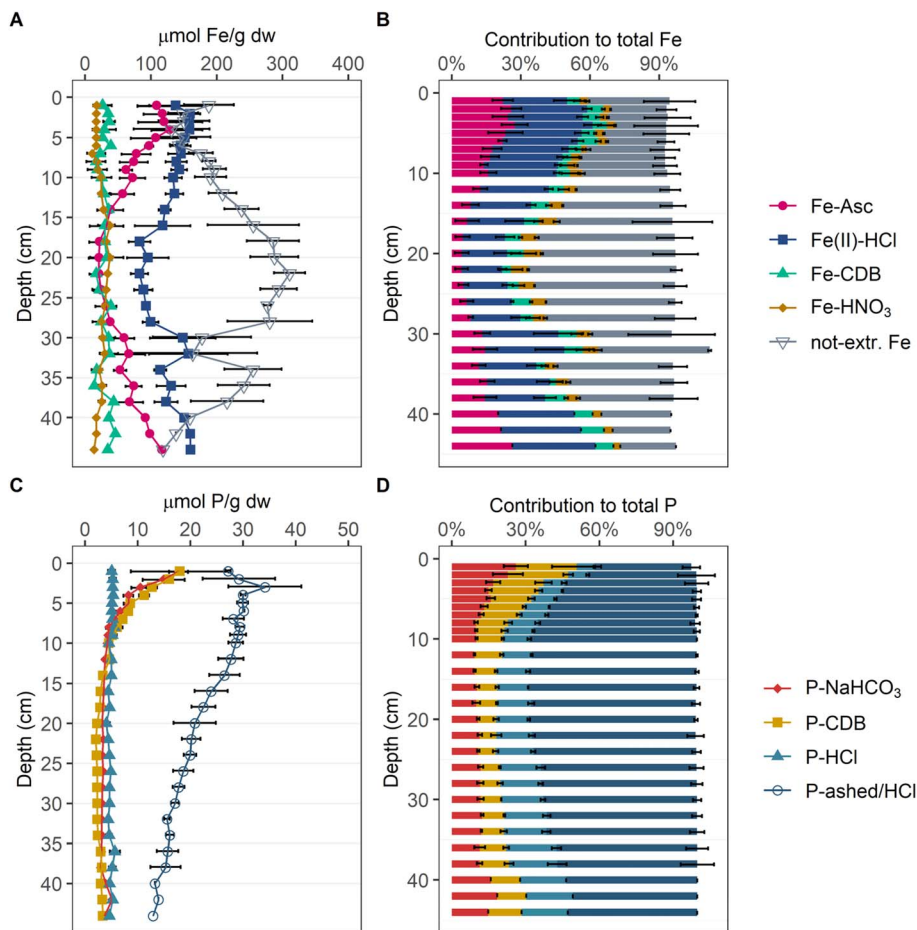


Fig. 2 Depth profiles of the Fe and P pools determined by SEP. Respective Fe content of the four extracted and the not extracted Fe pools in $\mu\text{mol}_{\text{Fe}}$ per g dw (A). Contribution of determined Fe pools to total Fe in averaged percentages (B). Respective P content of four extracted P pools in μmol_{P} per g dw (C). Contribution of determined P pools to total P in averaged percentages (D). For contents individual measurements at each depth were averaged and error bars denote their standard deviation. For the contribution to total contents, percentages of individual measurements were averaged, and error bars denote their standard deviation (1–36 cm: $n = 3$; 38 cm: $n = 2$; 40–44 cm: $n = 1$). Individual measurements are reported in Münch *et al.*¹

overexposure of the second-shell Fe signal of biotite in the EXAFS LCF could be partially compensated by fitting a higher fraction of Fe(III)–HA at the expense of CaFeP, because the Fe(III)–HA spectrum is characterized by a lower second-shell Fe signal. These considerations allow rationalizing why inclusion of Fe(II)aq. in the LCF occurs at the expense of the contribution of biotite and is accompanied by a higher CaFeP fraction at the expense of Fe(III)–HA. The small fractions of Fe(II)aq. included in the fits may thus account for differences in the average oxidation state and coordination of Fe in reference and sediment phyllosilicates or for a minor share of Fe(II) complexed by organic matter, adsorbed to other sediment constituents, or in other minor Fe(II) phases.

The LCF analysis of the EXAFS spectra returned consistently higher CaFeP and lower Fe(III)–HA, smectite, and pyrite fractions than the LCF analysis of the corresponding XANES spectra (data listed in ESI Table B3-1[†]), but the deviation for individual phases was typically below 10%. For further interpretation of the XAS-derived Fe speciation, the results of the LCF analysis of

the EXAFS and XANES range were averaged to avoid bias towards the similarity in local coordination environment or similarity in redox state and coordination geometry (data listed in ESI Table B3-2[†]).

The LCF-derived fractions of Fe phases for sediment samples from 1–36 cm depth are displayed in % of total Fe in Fig. 3A and as absolute Fe contents (fractions times total Fe contents) in Fig. 3B. Fe(III)–HA contributed $\sim 14\%$ to total Fe in the top 3 cm and decreased to below 1% of total Fe below 10 cm depth. CaFeP represented $\sim 29\%$ of total Fe in the top 7 cm and decreased to $\sim 17\%$ below 14 cm depth. Fe(II)aq. contribution to total Fe was approximately constant at, on average, $\sim 6\%$ of total Fe across the investigated sediment depth. The fraction of the spectra from the clay minerals smectite and biotite steadily increased from $\sim 20\%$ and $\sim 11\%$, respectively, at 1 cm depth, to $\sim 27\%$ and $\sim 14\%$, respectively, at 36 cm depth. The contribution of pyrite was lowest in the top 5 cm with a minimum of $\sim 20\%$ of total Fe at 2 cm depth, constant at around $\sim 25\%$ of total Fe at 5–7 cm and increased to a maximum of $\sim 46\%$ of total



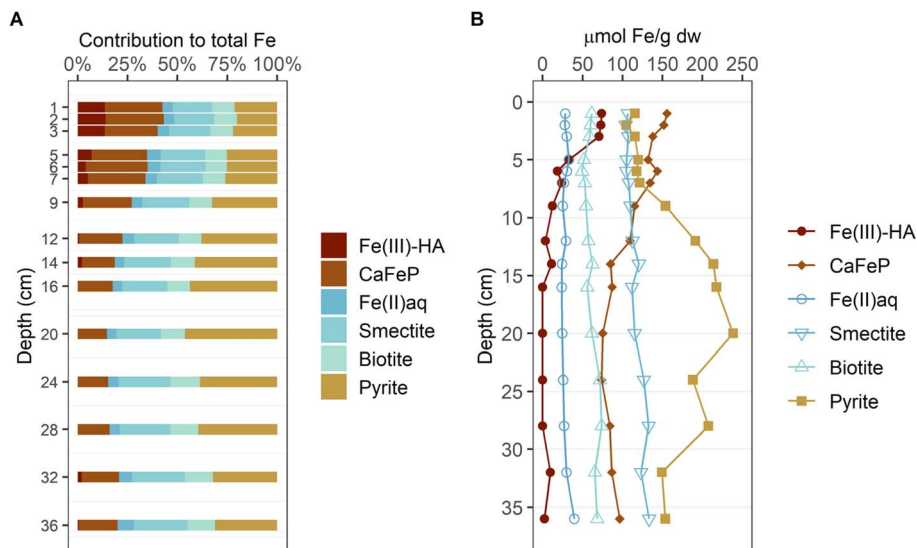


Fig. 3 Depth-resolved contribution of Fe reference phase to total Fe as determined by linear combination fitting (LCF) of the bulk Fe K-edge XAS. The given values represent averaged LCF of XANES and EXAFS (A). Concentration of each component as inferred from total Fe content and component contribution from LCF (B). The used references are grouped by similar coordination environment: highly amorphous mono- or oligomeric Fe(III) (red colors), Fe in oxygen-coordinated environments with varying degree of organization and little Fe–Fe coordination (blue colors) and pyrite (yellow). The LCF data is available in Tables ESI B3(1) and B3(2).[†]

Fe at 20 cm depth. Below that, pyrite content decreased again, but never contributed less than ~30% until the bottom of the core. In the top 5 cm of sediment CaFeP contributed most to total Fe, below 5 cm pyrite was the most abundant Fe phase.

3.3.2 Depth-resolved bulk XAS-measurements at the P K-edge. Bulk P K-edge XANES spectra recorded on sediment samples from 2, 8, 38 and 40 cm depth were virtually identical and resembled reference spectra of Mg-IP6 and DNA representative for organic phosphate mono- and diesters. Mg-IP6 represents the salt of Mg and phytate, the six valent anion of phytic acid (inositol hexaphosphate). Phytic acid is a plant-derived phosphate monoester and with a contribution of up to 60% the most abundant organic P species in soils.^{74,75} (ESI Fig. C1(1) and C2(1)).[†]

3.4 Scanning X-ray microscopy

The ground sediment from both, 2 cm (Fig. 4A) and 40 cm (Fig. 4B), depths, consisted of large fibrous aggregates, hundreds of micrometers in size, with uniform distribution of Fe, Ca and P. Al was also ubiquitously present in the aggregates at both depths but showed more variation in intensity compared to the other three elements (Fig. 4D, H, 5D, E(2), and F(2)). Within or on these aggregates, spherical particles with higher densities of Fe and S occurred, having a size of a few to a few tens of micrometers (Fig. 4B, and F). These particles were identified as framboidal pyrite based on morphology and elemental content in earlier scanning electron microscopy analyses.¹ At both depths, also particles were found with a size of tens of micrometers and containing predominately Fe, among the detectable elements (Fig. 4B–D, and F–H). At both depths, hotspots with S as the only detectable element, with sizes ranging from below ten to tens of micrometers (Fig. 4G, F,

5B, C, G, and K), suggested the presence of elementary or organic material locally enriched with S. Spherical particles with high Ca content with a size of tens of micrometers were encountered in material from 2 cm depth, while those particles were less abundant at 40 cm depth and Ca predominately occurred as part of the large aggregates (Fig. 4C, G, and 5). Next to the occurrence in the large aggregates, Si-rich particles with blurred contours, variable shapes and sizes ranging from tens to hundreds of micrometers were found at both sediment depths (Fig. 4D, H, 5D, E(2), and F(2)).

Based on the elemental maps Fe K-edge and P K-edge μ -XANES spectra were collected at spots representing the different features described above. A selection of positions of μ -XANES spectra are shown in Fig. 5.

