



Simultaneous electricity generation and tetracycline removal in continuous flow electrosorption driven by microbial fuel cells

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1 **Simultaneous electricity generation and tetracycline**
2 **removal in continuous flow electrosorption driven by**
3 **microbial fuel cells**

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8

9 A novel continuous flow electrosorption driven by microbial fuel cells (MFCs) was
10 developed for the first time to remove tetracycline, the second most commonly used
11 antibiotics, from a synthetic wastewater. The MFC-Sorption was proved to be
12 cost-effective without any external power consumption. Various operating parameters
13 including pH, electrolyte concentration, initial concentration of tetracycline, number
14 of MFCs connected in series and flow rate were investigated, and the adsorption
15 kinetics of tetracycline removal was studied. Three MFCs connected in series in the
16 continuous flow MFC-Sorption system reached an adsorption capacity of 23.12 mg
17 g⁻¹, higher than that with two MFCs (16.76 mg g⁻¹) and that with only one MFC
18 (14.16 mg g⁻¹). The performance was compared with that in the batch mode, and it
19 was confirmed that the continuous flow was practical for tetracycline removal. This
20 work confirmed that continuous flow electrosorption driven by MFCs was
21 cost-effective and environmentally friendly removal of tetracycline.

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22 **Introduction**

23 Antibiotics are widely used because of their vital roles on disease prevention,
24 modern agriculture and livestock industries.^{1,2} However, the excessive use of
25 antibiotics has led to some problems such as interfering the photosynthesis of aquatic
26 plants, disturbing the metabolism of native microbial communities.³ What's more,
27 antibiotic resistance in bacteria has become a great challenge for infection control all
28 over the world since even low levels of antibiotics may cause antibiotic resistant
29 bacteria in the aquatic environment.^{4,5}

30 Tetracycline, the second most commonly used antibiotics in the world, is hard to
31 be metabolized or absorbed completely in the human and animal bodies, with most
32 of the unchanged form being released in excreta. Thus, it has been detected in many
33 wastewater treatment plants as well as some surface water and groundwater, even in
34 pore water and overlying water.^{6,7} Due to special properties of tetracycline such as
35 wide application, high solubility in water, high residual toxicity and low
36 biodegradability, the abundant use of tetracycline in human and veterinary medicine
37 has a great effect on the quality of surface water and groundwater.⁸ Therefore, there
38 is a growing international concern about the abatement of these compounds to
39 decrease the potential impact on human health and environment.^{9,10}

40 Though significant researches have been carried out to remove tetracycline in
41 domestic, potable water and sludge, sound technologies are still very limited.¹¹ It
42 takes more than 3 months to achieve a 89% removal efficiency for swine wastewater
43 treatment by membrane bioreactor.¹² As one promising AOPs method for organics

44 mineralization, photocatalysis is not regarded as a cost-effective alternative due to
45 the need of UV light and the formation of more polar intermediates during
46 treatment.¹³ Though tetracycline could be removed by integration of ozonation and
47 biodegradation, it's found that the toxicity increased with treatment time.¹⁴ What's
48 more, tetracycline electro-oxidation on boron-doped diamond (BDD) anode has also
49 been investigated, but it is shown to be economically infeasible for practical
50 applications.¹⁵ Comparably, adsorption is a more practical and environmentally
51 friendly technology,¹⁶ and the recent studies on tetracycline removal have paid
52 attention to the use of various adsorbents including active sludge, soils, clay minerals,
53 carbonaceous adsorbents and bio-adsorbent.¹⁷⁻¹⁹

54 More recently, electrosorption has been developed as a more efficient adsorption
55 process for metal separation, desalination and organics removal due to its particular
56 advantages such as high capacity, no secondary waste and reversibility.²⁰⁻²² Operated
57 with a low voltage (0.5 - 1.5 V), the charged ions or polar molecules migrate to the
58 charged counter charge electrode, forming an electric double layer on the surface of
59 adsorption electrode, which enhances the adsorption capacity.²³ Thus, it can be easily
60 regenerated by removing the voltage or applying an inverse voltage on the
61 electrodes.²⁴⁻²⁶ To further improve the energy efficiency of the process, in our
62 previous work, an electrosorption system driven by microbial fuel cells
63 (MFC-Sorption) was first designed to remove phenol from wastewater in a batch
64 mode.²⁷ MFC is a promising technology for wastewater treatment and electricity
65 production at the same time,^{28,29} however, it is still far from industrial applications

66 due to low power densities. The design of MFC-Sorption can take full advantage of
67 the small amount of electricity generated by MFCs to achieve the electrosorption of
68 pollutants and energy recovery at the same time. Therefore, it would be of significant
69 environmental interest to explore the feasibility of this MFC-Sorption system for
70 effective removal of tetracycline.

71 In this work, this MFC-Sorption was studied for the first time to remove
72 tetracycline with continuous flow (Fig. 1), which would be more practical than bath
73 flow for real-world applications. Important operating parameters such as pH,
74 electrolyte concentration, initial tetracycline concentration, number of MFCs
75 connected in series, flow rate and so on were systematically investigated, which
76 would be predictive for some real wastewaters treatment application. The continuous
77 flow performance was also compared with that in batch mode. This research showed
78 that continuous flow MFC-Sorption system was effective and environmentally
79 friendly for tetracycline removal in synthetic wastewater.

80

81 **Results and discussion**

82 **Tetracycline removal in batch flow MFC-Sorption system**

83 Fig. 2A shows the adsorption capacity of tetracycline by MFC-Sorption and
84 traditional adsorption as well as their removal ratios under similar conditions in batch
85 mode. It was observed that the trend of adsorption capacity was similar in these two
86 systems, both increasing with time, which was consistent with the trend of phenol
87 removal in our previous work.²⁷ In the MFC-Sorption system, the adsorption capacity

88 of tetracycline was 69.3 mg g^{-1} , which increased by about 20% when compared with
89 that in traditional adsorption (57.6 mg g^{-1}). Such an enhancement was due to the
90 double electric layer on the anode surface.³⁰ The removal ratio of tetracycline in
91 MFC-Sorption system was also higher than that in the traditional sorption (insert Fig.
92 2A).

93 It is known that pH plays a significant role on the adsorption capacity since it is
94 related to the pK_a of tetracycline. Fig. 2B shows the effect of pH on adsorption
95 capacity and tetracycline removal ratio. It was observed that the adsorption capacity
96 of tetracycline decreased with pH decrease. When the solution pH was 3, the
97 adsorption capacity was 35.1 mg g^{-1} , and decreased slightly at pH 5, 7 and 9, but at
98 pH 11, it decreased noticeably by about 46% (19.0 mg g^{-1}). This phenomenon could
99 be explained as follows: when the solution pH was below 3.3, the tetracycline in
100 solution existed in the form of THC^{3+} , which was easy to be combined with the
101 negative charge on the surface of activated carbon fibers (ACF) electrode with cation
102 exchange to accomplish adsorption. The percentage of negative charge keeps rising
103 with the increase of solution pH, and there was less chance for cationic group on the
104 tetracycline to be combined with the negative charge on ACF electrode, leading to a
105 lower removal ratio.

