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1 **Novel Graphene Oxide Sponge synthesized by**
2 **Freeze-Drying Process for the Removal of**
3 **2,4,6-Trichlorophenol**

4 Jiali WANG^{1,2}, Xueli GAO^{1,2,*}, Yuhong Wang³, Congjie GAO^{1,2}

5 1. Key Laboratory of Marine Chemistry Theory and Technology, Ministry of
6 Education, Ocean University of China, Qingdao 266100, China

7 2. College of Chemistry and Chemical Engineering, Ocean University of China,
8 Qingdao 266100, China

9 3. National Center of Ocean Standards and Metrology, Tianjin 300112, China

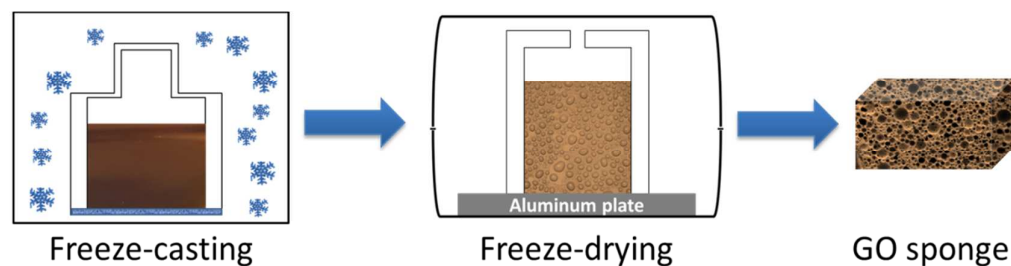
10 *Corresponding author; Tel/Fax.: +86-532-66782017;

11 Email address:wangjialiouc@gmail.com; gxl_ouc@126.com

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1 **Abstract:**

2 Graphene oxide (GO) spongy materials as environmental pollutant scavengers
3 have drawn great attentions owing to ultralarge surface area, unique spongy structure
4 and hydrogen-bonding interactions. Herein, novel GO sponge was synthesized by
5 improved Hummer's method followed freeze-drying process and its adsorption
6 capacity of 2,4,6-trichlorophenol (TCP) was investigated. The structural features of
7 GO sheets and GO sponge have been characterized by Fourier transform infrared
8 spectroscopy (FTIR), X-ray diffraction (XRD), Raman spectroscopy, X-ray
9 photoelectron spectroscopy (XPS), atomic force microscope (AFM) and scanning
10 electron microscopy (SEM). Specific surface area assessment and pore distribution
11 measurements were analyzed by Micromeritics ASAP 2020. Adsorption mechanism
12 and kinetic study of TCP on GO sheets and GO sponge were researched using a batch
13 equilibration method. The results suggest that GO sponge presented higher adsorption
14 capacity than GO sheets due to large specific surface areas and TCP had optimum
15 adsorption capacity on both GO sheets and GO sponge at pH 2.0~6.0. Adsorption
16 isotherms and kinetics curves of TCP on GO sheets and GO sponge were nonlinear,
17 indicating the homogeneous monolayer chemical adsorption.



19 **Keywords:** graphene oxide, freeze-drying, 2,4,6-trichlorophenol, adsorption

1 **1. Introduction**

2 Hazardous persistent aromatic materials produced by pharmaceutical,
3 petrochemical, dyestuff, pesticide, and other industries have been a severe
4 environmental concern. 2,4,6-Trichlorophenol (TCP) is a very common pollutant in
5 the aquatic environment due to its toxicity, stability, and bioaccumulation¹. Discharge
6 of TCP-contaminated wastewater into aquatic environment without adequate
7 treatment can pose adverse effects on human nervous system and cause many health
8 disorders². Numerous studies focused on the efficient removal of TCP from aqueous
9 solutions such as by photocatalysis³, adsorption⁴, and electrolysis⁵. Due to the
10 structural stability and persistence of TCP in the environment, adsorption has been
11 found to be superior to other techniques as a result of its low-cost, high efficiency and
12 easy operation.

13 Graphene (GN) is a two-dimensional structure consisting of sp^2 hybridized
14 carbons with only one atomic thickness, and has attracted significant attentions since
15 its discovery⁶. Graphene oxide (GO) is a highly oxidized form of chemically modified
16 GN that consists of single-atom-thick layer of GN sheets with carboxylic acid,
17 epoxide and hydroxyl groups in the plane⁷. Due to its atomic-level thickness, large
18 theoretical specific surface area, remarkable electronic and chemical properties,
19 potential environmental applications of GO as superior adsorbent have been
20 recognized for the removal of organic contaminants and metal ions in water. Gao et
21 al.⁸ revealed the high adsorption of tetracycline antibiotics on GO. Lin et al.⁹ found
22 that arginine-capped iron oxide/rGO nanocomposite is effective adsorbents for acid

1 dye removal. In addition, GO also have high adsorption affinity for heavy metals,
2 where the amount of active surface sites on GO is an important factor influencing the
3 adsorption of heavy metal ions. Previous works on Cd(II) and Co(II) adsorption onto
4 few-layered GO nanosheets¹⁰, and on Hg(II) adsorption to polypyrrole-reduced GO
5 composites¹¹ showed that GO composites have a strong adsorption affinity for metal
6 ions.

7 Spongy GN and GO materials are applied as environmental pollutant adsorbents
8 by utilizing the characteristics of ultralarge surface area, electrostatic interactions or
9 hydrogen bonds of oxygen-containing functional groups and strong π - π interaction on
10 the surface¹²⁻¹⁴. Spongy graphene (SG) has been made by reducing graphene oxide
11 platelets in suspension followed by shaping via moulding and heating and shows
12 highly efficient absorption of not only petroleum products and fats, but also toxic
13 solvents such as toluene and chloroform¹⁵. Liu et al.¹⁶ generated a three dimensional
14 (3D) graphene oxide sponge from a GO suspension through a simple centrifugal
15 vacuum evaporation method, and used them to remove both the methylene blue (MB)
16 and methyl violet (MV) dyes.

