



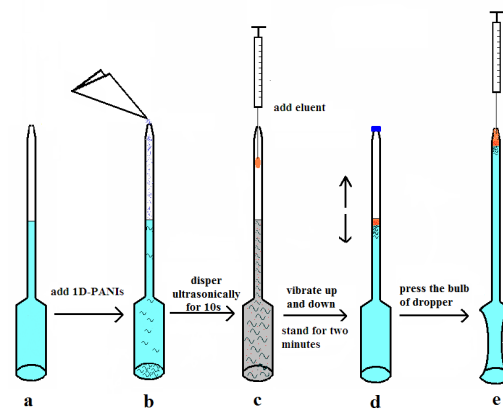
**A novel dispersive micro-solid phase extraction method  
combined with gas chromatography for analysis of  
organochlorine pesticides in aqueous samples**

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### Graphical Abstract

#### Highlights

1D-PANIs has good dispersibility and permeability.  
A plastic dropper was acted as both the extraction and preconcentration device.  
The proposed method is simple, inexpensive and highly efficient.



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4 1 **A novel dispersive micro-solid phase extraction method**  
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6 2 **combined with gas chromatography for analysis of**  
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8 3 **organochlorine pesticides in aqueous samples**  
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3 **Abstract:**  
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5 A novel dispersive micro-solid phase extraction (D- $\mu$ -SPE) method was  
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7 presented in this paper. The most attractive feature of this method is taking full  
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9 advantage of a plastic dropper and nanoscale one-dimensional polyaniline (1D-PANI).  
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11 The plastic dropper employed as both the extraction and preconcentration device. All  
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13 procedures needed in the common  $\mu$ -SPE were performed in a plastic dropper without  
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15 any additional devices, such as centrifuge, magnetic field and pressure blowing  
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17 concentrator. So the operating steps were simplified greatly. 1D-PANIs was used as  
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19 the sorbent. The huge surface and interconnected net-work structure of 1D-PANIs  
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21 endows it high dispersive adsorbability and good permeability for the adsorbed  
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23 analytes to be eluated. To test the feasibility of the proposed method, eight  
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25 organochlorine pesticides (OCPs) were selected as the model analytes for D- $\mu$ -SPE of  
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27 OCPs in water samples. A series of extraction parameters have been investigated  
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29 systematically. Under optimized conditions, the method showed the linear correlation  
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31 coefficients ( $r$ ) were better than 0.9971 and the limits of detection (LODs) for eight  
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33 OCPs ranged from 0.0121 to 0.0468  $\mu\text{g L}^{-1}$ . The intra- and inter-day relative standard  
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35 deviations (RSDs) were less than 11.9%. The recoveries of OCPs for three spiked  
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37 aqueous samples ranged from 73.6 to 107.0%. The novel D- $\mu$ -SPE is promising to be  
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39 an alternative sample preparation method for extracting apolar analytes in complex  
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41 sample matrices because of the simplicity, low cost and high extraction efficiency.  
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55 **Keywords:** Dispersive micro-solid phase extraction; Plastic dropper;  
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57 One-dimensional polyaniline; Organochlorine pesticides; Gas chromatography.  
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5 **1. Introduction:**  
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9 56 Sample preparation is an important step in trace analysis. The major problem  
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11 57 encountered is the separation of pollutants from matrix components. Owing to the low  
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13 58 concentration level of pollutants, inefficient extractions always cause loss of analytes  
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15 59 and make the analytes difficult to be determined. Hence, the development of efficient  
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17 60 extraction and preconcentration steps prior to accurate determination of trace level of  
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19 61 compounds has been explored in considerable depth over recent decades.  
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24 62 Generally, liquid-liquid extraction (LLE) [1] and solid-phase extraction (SPE)  
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26 63 [2,3] are commonly used steps in extraction and preconcentration of compounds from  
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28 64 various samples. SPE has distinguished from many other extraction techniques due to  
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30 65 the advantages such as lower cost, higher enrichment factor and less consumption of  
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32 66 organic solvents [4]. Although SPE is being applied broadly, it suffers from some  
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34 67 shortcomings such as solvent loss, large secondary wastes, a long procedure, and a  
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36 68 need for complex equipment. Dispersive micro-solid phase extraction (D- $\mu$ -SPE) is  
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38 69 categorized as a SPE technique. The D- $\mu$ -SPE exhibits some advantages over  
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40 70 traditional SPE, such as convenience for efficiency of recovery and reduced solvent  
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42 71 consumption [5,6]. Moreover, it is simple, economic and easy to perform [7,8].  
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44 72 Various sorbents can be employed in D- $\mu$ -SPE. Compared to traditional SPE sorbents,  
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46 73 nanomaterials possess large surface area and short diffusion route, which may result  
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48 74 in high extraction efficiency and rapid extraction dynamics. Hence,  
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50 75 nanomaterial-based D- $\mu$ -SPE as a novel SPE method aroused more and more concern  
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4 76 of researchers [9,10]. The nano-sorbent used in D- $\mu$ -SPE have been extended to  
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6 77 various forms of carbon [11] and conductive polymers (CPs) [12,13], etc. Conductive  
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9 78 polymers (CPs) are organic materials that possess an extended conjugated-system  
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11 79 associated with the polymer backbone, and have been known to have great application  
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14 80 potential in electroluminescent devices, electromagnetic shielding, sensors, corrosion  
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16 81 prevention, etc. [12,14-18]. Recent researches suggest CPs are also promising in  
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19 82 separation science owing to their versatile properties such as hydrophobicity,  
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21 83 large-conjugated structure, polar functional groups, ion exchange property, and so on  
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24 84 [12,19-21]. In a previous work, various forms of CP nanomaterials, including  
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26 85 zero-dimensional (0D) CPs nanoparticles and one-dimensional (1D) CPs (eg.  
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29 86 Nanofiber, nanotube, nanowire) were prepared and investigated as the sorbents in  
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31 87 D- $\mu$ -SPE [6,13,19-21]. It was found that 0D nanoparticles tended to agglomerate  
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34 88 closely after D- $\mu$ -SPE because of their high surface energy and solvophobic property  
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36 89 [14,22], which probably discounted the extraction efficiency. Furthermore, it is  
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39 90 difficult to eluate the target analytes from the tightly packed nanoparticles. 1D-CPs  
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41 91 are generally nano-scaled in diameter with lengths over several microns, and still  
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44 92 remain the advantage of high surface area. Although 1D-CPs also tend to agglomerate  
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47 93 in many solvents, the resulting interconnected network-like structure can offer better  
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49 94 permeability in comparison with fine powders and tightly packed nanoparticles  
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51 95 [23,24], and may obtain better extraction efficiency than 0D-CPs as the sorbents.