106 Fig. 2C shows the effect of electrolyte concentration on the adsorption capacity of
107 tetracycline, which increased with the electrolyte concentration from 44.1 mg g^{-1}
108 ($0.001 \text{ mol L}^{-1} \text{ Na}_2\text{SO}_4$) to 64.6 mg g^{-1} ($0.1 \text{ mol L}^{-1} \text{ Na}_2\text{SO}_4$). The enhancement of
109 adsorption could be attributed to the increase of current with higher electrolyte

110 concentrations. Moreover, the adsorption capacity and removal efficiency increased
111 with the initial concentration of tetracycline from 50 mg L⁻¹ to 200 mg L⁻¹. The
112 adsorption capacity increased from 39.9 mg g⁻¹ (50 mg L⁻¹) to 122.2 mg g⁻¹ (200 mg
113 L⁻¹), but the corresponding removal ratio decreased from 99.7% to 76.4% (Fig. 2D).
114 These results agreed well with literature in which the deionization was found
115 increased with the increase of initial NaCl concentration.³⁰ This fact also indicated
116 that in the case of treatment of highly-concentrated tetracycline (*e.g.*, > 200 mg L⁻¹),
117 several single MFC-Sorption systems connected in series would help to reach a better
118 effluent water quality to meet the discharge requirement since one MFC-sorption
119 could not achieve a satisfactory tetracycline removal efficiency.

120 **(Fig. 2)**

121

122 MFC-Sorption systems with different numbers of MFCs connected in series were
123 investigated since it would affect the performance of electrosorption due to varied
124 power supplies by the MFCs. As shown in Fig. 3A, the adsorption capacity increased
125 slightly with time, and the more MFCs connected in series, the higher adsorption
126 capacity was. For example, the adsorption capacity of tetracycline with 3 MFCs
127 connected in series was 69.3 mg g⁻¹, slightly higher than that for the single-MFC
128 system (62.5 mg g⁻¹), which was due to the higher current output as shown in Fig. 3B.
129 It showed that the current for 3 MFCs connected in series dropped from 2.1 mA to 0.4
130 mA, higher than the current which decreased from 1.4 mA to 0.1 mA in case of a
131 single MFC. This non-linear current relations with the numbers of MFCs connected in

132 series indicated that cell reversal behavior happened in the present MFC-Sorption
133 system, which was observed in many researches on MFCs connected in series.^{31,32}
134 Thus it was important to operate under similar conditions and sufficient substrate
135 concentration to improve the power performance since it was reported that the
136 imbalance in these MFCs would often cause cell reversal.³³ Fig. 3C presents the
137 power density output of the MFCs connected in series, which reached the maximum
138 at 752 mW m^{-2} (3 MFCs), 573 mW m^{-2} (2 MFCs), and 394 mW m^{-2} (1 MFC),
139 respectively.

140 **(Fig. 3)**

141

142 **Tetracycline removal in continuous flow MFC-Sorption system**

143 Similar to batch flow experiments, several tests were designed to explore the
144 adsorption of tetracycline in a continuous flow MFC-Sorption. Fig. 4A compares the
145 adsorption capacity with that in the traditional adsorption system under the same flow
146 rate of 12 mL min^{-1} . In both systems, the adsorption capacity increased with time, and
147 after reaching the maximum at 150 min, they gradually decreased with more feed
148 solution flowing into the reactor. It might be explained that ACF electrodes gradually
149 reached adsorption saturation with time, and the further increasing feed solution led to
150 the decrease of adsorption capacity of tetracycline. The maximum adsorption capacity
151 of tetracycline in the continuous flow MFC-Sorption system was 16.76 mg g^{-1} ,
152 increased by about 31% compared with the traditional adsorption (11.56 mg g^{-1}). This
153 result was in agreement with the case in the batch mode.

154 The effect of initial tetracycline concentration on electrosorption performance was
155 explored in the continuous flow MFC-Sorption system at the concentration of 20 mg
156 L⁻¹, 50 mg L⁻¹ and 100 mg L⁻¹. Fig. 4B presents the adsorption capacity and removal
157 ratio of tetracycline. It confirmed that the adsorption capacity was related to the initial
158 tetracycline concentration, and the higher the initial concentration, the higher
159 adsorption capacity achieved. The adsorption capacity were 11.27 mg g⁻¹ (20 mg L⁻¹),
160 16.76 mg g⁻¹ (50 mg L⁻¹) and 25.44 mg g⁻¹ (100 mg L⁻¹), respectively, which indicated
161 that a high initial tetracycline concentration was more favorable for tetracycline
162 adsorption due to a higher driving force of the concentration gradient.^{29,34} However,
163 the removal efficiency decreased with the initial tetracycline concentration, decreased
164 from 51.1% (20 mg L⁻¹) to 26.4% (100 mg L⁻¹), which were similar to the results in
165 the batch flow system.

166 It has been reported that the adsorption rate has a close relationship with the mass
167 transfer rates,³⁵ because a proper flow rate is important for the electrosorption of
168 tetracycline in continuous flow MFC-Sorption system.³⁶ When the flow rate is too
169 small, it limits the wastewater volume to be treated, while a very high flow rate
170 decreases the retention time of tetracycline on electrodes, leading to a low adsorption
171 capacity. Fig. 4C shows the adsorption capacities and removal ratios under three
172 different flow rates (6, 12 and 24 mL min⁻¹). It was observed that the trend of the three
173 curves was similar, increasing first with time and then dropped gradually. The
174 adsorption capacity and removal efficiency of tetracycline both decreased with the
175 increase of flow rate. This observation was consistent with literature.^{34,35} For example,

176 the electrosorption removal efficiency decreased from 94.2% to 65.8% with the
177 increase of flow rate from 1.0 to 4.5 L min⁻¹.³⁵ It was also noticed that the removal
178 efficiency of tetracycline in continuous flow MFC was lower than that in batch mode.
179 This limitation might be overcome by improving the electrosorption reactor design,
180 for example, shortening electrode spacing and increasing electrode area.³⁷

181

182 **(Fig. 4)**

183

184 The effect of MFCs connected in series (1 MFC, 2 MFCs and 3MFCs) on the
185 adsorption capacity and removal ratios of tetracycline were explored in continuous
186 mode. It showed that the adsorption of tetracycline increased with time and reached
187 the maximum after about 2 h, and then decreased with time (Fig. 5A). With more
188 MFCs connected in series, higher adsorption capacity and removal efficiency of
189 tetracycline were obtained. In the case of 1 MFC, the maximum adsorption capacity
190 was 14.16 mg g⁻¹, while with 3 MFCs connected in series as the power supply for the
191 electrosorption, the adsorption capacity was 23.12 mg g⁻¹, increasing by 63.31%.
192 Consequently, the removal ratio increased from 31.5% to 51.5% (insert Fig. 5A). The
193 variations of current in the MFC-Sorption system are presented in Fig. 5B. Similar to
194 the current changes of MFCs in batch flow system, the currents also increased initially
195 in the first 1 h and then dropped. The maximum current with 3 MFCs connected in
196 series was about 1.3 mA, increasing by about 85.7% compared with the single one
197 (0.7 mA).