17 Nowadays, there are many forming processes following colloidal suspension to
18 prepare spongy absorbing materials, such as slip casting, tape casting, screen printing,
19 centrifugal vacuum evaporation¹⁷⁻¹⁹. The common issues for these techniques are the
20 high fugitive organic contents and the problematic drying process because of the
21 presence of capillary force; the high organic content poses challenges in the removal
22 process, increases processing cost, and produces environmental hazards; the capillary

1 force generates drying stress and subsequently warping and cracking. However,
2 freeze-drying process is a simple technique to produce porous complex-shaped
3 inorganic composites or polymeric parts²⁰. It provides materials with a unique porous
4 structure, where the porosity is almost a direct replica of the frozen solvent crystals.
5 Proper control of the freezing conditions yields materials with elongated and
6 continuous porosity along the solidification direction. This unique structure endow
7 materials with excellent compressive strength, open porosity, high pore connectivity,
8 specific surface area and high adsorption ability²¹. Sun et al.²² fabricated the
9 hydrophobic CNT-GN aerogels by freeze-drying process and chemical reduction and
10 the aerogels possess ultrahigh oil-absorption capacity. Furthermore, freeze-drying
11 prevents defect formation by eliminating capillary force during drying and saves
12 tremendous effort in binder removal and also has the advantage of little cost and facile
13 utilization of non-toxic dispersing medium, such as water²³. As a result, a
14 shape-mouldable and nanoporous GO sponge is promisingly designed through
15 freeze-drying as a versatile and effective sorbent material.

16 In this work, we have demonstrated that the application of GO sponge materials as
17 environmental pollutant scavengers by utilizing unique porous structure with
18 ultralarge surface area, electrostatic interactions or hydrogen bonds of
19 oxygen-containing functional groups and π - π interaction on the surface. We prepared
20 GO sheets using improved Hummer's method and generated GO sponge from GO
21 suspension through freeze-drying process, and explored the potential application of
22 GO sponge to remove TCP. The structural features of GO sheets and GO sponge have

1 been characterized by Fourier transform infrared spectroscopy (FTIR), X-ray
2 diffraction (XRD), Raman spectroscopy, X-ray photoelectron spectroscopy (XPS),
3 Atomic Force Microscope (AFM) and scanning electron microscopy (SEM). Kinetic
4 study and adsorption mechanism on TCP adsorption were also investigated.

5

6 **2. Materials and methods**

7 **2.1 Materials**

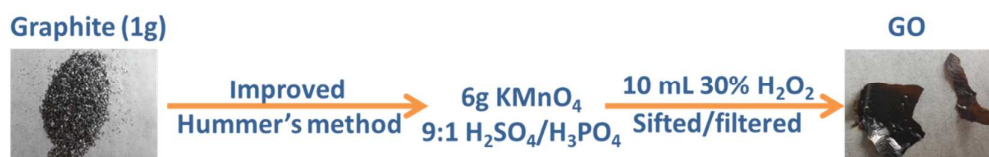
8 Graphite was purchased from Bay Carbon. H_2SO_4 (95 wt. %) and KMnO_4 were
9 all analytical grades and obtained from Beijing Chemical Works. H_3PO_4 (85 wt. %)
10 was analytical grade and supplied by Nanjing Chemical Reagent Co., Ltd. H_2O_2 (30
11 wt. %) and HCl (36~38 wt. %) was analytical grade and obtained from Sinopharm
12 Chemical Reagent Co., Ltd. TCP (97.4 wt. %) was purchased from Shanghai Zhenpin
13 Chemical Co., Ltd. NaOH and HNO_3 (98 wt. %) were analytical grade and obtained
14 from Sinopharm Chemical Reagent Co., Ltd. Methanol was HPLC grade and obtained
15 from Merck Chemicals (Shanghai) Co., Ltd.

16

17 **2.2 Synthesis of GO**

18 GO was synthesized from graphite powder according to the improved Hummer's
19 method²⁴ which is outlined in **Scheme 1**. The mixture of concentrated $\text{H}_2\text{SO}_4/\text{H}_3\text{PO}_4$
20 (90:10 mL) was added to a mixture of graphite flakes (1.0 g) and KMnO_4 (6.0 g). The
21 reaction was then heated to 50 °C and stirred for 12 h. The reaction was cooled to
22 room temperature and poured onto ice (~400 mL) with 30% H_2O_2 (10 mL). After

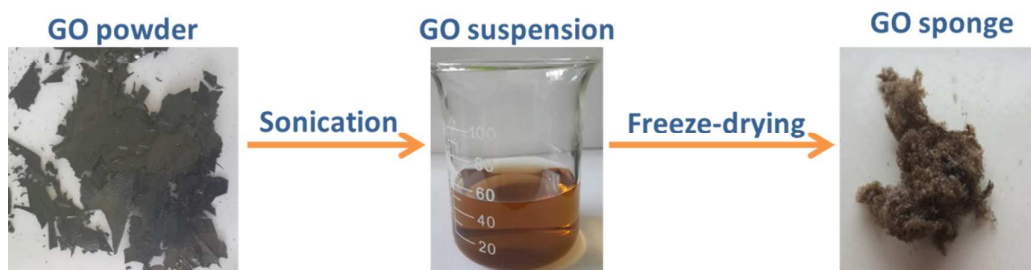
1 cooling, the mixture were centrifuged (7000 rpm for 4 h), and the supernatant was
2 decanted away. The remaining solid material was then washed in succession with 200
3 mL of water, 200 mL of 30% HCl solution. The products obtained by centrifugation
4 were vacuum-dried overnight at room temperature.



6 **Scheme 1** Synthesis of GO sheets

7 **2.3 Preparation of GO sponge via freeze dry method**

8 GO was re-dissolved in deionized water and ultrasonicated for 1 h at 200 W. The
9 dispersion was centrifuged at 7000 rpm, and GO dispersion was obtained by
10 discarding the precipitate. GO dispersion was placed on a metal die, which was then
11 transferred to a cold source with constant temperature. Freezing lasted for 2h until the
12 suspension was fully solidified. Temperatures above -50°C can be achieved directly
13 by applying cold traps to the FD-1A-80 freeze dryer (Boyikang corp., China). After
14 GO were completely frozen, the samples were placed in a freeze dryer for 48 h,
15 generating black GO sponges (**Scheme 2**).