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54 96 In this work, 1D-PANI was chosen as the sorbent, and eight organic chloride  
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56 97 pesticides (OCPs) were selected as model analytes. A novel D- $\mu$ -SPE method was  
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4 98 developed in a plastic dropper which was acted as both extraction and  
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6 99 preconcentration devices. No complex equipments, such as centrifuge, magnetic field  
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9 100 and pressure blowing concentrator, were needed. 1D-PANIs was dispersed to extract  
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11 101 OCPs in water loaded in a plastic dropper under ultrasonic radiation. After extraction,  
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13 102 1D-PANIs could be agglomerated with the addition of the eluent (n-hexane) which  
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16 103 owns lower density than water. The agglomerated 1D-PANI located between the  
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19 104 aqueous phase and organic phase (n-hexane) at the narrow end of the plastic dropper.  
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21 105 The plastic dropper was vibrated up and down for several seconds and stand for a  
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24 106 while, and then the eluent floated again on the gathered 1D-PANIs. The eluent was  
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26 107 pressed up to the tip end of the plastic dropper and withdrawn to be analyzed by gas  
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29 108 chromatography (GC). Some parameters which influenced the extraction efficiency  
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31 109 were investigated. Hence, a 1D-PANI based D- $\mu$ -SPE was developed and applied for  
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34 110 monitoring eight OCPs in some environmental water samples. Neither centrifugation  
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37 111 nor magnetic separation was needed. The novel D- $\mu$ -SPE is promising to be an  
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39 112 alternative sample preparation method for extracting apolar analytes in complex  
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41 113 sample matrices because of the simplicity, low cost and short sample preparation  
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44 114 time.