198

199

(Fig. 5)

200

201 Comparison of MFC-Sorption in batch and continuous flow

202 In order to compare the adsorption properties in batch and continuous flow systems,
203 relative parameters after 2 h removal of tetracycline were evaluated, as indicated in
204 Figs. 3A and 4A. The adsorption capacities in batch flow system were higher than
205 those in the continuous flow MFC-Sorption system. For example, with 3 MFCs
206 connected in series, the adsorption capacity in batch flow was 69.3 mg g⁻¹, much
207 higher than that in continuous flow (23.1 mg g⁻¹). Similarly, the tetracycline removal
208 efficiency in batch flow (93.16%) was larger than that in continuous flow (51.48%)
209 However, it should be noted that the treated volume in the continuous flow system
210 (1440 mL) was more than 3 times that of the batch one (400 mL), indicating that
211 continuous flow mode would be more effective for practical application.

212

213 MFC-Sorption adsorption kinetics

214 The pseudo-first-order kinetic and pseudo-second-order kinetic models were applied
215 to fit the adsorption of tetracycline by MFC-Sorption in batch flow.³⁸ The two model
216 equations are as follows:

217 Pseudo-first-order kinetics:

$$218 \ln(q_e - q_t) = \ln q_e - k_1 t \quad (1)$$

219 Pseudo-second-order kinetics:

$$220 \quad \frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (2)$$

221 where q_e and q_t (in $\text{mg g}^{-1}\text{ACF}$) are the amount of adsorbed tetracycline at equilibrium
 222 and time t (min), respectively. k_1 (min^{-1}) and k_2 ($\text{g mg}^{-1} \text{min}^{-1}$) are the
 223 pseudo-first-order and pseudo-second-order adsorption rate constant, respectively.

224 Figs. 6A and 6B display the tetracycline MFC-Sorption kinetics with different
 225 numbers of MFCs connected in series. The experimental results were fitted by the
 226 pseudo-first-order kinetic model (Fig. 6A) and pseudo-second-order kinetic model
 227 (Fig. 6B), respectively. Table 1 shows the relevant parameters of the kinetics in the
 228 above two systems. The results suggested that the adsorption capacity and the rate
 229 constant of tetracycline increased with the MFCs in series from 1 to 3 in both models.
 230 It was found that the pseudo-second-order kinetics model could fit the experimental
 231 results better than the pseudo-first-order kinetics equation due to the higher value of
 232 the correlation coefficients. A better criterion for the assessment of experimental
 233 results is a parameter called percent relative deviation modulus, P ,³⁹ or normalized
 234 percent deviation,⁴⁰ as shown in equation (5):

$$235 \quad P = \left(\frac{1}{N} \sum_{i=1}^N \frac{|q_{e(\text{expt})} - q_{e(\text{pred})}|}{q_{e(\text{expt})}} \right) \times 100\% \quad (3)$$

236 where N is the number of experimental data points, $q_{e(\text{expt})}$ is the experimental q_e , and
 237 $q_{e(\text{pred})}$ is the corresponding predicted q_e according to the proposed kinetic equation. It
 238 is obvious that the lower the P value, the better is the fit, and it is generally accepted
 239 that when the P value is less than 5, the fit is considered excellent.³⁹ Thus it is clear
 240 that in the present work the pseudo-second-order kinetics model fitted much better

241 when comparing these two models, which was in agreement with the results using
242 correlation coefficients.

243 (Fig. 6)

244

245 **Materials and methods**

246 **MFC configuration and operation**

247 A series of MFCs without membranes were used with a total volume of 14 mL. All of
248 the anode of MFCs were made by carbon cloth (Jilin Shenzhou Carbon Fiber Co.,
249 Ltd., China) with a surface area of 7 cm², and the cathode (7 cm²) were made by the
250 same material with a 0.2 mg cm⁻² Pt loading. A 1000 Ω external resistance was fixed
251 to connect with the anode and cathode. In the beginning, the MFC was inoculated
252 with anaerobic sludge with about 20%-30% inoculum to reactor liquid volume
253 fraction, 50 mM phosphate buffer solution (PBS) containing KCl 0.13 g L⁻¹, NH₄Cl
254 0.31 g L⁻¹, Na₂HPO₄·12H₂O 10.36 g L⁻¹ and NaH₂PO₄·2H₂O 3.32 g L⁻¹; 5 mL L⁻¹
255 vitamins; 12.5 mL L⁻¹ trace minerals and 1 g L⁻¹ glucose as carbon source. All of the
256 MFCs were operated in a 30±0.5°C temperature-controlled biochemical incubator.

257

258 **MFC-Sorption system**

259 The continuous flow MFC-Sorption system (Fig. 1a) consisted of four parts: MFCs
260 connected in series as the power supply, electrosorption unit, an external resistance
261 used to measure the current of system and a peristaltic pump to control the flow rate
262 of the feed solution. In the electrosorption unit (10×4×10 cm), both the anode and

263 cathode (10×10 cm) were ACFs, a certain concentration of tetracycline with the
 264 supporting electrolyte of sodium sulphate were used as the synthetic wastewater.
 265 MFCs were connected in series in order to obtain a suitable voltage output for the
 266 electrosorption unit. For the batch flow MFC-Sorption system (Fig. 1b), it was very
 267 similar with that of the flow system using the same reactor materials and sizes except
 268 the absence of the peristaltic pump and the flow of the feed solution.

269 **(Fig. 1)**

270

271 **Analysis**

272 The current of this system was calculated from $I=U_1/R$, where U_1 represents the
 273 voltage of the external resistance, and power density could be calculated from $P=IU_2$,
 274 where U_2 represents the voltage applied to the adsorption process.

275 The concentration of tetracycline was determined from the experimentally derived
 276 standard curves and the absorbance of tetracycline was obtained by a UV-Vis
 277 spectrophotometer (VI-1501, Tianjin Gangdong Instrument Analysis Instrument Co.,
 278 Ltd) at the wavelength of 357 nm. The adsorption capacity (q) and the removal ratio
 279 (R) of tetracycline in the electrosorption unit were calculated from the following two
 280 equations, respectively:

$$281 \quad \text{Adsorption capacity (mg g}^{-1}\text{)} \quad q = \frac{(c_0 - c_t)V}{m} \quad (4)$$

$$282 \quad \text{Removal ratio} \quad R\% = \frac{c_0 - c_t}{c_0} \times 100\% \quad (5)$$

283 Where c_0 , c_t (mg L⁻¹) represent the initial concentration and the given time
 284 concentration of tetracycline, respectively, V represents the available volume of the

285 reactor (400 mL), and m represents the mass of the ACF electrode in the
286 electrosorption unit (g).