17 **Scheme 2** Preparation of GO sponges

18 **2.4 Characterization of GO sheets and GO sponge**

1 The Fourier transform infrared (FTIR) spectroscopy was performed using Tensor
2 27 spectrometer (Bruker, Germany) using the KBr pellet method at transmission
3 reflectance (TR) mode. The FTIR spectra were scanned in the wavenumbers range
4 from 650 to 4000 cm^{-1} during 64 scans, with 2 cm^{-1} resolution. X-ray powder
5 diffraction (XRD) patterns were recorded using D8 ADVANCE diffractometer
6 (Bruker, Germany), equipped with a Cu $K\alpha$ radiation source ($\lambda = 1.5418 \text{ \AA}$). Raman
7 scattering was performed on NEXUS 670 Microscopy (Thermo Nicolet, USA) with
8 an excitation laser of 532 nm. The X-ray photoelectron spectroscopy (XPS)
9 measurement was performed on ESCALAB 250 spectrophotometer (Thermo Fisher
10 SCIENTIFIC, USA) with an achromatic Mg/Al X-ray source at 450W. Atomic force
11 microscopy (AFM) images were recorded using Multimode-V microscope (Veeco,
12 USA) in contact mode. AFM samples were prepared by drop casting the GO
13 suspension in water onto freshly cleaved mica surfaces and dried under room
14 temperature. The morphologies of GO sheets and GO sponge were observed using
15 S-4800 scanning electron microscopy (SEM) (Hitachi, Japan) at a 15 kV accelerating
16 voltage. The samples were coated with gold under an argon purge before the
17 characterization. Specific surface area assessment and pore distribution measurements
18 were carried out by N_2 adsorption-desorption analysis at 77 K using ASAP 2020
19 (Micromeritics, USA).

20 **2.5 Adsorption experiments of TCP on GO sheets and GO sponge**

21 **2.5.1 Effect of initial solution pH**

22 Batch adsorption experiments were carried out by using the GO sheets and GO

1 sponge as the adsorbents. All batch adsorption experiments were performed in
2 triplicates on a SHA-C shaker (Changzhou, China) with a shaking speed of 150 rpm
3 until the system reached equilibrium. 0.015 g of GO sheets was firstly dispersed in 15
4 mL of water and ultrasonicated for 15 min at 200 W for full suspension of GO. The
5 pH effect experiments were respectively conducted by mixing 0.015 g of GO sponge
6 or 15 mL of GO suspension with TCP into 30 mL mixture solution. The initial
7 concentrations of TCP is $5.0 \text{ mg}\cdot\text{L}^{-1}$. After equilibrium was reached, the suspension
8 was centrifuged at 5000 rpm for 15 min to separate liquid from solid phases, and the
9 concentrations of solutes in the supernatant phase were determined by HPLC
10 (LC-2010HT, Shimadzu, Japan). Controls were also prepared identically in triplicate
11 but contained no adsorbents, which were simultaneously run to assess loss of solutes.
12 Results showed that no significant loss of solutes was observed, indicating that
13 microbial degradation, volatilization, or adsorption to the glass walls were negligible
14 during the adsorption experiments. The adsorbed mass was calculated from the
15 differences between the initial and final equilibrium concentrations, according to the
16 following equation:

$$17 \quad Q = \frac{(C_0 - C_e)V}{m} \quad (1)$$

18 where C_0 is the initial TCP concentration ($\text{mg}\cdot\text{L}^{-1}$), C_e is the TCP equilibrium
19 concentration ($\text{mg}\cdot\text{L}^{-1}$), V is the volume of TCP solution (L), and m is the mass of GO
20 sheets and GO sponge (g).

21 **2.5.2 Adsorption equilibrium study**

22 Adsorption of TCP (1.0, 2.0, 5.0, 10.0, 15.0 and 20.0 $\text{mg}\cdot\text{L}^{-1}$) were carried out

1 using a batch adsorption approach as mentioned below. The suspended solution pH
2 was adjusted to 5.0 ± 0.1 by addition of 0.1 M HNO₃ or 0.1 M NaOH. After
3 equilibrium was reached, the initial and final TCP concentrations was detected by
4 HPLC and the uptake was calculated based on Eq. (1) as mentioned above.

5 **2.5.3 Adsorption kinetics study**

6 Adsorption kinetics study on GO sponge or GO suspension was carried out with
7 the initial TCP concentration of $10.0 \text{ mg}\cdot\text{L}^{-1}$ at 298 K and pH 5.0 as mentioned above,
8 respectively. About 1.0 mL of the solution was then taken out at desired time intervals
9 to analyze the current TCP concentration. Meanwhile, the same volume of pure water
10 was added into the bulk solution to keep the volume constant. The uptake at time t_i ,
11 $q(t_i)$ ($\text{mg}\cdot\text{g}^{-1}$) was calculated using the following equation:

$$12 \quad q(t_i) = \frac{(C_0 - C_{t_i})V_0 - \sum_2^{i-1} C_{t_{i-1}} V_s}{m} \quad (2)$$

13 where C_0 is the initial concentration of TCP ($\text{mg}\cdot\text{L}^{-1}$) and C_{t_i} is the TCP concentration
14 at time t_i ($\text{mg}\cdot\text{L}^{-1}$). V_0 is the volume of the mixed solution (L) and V_s is the volume of
15 the sample solution taken out each time for TCP concentration analysis (L). In this
16 equation, V_s is equal to 1.0 mL. Finally, m represents the mass of the adsorbent (g).

17

18 **3. Results and discussion**

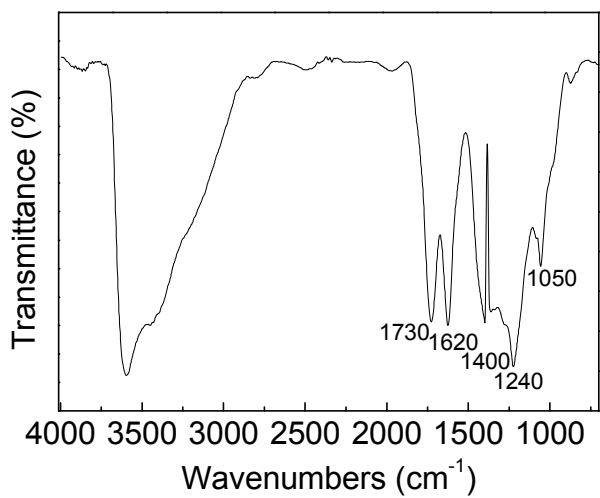
19 **3.1 Characterization of GO sheets and GO sponge**

20 The FTIR spectra of GO was shown in **Fig. 1**. The peak at 1620 cm^{-1} was
21 corresponding to C=C stretching vibration of the sp^2 carbon skeletal network.
22 Oxygen-containing functional groups such as COOH (1730 cm^{-1}), C–OH (1400, 1240