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## 49 116 **2. Experimental**

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### 54 118 **2.1 Instruments**

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4 120 Chromatographic analysis was carried out on an Agilent 6890N GC (Agilent  
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6 121 Technologies, USA) with a micro-cell electron capture detector ( $\mu$ -ECD). Analytes  
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9 122 were separated by a DB-17 fused silica capillary column (30m $\times$ 0.25mm I.D., 0.25  $\mu$ m  
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11 123 film thickness, J & W Scientific, Folsom, CA, USA). The column oven temperature  
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14 124 was initially set at 150  $^{\circ}$ C, held on 1 min, and then raised to 280  $^{\circ}$ C at 8  $^{\circ}$ C min $^{-1}$  and  
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16 125 hold on 8 min. The ECD temperature was set at 300  $^{\circ}$ C. The inlet was set at 280  $^{\circ}$ C, in  
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19 126 splitless mode. Nitrogen (purity 99.9995%) was used as carrier gas at the flow rate of  
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21 127 0.5 mL min $^{-1}$  and make up gas at 60 mL min $^{-1}$ . Ultrasonic instrument KQ-100DE was  
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24 128 purchased from Kunshan Ultrasonic Instrument (Jiangsu, China).

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26 129 The 1D-PANIs were characterized by JEOL JSM-6500F scanning electron  
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29 130 microscope (SEM, Jeol, Japan).

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## 33 34 132 **2.2 Reagents and materials**

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39 134 Aniline and ammonium persulfate ((NH<sub>4</sub>)<sub>2</sub>S<sub>2</sub>O<sub>8</sub>) were purchased from Sinopharm  
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41 135 Chemical Reagent (Shanghai, China) and were used directly without further  
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44 136 purification. N-hexane (Pesticide analysis grade) was ordered from Tedia (Fair Lawn,  
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46 137 New Jersey, USA). Organochlorine pesticides  $\alpha$ ,  $\beta$ ,  $\gamma$  and  $\delta$ -HCH; p, p'-DDE; p,  
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49 138 p'-DDD; p, p'-DDT; o, p'-DDT was purchased from Sigma-Aldrich (St. Louis, MO,  
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51 139 USA). Sodium chloride from Zhanyun Chemical Co., Ltd. (Shanghai, China) was  
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54 140 used in the experiment. Ultrapure water was purified on a Mill-Q water purification  
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56 141 system (Millipore Corporation, Billerica, MA, USA).



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4 142 The stock solutions containing  $20 \mu\text{g mL}^{-1}$  of OCPs were prepared in HPLC  
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6 143 grade methanol. A series of standard solutions were daily prepared by appropriate  
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9 144 dilution from the stock solution with ultrapure water, and all of which were stored at  
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11 145  $4 \text{ }^{\circ}\text{C}$  in a refrigerator.  
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### 15 16 147 **2.3 Sample preparation**

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21 149 Danjiang Kou water samples were collected from Danjiang Kou reservoir  
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24 150 (Danjiang Kou, China). Ponding water samples were collected from a vegetable plot  
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26 151 (Wuhan, China). Lake water samples were collected from South Lake (Wuhan, China),  
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29 152 and all of them were kept at  $4 \text{ }^{\circ}\text{C}$ . These real water samples were filtered through a  
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31 153  $0.45 \mu\text{m}$  polyether sulfone membrane which was purchased from Wuhan Shenshi  
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34 154 Chemical Industry Co. Ltd (Wuhan, China) prior to analysis.  
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### 37 38 39 156 **2.5 Proposed D- $\mu$ -SPE procedure**

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44 158 Preparation and characterization of 1D-PANIs are described in the  
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46 159 Supplementary Information (SI) [14,25,26]. The proposed D- $\mu$ -SPE procedure is  
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49 160 roughly illustrated by Fig. 1. Typically,  $3.0 \text{ mg}$  1D-PANIs powder used as the sorbent  
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51 161 was added to  $6.5 \text{ mL}$  aqueous samples spiked with variable known amounts of OCPs  
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54 162 in a  $3 \text{ mL}$  plastic dropper (containing  $15\% \text{ NaCl}$ ). The mixture was sonicated for  
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56 163 about 10 seconds to form a homogeneous dispersion solution due to the high  
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4 164 dispersity of nanoscale 1D-PANIs, and then 1D-PANIs was agglomerated after adding  
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6 165 65  $\mu\text{L}$  n-hexane as the eluent. The eluent floated up on the gathered 1D-PANIs  
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9 166 because the density of eluent is lower than water. Vibrated the plastic dropper up and  
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11 167 down adequately, and then let it stand for two minutes. The eluent refloated up on the  
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13 168 surface of 1D-PANIs, ascending at the tip of the plastic dropper by pressing the bulb  
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16 169 end of the plastic dropper. 1  $\mu\text{L}$  was withdrawn by a microsyringe for GC analysis.  
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### 21 171 **3. Results and discussion**

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#### 26 173 **3.1 Application to the proposed D- $\mu$ -SPE**