3.4.1 Fe K-edge μ -XANES data. None of the Fe μ -XANES spectra from 2 cm depth (Table 3a, spectra in Fig. 6A, LCF results in Fig. 7A and ESI Table B4(1)).[†] unequivocally matched a single reference spectrum. Three different types of Fe spectra were frequently encountered on multiple maps (t-T1-Fe to t-T3-Fe). Type 1 (t-T1-Fe) was most abundant, representing 43% of the Fe μ -XANES spectra, which were found on 5 maps. Fe μ -XANES spectra of this type were mainly collected on the large aggregates outside particles with distinct higher elemental densities (Fig. 5A, and B). This also applied to Type 3 (t-T3-Fe, Fig. 5A–C), even though Fe and S densities were generally a bit higher at collection positions of t-T3-Fe compared to t-T1-Fe. t-T3-Fe was found on 4 maps but represented only 8% of the spectra. Using LCF, the spectra t-T1-Fe and t-T3-Fe could be described using the same reference spectra as for the bulk Fe XAS data, as a mix of CaFeP, Fe(II)aq., smectite, biotite and pyrite: t-T1-Fe was described by the highest fractions of CaFeP and Fe(III), and t-T3-Fe by a higher share of ~50% pyrite. Type 2



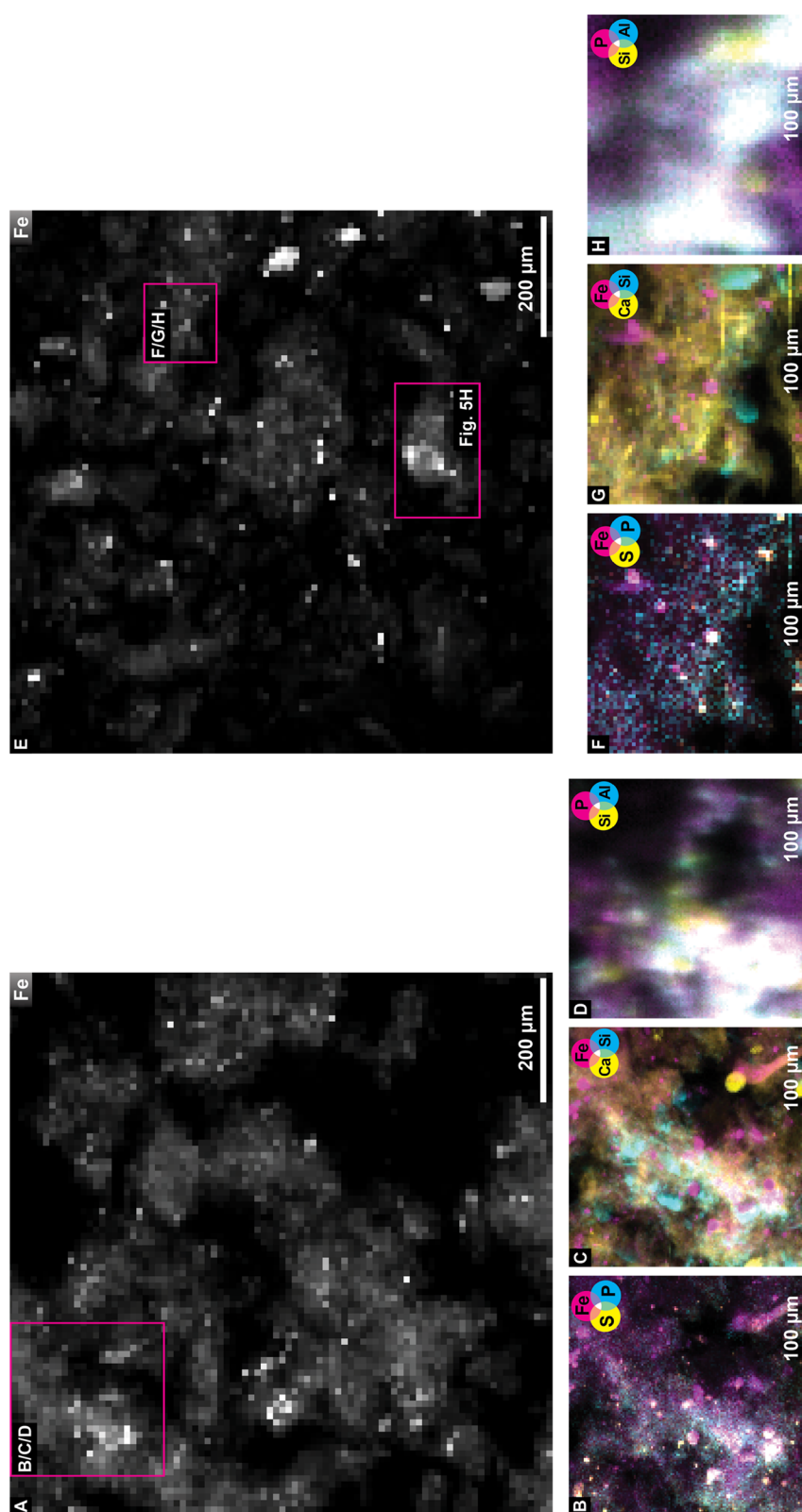


Fig. 4 μ -XRF maps collected on samples from 2 cm depth (A–D) and 40 cm depth (E–H), at 7200 eV for Fe (A–C and E–G) and at 2400 eV for P (D and H), with a spatial resolution of 10 μ m (A and E), and 2 μ m (B–D and F–H). The identities of the maps are listed in ESI Table D2(2).†



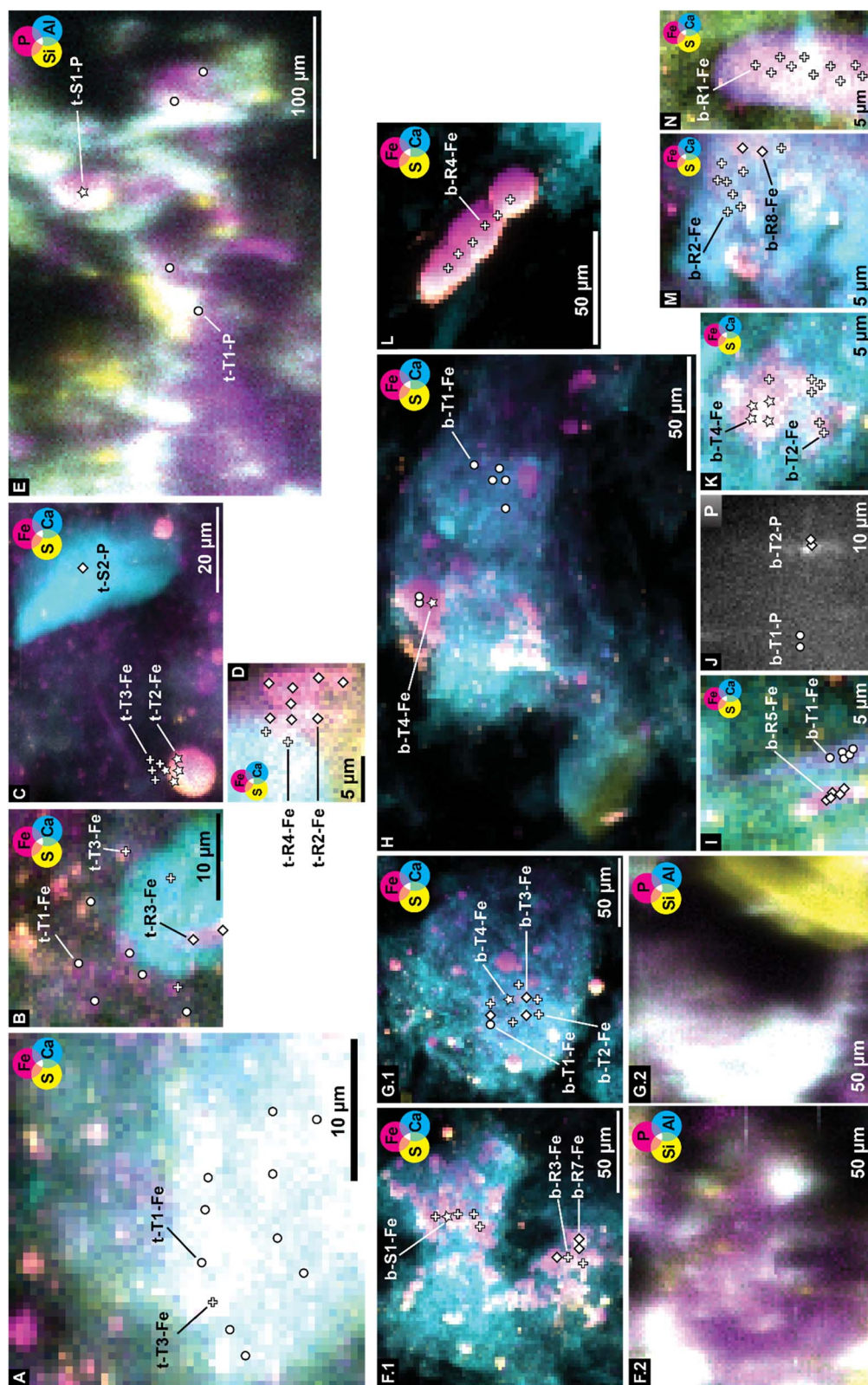


Fig. 5 μ -XRF maps collected on samples from 2 cm depth (A–E) and 40 cm depth (F–L), at 7200 eV for Fe (A–D, F(1), G(1), H, I, K, L, M, and N) and at 2400 eV for P (E, F(2), G(2), and J). The identities of the maps are listed in ESI Table D2(2). † Markers and labels refer to positions of μ -XANES spectra (listed in Tables 3 (Fe) and 4 (P)), identity in ESI Table D2(1), ‡ spectra in Fig. 6 (Fe) and 8 (P), composition according to LCF in Fig. 7 (Fe) and 9 (P)).