1 cm^{-1}) and C–O–C (1050 cm^{-1}) were clearly visible. XRD pattern of GO was recorded
2 in **Fig. 2**. The feature diffraction peak of GO at 11.107° was observed as the AB
3 stacking order with layer-to-layer distance (d-spacing) of 7.959 nm. Compared with
4 graphite (0.341 nm , $2\theta=26.08^\circ$)⁷, the layer-to-layer distance of GO was larger because
5 of the functional groups (such as epoxy and hydroxyl groups) on basal plane of GO
6 sheet and the intercalated water molecules between layers. The Raman spectra of GO
7 was presented in **Fig. 3**. The D-band at 1350 cm^{-1} is related to the order/disorder
8 degree from a breathing κ -point phonon of A_{1g} symmetry and the G-band at 1595 cm^{-1} ,
9 which is an indicator of the stacking structure, is assigned to the E_{2g} phonon of sp^2
10 hybridized carbon atoms²⁵. The general Raman spectrum of graphite is demonstrated
11 to have a strong G peak at 1570 cm^{-1} ²⁶. Upon oxidation of graphite, the G-band was
12 shifted toward longer wavenumber due to the formation of GO with oxygenated
13 functional groups on its basal plane and at the edges. Besides, the I_D/I_G ratio of GO
14 was 0.939, which indicated sp^2 hybridized carbons were converted to sp^3 hybridized
15 carbons due to generation of -OH, -COOH and epoxide groups during oxidation^{27,28}.
16 **Fig. 4** showed the XPS spectra of GO. The C1s spectrum could be deconvoluted into
17 three peak components with binding energies at 284.6, 286.0 and 288.5 eV, attributed
18 to C-C, C-O and C=O species, respectively²⁹. The XPS data in **Fig. 4(A)** showed that
19 about 48.81% of carbon was not oxidized, 41.66% had C-O bond (representing
20 hydroxyl and epoxide groups), 9.54% had COOH bond, and the O/C ratio was 0.61.
21 The AFM image and its corresponding height profile were presented in **Fig. 5**. GO
22 shows a height of around 0.737 nm, suggesting a single-layer nanosheet. The

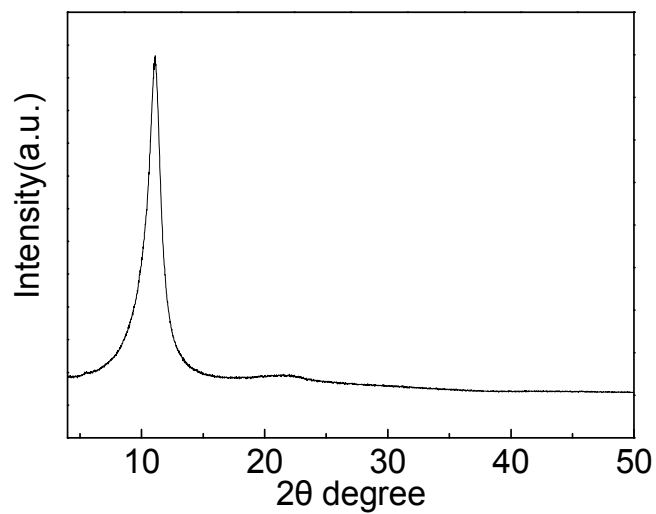
1 microscopic topographies of GO sheets and GO sponge were analyzed using SEM
2 and shown in **Fig. 6**. Changes were obvious in the SEM images of GO sheets and GO
3 sponge. The GO sheets presented the sheet-like structure (**Fig. 6A**), while GO sponge
4 presented spongy and foam-like structure owing to freeze-drying process (**Fig. 6B**).
5 Thus, the results of FTIR, XRD, Raman spectroscopy, XPS, AFM measurements
6 strongly prove the successful preparation of GO and SEM demonstrates the unique
7 structure of GO sponge compared with GO sheets.

8 The porous property of GO sheets and GO sponge was investigated by nitrogen
9 adsorption-desorption tests in **Fig. 7**. The adsorption/desorption isotherms both
10 showed a characteristic H4 hysteresis loop, indicating the presence of mesopores. The
11 surface area of GO sheets was $60 \text{ m}^2 \cdot \text{g}^{-1}$ by fitting the isotherms to the
12 Brunauer-Emmett-Teller (BET) model, which were comparable to some reported
13 works³⁰, while that of GO sponge was $189 \text{ m}^2 \cdot \text{g}^{-1}$. Therefore, GO sponge presented
14 the remarkable higher surface area than GO sheets. The pore size distribution curves
15 determined by the Barrett–Joyner–Halenda (BJH) method suggested that much of the
16 pore volume of GO sheets and GO sponge lay in the pores with a diameter of 3.3–100
17 nm. The pore size distribution curve of GO sheets showed one sharp peaks at 3.7 nm,
18 whereas that of GO sponge displayed a sharp peak at 3.8 nm and a broad peak at 16.4
19 nm. In addition, the pore volumes of GO sheets and GO sponge were 0.392 and 0.683
20 $\text{cm}^3 \cdot \text{g}^{-1}$, respectively. Overall, the pore size analyses are consistent with the SEM
21 measurements, and all the results implied that GO sponge possessed ultralarge surface
22 area and higher porosity.



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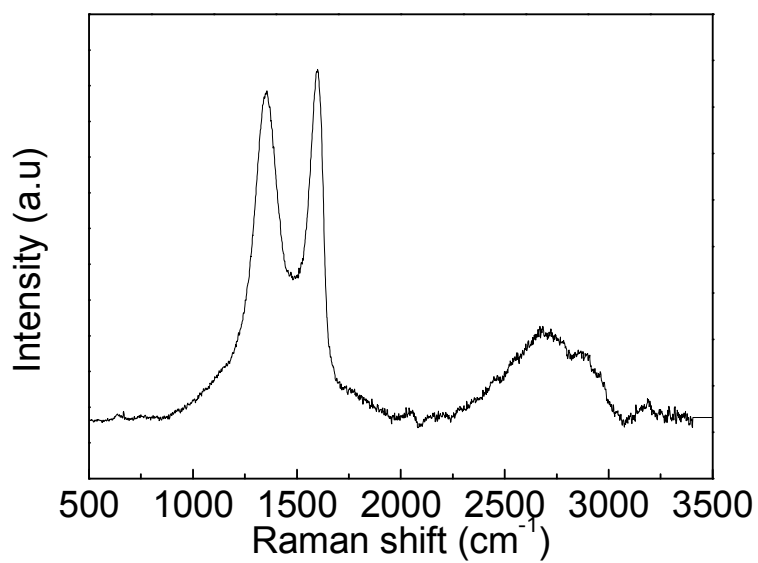
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Fig. 1 FTIR spectra of GO