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31 175 Due to the large surface area and high affinity towards many organic compounds,  
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33 176 1D-PANIs have been recently exploited as the SPE sorbents [24]. In the present case,  
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36 177 the time-consuming procedures, such as centrifugation, elution and even magnetic  
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39 178 separation were left out. A plastic dropper were used as a water sample container, and  
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41 179 the high dispersive adsorption capacity and good permeability of interconnected  
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44 180 network-like structure of the 1D-PANIs (Fig. 1s, see SI) were taken full advantage.  
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46 181 Here, eight OCPs ( $\alpha$ -HCH,  $\beta$ -HCH,  $\gamma$ -HCH,  $\delta$ -HCH, P,P'-DDE, P,P'-DDD, O,P'-DDT,  
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49 182 P,P'-DDT) were chosen as model compounds to verify the feasibility of proposed  
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51 183 D- $\mu$ -SPE method. The parameters that might affect the performance of the extraction,  
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54 184 such as sample volume, 1D-PANIs amount, extraction time (see SI), ionic strength  
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57 185 (see SI), and desorption conditions, etc. were investigated as follows.  
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6 187 **3.1.1. Effect of sample volume**

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10 189 It is evident that an increase in sample volume enhances the amount of the target  
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12 190 analytes adsorbed onto the sorbents and transferred to the eluent, which improves the  
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14 191 sensitivity and extraction efficiency. Firstly, different volumes of plastic dropper  
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16 192 should be considered in our study because it influenced the sample volume. The  
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18 193 results showed that the water sample in 1 mL plastic dropper is too small to bring  
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20 194 about low extraction efficiency. 5 mL plastic dropper is so big that 1D-PANIs can not  
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22 195 gathered with a certain thickness at the narrow end. Hence, the eluent can not form a  
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24 196 drop on the gathered 1D-PANIs. The enrichment effect decreased greatly. Only 3 mL  
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26 197 plastic dropper meet the above experiment requirements well. Hence, 3 mL plastic  
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28 198 dropper was selected as the extraction device. The influence of sample volume on the  
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30 199 extraction efficiency was studied in the range of 4.5-6.5 mL. As shown in fig.2s (see  
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32 200 SI). The results demonstrated that the analytical signal increases with the sample  
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34 201 volume in the range of 4.5-6.5 mL. This may be because increasing sample volume  
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36 202 leads to the increase of analytical signals. However, the 3 mL plastic dropper can not  
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38 203 hold more water. Consequently, 6.5 mL water sample was selected as the optimal  
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40 204 volume of sample.

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44 206 **3.1.2 Sorbent amount**

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48 208 To achieve a high recovery, different amounts of 1D-PANIs ranging from 1 to  
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4 209 5 mg were applied to extract OCPs in water samples (Fig.3s, see SI). The result shows  
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6 210 that the recovery increased with the increase of the 1D-PANIs mass from 1 to 3 mg,  
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9 211 and then decreased with the increase of 1D-PANI mass from 4 to 5 mg. The reason is  
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11 212 that the amount of 1D-PANIs is less than 3mg, less target analytes would be adsorbed  
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14 213 by 1D-PANIs and further eluated by the eluent (n-hexane). Hence the extraction  
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16 214 efficiency decreased. If the amount of 1D-PANIs is more than 3mg, owing to the  
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19 215 narrow neck of the dropper, 1D-PANIs would stack much higher, and the eluent  
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21 216 (n-hexane) could not permeate 1D-PANIs to eluate the target analytes adsorbed by  
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24 217 1D-PANIs well. So the extraction efficiency decreased. Thus, 3 mg 1D-PANIs was  
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26 218 selected and employed in the following experiment.

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### 220 **3.1.3 Desorption conditions**

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222 The desorption time was optimized by increasing the vibrating time from 10s to  
223 60s, and then letting it stand for 2 mins. The result indicated that 30s vibrating time  
224 and two mins were enough to elute the extracted OCPs from the 1D-PANIs sorbent.  
225 Selecting an appropriate eluent is essential for D- $\mu$ -SPE. It should have low solubility  
226 in water, lower density than water, high affinity for the analytes, and good  
227 chromatographic behavior. Five solvents with these properties were tested: n-hexane,  
228 cyclohexane, n-octane, o-xylene and p-xylene. As shown in Fig.5s (see SI), n-hexane  
229 got the best extraction efficiency, followed by cyclohexane, p-xylene, n-octane and  
230 o-xylene. It may be because the solubilities of the target analytes in n-hexane is the  
231 biggest among the selected eluents. Hence, n-hexane was selected as the optimal

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4 232 eluent in the following experiments.