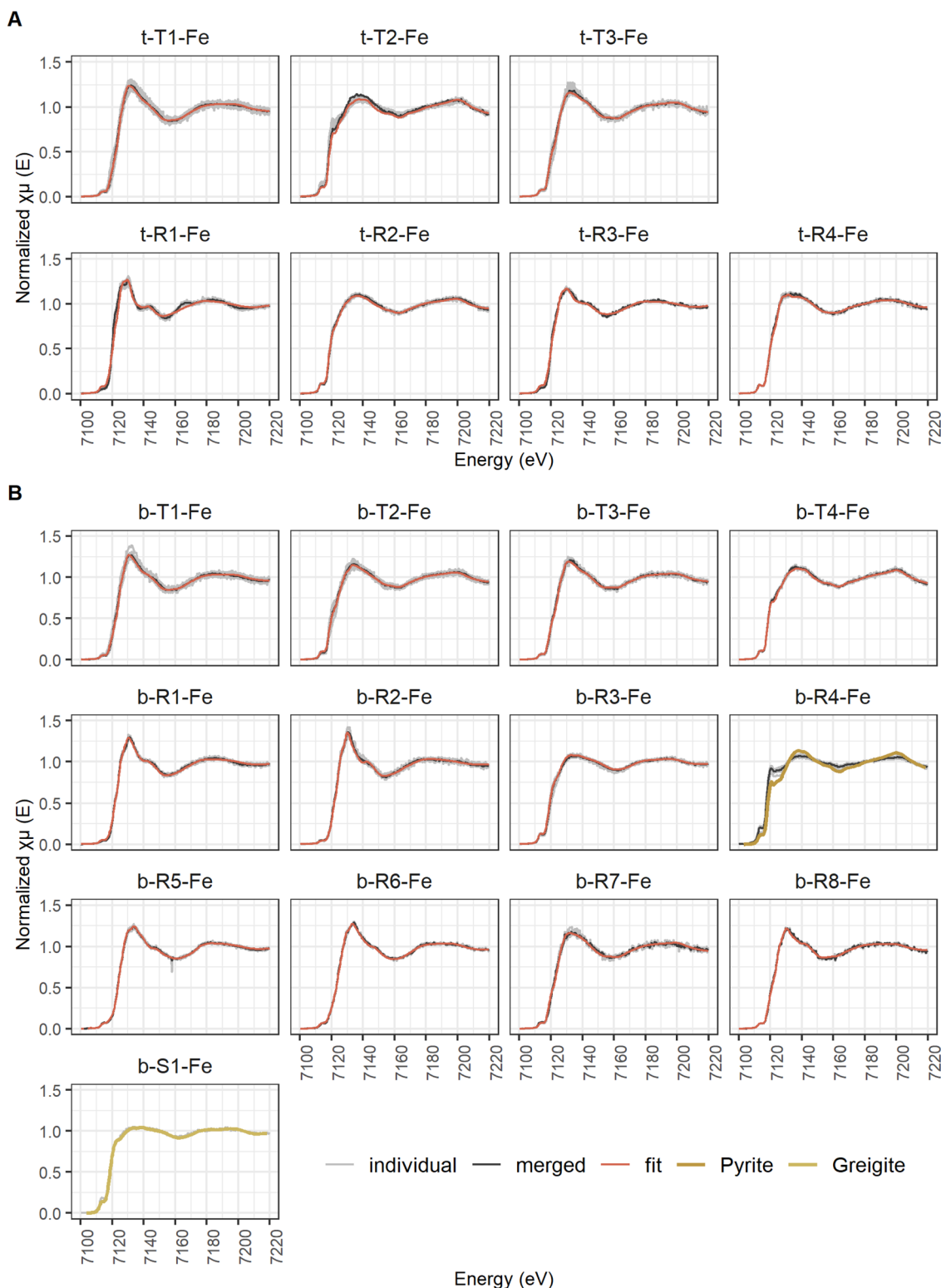


Fig. 6 Classified Fe μ -XANES spectra showing individual spectra (individual), the merged spectra (merged) and the LCF spectra (fit) for the samples from 2 cm (A) and 40 cm (B) sediment depth. For the sample at 40 cm depth, the spectra of Region 4 (b-R4-Fe) are compressed, but their shape clearly indicates pyrite and the spectrum of Spot 1 (b-S1-Fe) is almost identical to the greigite reference spectrum.

(t-T2-Fe), the second most abundant type of Fe μ -XANES spectra at 2 cm depth, accounted for 26% of the measured spectra and was found on 4 maps, mainly on areas with high relative Fe and

S densities and on spherical particles with high S density (Fig. 5C). The spectrum t-T2-Fe closely resembles the spectrum of pyrite, and was considered to mainly represent pyrite. Based



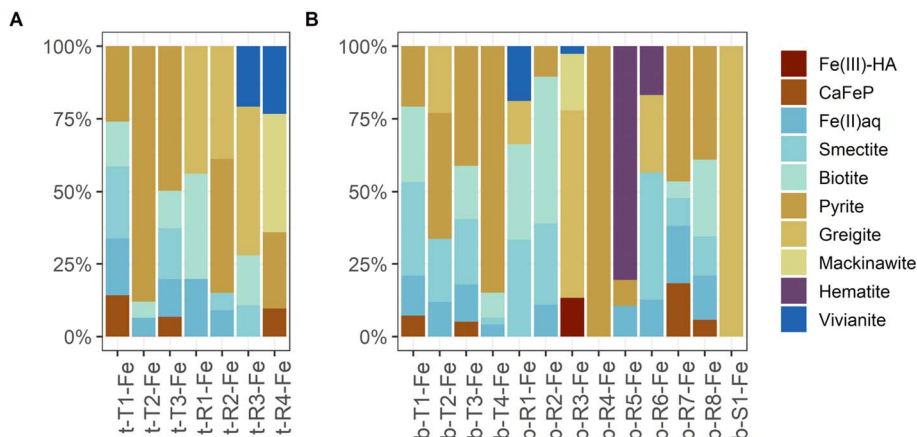


Fig. 7 Normalized LCF results for aggregated μ -XANES type, region and spot spectra collected on sediment samples from 2 cm (A) and 40 cm (B) depth. The derivation and occurrence of the aggregated spectra are listed in Table 3. Complete LCF results are listed in ESI B3.†

on LCF, the difference between the three type spectra can be predominately assigned to a decrease in the contribution of pyrite in the sequence Type 2 > Type 3 > Type 1, mostly on the expense of Fe(III)-containing components. All other spectra were grouped into four regions (t-R1-Fe to t-R4-Fe), each located on a different map and characterized by the same spectral signature at multiple points. For each region, additional reference spectra were required for the LCF. Region 1 (t-R1-Fe) had a distinct spectrum which can be attributed to the high contributions of biotite and greigite (but one peak between 7160 and 7170 eV could not be fitted with any available reference). Region 2 (t-R2-Fe) was represented by over 80% S-bound Fe (46% pyrite; 39% greigite) in the LCF, and was found on an area of high S density without distinct shape (Fig. 5D). Regions 3 and 4 (t-R3-Fe and t-R4-Fe) both yielded the best fits when using ~20% of vivianite and were both collected on areas with co-occurrence of Fe and Ca (Fig. 5B, and D).

Although the same number of Fe μ -XANES spectra were recorded at both depths, the diversity of spectra was larger at 40 cm depth (Table 3b, spectra in Fig. 6B, LCF results in Fig. 7B and ESI Table B4(1)†). Four types of Fe spectra were frequently encountered on multiple maps (b-T1-Fe to b-T4-Fe). Type 1 (b-T1-Fe) and Type 2 (b-T2-Fe) were most abundant, both accounting for ~20% of the measured spectra and found on 4 and 3 maps respectively. Spectrum b-T1-Fe was, as t-T1-Fe at 2 cm depth, mainly found as part of the large aggregates where no element showed pronounced relative densities (Fig. 5G(1), H, and I), b-T2-Fe was both found as part of the aggregates and in areas with slightly higher S abundance (Fig. 5G(1) and K). Type 3 (b-T3-Fe) and Type 4 (b-T4-Fe) were less abundant, accounting for 7% and 6% of the measured spectra and found on 2 and 3 maps, respectively. While b-T3-Fe was mainly part of the matrix of large aggregates (Fig. 5G(1)), b-T4-Fe was next to the large aggregates (Fig. 5G(1)) also obtained from Fe and S hotspots (Fig. 5H, and K). Based on LCF, the composition of the frequently encountered Fe spectra is similar at 40 cm depth and at 2 cm depth. The most abundant types of Fe spectra at both depths (t-T1-Fe and b-T1-Fe) had the highest abundances of

CaFeP, next to Fe(II)aq., smectite, biotite and pyrite. b-T1-Fe, however, had higher clay content with in total 72% clays and Fe(II)aq. and slightly less pyrite than t-T1-Fe. At both depths, the LCF -results of Type 3 (t-T3-Fe and b-T3-Fe) were similar to Type 1 (t-T1-Fe and b-T1-Fe), but showed higher pyrite content, mainly on the expense of CaFeP. At both depths, one of the frequently encountered types of Fe spectra essentially represented pyrite (t-T2-Fe and b-T4-Fe). At 2 cm depth, this was the second most abundant type, which was not the case at 40 cm. At 40 cm depth the second most abundant type, b-T2-Fe, was described by ~66% S-bound Fe, of which ~23% were greigite, next to pyrite, Fe(II)aq. and smectite. At 40 cm depth, there was a high diversity of Fe spectra encountered on just one map (b-R1-Fe to b-R8-Fe) and one unique Fe spectrum (b-S1-Fe). Region 7 (b-R7-Fe) and Region 8 (b-R8-Fe) were variations of b-T1-Fe, both containing CaFeP, but with considerably more pyrite according to LCF. b-R8-Fe was indeed located on an area of elevated relative densities of Fe and Ca (Fig. 5M). But b-R7-Fe, even though having the highest CaFeP contribution (~18%) measured by focused beam on both samples, was located on an area with an agglomeration of ~10 micrometer-sized Fe and S hotspots (Fig. 5F(1)). This agglomeration resembled clusters of euhedral pyrite observed by SEM on the same sample in an earlier study.¹ Indeed, the other two groups of Fe μ -XANES spectra located on this area of Fe and S agglomerations (Fig. 5F(1)), Region 3 (b-R3-Fe) and Spot 1 (b-S1-Fe), were characterized by high abundance of iron sulfides according to LCF. b-R3-Fe closely resembled the greigite reference spectrum, but also required mackinawite, Fe(III)-HA and vivianite for a satisfying reproduction by LCF. b-S1-Fe was identical to the greigite reference spectrum. Region 1 (b-R1-Fe) and Region 2 (b-R2-Fe) showed the characteristic pre and post white line shoulders of biotite. Accordingly, LCF indicated that most Fe in those regions was associated to phyllosilicate minerals. Nevertheless, LCF of b-R1-Fe yielded the highest vivianite contribution of all spectra measured at 40 cm depth. While b-R1-Fe was located on a rod-shaped particle of increased relative Fe density (Fig. 5N), b-R2-Fe was part of the matrix of a large aggregate



(Fig. 5M). LCF of Region 5 (b-R5-Fe) and Region 6 (b-R6-Fe) indicated the presence of hematite, with b-R5-Fe being composed of ~80% hematite and located on Fe hotspot without occurrence of other detectable elements (Fig. 5I). The spectrum at Region 4 (b-R4-Fe) showed the features of the pyrite reference spectrum but the features were damped. b-R4-Fe was located on a rod shaped area of high relative densities of Fe and S, indicating that the features are damped due to self-absorption on a spot with high Fe concentration (Fig. 5L).

3.4.2 P K-edge micro XANES data. In contrast to the bulk P XANES spectra, some diversity was noticed among the P μ -XANES spectra. 85% of spectra of the sample from 2 cm depth (Table 4a, spectra in Fig. 8A, LCF results in Fig. 9A and ESI Table C4(1)[†]), were virtually identical. Classified as Type 1 (t-T1-P), they could be found across 4 maps as part of the large aggregates (Fig. 5E). LCF yielded 53% Mg-IP6 and 40% DNA, along with 6% apatite. The spectra each belonged to a singular spot. Spot 1 (t-S1-P) was characterized by a high fraction of Fe-coordinated P and was located on a hotspot of high P density (Fig. 5E). Spot 2 (t-S2-P) contained exclusively apatite. Spectra resembling pure apatite were also collected at multiple locations in the sample from 40 cm depth (Table 4b, spectra in Fig. 8B, LCF results in Fig. 9B and ESI Table C4(1)[‡]). They

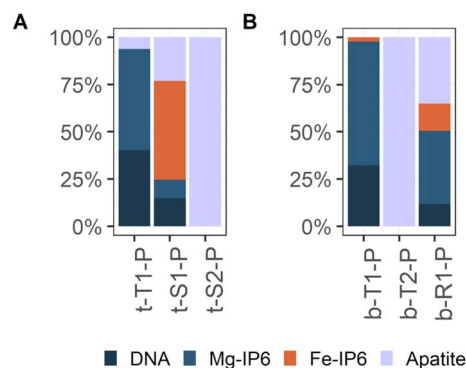


Fig. 9 Normalized LCF results of the sub-micron P phases detected by μ -XANES at 2 cm (A) and 40 cm (B) depth. Occurrence of each phase is listed in Table 4. Values are listed in ESI C3.[†]

represented the second most abundant type of P spectra, Type 2 (b-T2-P), and were mostly found on hotspots of relatively high P density (Fig. 5J). The most abundant type of spectra, Type 1 (b-T1-P), was also part of the large aggregates (Fig. 5J) and LCF gave fractions of 66% Mg-IP6, 32% DNA and 2% Fe-IP6. All the other spectra collected at 40 cm depth were identical within the range of noise and constituted Region 1 (b-R1-P). According to LCF, b-R1-P was a mixture of organic and mineral P with 51% purely organic P (Mg-IP6 and DNA), 35% apatite and 14% Fe-IP6.