287

288 **Conclusions**

289 A novel continuous flow MFC-Sorption system was developed for the first time in our
290 study, which achieved a tetracycline removal ratio of 51.48%, and the results showed
291 that high concentration of electrolyte, initial tetracycline concentration and low value
292 of pH could enhance the adsorption capacity of tetracycline in the synthetic
293 wastewater. This work indicated that the adsorption capacity of tetracycline increased
294 with the increased numbers of MFCs connected in series, proving that the higher
295 voltage and current could accelerate the removal efficiency of tetracycline. At pH 3,
296 flow rate of 6 mL/min, and 3 MFCs connected in series, the removal efficiency for 20
297 mg/L tetracycline was the best. The adsorption process of tetracycline fitted the
298 pseudo-second-order kinetics model better than the pseudo-first-order kinetics
299 equation. The continuous flow MFC-Sorption system without an external power
300 supply was proven to be an efficient and energy saving process for the removal of
301 tetracycline in synthetic wastewater.

302

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307

308 **Notes and references**

- 309 1 R. E. Alock, A. Sweetman and K. C. Jones, *Chemosphere*, 1999, **38**, 2247-2262.
- 310 2 A. K. Sarmah, M. T. Meyer and A. B. A. Boxall, *Chemosphere*, 2006, **65**, 725-759.
- 311 3 H. Y. Sun, X. Shi, J. D. Mao and D. Q. Zhu, *Environ. Toxicol. Chem.*, 2010, **29**,
- 312 1934-1942.
- 313 4 T. Schwartz, W. Kohnen, B. Jansen and U. Obst, *FEMS. Microbiol. Ecol.*, 2003, **43**,
- 314 325-335.
- 315 5 D. M. Cheng, X. H. Liu, L. Wang, W. W. Gong, G. N. Liu and W. J. Fu, *Sci. Total.*
- 316 *Environ.*, 2014, **476-477**, 266-75.
- 317 6 X. S. Miao, F. Bishay, M. Chen and C. D. Metcalfe, *Environ. Sci. Technol.*, 2004,
- 318 **38**, 3533-3541.
- 319 7 K. G. Karthikeyan and M. T. Meyer, *Sci. Total. Environ.*, 2006, **361**, 196-207.
- 320 8 Y. A. Ouaisa, M. Chabani, A. Amrane and A. Bensmaili, *J. Environ. Chem. Eng.*,
- 321 2014, **2**, 177-184.
- 322 9 M. Klavarioti, D. Mantzavinos and D. Kassinos, *Environ. Int.*, 2009, **35**, 402-417.
- 323 10 I. Sirés and E. Brillas, *Environ. Int.*, 2012, **40**, 212-229.
- 324 11 X. R. Jing, Y. Y. Wang, W. J. Liu, Y. K. Wang and H. Jiang, *Chem. Eng. J.*, 2014,
- 325 **248**, 168-174.
- 326 12 N. Prado, J. Ochoa and A. Amrane, *Bioresour. Technol.*, 2009, **100**, 3769-3774.
- 327 13 X. D. Zhu, Y. J. Wang, R. J. Sun and D. M. Zhou, *Chemosphere*, 2013, **92**,
- 328 925-932.

- 329 14 C. V. Gómez-Pacheco, M. Sánchez-Polo, J. Rivera-Utrilla and J. López-Peñalver,
330 *Chem. Eng. J.*, 2011, **178**, 115-121.
- 331 15 C. I. Brinzila, M. J. Pacheco, L. Ciriaco, R. C. Ciobanu and A. Lopes, *Chem. Eng.*
332 *J.*, 2012, **209**, 54-61.
- 333 16 X. R. Jing, Y. Y. Wang, W. J. Liu, Y. K. Wang and H. Jiang, *Chem. Eng. J.*, 2014,
334 **248**, 168-174.
- 335 17 F. Yu, J. Ma and S. Han, *Sci. Rep.*, 2014, **4**, 5326.
- 336 18 J. G. Yu, X. H. Zhao, H. Yang, X. H. Chen, Q. Q. Yang and L. Y. Yu, *Sci. Total.*
337 *Environ.*, 2014, **482-483**, 41-51.
- 338 19 Y. Zhuang, F. Yu, J. Ma, and J. Chen, *RSC. Adv.*, 2015, **5**, 27964.
- 339 20 Y. Liu, C. Nie, X. Liu, X. Xu, Z. Sun and L. Pan, *RSC. Adv.*, 2015, **5**, 15205.
- 340 21 C. O. Ania and F. Béguin, *Water Res.*, 2007, **41**, 3372-3380.
- 341 22 Z. L. Chen, C. Y. Song, X. W. Sun, H. F. Guo and G. D. Zhu, *Desalination*, 2011,
342 **267**, 239-243.
- 343 23 Y. Kong, W. Li, Z. L. Wang, C. Yao and Y. X. Tao, *Electrochem. Commun.*, 2013,
344 **26**, 59-62.
- 345 24 E. Ayranci and B. E. Conway, *J. Electroanal. Chem.*, 2001, **513**, 100-110.
- 346 25 L. Li, L. Zou, H. Song and G. Morriss, *Carbon*, 2009, **47**, 775-781.
- 347 26 E. Bayram and E. Ayranci, *Sep. Purif. Technol.*, 2012, **86**, 113-118.
- 348 27 J. Yang, M. H. Zhou, Y. Y. Zhao, C. Zhang and Y. S. Hu, *Bioresour. Technol.*,
349 2013, **150**, 271-277.
- 350 28 H. Wang, J. Park and Z. Ren, *Environ. Sci. Technol.*, 2015, **49**, 3267.

- 351 29 M. H. Zhou, M. L. Chi, H. Y. Wang and T. Jin, *Biochem. Eng. J.*, 2012, **60**,
352 151-155.
- 353 30 C. J. Feng, C. H. Hou, S. H. Chen and C. P. Yu, *Chemosphere*, 2013, **91**, 623-628.
- 354 31 S. E. Oh and B. E. Logan, *J. Power Sources*, 2007, **167**, 11-17.
- 355 32 I. Ieropoulos, J. Greenman and C. Melhuish, *Bioelectrochemistry*, 2010, **78**, 44-50.
- 356 33 H. C. Boghani, G. Papaharalabos, I. Michie, K. R. Fradler, R. M. Dinsdale, A. J.
357 Guwy, I. Ieropoulos, J. Greenman and G. C. Premier, *J. Power Sources*, 2014, **269**,
358 363-369.
- 359 34 Y. Zhao, X. M. Hu, B. H. Jiang and L. Li, *Desalination*, 2014, **336**, 64-71.
- 360 35 M. Mossad and L. Zou, *J. Hazard. Mater.*, 2012, **213-214**, 491-497.
- 361 36 X. Z. Wang, L. K. Pan, M. G. Li, Y. W. Chen, R. M. Cheng and S. M. Huang, *Surf.*
362 *Rev. Lett.*, 2007, **14**, 135-139.
- 363 37 S. Y. Wang, X. N. Li, Y. B. Zhang, X. Quan, S. Chen, H. T. Yu and H. M. Zhao,
364 *Chin. Sci. Bull.*, 2014, **59**, 2890-2897.
- 365 38 M. Mossad and L. Zou, *Chem. Eng. J.*, 2013, **223**, 704-713.
- 366 39 C. J. Lomauro, A. S. Bakshi and T. P. Labuza, *Lebensm-Wiss. Technol.*, 1985, **18**,
367 111-117.
- 368 40 R. -S. Juang, R. -L. Tseng, F. -C. Wu and S. -H. Lee, *Sep. Sci. Technol.*, 1996, **31**,
369 1915-1931.
- 370