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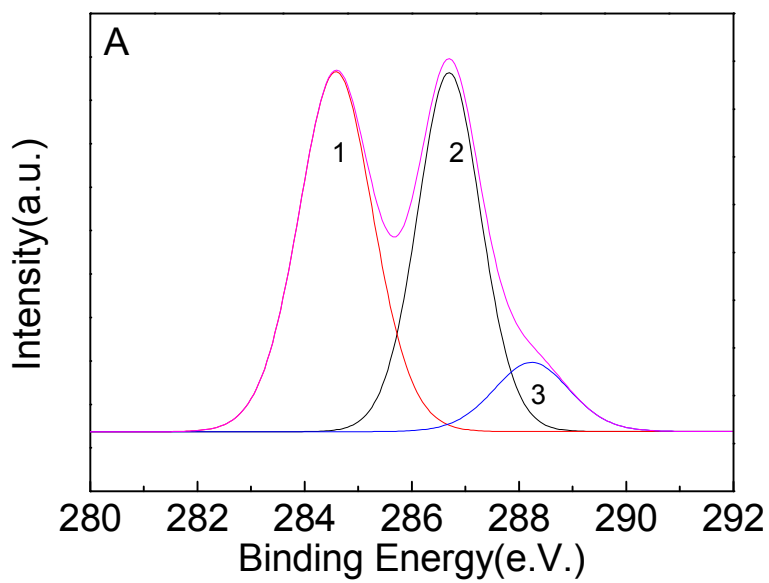
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Fig. 2 XRD spectra of GO

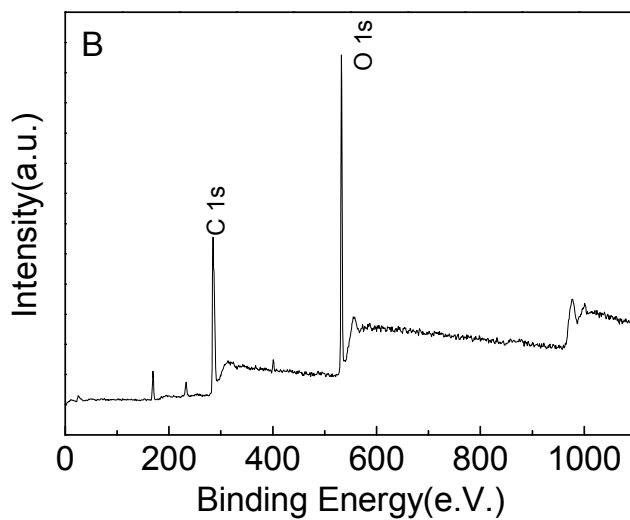


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Fig. 3 Raman spectra of GO



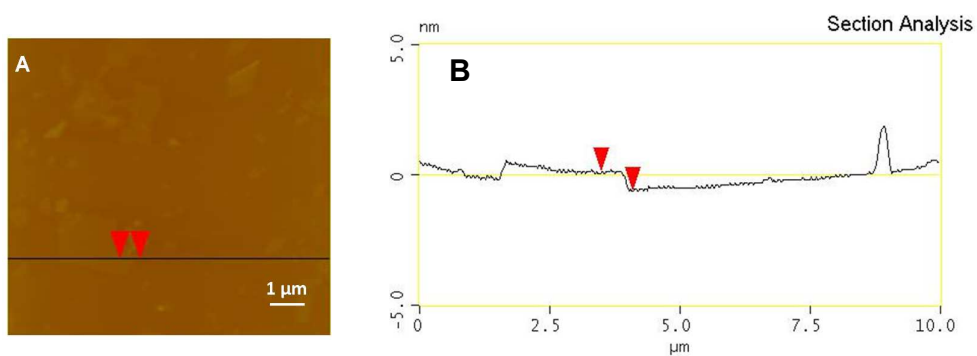
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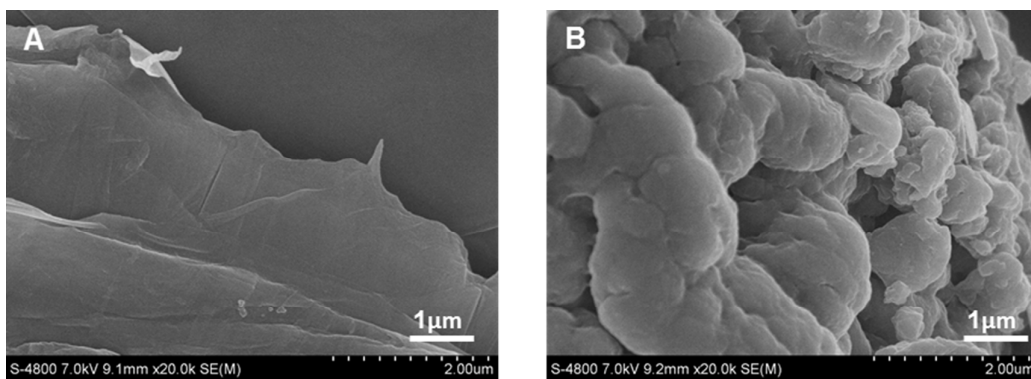
Fig. 4 C1s XPS spectrum (A) and full survey (B) of GO