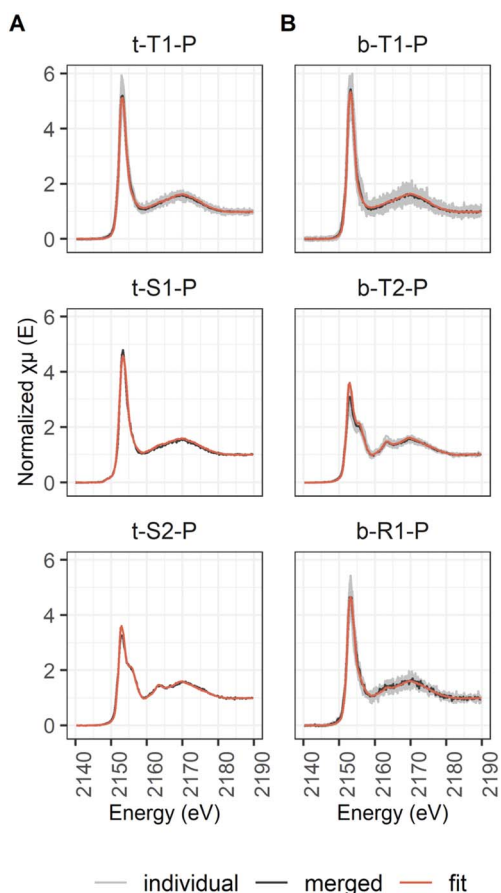


Fig. 8 Classified P μ -XANES spectra showing individual spectra (light grey), the merged spectra (dark grey) and the LCF spectra (orange) for the samples from 2 cm (A) and 40 cm (B) sediment depth.

3.5 Correlations of total element contents, results from sequential extractions and X-ray spectroscopy

Correlation matrices between elemental contents and SEP results for sediment samples above 10 cm show distinct differences to those below 10 cm. In contrast to deeper sediment layers, Ca, Mn, Fe and P are stronger correlated with each other and to SEP-derived contents in the top 10 cm of the sediment. In the top 10 cm, Fe-Asc and Fe(II)-HCl are strongly correlated with total Fe contents, indicating that the variability in Fe content is connected to variations in these two SEP fractions. Furthermore, the total Fe contents are correlated to Ca, Mn and P, which is not the case for the latter two elements in the deeper layers. Thus, the strong correlation between Mn and P contents is an exclusive feature of the top 10 cm of the sediment. Another remarkable and differentiating feature are the correlations of P-NaHCO₃ and P-CDB with other variables in the top 10 cm: P-NaHCO₃ and P-CDB are strongly correlated with the total P content, and they are stronger correlated with the Fe, Ca and Mn contents compared to the deeper sediments. In the deeper sediment layers, next to Ca, Fe is correlated with S and the Fe-SEP fractions Fe-HNO₃ and not extractable Fe. The stronger correlation between S with Fe-HNO₃ and not extractable Fe distinguishes the lower sediment layers from the top 10 cm. For P, the strong correlation with TON and the P-ashed/HCl fraction is only observed in the lower sediment layer.

The correlation of LCF-based Fe species with element contents as well as SEP-based Fe and P fractions can be found in Fig. 10C. Of all LCF-based Fe species, Fe(III)-HA and CaFeP showed the most positive correlations with other variables and they are also strongly correlated with each other. In particular,



Fe(III)-HA and CaFeP strongly correlate with P, Mn and Fe-Asc. Also all SEP-based P fractions are positively correlated with Fe(III)-HA and CaFeP, but correlation is strongest with P-NaHCO₃ and P-CDB. The LCF-derived pyrite-content is strongly correlated with the S contents and shows strong and medium

correlations with the HNO₃ and not-extractable Fe pool, respectively. The contents of smectite and biotite, obtained by LCF, are strongly correlated with Al as well as to each other, supporting their interpretation as proxies for phyllosilicate-Fe in the sediment.



Fig. 10 (A, and B) Pearson correlations of total elemental contents (black labels), Fe pools according to Fe-SEP (blue labels), P pools according to P-SEP (orange labels) in (A) the top 10 cm of sediment ($n = 30$ samples) and (B) below 10 cm depth ($n = 44$ samples). (C) Pearson correlations including the Fe species according to Fe K-edge bulk XAS LCF (green labels) across the entire investigated depth ($n = 15$ samples). Correlations were calculated using contents in $\mu\text{mol per g dw}$ and are significant with p -values < 0.001 (***), < 0.01 (**) and < 0.05 (*). Insignificant correlations ($p \geq 0.05$) are not shown.



4 Discussion

4.1 Reducible Fe(III) phases and associated P

The subject of this study is the shallow peat lake Terra Nova in which Fe treatment induced seasonal peaks of internal P loading. An earlier investigation of the system suggested that Fe addition led to the build-up of a highly redox sensitive OM-associated Fe pool, which can bind considerable amounts of P but tends to release this P during periods of low O₂ concentrations in the bottom water. This reducible Fe pool, mainly extracted by Fe-Asc is pivotal for the surface water quality of Terra Nova,¹ and elucidating its nature was the main objective of this study.

Correlations between Fe-Asc and CaFeP as well as with Ca, Fe, P contents indicate that oxidation of Fe²⁺ in the top sediment results in the co-precipitation of these elements. The correlation of Ca, Fe and P is reflected in concurrently decreasing contents with depth in the top 10 cm of sediment. Presence of dissolved Ca and Si and P/Fe ratios >0.5 in the porewaters in the corresponding depth interval (ESI Fig. E(1)†) resemble solution compositions that lead to the precipitation of CaFeP.⁵⁷ This implies that Fe redox cycling in these sediments does not lead to the precipitation of pure Fe (oxyhydr)oxides such as ferrihydrite or lepidocrocite but to the formation of an amorphous co-precipitate. Precipitation of Fe together with Ca and P in aquatic sediments has been previously reported by Hyacinthe and van Cappellen⁷⁶ in a heavily bioturbated freshwater site in the Scheldt estuary (Belgium/The Netherlands). They found that on average 77% of Fe(III) in the top 14 cm of sediment was present in an authigenic amorphous Fe–Ca–Mn–P phase, which was extractable by the same ascorbic acid extraction we used in the Fe-SEP. This finding is in line with the correlation of Ca, Fe, P as well as Fe-Asc and CaFeP with Mn contents in the top 10 cm of the sediments from lake Terra Nova (Fig. 10A). Hyacinthe and van Cappellen⁷⁶ further reported that Mn-contents decreased from freshly formed Fe–Ca–Mn–P precipitates to older precipitates at larger depth. Our results are in line with those reported by Hyacinthe and van Cappellen⁷⁶ and hence support the presence of authigenic CaFeP co-precipitates in the surface sediment of Terra Nova, which become Mn depleted when buried deeper in the sediment: in sediments for Terra Nova the Fe and Mn contents were correlated in sections of the fibrous aggregates without pronounced enrichment of pyrite particles and the Mn/Fe ratio was larger in 2 cm depth compared to 40 cm depth (ESI F†). Pearson correlation further showed that Fe(III)–HA fractions were closely correlated with CaFeP (LCF), Fe and P fractions mobilizable by reduction (SEPs), and total Ca, Mn, and P contents across the entire sediment depth (Fig. 10C). This offers additional evidence for the presence of a Fe–Ca–Mn–P phase, but also indicates that part of the reducible Fe(III) is coordinated with functional groups of the OM. Formation of complexes between Fe and functional groups of OM is key regarding the influence of OM on the formation of Fe(III)–precipitates. Organic matter can limit Fe(III) polymerization to monomeric or oligomeric Fe-OM complexes or ferrihydrite, depending on the molar OC/Fe

ratio and pH.^{27,77–79} In laboratory experiments on the oxidation and precipitation of Fe in the presence of Leonardite humic acid at pH 6.5 and at OC/Fe from ~10 to 50, Beauvois *et al.*⁸⁰ observed the formation of oligomeric Fe(III) and ferrihydrite-like Fe(III) nanoparticles embedded in OM, with a decreasing share of ferrihydrite at increasing OC/Fe. For OC/Fe from 50 to 500, at neutral to acidic pH, Karlsson and Persson²⁷ reported a shift to predominantly monomeric Fe(III) bound in chelate complexes. At OC/Ca ratios <3.8, Beauvois *et al.*⁸⁰ found Ca to act as a coagulant promoting the formation of large, branched organic networks in which Fe(III) nano-aggregates of ~6 nm size were distributed. At higher OC/Ca, on the other hand, they observed the development of smaller Fe-OM-Ca aggregates. Considering contents of ~30 mol kg⁻¹ TOC, 0.5 mol kg⁻¹ Fe, and 0.6 mol kg⁻¹ Ca in the sediments from Lake Nova, an overall OC/Ca ratio of ~50 and an overall OC/Fe ratio of ~60 can be estimated. Taking into account that CaFeP and Fe(III)–HA accounted for ~1/3 of the total Fe, an OC/Fe ratio of ~180, can be estimated for the Fe(III) represented by these two LCF references. Based on all aforementioned considerations, we therefore conclude that Fe(III)–HA and CaFeP identified by LCF do not constitute two distinctive Fe phases in our peat lake samples, but that Fe(III)–HA and CaFeP together describe a co-precipitate of Ca, P, and mono- and oligomeric Fe(III) that is also coordinated with OC. This mixed inorganic–organic co-precipitate is finely distributed within the OM matrix of the peat sediment.

Indeed, X-ray micro-microscopy supports the presence of Fe–Ca–P oligomers and nanoparticles associated with the OM matrix. At both investigated depths, the majority of the measured μ -XAS spectra (t-T1-Fe and b-T1-Fe) contained CaFeP (Fig. 7), all of them found on large aggregates with uniform distribution of Fe, Ca and P (Fig. 4 and 5). These spots exhibited relatively low densities of these elements, in contrast to spots with high local densities for which LCF pointed to high abundance of Fe sulfides or apatite, respectively. That is, the inorganic–organic precipitates occur finely dispersed within the OM matrix. Consequently, it was not possible to discern and characterize individual particles when analyzing the material with a spatial resolution of several hundred nm. Notably, LCF of most μ -XANES spectra returned higher fraction of Fe(II)aq. than LCF analyses of bulk XANES and EXAFS spectra, but did not include the Fe(III)–HA reference. This could be due to lower sensitivity in the microanalyses constrained to the XANES region, but could indicate that the orders of magnitude higher flux density of the focused X-ray beam induced photoreduction of organically complexed Fe(III).