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6 233 After selecting n-hexane as the eluent, its volume should be optimized. At its low  
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9 234 volume, n-hexane can not eluate analytes from the sorbent completely. While, at its  
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11 235 high volume, the concentration of analytes eluated by n-hexane decreased, which will  
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14 236 result in a reduced extraction efficiency. To optimize the volume of acetonitrile, the  
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16 237 effect of its volume on the extraction efficiency was investigated in the range of  
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19 238 55-100  $\mu\text{L}$ . As shown in Fig. 6s (see SI), extraction efficiency increased with the  
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21 239 increase of n-hexane volume from 55 to 65  $\mu\text{L}$ , and decreased with the increase of  
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24 240 n-hexane volume up to 100  $\mu\text{L}$ . On the basis of these results, 65  $\mu\text{L}$  was chosen as the  
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26 241 optimal volume of the desorption solvent.

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### 30 31 243 **3.2 Analytical performance**

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36 245 To evaluate the proposed method, experiments with regard to the linearity, limit  
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38 246 of detection (LOD) and limit of quantification (LOQ), precision were performed.

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41 247 The calibration was constructed by plotting peak area versus concentration. As  
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44 248 shown in Table 1, satisfactory correlation coefficients ( $R^2$ ) ranging from 0.9971 to  
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46 249 0.9993 are obtained for OCPs in the concentration range of 0.05-50  $\mu\text{g L}^{-1}$  for  $\alpha$ -HCH,  
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49 250  $\beta$ -HCH and  $\gamma$ -HCH, 0.1-100  $\mu\text{g L}^{-1}$  for  $\delta$ -HCH and p,p'-DDT, 0.15-100  $\mu\text{g L}^{-1}$  for  
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51 251 p,p'-DDE and o,p'-DDT, 0.2-100  $\mu\text{g L}^{-1}$  for p,p'-DDD. The LOD and LOQ are  
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54 252 considered as the analyte's minimum concentration that can be identified and  
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56 253 quantified by the method, and they are calculated at a concentration at which  
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4 254 signal-to-noise ratios are equal to 3 and 10, respectively. As shown in Table 1, the  
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6 255 LODs and LOQs for eight OCPs are found to be 0.0074-0.0468  $\mu\text{g L}^{-1}$  and  
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9 256 0.0244-0.1544  $\mu\text{g L}^{-1}$ , respectively. Precision was evaluated by measuring intra- and  
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11 257 interday relative standard deviations (RSDs). The results obtained are shown in Table 1.  
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14 258 RSDs of intra- and interday ranging from 0.6 to 6.6% and from 4.2 to 11.9% are  
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16 259 obtained, respectively.  
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### 261 **3.3 Comparison of the proposed method with other published methods**

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263 The proposed method was compared with  $\text{C}_{18}$  based SPE ( $\text{C}_{18}$ -SPE),  
264 polydimethylsiloxane/divinylbenzene based solid phase microextraction (PDMS/DVB  
265 -SPME), graphene oxide based SPE (GO-SPE), Electrodeposited polyaniline based  
266 solid-phase microextraction (ELC-PANI-SPME), dual stir bar sorptive extraction  
267 (SBSE) and directly suspended droplet microextraction (DSDME) method. And the  
268 results are shown in Table 2. It is obvious that the LODs of the proposed method are  
269 lower than that of  $\text{C}_{18}$ -SPE and DSDME. The sample volume is smaller and the  
270 extraction time is shorter than all the other mentioned methods. Furthermore, all  
271 procedures of the proposed method can be performed in a plastic dropper, without  
272 additional centrifuge and pressure blowing concentrator, so the cost is low. The high  
273 dispersity and good permeability of 1D-PANIs made the extraction and elution steps  
274 easy, so the sample preparation time is short. The extraction recovery is high because  
275 of no solvent loss. The proposed method is promising to be an alternative of D- $\mu$ -SPE

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4 276 method for extracting apolar analytes in complex sample matrices because of high  
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6 277 extraction efficiency, low cost and short sample preparation time.  
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### 11 279 **3.4 Application to real water samples**