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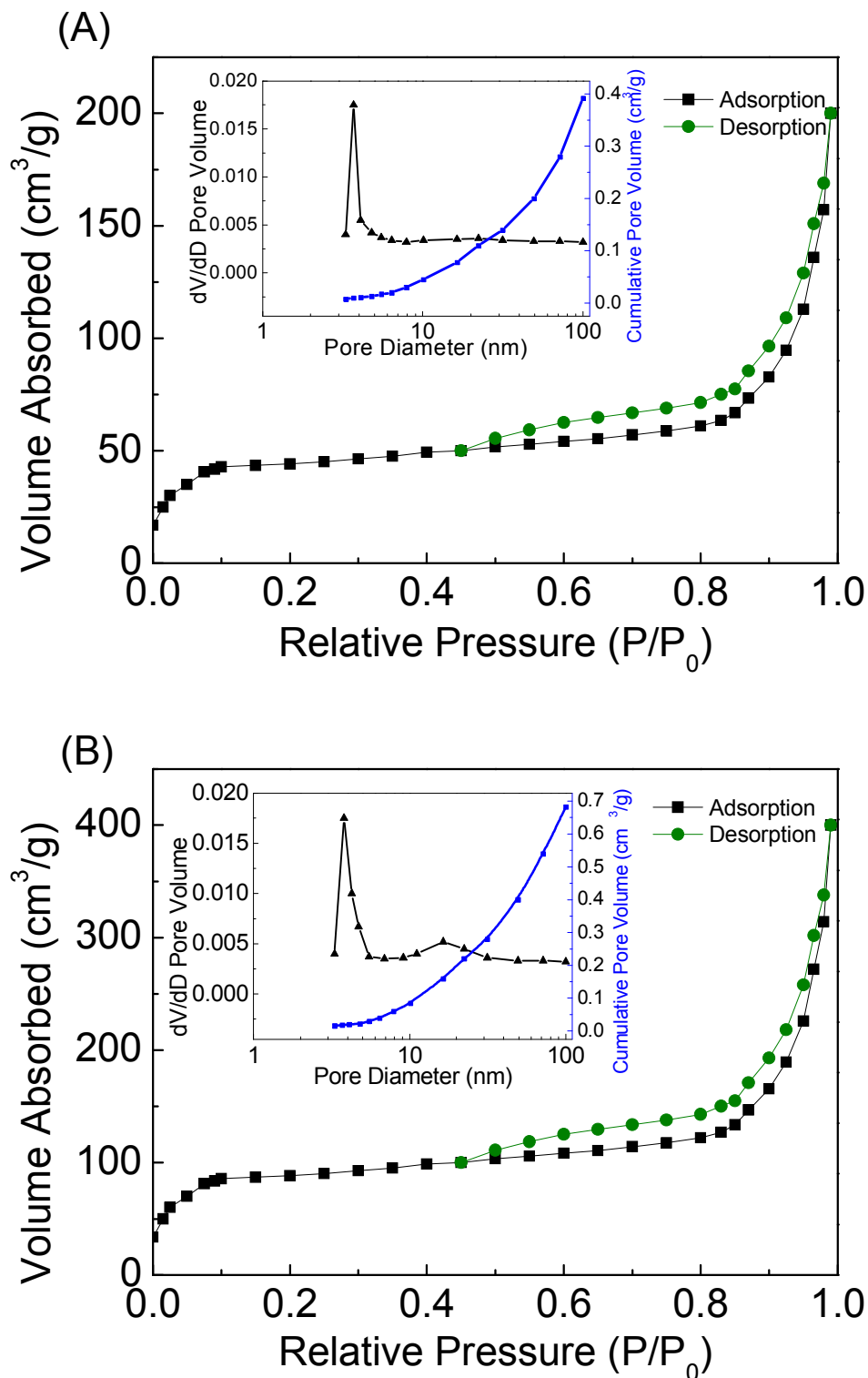
Fig.5 AFM image (A) and height profile (B) of a single layer of GO



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Fig. 6 SEM images of GO sheets (A) and GO sponge (B)



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Fig. 7 N_2 adsorption-desorption isotherms and pore size distribution curves (inset) of

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GO sheets (A) and GO sponge (B)

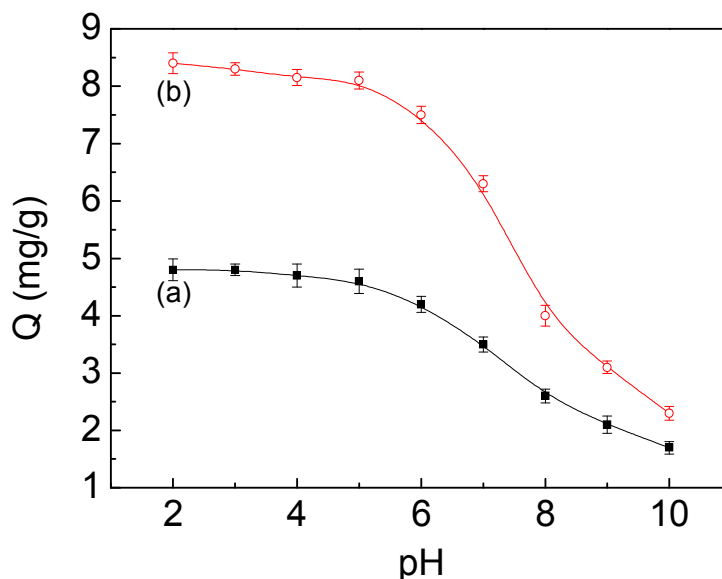
1 3.2 Adsorption study

2 3.2.1 Effect of pH on adsorption

3 The pH of aqueous solution was an important parameter that determined the
4 adsorption capacity and the results were illustrated in **Fig. 8**. The adsorption of TCP
5 on both GO sheets and GO sponge decreased slightly over pH 2.0–6.0, decreased
6 rapidly at pH>6.0 and then decreased slightly, consistent to the previous study on
7 adsorption of TCP³¹. Considering pKa value of TCP (6.15), it appears that the greater
8 adsorption at low pH may be due to higher content of neutral TCP. An increase in
9 solution pH increased the fraction of negatively charged TCP, rendering the decrease
10 of TCP adsorption. Besides, the introduction of many oxygen-containing groups
11 endow GO with negative charges. Some of the anionic groups would be deprotonized
12 at higher pH resulting in enhanced negative charge. Thus, electrostatic repulsion of
13 anionic TCP with negatively charged surfaces of GO makes adsorption unfavorable.
14 Therefore, the optimum pH range for TCP adsorption onto GO sheets and GO sponge
15 adsorbents was 2.0-6.0.

16 GO sheets and GO sponge both have good adsorption capacity for TCP because
17 the oxygen-containing functional groups of GO are facile to bind TCP due to
18 electrostatic interactions or hydrogen bonds, and the aromatic matrix of GO tends to
19 TCP by π - π stacking interactions³². However, the adsorption of TCP on GO sponge
20 was almost twice higher than that of GO sheets, which may be resultant from its
21 spongy structure. GO sponge has ultralarge surface area and high porosity due to
22 porous structure, which has been already observed directly under SEM as mentioned

1 above. Therefore, GO sponge could be more facile to form parallel π - π stacking
 2 interactions with TCP in water and provides more inner interplanar sites for
 3 electrostatic or hydrogen-bonding interactions. The large adsorption area of GO
 4 sponge has thus effectively improved the TCP adsorption.