In contrast to what Fe K-edge XAS results suggest, the P K-edge XAS data indicate that the majority of P in the Terra Nova sediment is present as organic P esters. The bulk P K-edge XAS is exclusively fitted by inositol phosphate and DNA in the LCF with no indication of considerable contributions of an inorganic P phase (ESI C2 and C3†). Organic P esters thus constitute an important P pool in the Terra Nova sediment. While according to P-SEP the majority of sediment P is indeed P structurally bound in OM, extracted as P-ashed/HCl (Fig. 2C), this pool accounts for at most 70% of the total P (Fig. 2D). In the



surface sediment, up to 51% of sediment P is extracted as P-NaHCO₃ and P-CDB. According to the used P-SEP, P-NaHCO₃ represents P bound to OM over metal cation bridges^{46,53} and P-CDB represents P bound to Fe (oxyhydr)oxides, CaCO₃ and Fe-phosphates.^{45,47,81} Indeed, correlation analysis suggests that NaHCO₃ and CDB extract P bound to the Fe–Ca–P-OM coprecipitates. P-NaHCO₃ and P-CDB show almost identical depth profiles (Fig. 2C), correlate to each other (Fig. 10A and B) and to Fe(III)–HA, CaFeP and Fe-Asc (Fig. 10C) over the entire investigated sediment depth. Further, P-concentrations in the SEP-extracts were measured by colorimetry,¹ which is exclusively sensitive to *ortho*-phosphate. These findings suggest that bulk P K-edge XAS underestimates inorganic P content. P K-edge XANES analysis has previously been reported to have its limitations for determining the P speciation in peat soils. Kruse and Leinweber⁸² found that P-XANES spectra did not show a pronounced change upon extractions targeting different P phases. Successive removal of P pools was not reflected in distinct spectral changes as all spectra were dominated by organic P, which lacks marked pre- or post-edge features and obscures the signal from other phases. Even distinguishing organic and inorganic phosphate groups complexed to the same solid phases is difficult by P K-edge,⁶⁷ as the corresponding XANES spectra differ only marginally. Even if inorganic PO₄ features a higher white line intensity than inositol phosphate,⁶⁷ the intensity of the white line is particularly sensitive to self-absorption and, consequently, the contribution of inorganic P can be underestimated at the expense of organic P when P XANES spectra are collected in fluorescence mode. These methodological constraints can explain why inorganic P was not included in the LCF of bulk sample spectra. Nevertheless, μ -XAS analysis revealed local variations in P speciation. The characteristic pre-edge feature of Fe-bound P was detected on several spots (Fig. 5 and 8), and was reproduced by the Fe-IP6 reference in the LCF (Fig. 9, ESI C3[†]). Further, apatite particles were unequivocally detected (Fig. 5, 8 and 9), in line with 14% (on average) HCl-extractable P across the entire investigated depth (Fig. 2C and D) that is interpreted as Ca–phosphate P. Despite the limitations of distinguishing organic and inorganic phosphate as delineated above, the results of P K-edge XAS thus further point towards the presence of Fe–Ca–P-OM coprecipitates, and also suggest that organic phosphate groups in addition to carboxyl or phenolic groups might contribute to Fe(III) binding by OM.

4.2 Other Fe pools

4.2.1 Fe sulfides. Sulfidation of Fe(III) phases is pivotal in the coupling of Fe and P cycling in aquatic sediments, as it removes Fe from the redox cycle and makes it unavailable for P binding in sediments.^{15,19} Presence of pyrite in the sediments from Terra Nova, deduced from sequential extractions and SEM analysis,¹ is confirmed by XAS analyses. Pyrite dominates the bulk Fe EXAFS signal over the entire sediment depth (ESI Fig. B2(1)[†]), and accounts for 20% to 46% of total Fe according to LCF (Fig. 3, ESI B3[†]). These fractions exceed those estimated from Fe-SEP as only a maximum ~7% of total Fe were extracted

as Fe-HNO₃ (Fig. 2A and B). As the EXAFS spectrum of pyrite is very distinctive, the fractions obtained by LCF can be considered to be reliable. Hence, it is more likely, that pyrite was not completely extracted by HNO₃ during the extraction step, even though HNO₃ has been reported to fully extract pyrite with high selectivity.^{21,39} While there is strong correlation between LCF-derived pyrite and S contents (Fig. 10C), LCF-derived pyrite and S contents are not only correlated with Fe-HNO₃ but also with not-extracted Fe (Fig. 10C), suggesting that part of pyrite was not extracted by concentrated HNO₃ or preceding extractions steps and remained unextracted.

High abundance of organic matter and the embedding of pyrite into the OM matrix might retard or prevent pyrite dissolution in HNO₃. Most of the μ -XRF maps at both investigated depths show areas of high density collocated Fe and S, often also in the characteristic spherical shape of framboidal pyrite (Fig. 10), which is confirmed by the dominance of pyrite in the LCF of the corresponding spectra (t-T2-Fe, b-R4-Fe, Fig. 6, 5C, L, and 7). However, also the LCF of two most abundant types of Fe μ -XANES spectra, t-T1-Fe and b-T1-Fe (Table 3, and Fig. 6) collected from OM aggregates with low Fe density (Fig. 5), indicated up to 25% pyrite (Fig. 7). This suggests that an abundant part of pyrite is embedded in the organic matrix in finely dispersed form. Oxidative dissolution by HNO₃ can be conceived as an electrochemical reaction⁸³ whereby the cathodic reduction of nitrate could be slowed down when pyrite is encapsulated by OM, in analogy to the retarding effect of the formation of a passivating layer.⁸⁴ As the oxidation of pyrite is autocatalytic and formation of Fe³⁺ ion accelerates the oxidation of pyrite⁸⁵ scavenging of Fe³⁺ ions by organic matter *via* reduction or complexation can impede the progress of pyrite dissolution. That is, two hours of HNO₃ extraction might have been too short to completely extract pyrite from the OM-rich sediments studied here.

Next to pyrite, mackinawite (FeS) and greigite (Fe₃S₄) were detected in the sediment samples by μ -XANES analyses. The three sulfide phases occurred in vicinity of each other, from a mix of mainly mackinawite and greigite (b-R3-Fe), to pure greigite (b-S1-Fe), to ~50% pyrite (b-R7-Fe) (spectra: Fig. 6B, LCF results: Fig. 7B, ESI B4[†]) in the sample from 40 cm depth. The spectra were collected from an area with elevated Fe and S densities extending over tens of micrometers and looking like an assembly of euhedral pyrite (Fig. 5F(1)). Also in the sample from 2 cm depth, mackinawite and greigite were occasionally detected but only in small areas with the size of few micrometers (region t-R4-Fe containing ~40% mackinawite and ~25% pyrite, region t-R2-Fe containing ~40% greigite and ~50% pyrite, spectra: Fig. 6A, LCF results: Fig. 7A, ESI B4[†]). Mackinawite and greigite can be conceived as intermediates in the formation of pyrite upon the sulfidation of Fe(III) phases. That is, initially FeS is formed as the primary product of the reaction of dissolved sulfide with Fe(III) (oxyhydr)oxides⁸⁶ and greigite can be the intermediate on the polysulfide pathway of pyrite formation from mackinawite.⁸⁷ Hence, the presence of these two sulfides might bear witness to the active formation of pyrite in the whole sediment of Terra Nova, which is most pronounced in layers below 10 cm depth, where pyrite contents increase and pyrite becomes the dominant phase based on bulk LCF analysis (Fig. 3).



4.2.2 Phyllosilicate minerals. LCF and Fe-SEP also yield inconsistent results for the silicate-bound Fe contents in the sediment. According to LCF of bulk Fe XAS spectra, about a third of the total Fe occurs as phyllosilicate minerals (represented by ~10% Fe in biotite and ~20% Fe in smectite) over the entire investigated sediment depth. In contrast, Fe, not extracted by Fe-SEP and conceived to represent non-reactive Fe in silicates, accounts for up to 60% of total Fe and, hence, exceeds the fraction of Fe based on SEP by a factor of about two. The contribution of pyrite to the not-extracted Fe pool can partially explain this discrepancy. However, the combined relative fractions of Fe-HNO₃ and not extracted Fe give consistently lower values compared to the sum of pyrite and clay-bound Fe fractions based on LCF. A likely explanation for the remaining discrepancy is that part of the silicate-bound Fe pool is included in other extraction steps. Loss of up to 20% of Fe from smectite by CDB extraction has been reported in literature⁸⁸ and has been attributed to rearrangement of the silicate structure upon reduction of present Fe(III).⁸⁹ Further, CDB was reported to dissolve Fe from biotite, chlorite, muscovite and illite.⁹⁰ Also HCl can dissolve clays⁹¹ while dissolution rates increase for smaller particles.⁹² This suggests, that part of the HCl extracted Fe(II) might originate from phyllosilicates and not from authigenic Fe(II) sulfides or carbonates. This also reconciles the ubiquitous presence of Fe(II)-HCl, with contribution to total Fe between 18% and 36% across the entire investigated sediment depth (Fig. 2), while the corresponding Fe(II) phases could not or only sparsely be detected by XAS analyses. This finding is another illustration of the complementary value of spectroscopic analyses assisting the interpretation of sequential extractions, in particular in samples with atypical matrix or from exceptional environmental settings.

4.3 Implications for sediment P dynamics

Combining results from sequential extractions with elemental and spectroscopic analyses leads to the conclusion that the addition of Fe salts to lake Terra Nova has resulted in the formation of multi-elemental co-precipitates containing Fe, Ca, P and minor amounts of Mn, which form a continuum of oligomers and nanoparticles embedded in the OM matrix and involving inner-sphere complexes of Fe with functional groups of the OM. These precipitates are likely authigenic and have been formed as a consequence of redox cycling of Fe in the sediments. During dosing of FeCl₃, Fe³⁺ hydrolysis caused the formation of large flocs in the water column, which settled on top of the sediment in the form of Fe hydroxides.⁹³ The transformation of these primary products of Fe addition into multi-elemental precipitates upon redox cycling in the sediment has consequences for the P binding and release in the sediment.