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16 281 The proposed method was applied to determine the  $\alpha$ ,  $\beta$ ,  $\gamma$  and  $\delta$ -HCH; p,  
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18 282 p'-DDE; p, p'-DDD; p, p'-DDT; o, p'-DDT OCPs in three kinds of real water samples  
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21 283 including Danjiang Kou water, ponding water and lake water from South Lake of  
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24 284 Wuhan, China. The water samples were filtered before analysis and the subsequent  
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26 285 results are shown in Table 3. As recorded in this table,  $0.056 \mu\text{g L}^{-1}$   $\beta$ -HCH and  $0.187$   
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28 286  $\mu\text{g L}^{-1}$   $\delta$ -HCH were detected in the Danjiang Kou water,  $0.012 \mu\text{g L}^{-1}$   $\gamma$ -HCH,  $0.151$   
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30 31  $\mu\text{g L}^{-1}$   $\beta$ -HCH and  $0.128 \mu\text{g L}^{-1}$   $\delta$ -HCH were detected in the Lake water but no  
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34 288 analytes were found in ponding water. In order to evaluate the feasibility of the  
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36 289 improved method in real samples, they were spiked with three different kinds of  
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38 290 concentrations of the target analytes, including  $0.2 \mu\text{g L}^{-1}$ ,  $1.0 \mu\text{g L}^{-1}$  and  $10.0 \mu\text{g L}^{-1}$ .  
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41 291 The result demonstrated good recoveries ranging from 73.6% to 107.0%. The  
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44 292 chromatograms of the blank and spiked South Lake water using the proposed  
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46 293 D- $\mu$ -SPE method are shown in Fig. 2.  
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### 51 295 **4. Conclusions**

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56 297 In summary, 1D-PANIs was synthesized by oxidative polymerization method.  
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4 298 The interconnected network-like structure endowed the nano-material with good  
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6 299 adsorbability and permeability. A plastic dropper was employed as both the extraction  
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9 300 and preconcentration device. The complex equipment, such as a centrifuge or a  
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11 301 pressure blowing concentrator was not needed. The study demonstrated that the  
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13 302 proposed 1D-PANI-based D- $\mu$ -SPE was simple, rapid, and effective method for  
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16 303 sample preparation. Furthermore, the work would considerably expand the application  
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19 304 of 1D-CPs in sample preparation,  
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24 306 **References**

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14 346 **Acknowledgement**

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18 348 21177048) is greatly appreciated.

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24 350 **Figure captions**

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29 352 **Fig.1.** Schematic procedure of the proposed D- $\mu$ -SPE: (a) 6.5 mL water sample were

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31 353 added in a 3 mL plastic dropper. (b) 3 mg 1D-PANIs was added into the plastic

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33 354 dropper and dispersed ultrasonically for 10 seconds (c) 65  $\mu$ L n-hexane was

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35 355 injected into 6.5 mL water sample with a microsyringe and the plastic dropper

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37 356 was vibrated up and down 30s (d) 1D-PANIs was agglomerated and the eluent

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39 357 (n-hexane) floated up on it by letting the plastic dropper stand for a several

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41 358 minutes. (e) The eluent was ascended to the tip of the plastic dropper by

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43 359 pressing the bulb of the plastic dropper and was withdrawn by a microsyringe

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45 360 for analysis by GC.

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49 362 **Fig.2.** Chromatograms of eight OCPs from blank (a) and spiked South Lake water (b)

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51 363 extracted by the proposed D- $\mu$ -SPE method. The sample solution was spiked at

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4 364 0.5  $\mu\text{g L}^{-1}$  for eight OCPs. Peaks: (1)  $\alpha$ -HCH; (2)  $\gamma$ -HCH; (3)  $\beta$ -HCH; (4)

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6 365  $\delta$ -HCH; (5) p,p'-DDE; (6) p,p'-DDD; (7) o,p'-DDT; (8) p,p'-DDT

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11 367 **Table captions**

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14 368 **Table 1** Analytical performance data for eight OCPs by the proposed D- $\mu$ -SPE