5
 6 **Fig. 8** Effect of pH on the adsorption of TCP on GO sheets (a) and GO sponge (b)

7 3.2.1 Adsorption isotherms

8 The adsorption isotherms of the GO sheets and GO sponge for TCP were
 9 presented in **Fig. 9** and the isotherms fitting parameters were listed in **Table 1**. The
 10 equilibrium data were fitted by Langmuir and Freundlich and the equations are given
 11 below:

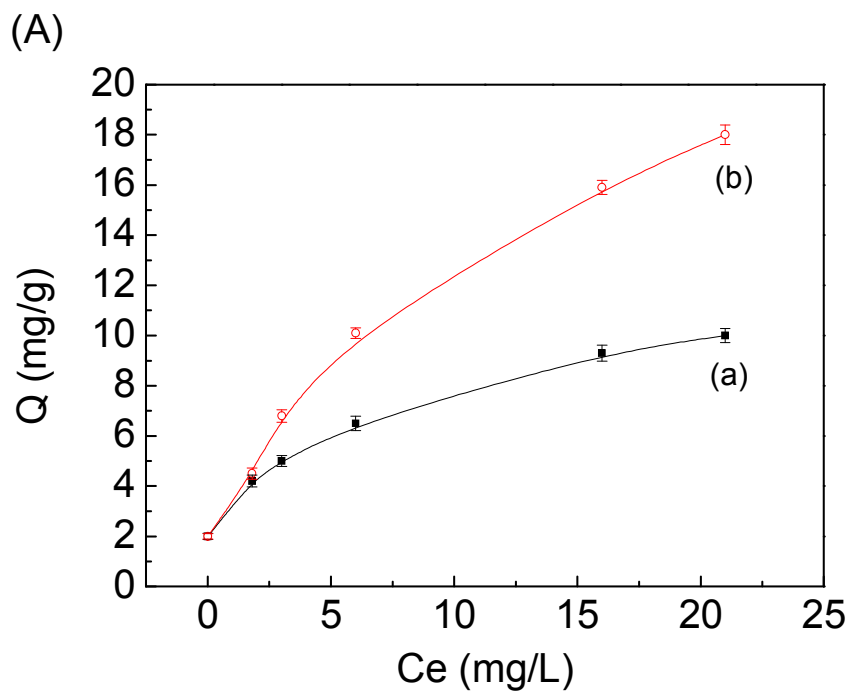
$$12 \quad \frac{1}{Q_e} = \frac{1}{Q_m} + \frac{1}{K_L Q_m C_e} \quad (3)$$

$$13 \quad \ln Q_e = \frac{\ln C_e}{n} + \log k_f \quad (4)$$

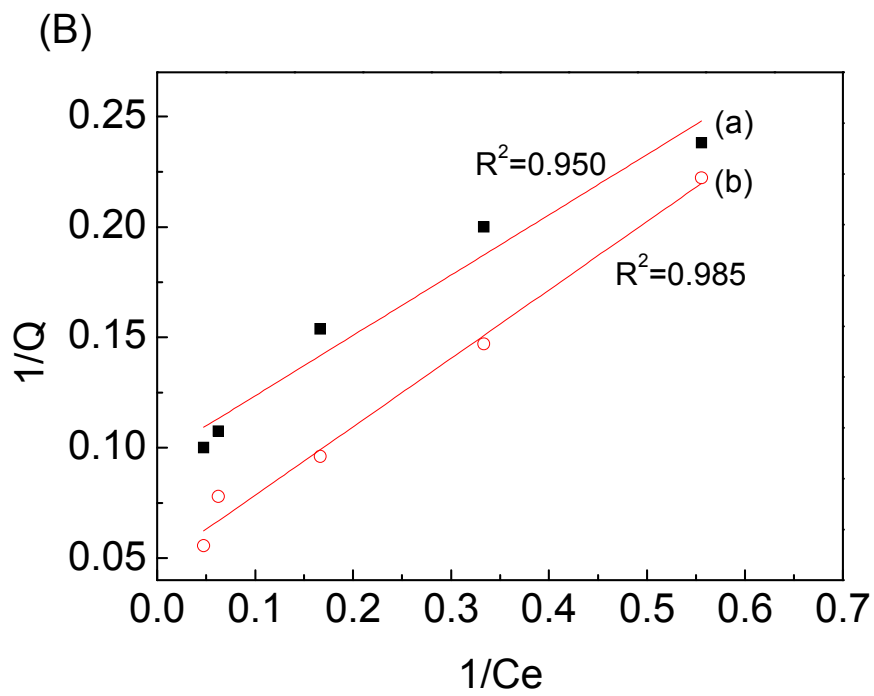
14 where Q_m is the theoretical maximum adsorption capacity per unit weight of the

1 adsorbent ($\text{mg}\cdot\text{g}^{-1}$), K_L and K_F are adsorption constants of Langmuir and Freundlich
2 models ($\text{L}\cdot\text{mg}^{-1}$), respectively, and n is the Freundlich linearity index.

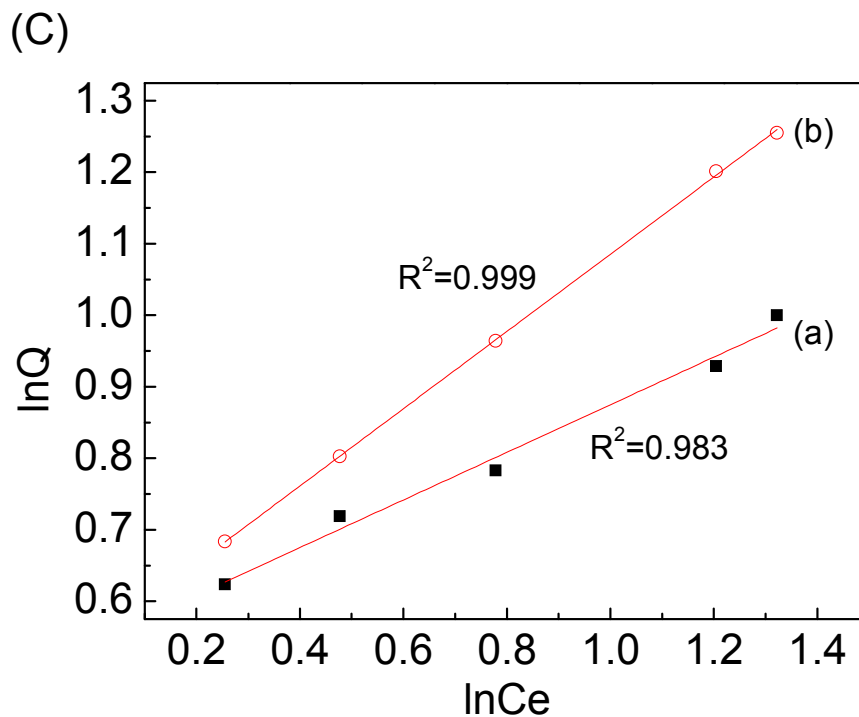
3 From **Fig. 9**, TCP uptakes of the GO sheets and GO sponge increased nonlinearly
4 with increasing the initial TCP concentrations and both isotherms of GO sheets and
5 GO sponge fitted Freundlich equation well. Interestingly, the TCP uptake of GO
6 sponge was much more than that of GO sheets. It reveals that the adsorption sites on
7 GO sponge with higher specific surface areas are sufficient and have strong
8 interactions with TCP. The adsorption capacity relies on the amount of TCP
9 transported from the bulk solution to the surfaces of the adsorbent at lower initial
10 concentrations. The isothermal adsorption behavior of GO sheets and GO sponge both
11 obey Freundlich equation, which reveals that a homogeneous monolayer adsorption is
12 dominant in both GO sheets and GO sponge. Based on previous literatures, GN
13 binding to adsorbates through parallel π - π stacking interactions usually forms
14 multilayer adsorption³². However, π - π stacking interaction was not likely a primary
15 cause of TCP adsorption on GO sheets and GO sponge. GO dominantly show
16 monolayer adsorption because oxygen-containing functional groups of GO are facile
17 to bind TCP due to electrostatic interactions or hydrogen bonds.