371 **Figure captions**

372

373 **Fig. 1.** The schematic diagram of a continuous (a) and batch (b) flow MFC-Sorption
374 apparatus.

375

376 **Fig. 2.** Operating parameters (power supply, pH, electrolyte concentration and initial
377 concentration of tetracycline) in batch flow MFC-Sorption system.

378

379 **Fig. 3.** Effect of MFCs connected in series on the adsorption of tetracycline (A),
380 current (B) and power density (C) in batch flow MFC-Sorption system.

381

382 **Fig. 4.** Operating parameters (power supply, initial concentration of tetracycline and
383 flow rate) in continuous flow MFC-Sorption system.

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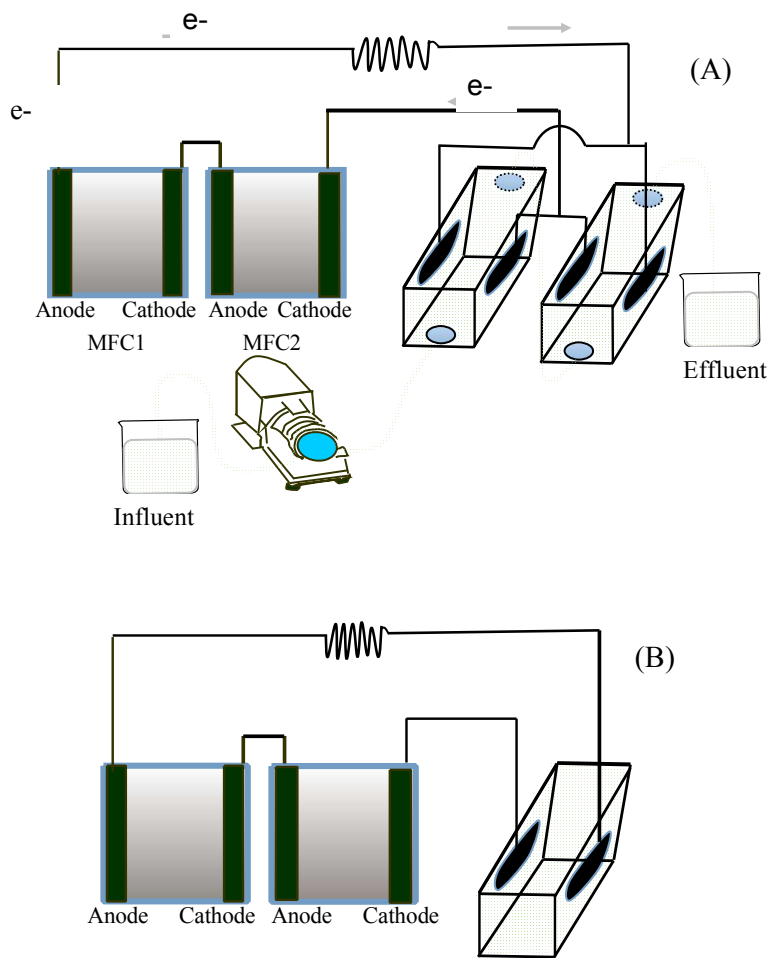
385 **Fig. 5.** Effect of MFCs connected in series on the adsorption of tetracycline (A),
386 current (B) in continuous flow MFC-Sorption system.

387

388 **Fig. 6.** The kinetics of batch flow MFC-Sorption system with different MFCs
389 connected in series.

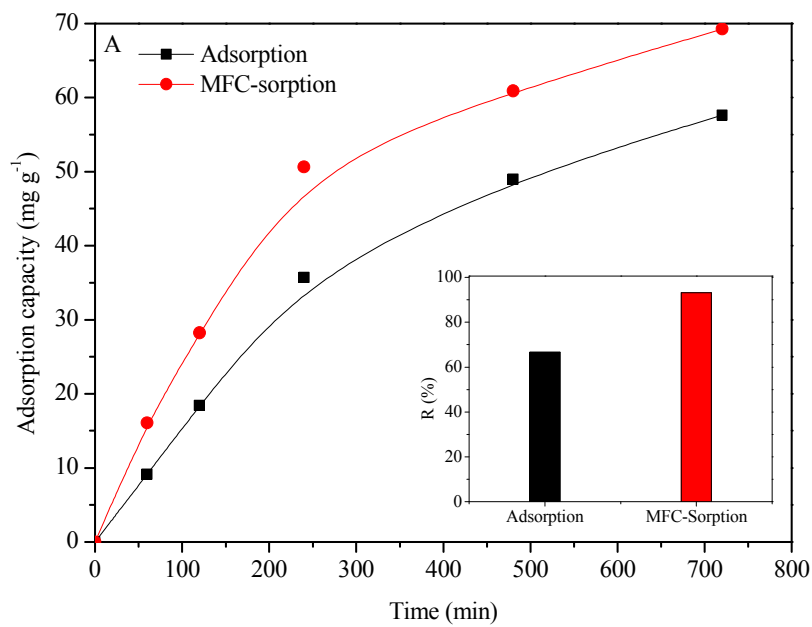
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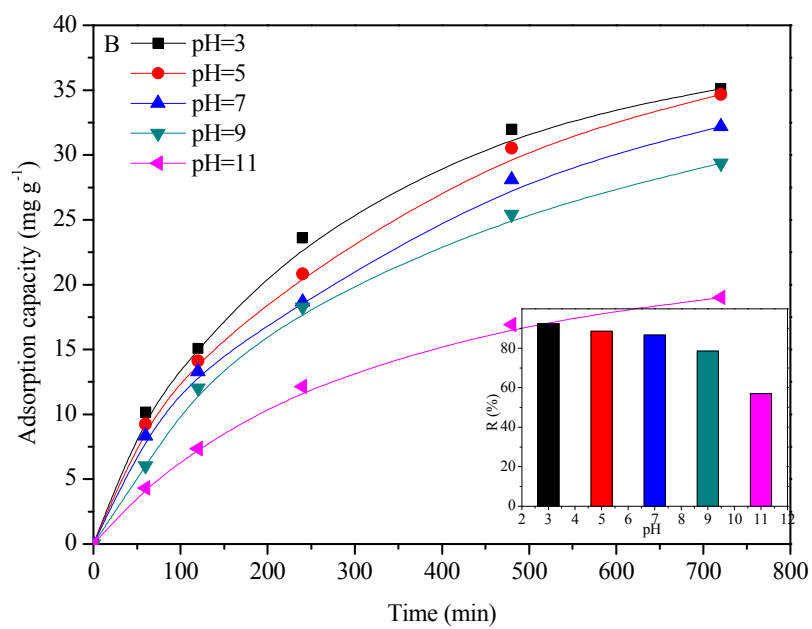