15  
16 369 method

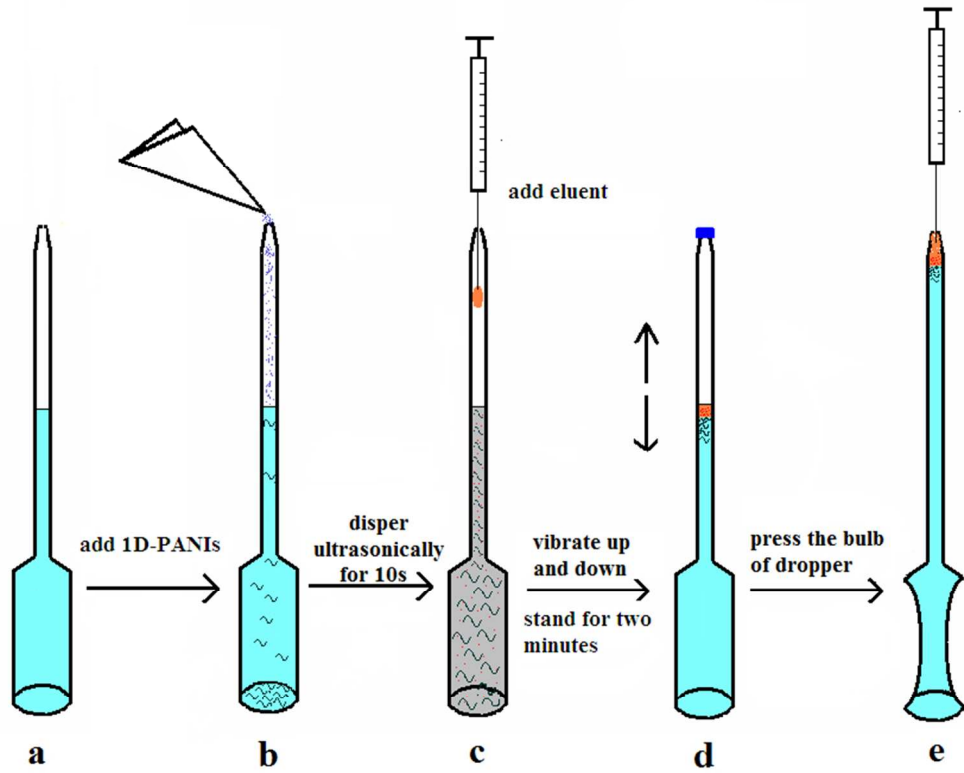
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19 370 **Table 2** Comparison of the proposed D- $\mu$ -SPE method with other published methods

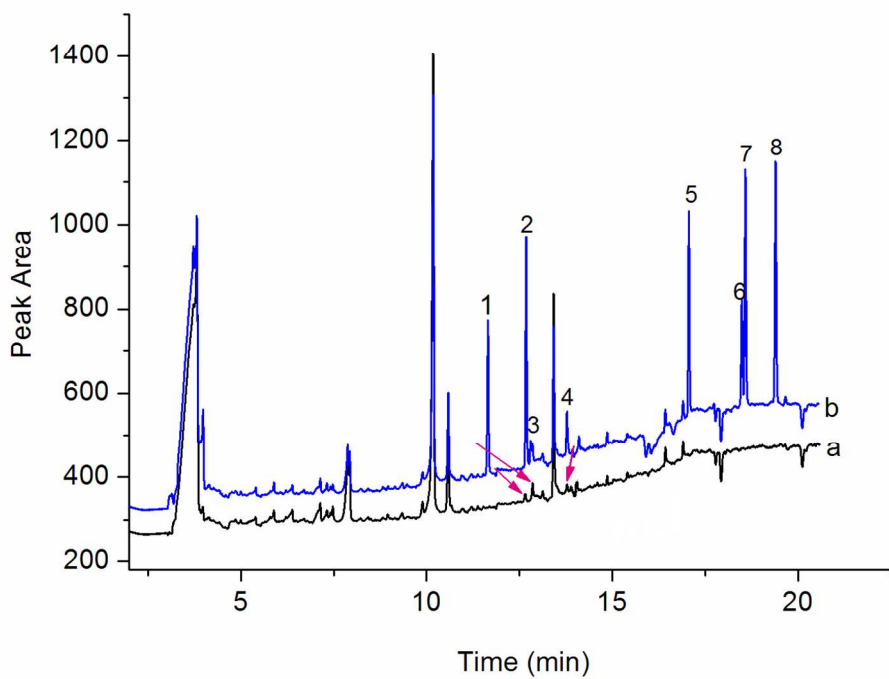
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21 371 **Table 3** Relative recoveries of eight OCPs from three spiked environmental water

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**Table 1 Analytical performance data for eight OCPs by the proposed D- $\mu$ -SPE method**

Analyte	Linear range ( $\mu\text{g L}^{-1}$ )	$R^2$	LOD <sup>a</sup> ( $\mu\text{g L}^{-1}$ )	LOQ <sup>b</sup> ( $\mu\text{g L}^{-1}$ )	RSD (%)	
					Intra-day	Inter-day
$\alpha$ -HCH	0.05-50	0.9992	0.0126	0.0416	6.6	11.9
$\gamma$ -HCH	0.05-50	0.9981	0.0074	0.0244	1.5	4.2
$\beta$ -HCH	0.05-50	0.9981	0.0121	0.0399	0.6	5.6
$\delta$ -HCH	0.1-100	0.9971	0.0174	0.0574	5.2	6.2
p,p'-DDE	0.15-100	0.9988	0.0315	0.1040	6.5	8.5
p,p'-DDD	0.2-100	0.9996	0.0468	0.1544	3.8	5.3
o,p'-DDT	0.15-100	0.9989	0.0306	0.1009	2.0	8.3
p,p'-DDT	0.1-100	0.9993	0.0260	0.0858	1.3	6.6

<sup>a</sup>Limit of detection, S/N = 3; <sup>b</sup>Limit of quantification, S/N = 10.