The formation of co-precipitates can enhance the efficiency of Fe addition for P binding in the sediment. Ratios between P and Fe in the corresponding Fe pools obtained by sequential extractions reached values up to about 0.3 in the top sediments of Terra Nova.¹ Fe(III)-OM complexes and CaFeP can have P/Fe ratios of up to 1.^{31,57} Hence, the formation of Fe-Ca-P-OM co-precipitates can constitute an efficient sink for P in sediments

of Fe treated lakes. However, a threshold P/Fe ratio of 0.12 was reported, at which, in aquatic sediments, the retention of P *via* adsorption is insufficient to prevent a significant increase in benthic P fluxes during temporal bottom water anoxia.⁹⁴⁻⁹⁶ Consequently, the P/Fe ratios in Terra Nova sediment indicate that the amount of P bound in Fe(III)-OM complexes and CaFeP could further increase in the future. Nevertheless, the current ratio already exceeds the threshold value beyond which release of P into porewater upon Fe(III) reduction under summer low oxygen conditions cannot be attenuated by adsorption to remaining Fe(III) phases. Further, analyses of P XANES spectra suggest that part of the Fe(III) is coordinated with organic P groups. As organically-bound P, mostly in the form of phosphate mono- and diesters, constitutes the largest P pool in the sediment, further research would be necessary to inquire possible competition of organic P for inorganic P adsorption and complexation with Fe in these sediments.

Although formation of co-precipitates is an efficient pathway for P binding, these precipitates might be particularly sensitive to P mobilization upon changes in redox conditions when bottom water hypoxia emerge. Freshly precipitated Fe(III) phosphates exhibit a high reactivity towards microbial reduction and the similarity of kinetic parameters with those obtained in estuarine sediments containing Fe(III) phosphate precipitates suggests, that those are the preferential electron acceptor for Fe reducing bacteria in these sediments.⁹⁷ Furthermore, linkage of Fe(III) (oxyhydroxides) with OM enhances microbial and chemical Fe(III) reduction^{98,99} and, in general, the small particle size and its wide distribution is expected to increase the reactivity of the Fe(III) containing co-precipitates in the sediment of Lake Terra Nova. In conclusion, formation of co-precipitates with Fe and Ca in the presence of OM presents an effective P sink in the sediments with respect to P/Fe ratio, but the product is eminently prone to P release under reducing conditions.

5 Conclusion

In organic-rich dynamic systems such as peat sediments, the classical model of Fe and P diagenesis, based on oxidative precipitation and reductive dissolution of Fe (oxyhydr)oxides does not apply. Addition of Fe and its redox cycling in the peat sediment of lake Terra Nova does not result in the formation of distinctive Fe(III) (oxyhydr)oxides in the top part of the sediment. Instead, the presence of OM and Ca inhibits polymerization and particle growth leading to finely dispersed solids intimately linked to the OM matter. Fe(III) produced by Fe(II) oxidation forms a continuum of highly amorphous Fe-Ca-(Mn)-P-OM co-precipitates, which are highly redox sensitive and responsible for the rapid release of P upon emerging bottom water hypoxia.

Methodologically, this study shows that X-ray spectroscopy provides significant additional information next to SEPs, with scanning X-ray microscopy as a valuable asset for qualitative insights into the presence and spatial distribution of individual species. Our results suggest that Fe-Asc represents a good estimate of the most redox-sensitive amorphous Fe(III) pool in organic-rich sediments and that P-NaHCO₃ and P-CDB together



yield an estimate of P associated with nanoscale inorganic-organic Ca-Fe-P-OM co-precipitates. On the other hand, the study demonstrates that concentrated HNO₃ does not quantitatively extract pyrite in the assigned reaction time and may therefore lead to its underestimation. Furthermore, Fe release from phyllosilicate minerals contributes to the HCl-extractable Fe, and this fraction should therefore not be interpreted exclusively as Fe(II) incorporated in authigenic phases. SEPs can thus deliver an indication of Fe and P pool reactivities in organic rich systems, but should be interpreted as such and not as discrete phases.

In conclusion, our study shows that organic-rich systems should not be approached using models and methods developed for organic poor systems, and that there is a need for new models and more fundamental research on OM-associated Fe phases to be able to reliably predict Fe and P dynamics in organic rich sediments.

Data availability

Data collected for this article including results of chemical analysis, correlation calculations and XAS spectra are available at Zenodo at <https://doi.org/10.5281/zenodo.12610105>. Data supporting this article including linear combination fit results and map and spectra identities are listed in the ESI.† Data of XAS reference spectra are available *via* the respective publication, or with the person who collected them, as listed in the ESI.‡ XAS raw data will be available from 2025 on under <https://doi.esrf.fr/10.15151/ESRF-ES-845384380>.

Author contributions

Melanie Münch: conceptualization, investigation, formal analysis, visualization, writing – original draft, writing – review & editing. Andreas Voegelin: conceptualization, methodology, investigation, formal analysis, supervision, writing – review & editing. Luis Carlos Colocho Hurtarte: methodology, investigation, formal analysis, resources, Jörg Göttlicher: investigation, formal analysis, resources. Thilo Behrends: conceptualization, funding acquisition, project administration, methodology, supervision, investigation, formal analysis, writing – review & editing.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This research was part of the P-Trap project funded by the European Union's Horizon 2020 Research and Innovation Program under the Marie Skłodowska-Curie grant agreement no. 813438. We acknowledge the European Radiation Facility (ESRF) for provision of synchrotron radiation facilities under proposal number ES-1169 and the staff of beamline ID21 for their great support. We would like to thank Dr Edward D. Burton (Southern Cross University, Australia) and Prof. Jörg Prietzel

(Technical University of Munich, Germany) for providing XAS reference spectra, and MSc Niccolò Pesenti for helping with the XAS measurements. We further thank three anonymous reviewers for their valuable comments which improved the manuscript.

References

- 1 M. A. Münch, R. van Kaam, K. As, S. Peiffer, G. T. Heerdt, C. P. Slomp, *et al.*, Impact of iron addition on phosphorus dynamics in sediments of a shallow peat lake 10 years after treatment, *Water Res.*, 2023, 120844, available from, <https://www.sciencedirect.com/science/article/pii/S0043135423012848>.
- 2 R. W. Sterner, On the phosphorus limitation paradigm for lakes, *Int. Rev. Hydrobiol.*, 2008, **93**(4-5), 433–445.
- 3 D. W. Schindler, R. E. Hecky, D. L. Findlay, M. P. Stainton, B. R. Parker, M. J. Paterson, *et al.*, Eutrophication of lakes cannot be controlled by reducing nitrogen input: Results of a 37-year whole-ecosystem experiment, *Proc. Natl. Acad. Sci. U. S. A.*, 2008, **105**(32), 11254–11258.
- 4 W. J. Brownlie, M. A. Sutton, D. S. Reay, K. V. Heal, L. Hermann, C. Kabbe, *et al.*, Global actions for a sustainable phosphorus future, *Nat. Food*, 2021, **2**(2), 71–74.
- 5 M. Shore, P. Jordan, P. E. Mellander, M. Kelly-Quinn, K. Daly, J. T. Sims, *et al.*, Characterisation of agricultural drainage ditch sediments along the phosphorus transfer continuum in two contrasting headwater catchments, *J. Soils Sediments*, 2016, **16**(5), 1643–1654, DOI: [10.1007/s11368-015-1330-0](https://doi.org/10.1007/s11368-015-1330-0).
- 6 P. Geranmayeh, K. M. Johannesson, B. Ulén and K. S. Tonderski, Particle deposition, resuspension and phosphorus accumulation in small constructed wetlands, *Ambio*, 2018, **47**(1), 134–145, DOI: [10.1007/s13280-017-0992-9](https://doi.org/10.1007/s13280-017-0992-9).
- 7 P. Kleinman, A. Sharpley, A. Buda, R. McDowell and A. Allen, Soil controls of phosphorus in runoff: Management barriers and opportunities, *Can. J. Soil Sci.*, 2011, **91**(3), 329–338, DOI: [10.4141/cjss09106](https://doi.org/10.4141/cjss09106).
- 8 P. J. A. Withers and H. P. Jarvie, Delivery and cycling of phosphorus in rivers: A review, *Sci. Total Environ.*, 2008, **400**(1), 379–395, available from, <https://www.sciencedirect.com/science/article/pii/S0048969708008139>.
- 9 M. Søndergaard, J. P. Jensen and E. Jeppesen, Role of sediment and internal loading of phosphorus in shallow lakes, *Hydrobiologia*, 2003, **506**(1), 135–145.
- 10 A. D. Steinman, and B. M. Spears, *Internal Phosphorus Loading in Lakes: Causes, Case Studies, and Management*, J. Ross Publishing, 2020.
- 11 B. M. Spears, D. P. Hamilton, Y. Pan, C. Zhaosheng and L. May, Lake management: is prevention better than cure?, *Inland Waters*, 2022, **12**(1), 173–186.
- 12 M. Søndergaard, R. Bjerring and E. Jeppesen, Persistent internal phosphorus loading during summer in shallow eutrophic lakes, *Hydrobiologia*, 2013, **710**(1), 95–107.