(Fig. 1)

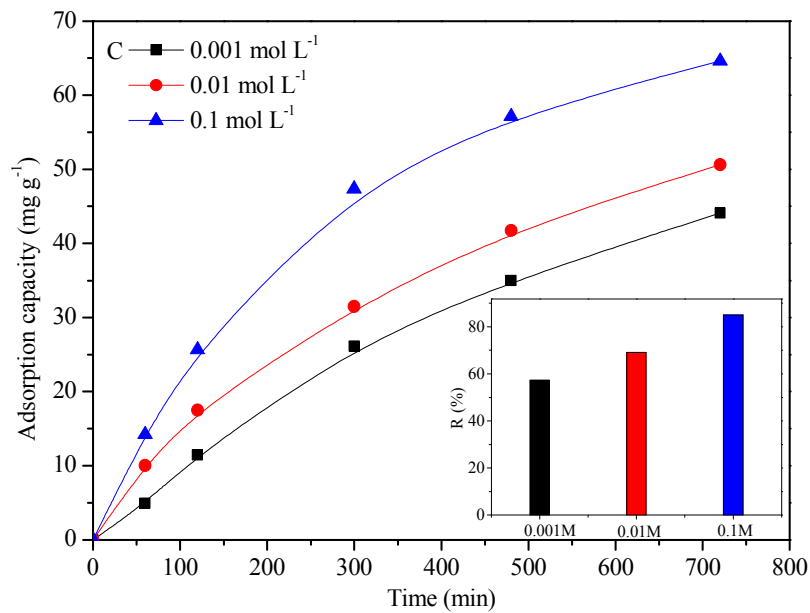
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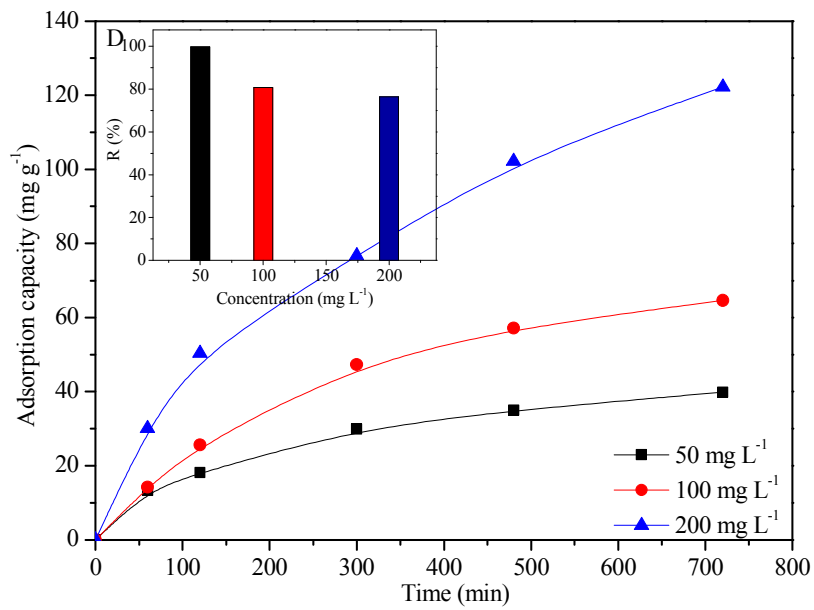
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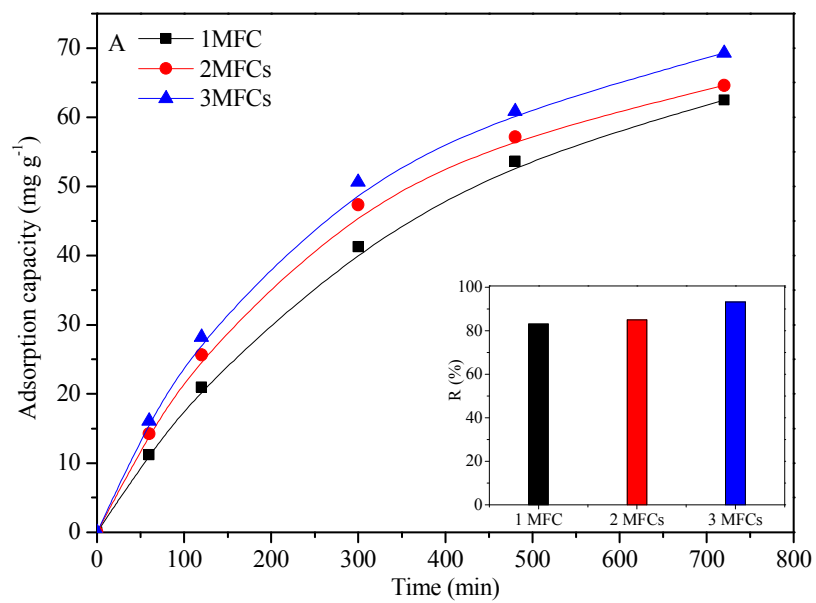
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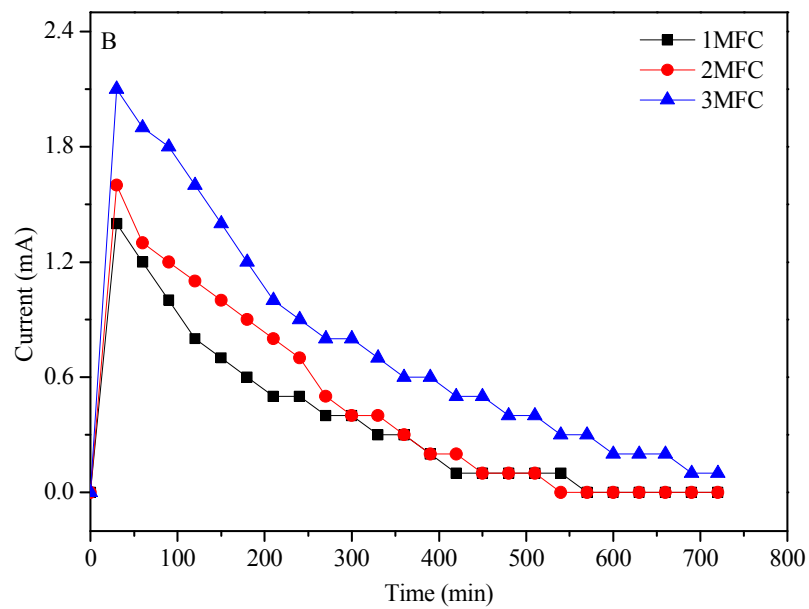
(Fig. 2)