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2 **Fig. 9** (A) Adsorption isotherms, (B) Langmuir and (C) Freundlich model fitting for

3

TCP on GO sheets (a) and GO sponge (b)

Adsorbents	Langmuir			Freundlich		
	Q_m ($\text{mg}\cdot\text{g}^{-1}$)	K_l ($\text{L}\cdot\text{mg}^{-1}$)	R^2	K_f ($\text{L}\cdot\text{mg}^{-1}$)	n	R^2
GO sheets	10.39	0.352	0.950	3.38	2.738	0.983
GO sponge	21.06	0.153	0.985	3.89	2.97	0.999

4 **Table 1** Langmuir and Freundlich adsorption isotherms fitting parameters of TCP on

5

GO sheets and GO sponge

6 **3.2.3 Adsorption kinetics**

7

Aside from the adsorption equilibrium study, adsorption kinetics was also carried

8

out to establish the time course of TCP uptakes on GO sheets and GO sponge. The

9

typical experimental results of TCP adsorption versus time are shown in **Fig. 10**. The

1 adsorption at medium and low initial concentration achieved equilibrium in a short
2 term, and reached a plateau at adsorption time of 30 min for GO sheets and 12 min for
3 GO sponges, respectively. To further discuss the adsorption mechanisms, pseudo-first
4 and pseudo-second order models were applied to study the experimental data. The
5 non-linear forms of aforementioned kinetics models are presented below, respectively:

$$6 \quad \ln(Q_e - Q_t) = \ln Q_e - k_1 t \quad (4)$$

$$7 \quad \frac{t}{Q_t} = \frac{1}{k_2 Q_e^2} + \left(\frac{1}{Q_e}\right)t \quad (5)$$

8 where Q_e and Q_t are the amount of TCP adsorbed onto adsorbents ($\text{mg}\cdot\text{g}^{-1}$) at
9 equilibrium and at time t (min), respectively. k_1 (min^{-1}) and k_2 ($\text{g}\cdot\text{mg}^{-1}\cdot\text{min}^{-1}$) are the
10 rate constants of pseudo-first and pseudo-second order models, respectively.

11 The results were also listed in **Table 2**. It was found that the correlation
12 coefficients of pseudo-second order model for GO sheets and GO sponge were both
13 near to 1.0. It could be inferred that chemical adsorption may be involved in the
14 adsorption process. The strong adsorption of hydroxyl-substituted TCP indicated that
15 its adsorption reaction was related to the presence of hydroxyl group. It was well
16 known that GO has less p-electron and more O-containing groups on its surface. Thus,
17 hydroxyl groups of TCP could interact with O-containing groups on GO through
18 hydrogen-bonding interactions.

19

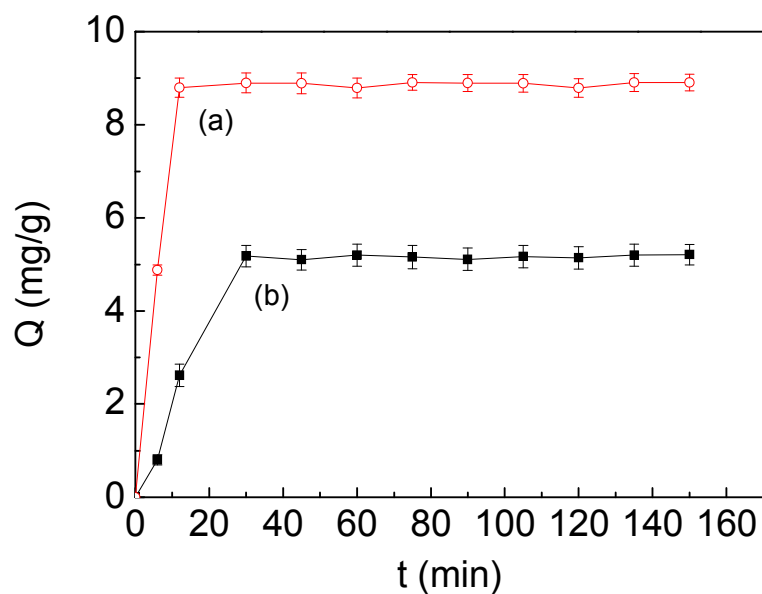


Fig. 10 Adsorption kinetics of TCP on GO sheets (a) and GO sponge (b)

Adsorbents	Pseudo first-order model		Pseudo second-order model	
	k_1 (min^{-1})	R^2	k_2 ($\text{mg}\cdot\text{g}^{-1}\cdot\text{min}^{-1}$)	R^2
GO sheets	0.042	0.901	0.1186	0.985
GO sponges	0.15	0.944	0.322	0.997

Table 2 Adsorption kinetic parameters of TCP on GO sheets and GO sponge

4. Conclusions

In this study, GO sponge as novel adsorbents had successfully been prepared via freeze-drying process. Spongy and foam-like structure of GO sponge was endowed with ultralarge surface areas and showed better adsorption capacity of TCP than GO sheets due to more sufficient and stronger hydrogen-bonding interactions of oxygen-containing functional groups on the surface. The results suggest that TCP had

1 optimum adsorption capacity on both GO sheets and GO sponge at pH 2.0~6.0.
2 Adsorption isotherms and kinetics curves of TCP on GO sheets and GO sponge were
3 nonlinear, indicating the homogeneous monolayer chemical adsorption. In other
4 words, except for hydrophobic interaction, hydrogen-bonding interactions were
5 involved in adsorption. Therefore, GO sponge could be considered as an excellent and
6 economical adsorption material for the removal of pollutants.

7

8 **Acknowledgements**

9 The authors gratefully acknowledge financial support from Ministry of Science and
10 Technology of China (Grants Nos. 2012AA03A602 and 2012AA021505), the Key
11 Laboratory of Marine Chemistry Theory and Technology, Ministry of Education (No.
12 841313040) and Shandong Major Project of Science and Technology (2012CX7301).

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