- 13 E. S. Bakker, E. Van Donk and A. K. Immers, Lake restoration by in-lake iron addition: a synopsis of iron impact on aquatic organisms and shallow lake ecosystems, *Aquat. Ecol.*, 2016, **50**(1), 121–135.
- 14 A. J. P. Smolders, L. P. M. Lamers, L. Echet, G. Van Der Velde and J. G. M. Roelofs, Internal eutrophication: How it works and what to do about it – A review, *Chem. Ecol.*, 2006, **22**(2), 93–111.
- 15 W. Einsele, Über die Beziehungen des Eisenkreislaufs zum Phosphatkreislauf im eutrophen See, *Arch. Hydrobiol.*, 1936, **29**, 664–686.
- 16 C. H. Mortimer, The Exchange of Dissolved Substances Between Mud and Water in Lakes, *J. Ecol.*, 1941, **29**(2), 280–329, <https://www.jstor.org/stable/2256395>.
- 17 P. Campbell and T. Torgersen, Maintenance of Iron Meromixis by Iron Redeposition in a Rapidly Flushed Monimolimnion, *Can. J. Fish. Aquat. Sci.*, 1980, **37**(8), 1303–1313.
- 18 A. Kleeberg, C. Herzog and M. Hupfer, Redox sensitivity of iron in phosphorus binding does not impede lake restoration, *Water Res.*, 2013, **47**(3), 1491–1502.
- 19 A. Kleeberg, A. Köhler and M. Hupfer, How effectively does a single or continuous iron supply affect the phosphorus budget of aerated lakes?, *J. Soils Sediments*, 2012, **12**(10), 1593–1603.
- 20 R. A. Berner, Sedimentary pyrite formation, *Am. J. Sci.*, 1970, **268**, 1–23.
- 21 P. Kraal, E. D. Burton and R. T. Bush, Iron monosulfide accumulation and pyrite formation in eutrophic estuarine sediments, *Geochim. Cosmochim. Acta*, 2013, **122**, 75–88.
- 22 J. Gerke, Phosphate adsorption by humic/Fe-oxide mixtures aged at pH 4 and 7 and by poorly ordered Fe-oxide, *Geoderma*, 1993, **59**(1–4), 279–288.
- 23 C. Mikutta and R. Kretzschmar, Spectroscopic evidence for ternary complex formation between arsenate and ferric iron complexes of humic substances, *Environ. Sci. Technol.*, 2011, **45**(22), 9550–9557.
- 24 A. Sundman, T. Karlsson, S. Sjöberg and P. Persson, Impact of iron–organic matter complexes on aqueous phosphate concentrations, *Chem. Geol.*, 2016, **426**, 109–117.
- 25 K. Lalonde, A. Mucci, A. Ouellet and Y. Gélinas, Preservation of organic matter in sediments promoted by iron, *Nature*, 2012, **483**(7388), 198–200, DOI: [10.1038/nature10855](https://doi.org/10.1038/nature10855).
- 26 T. Karlsson and P. Persson, Coordination chemistry and hydrolysis of Fe(III) in a peat humic acid studied by X-ray absorption spectroscopy, *Geochim. Cosmochim. Acta*, 2010, **74**(1), 30–40.
- 27 T. Karlsson and P. Persson, Complexes with aquatic organic matter suppress hydrolysis and precipitation of Fe(III), *Chem. Geol.*, 2012, **322–323**, 19–27.
- 28 L. K. Thomasarrigo, J. M. Byrne, A. Kappler and R. Kretzschmar, Impact of Organic Matter on Iron(II)-Catalyzed Mineral Transformations in Ferrihydrite-Organic Matter Coprecipitates, *Environ. Sci. Technol.*, 2018, **52**(21), 12316–12326, DOI: [10.1021/acs.est.8b03206](https://doi.org/10.1021/acs.est.8b03206).
- 29 J. Dolfing, W. Chardon and J. Japenga, Association between colloidal iron, aluminum, phosphorus and humic acids, *Soil Sci.*, 1999, **164**(3), 171–179.
- 30 C. Moens and E. Smolders, Suwannee River Natural Organic Matter concentrations affect the size and phosphate uptake of colloids formed by iron oxidation, *Geochim. Cosmochim. Acta*, 2021, **312**, 375–391.
- 31 J. Gerke, Humic (organic matter)-Al(Fe)-phosphate complexes: An underestimated phosphate form in soils and source of plant-available phosphate, *Soil Sci.*, 2010, **175**(9), 417–425.
- 32 I. Guardado, O. Urrutia and J. M. Garcia-Mina, Size Distribution, Complexing Capacity, and Stability of Phosphate–Metal–Humic Complexes, *J. Agric. Food Chem.*, 2007, **55**, 408–413.
- 33 J. Riggle and R. von Wandruszka, Binding of inorganic phosphate to dissolved metal humates, *Talanta*, 2005, **66**(2), 372–375, available from, <https://www.sciencedirect.com/science/article/pii/S0039914004006605>.
- 34 C. Yu, J. J. Virtasalo, T. Karlsson, P. Peltola, P. Österholm, E. D. Burton, *et al.*, Iron behavior in a northern estuary: large pools of non-sulfidized Fe (II) associated with organic matter, *Chem. Geol.*, 2015, **413**, 73–85.
- 35 I. Guardado, O. Urrutia and J. M. Garcia-Mina, Some structural and electronic features of the interaction of phosphate with metal–humic complexes, *J. Agric. Food Chem.*, 2008, **56**(3), 1035–1042.
- 36 D. W. O'Connell, N. Ansems, R. K. Kukkadapu, D. Jaisi, D. M. Orihel, B. J. Cade-Menun, *et al.*, Changes in Sedimentary Phosphorus Burial Following Artificial Eutrophication of Lake 227, Experimental Lakes Area, Ontario, Canada, *J. Geophys. Res.: Biogeosci.*, 2020, **125**(8), e2020JG005713.
- 37 K. C. Ruttenberg, Development of a sequential extraction method for different forms of phosphorus in marine sediments, *Limnol. Oceanogr.*, 1992, **37**(7), 1460–1482.
- 38 R. Psenner and R. Puckso, Phosphorus fractionation: Advantages and limits of the method for the study of sediment P origins and interactions, *Arch. Hydrobiol.*, 1988, **30**, 43–59.
- 39 S. R. Claff, L. A. Sullivan, E. D. Burton and R. T. Bush, A sequential extraction procedure for acid sulfate soils: Partitioning of iron, *Geoderma*, 2010, **155**(3–4), 224–230, DOI: [10.1016/j.geoderma.2009.12.002](https://doi.org/10.1016/j.geoderma.2009.12.002).
- 40 S. W. Poulton and D. E. Canfield, Development of a sequential extraction procedure for iron: Implications for iron partitioning in continentally derived particulates, *Chem. Geol.*, 2005, **214**(3–4), 209–221.
- 41 R. Raiswell, H. P. Vu, L. Brinza and L. G. Benning, The determination of labile Fe in ferrihydrite by ascorbic acid extraction: Methodology, dissolution kinetics and loss of solubility with age and de-watering, *Chem. Geol.*, 2010, **278**(1–2), 70–79, DOI: [10.1016/j.chemgeo.2010.09.002](https://doi.org/10.1016/j.chemgeo.2010.09.002).
- 42 I. C. Regelink, A. Voegelin, L. Weng, G. F. Koopmans and R. N. J. Comans, Characterization of Colloidal Fe from Soils Using Field-Flow Fractionation and Fe K - Edge X -



- ray Absorption Spectroscopy, *Environ. Sci. Technol.*, 2014, **48**(8), 4307–4316.
- 43 L. L. Baker, D. G. Strawn, K. L. Vaughan and P. A. McDaniel, XAS study of Fe mineralogy in a chronosequence of soil clays formed in basaltic cinders, *Clays Clay Miner.*, 2010, **58**, 772–782.
- 44 W. Li, S. R. Joshi, G. Hou, D. J. Burdige, D. L. Sparks and D. P. Jaisi, Characterizing phosphorus speciation of Chesapeake Bay sediments using chemical extraction, ³¹P NMR, and X-ray absorption fine structure spectroscopy, *Environ. Sci. Technol.*, 2015, **49**(1), 203–211.
- 45 P. Kraal, N. Dijkstra, T. Behrends and C. P. Slomp, Phosphorus burial in sediments of the sulfidic deep Black Sea: Key roles for adsorption by calcium carbonate and apatite authigenesis, *Geochim. Cosmochim. Acta*, 2017, **204**, 140–158, DOI: [10.1016/j.gca.2017.01.042](https://doi.org/10.1016/j.gca.2017.01.042).
- 46 D. S. Baldwin, The phosphorus composition of a diverse series of Australian sediments, *Hydrobiologia*, 1996, **335**(1), 63–73.
- 47 C. P. Slomp, E. H. G. Epping, W. Helder and W. Van Raaphorst, A key role for iron-bound phosphorus in authigenic apatite formation in North Atlantic continental platform sediments, *J. Mar. Res.*, 1996, **54**(6), 1179–1205.
- 48 S. R. Claff, E. D. Burton, L. A. Sullivan and R. T. Bush, Effect of sample pretreatment on the fractionation of Fe, Cr, Ni, Cu, Mn, and Zn in acid sulfate soil materials, *Geoderma*, 2010, **159**(1–2), 156–164, DOI: [10.1016/j.geoderma.2010.07.007](https://doi.org/10.1016/j.geoderma.2010.07.007).
- 49 J. van Dijk, M. Amesz, M. Ouboter, L. Moria, G. ter Heerdt, and T. van der Wijngaart, *Aktualisatie KRW-Waterlichamen AGV: Maatregelenprogramma 2022-2027*, Amsterdam, 2020.
- 50 N. A. G. M. van Helmond, T. Jilbert and C. P. Slomp, Hypoxia in the Holocene Baltic Sea: Comparing modern *versus* past intervals using sedimentary trace metals, *Chem. Geol.*, 2018, **493**(June), 478–490, DOI: [10.1016/j.chemgeo.2018.06.028](https://doi.org/10.1016/j.chemgeo.2018.06.028).
- 51 D. R. Lovley and E. J. P. Phillips, Rapid Assay for Microbially Reducible Ferric Iron in Aquatic Sediments, *Appl. Environ. Microbiol.*, 1987, **53**(7), 1536–1540.
- 52 J. A. McKeague and J. H. Day, Dithionite and oxalate-extractable Fe and Al as aids in differentiating various classes of soils, *Can. J. Soil Sci.*, 1966, **46**(1), 13–22, DOI: [10.4141/cjss66-003](https://doi.org/10.4141/cjss66-003).
- 53 C. T. Parsons, F. Rezaeezhad, D. W. O'Connell and P. Van Cappellen, Sediment phosphorus speciation and mobility under dynamic redox conditions, *Biogeosciences*, 2017, **14**(14), 3585–3602.
- 54 M. Salomé, M. Cotte, R. Baker, R. Barrett, N. Benseny-Cases, G. Berruyer, *et al.*, The ID21 Scanning X-ray Microscope at ESRF, *J. Phys.: Conf. Ser.*, 2013, **425**(18), 182004, DOI: [10.1088/1742-6596/425/18/182004](https://doi.org/10.1088/1742-6596/425/18/182004).
- 55 F. Werner, C. W. Mueller, J. Thieme, A. Gianoncelli, C. Rivard and C. Höschel, Micro-scale heterogeneity of soil phosphorus depends on soil substrate and depth, *Sci. Rep.*, 2017, **7**(1), 3203.
- 56 B. Ravel and M. Newville, ATHENA, ARTEMIS, HEPHAESTUS: Data analysis for X-ray absorption spectroscopy using IFEFFIT, *J. Synchrotron Radiat.*, 2005, **12**, 537–541, DOI: [10.1107/FS0909049505012719&partnerID=40&md5=450ffc75c60d60335c23ce51c2dd2635](https://doi.org/10.1107/FS0909049505012719&partnerID=40&md5=450ffc75c60d60335c23ce51c2dd2635).
- 57 A. C. Senn, R. Kaegi, S. J. Hug, J. G. Hering, S. Mangold and A. Voegelin, Composition and structure of Fe(III)-precipitates formed by Fe(II) oxidation in water at near-neutral pH: Interdependent effects of phosphate, silicate and Ca, *Geochim. Cosmochim. Acta*, 2015, **162**, 220–246.
- 58 J. Frommer, A. Voegelin, J. Dittmar, M. A. Marcus and R. Kretzschmar, Biogeochemical processes and arsenic enrichment around rice roots in paddy soil: results from micro-focused X-ray spectroscopy, *Eurasian J. Soil Sci.*, 2011, **62**(2), 305–317.
- 59 M. Ma, P. Overvest, A. Hijlkema, S. Mangold, C. McCammon, A. Voegelin, *et al.*, Phosphate burial in aquatic sediments: Rates and mechanisms of vivianite formation from mackinawite, *Chem. Eng. J. Adv.*, 2023, **16**, 100565.
- 60 X. Chen, X. Zhang, J. Wan, Z. Wang and Y. Qian, Selective fabrication of metastable greigite (Fe₃S₄) nanocrystallites and its magnetic properties through a simple solution-based route, *Chem. Phys. Lett.*, 2005, **403**(4–6), 396–399.
- 61 E. D. Burton, R. T. Bush, L. A. Sullivan, R. K. Hocking, D. R. G. Mitchell, S. G. Johnston, *et al.*, Iron-Monosulfide Oxidation in Natural Sediments: Resolving Microbially Mediated S Transformations Using XANES, Electron Microscopy, and Selective Extractions, *Environ. Sci. Technol.*, 2009, **43**(9), 3128–3134, DOI: [10.1021/es8036548](https://doi.org/10.1021/es8036548).
- 62 M. Wolthers, S. J. Van Der Gaast, and D. Rickard. The structure of disordered mackinawite. *2003***388**(11–12):2007–2015. Available from: doi: DOI: [10.2138/am-2003-11-1245](https://doi.org/10.2138/am-2003-11-1245).
- 63 A. Al-Borno and M. B. Tomson, The temperature dependence of the solubility product constant of vivianite, *Geochim. Cosmochim. Acta*, 1994, **58**(24), 5373–5378.
- 64 J. Wielinski, J. Jimenez-Martinez, J. Göttlicher, R. Steininger, S. Mangold, S. J. Hug, *et al.*, Spatiotemporal Mineral Phase Evolution and Arsenic Retention in Microfluidic Models of Zerovalent Iron-Based Water Treatment, *Environ. Sci. Technol.*, 2022, **56**(19), 13696–13708, DOI: [10.1021/acs.est.2c02189](https://doi.org/10.1021/acs.est.2c02189).
- 65 L. C. Colocho Hurtarte, L. F. Souza-Filho, W. Oliveira Santos, L. Vergütz, J. Prietzel and D. Hesterberg, Optimization of Data Processing Minimizes Impact of Self-Absorption on Phosphorus Speciation Results by P K-Edge XANES, *Soil Syst.*, 2019, **3**(3), 61.
- 66 Z. He, C. W. Honeycutt, T. Zhang and P. M. Bertsch, Preparation and FT-IR characterization of metal phytate compounds, *J. Environ. Qual.*, 2006, **35**(4), 1319–1328.
- 67 J. Prietzel, G. Harrington, W. Hausler, K. Heister, F. Werner and W. Klysubun, Reference spectra of important adsorbed organic and inorganic phosphate binding forms for soil P speciation using synchrotron-based K-edge XANES spectroscopy, *J. Synchrotron Radiat.*, 2016, **23**(2), 532–544, DOI: [10.1107/S1600577515023085](https://doi.org/10.1107/S1600577515023085).
- 68 H. L. C. Colocho, H. C. Santana Amorim, J. Kruse, J. Criginski Cezar, W. Klysubun and J. Prietzel, A Novel Approach for the Quantification of Different Inorganic and