**Table 2 Comparison of the proposed D- $\mu$ -SPE method with other published methods**

Analytical method	Sample volume (mL)	Extraction time (min)	Correlation coefficient	LOD ( $\mu\text{g L}^{-1}$ )	Detector	Reference
C18-SPE	2000	>60	>0.999	0.2-40	GC-ECD	20
PDMS/DVB	18	>50	0.996-0.999	0.0038-0.011	GC-MS	6
GO-SPE	200	>96	0.982-0.999	0.0015-0.0094	GC-MS	21
ELC-PANI-SPME	10	>46	0.979-0.999	0.0001-0.0016	GC-ECD	13
SBSE-TD-LTM	20	>60	0.993-0.999	0.0014-0.006	GC-MS	22
DSDME	5	15	0.993-0.999	0.04-0.1	GC-ECD	23
D- $\mu$ -SPE	6.5	3	0.997-0.999	0.0074-0.0468	GC-ECD	This method

Table 3 Relative recoveries of eight OCPs from three spiked environmental water sample

	Danjiangkou water (%±RSD, n=3)				Ponding water (%±RSD, n=3)				Lake water (%±RSD, n=3)			
	0	0.2 µg <sup>-1</sup>	1.0 µg <sup>-1</sup>	10 µg <sup>-1</sup>	0	0.2 µg <sup>-1</sup>	1.0 µg <sup>-1</sup>	10 µg <sup>-1</sup>	0	0.2 µg <sup>-1</sup>	1.0 µg <sup>-1</sup>	10 µg <sup>-1</sup>
α-HCH	ND <sup>a</sup>	90.0±4.8	94.0±3.1	91.6±3.4	ND <sup>a</sup>	85.8±5.2	99.6±6.3	94.3±4.1	ND <sup>a</sup>	84.7±8.9	81.4±5.2	99.5±4.3
γ-HCH	ND <sup>a</sup>	95.1±5.6	98.3±3.3	89.8±2.9	ND <sup>a</sup>	88.8±4.5	107.0±4.9	86.7±3.5	0.012	83.9±7.4	89.7±5.6	93.9±4.9
β-HCH	0.056	76.6±5.1	83.7±7.6	103.3±4.8	ND <sup>a</sup>	84.1±6.0	82.9±3.7	81.4±4.9	0.151	86.9±5.3	86.6±4.1	87.4±3.9
§-HCH	0.187	73.6±2.0	106.0±3.8	92.7±3.3	ND <sup>a</sup>	90.4±5.2	81.1±4.1	92.4±4.6	0.128	79.5±5.7	94.6±2.1	88.6±4.8
p,p'-DDE	ND <sup>a</sup>	104.0±5.3	96.0±3.2	86.5±4.6	ND <sup>a</sup>	99.6±4.9	94.1±3.8	81.3±6.6	ND <sup>a</sup>	103.4±6.0	90.1±7.2	93.3±5.1
p,p'-DDD	ND <sup>a</sup>	100.9±5.1	91.8±5.8	86.2±6.6	ND <sup>a</sup>	85.2±7.8	80.7±5.8	99.1±8.1	ND <sup>a</sup>	87.6±8.3	95.4±7.2	90.4±5.9
o,p'-DDT	ND <sup>a</sup>	98.6±5.6	86.6±7.2	83.3±4.3	ND <sup>a</sup>	105.9±10.2	93.0±8.9	79.9±7.5	ND <sup>a</sup>	106.1±11.3	94.2±10.5	90.3±9.9
p,p'-DDT	ND <sup>a</sup>	90.2±6.1	88.7±7.7	85.3±8.2	ND <sup>a</sup>	91.5±5.3	80.1±8.0	82.4±7.6	ND <sup>a</sup>	98.9±8.4	83.8±7.3	91.2±9.6

<sup>a</sup>ND: not detected;1  
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