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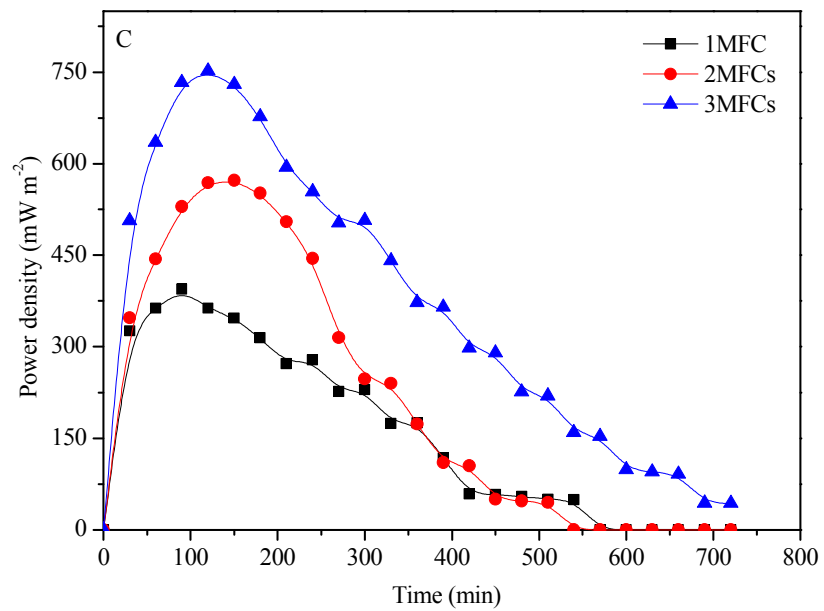
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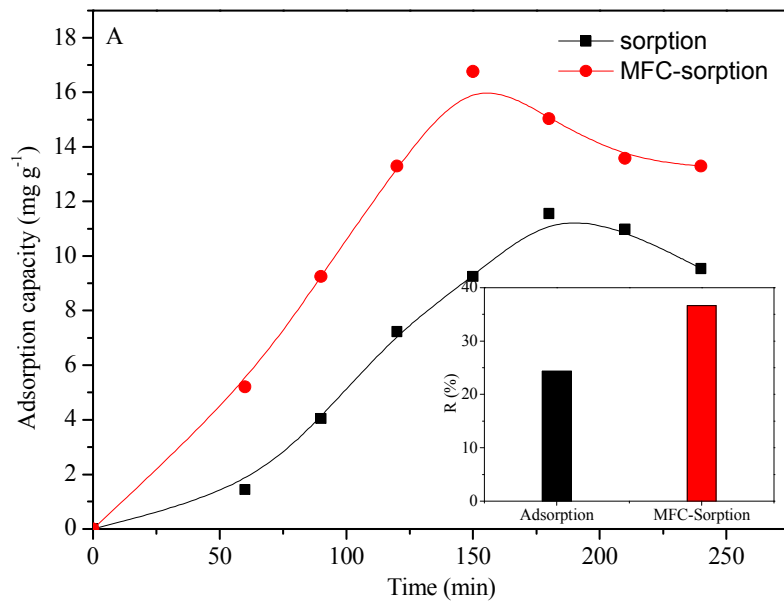
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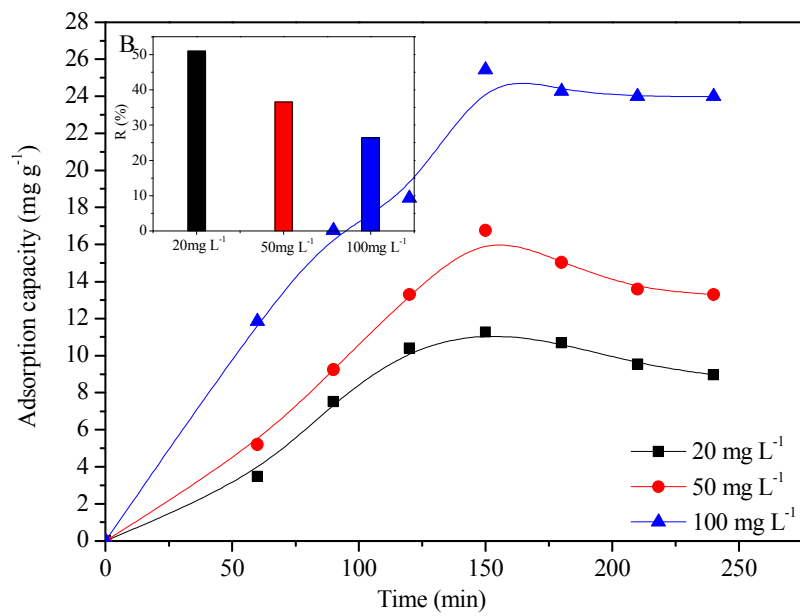
(Fig. 3)