- Organic Phosphorus Compounds in Environmental Samples by P L2,3-Edge X-ray Absorption Near-Edge Structure (XANES) Spectroscopy, *Environ. Sci. Technol.*, 2020, **54**(5), 2812–2820, DOI: [10.1021/acs.est.9b07018](https://doi.org/10.1021/acs.est.9b07018).
- 69 V. A. Solé, E. Papillon, M. Cotte, P. Walter and J. Susini, A multiplatform code for the analysis of energy-dispersive X-ray fluorescence spectra, *Spectrochim. Acta, Part B*, 2007, **62**(1), 63–68.
- 70 J. Schindelin, C. T. Rueden, M. C. Hiner and K. W. Eliceiri, The ImageJ ecosystem: An open platform for biomedical image analysis, *Mol. Reprod. Dev.*, 2015, **82**(7–8), 518–529.
- 71 L. K. Huang and M. J. J. Wang, Image thresholding by minimizing the measures of fuzziness, *Pattern Recogn.*, 1995, **28**(1), 41–51.
- 72 F. E. Harrell Jr, *Harrell Jr M. F. E. Package 'hmisc'. CRAN2018*, 2019, pp. 235–236.
- 73 J. Babej, R. Červenka, J. Kuta, K. Brabec and Z. Máčka, Organic and inorganic carbon storage in riparian zones of Central European rivers—A comparison of methods for their determination, *Quat. Int.*, 2023, **681**, 44–51.
- 74 L. Celi, and E. Barberis, *Abiotic reactions of inositol phosphates in soil*, 2007.
- 75 B. L. Turner, M. J. Papházy, P. M. Haygarth and I. D. Mckelvie, Inositol phosphates in the environment, *Philos. Trans. R. Soc. London, Ser. B*, 2002, **357**(1420), 449–469.
- 76 C. Hyacinthe and P. van Cappellen, An authigenic iron phosphate phase in estuarine sediments: composition, formation and chemical reactivity, *Mar. Chem.*, 2004, **91**(1–4), 227–251.
- 77 A. Vilgé-Ritter, J. Rose, A. Masion, J. Y. Bottero and J. M. Lainé, Chemistry and structure of aggregates formed with Fe-salts and natural organic matter, *Colloids Surf., A*, 1999, **147**(3), 297–308.
- 78 H. Guénet, M. Davranche, D. Vantelon, J. Gigault, S. Prévost, O. Taché, *et al.*, Characterization of iron–organic matter nano-aggregate networks through a combination of SAXS/SANS and XAS analyses: impact on As binding, *Environ. Sci.: Nano*, 2017, **4**(4), 938–954.
- 79 D. Vantelon, M. Davranche, R. Marsac, C. La Fontaine, H. Guénet, J. Jestin, *et al.*, Iron speciation in iron–organic matter nanoaggregates: a kinetic approach coupling QuickEXAFS and MCR-ALS chemometrics, *Environ. Sci.: Nano*, 2019, **6**(8), 2641–2651.
- 80 A. Beauvois, D. Vantelon, J. Jestin, C. Rivard, M. Bouhnik-Le Coz, A. Dupont, *et al.*, How does calcium drive the structural organization of iron–organic matter aggregates? A multiscale investigation, *Environ. Sci.: Nano*, 2020, **7**(9), 2833–2849.
- 81 N. Dijkstra, C. P. Slomp and T. Behrends, Vivianite is a key sink for phosphorus in sediments of the Landsort Deep, an intermittently anoxic deep basin in the Baltic Sea, *Chem. Geol.*, 2016, **438**, 58–72, DOI: [10.1016/j.chemgeo.2016.05.025](https://doi.org/10.1016/j.chemgeo.2016.05.025).
- 82 J. Kruse and P. Leinweber, Phosphorus in sequentially extracted fen peat soils: AK-edge X-ray absorption near-edge structure (XANES) spectroscopy study, *J. Plant Nutr. Soil Sci.*, 2008, **171**(4), 613–620.
- 83 J. R. Flatt and R. Woods, A voltammetric investigation of the oxidation of pyrite in nitric acid solutions: relation to treatment of refractory gold ores, *J. Appl. Electrochem.*, 1995, **25**(9), 852–856.
- 84 H. Long and D. G. Dixon, Pressure oxidation of pyrite in sulfuric acid media: a kinetic study, *Hydrometallurgy*, 2004, **73**(3–4), 335–349.
- 85 C. O. Moses, D. K. Nordstrom, J. S. Herman and A. L. Mills, Aqueous pyrite oxidation by dissolved oxygen and by ferric iron, *Geochim. Cosmochim. Acta*, 1987, **51**(6), 1561–1571.
- 86 D. T. Rickard, Kinetics and mechanism of the sulfidation of goethite, *Am. J. Sci.*, 1974, **274**(8), 941–952.
- 87 S. Hunger and L. G. Benning, Greigite: a true intermediate on the polysulfide pathway to pyrite, *Geochem. Trans.*, 2007, **8**, 1–20.
- 88 J. W. Stucki, D. C. Golden and C. B. Roth, Effects of Reduction and Reoxidation of Structural Iron on the Surface Charge and Dissolution of Dioctahedral Smectites, *Clays Clay Miner.*, 1984, **32**(5), 350–356, DOI: [10.1346/CCMN.1984.0320502](https://doi.org/10.1346/CCMN.1984.0320502).
- 89 A. Neumann, S. Petit and T. B. Hofstetter, Evaluation of redox-active iron sites in smectites using middle and near infrared spectroscopy, *Geochim. Cosmochim. Acta*, 2011, **75**(9), 2336–2355, available from, <https://www.sciencedirect.com/science/article/pii/S0016703711000883>.
- 90 M. A. Arshad, R. J. St. Arnaud and P. M. Huang, Dissolution of trioctahedral layer silicates by ammonium oxalate, sodium dithionite-citrate-bicarbonate, and potassium pyrophosphate, *Can. J. Soil Sci.*, 1972, **52**(1), 19–26, DOI: [10.4141/cjss72-002](https://doi.org/10.4141/cjss72-002).
- 91 M. E. Tuccillo, I. M. Cozzarelli and J. S. Herman, Iron reduction in the sediments of a hydrocarbon-contaminated aquifer, *Appl. Geochem.*, 1999, **14**(5), 655–667, available from, <https://www.sciencedirect.com/science/article/pii/S0883292798000894>.
- 92 T. S. Peretyazhko, J. M. Zachara, R. K. Kukkadapu, S. M. Heald, I. V. Kutnyakov, C. T. Resch, *et al.*, Perchnetate (TcO₄[−]) reduction by reactive ferrous iron forms in naturally anoxic, redox transition zone sediments from the Hanford Site, USA, *Geochim. Cosmochim. Acta*, 2012, **92**, 48–66, available from, <https://www.sciencedirect.com/science/article/pii/S0016703712003511>.
- 93 G. ter Heerdt, *Ijzersuppletie in Laagveenplassen – Rapport 2012-43*, STOWA, STOWA, Amersfoort, Nederlands, 2012, available from, <https://www.stowa.nl/>.
- 94 I. Hilbrandt, V. Lehmann, F. Zietzschmann, A. S. Ruhl and M. Jekel, Quantification and isotherm modelling of competitive phosphate and silicate adsorption onto micro-sized granular ferric hydroxide, *RSC Adv.*, 2019, **9**(41), 23642–23651, DOI: [10.1039/C9RA04865K](https://doi.org/10.1039/C9RA04865K).
- 95 E. Smolders, E. Baetens, M. Verbeeck, S. Nawara, J. Diels, M. Verdievel, B. Peeters, W. De Cooman and S. Baken, Internal Loading and Redox Cycling of Sediment Iron



- Explain Reactive Phosphorus Concentrations in Lowland Rivers, *Environ. Sci. Technol.*, 2017, **51**(5), 2584–2592.
- 96 L. Xia, *The Application of Iron-Rich Sorbents to Mitigate Phosphorus Release from Sediments in Lowland Rivers*, KU Leuven, Leuven, 2022.
- 97 C. Hyacinthe, S. Bonneville and P. Van Cappellen, Reactive iron(III) in sediments: chemical *versus* microbial extractions, *Geochim. Cosmochim. Acta*, 2006, **70**(16), 4166–4180.
- 98 J. Chen, B. Gu, R. A. Royer and W. D. Burgos, The roles of natural organic matter in chemical and microbial reduction of ferric iron, *Sci. Total Environ.*, 2003, **307**(1–3), 167–178.
- 99 A. Fritzsche, J. Bosch, M. Sander, C. Schroder, J. M. Byrne, T. Ritschel, *et al.*, Organic matter from redoximorphic soils accelerates and sustains microbial Fe(III) reduction, *Environ. Sci. Technol.*, 2021, **55**(15), 10821–10831.