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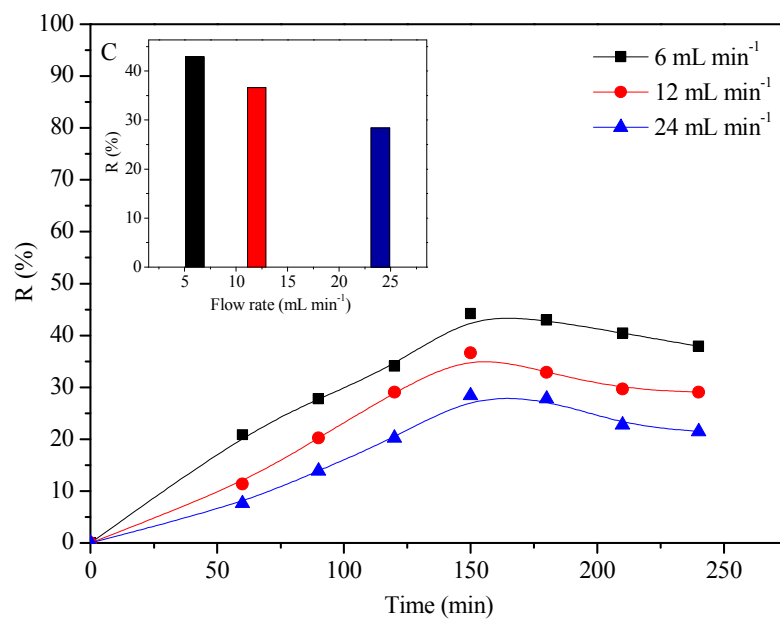
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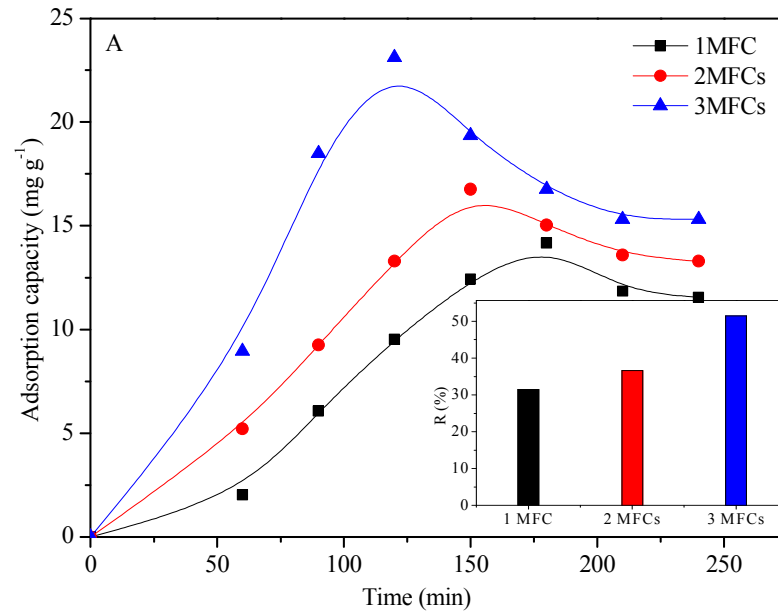
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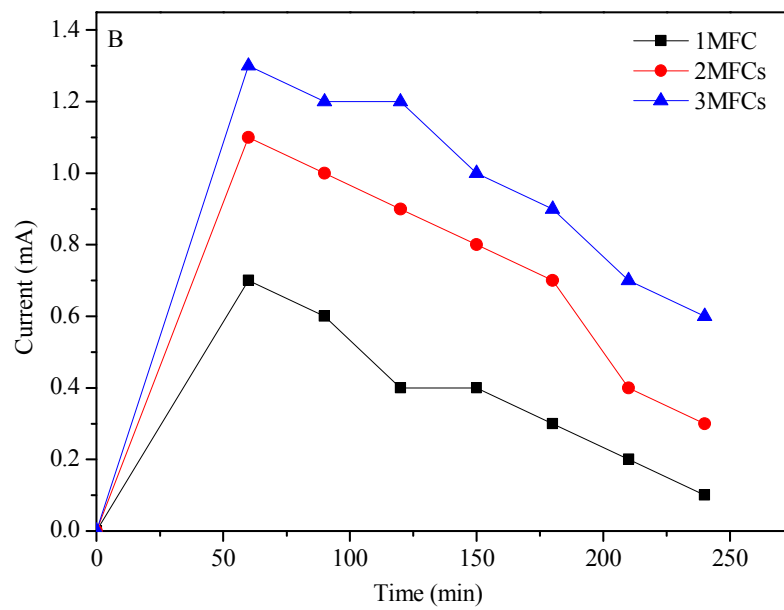
(Fig. 4)

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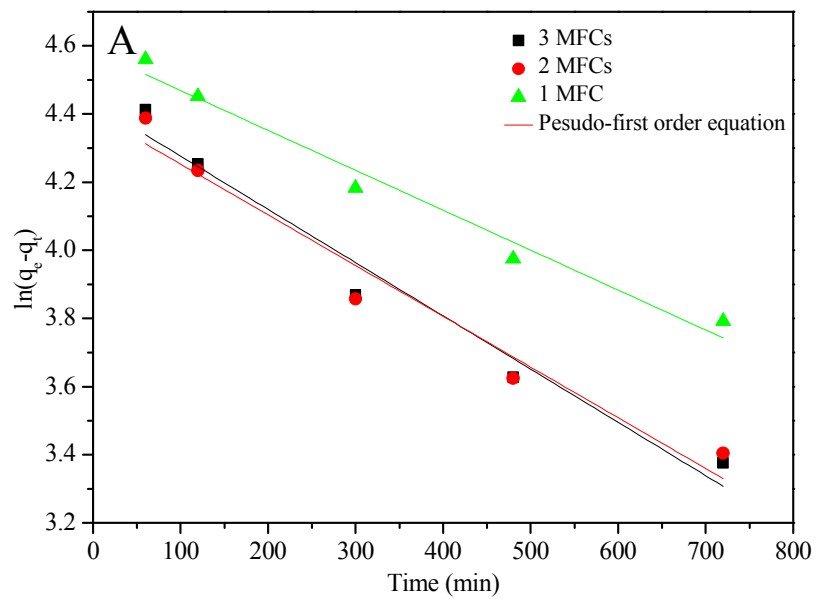
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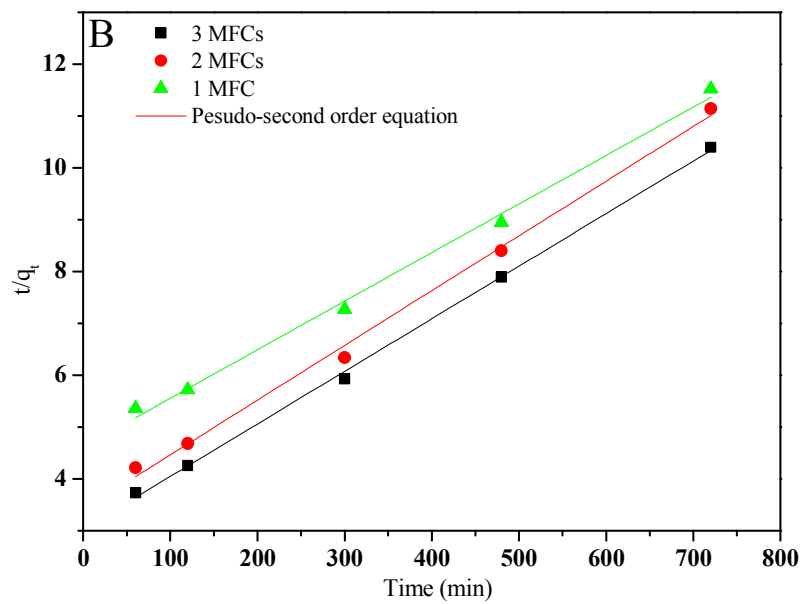
(Fig. 5)

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(Fig. 6)

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510 **Table 1**

511 MFC-Sorption kinetic parameters of tetracycline removal under different
 512 experimental conditions.

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	Pseudo-first-order model				Pseudo-second-order model			
	q_e	k_1	R^2	P	q_e	k_2	R^2	P
	$(10^{-3} \text{ min}^{-1})$				$(10^{-5} \text{ g mg}^{-1} \text{ min}^{-1})$			
3 MFCs	84.12	1.56	0.960	21.4	98.52	3.4	0.999	1.16
2 MFCs	81.55	1.49	0.950	26.3	94.70	3.27	0.995	2.13
1 MFC	98.09	1.17	0.969	21.1	106.84	1.90	0.994	2.84

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