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1 **Oxidation and reduction performance of 1,1,1-trichloroethane in**
2 **aqueous solution by means of a combination of persulfate and**
3 **zero-valent iron**

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16 **Abstract**

17 In this study, the degradation performance of 1,1,1-trichloroethane (TCA) involving both
18 oxidation and reduction processes was investigated with an application of the persulfate-ZVI
19 (zero-valent iron) system, in which it is generally believed that $\text{SO}_4^{\cdot-}$ -induced oxidation was
20 responsible for pollutants removal. The study was conducted with persulfate and
21 un-pretreated ZVI through batch experiments. The results showed that TCA was stable in the
22 presence of ZVI alone within 12 h and degraded with the addition of persulfate. TCA
23 degradation efficiency was found to increase with increasing persulfate concentrations, but to
24 decrease with increasing ZVI dosage. A two-stage process involving persulfate oxidation and
25 ZVI reduction was developed during TCA degradation. The addition of isopropanol and
26 *tert*-butyl alcohol proved the existence of sulfate and hydroxyl radicals during the 1st-stage
27 (0~2 h), which were absent in the 2-nd stage (2~12 h) when persulfate was exhausted. The
28 degradation performance of carbon tetrachloride, a reduction probe compound, was evidence
29 of the persulfate-ZVI system involving an enhanced ZVI reduction, and which was mainly
30 responsible for TCA degradation in the 2nd-stage. 1,1-Dichloroethane was the only
31 confirmed intermediate emerging during the 2nd-stage.

32 **Key words:** 1,1,1-Trichloroethane; Persulfate-ZVI system; Persulfate oxidation; Enhanced
33 ZVI reduction

34 1. Introduction

35 1,1,1-Trichloroethane (TCA), known as a chlorinated solvent, is a commonly identified
36 groundwater contaminant that has been detected in at least 50% of the sites listed on the U.S.
37 Environmental Protection Agency (USEPA) National Priorities List.¹ The appearance of TCA
38 in soils and groundwater has elicited much attention because of its recalcitrant characteristic
39 and potential to cause liver, nervous system and circulatory system problems from the
40 long-term exposure. Hence, the USEPA maximum contaminant level of TCA in drinking
41 water has been set at 0.2 mg L⁻¹.²

42 In situ chemical oxidation (ISCO) has become a widely used technology for the
43 remediation of groundwater contaminated by chlorinated solvents. Persulfate, one of the
44 strongest oxidants with a redox potential (E^0) of 2.01 V, has recently received considerable
45 attention for ISCO. In most case, various methods, including heat, ultraviolet irradiation,
46 transition metals, hydrogen peroxide, alkaline pH, etc., are used in order to activate persulfate
47 and generate the sulfate radical ($SO_4^{\cdot-}$, $E^0 \approx 2.6$ V) and other reactive species.^{3,4} Transition
48 metals such as Fe^{2+} have been proved to be effective to activate persulfate for the degradation
49 of most prevalent organic contaminants (eq 1).⁵⁻⁷ However, it has been found that excess Fe^{2+}
50 can further consume the produced $SO_4^{\cdot-}$ which results in a reduction of Fe^{2+} activation (eq 2).
51 Therefore, it is important to maintain an appropriate concentration of Fe^{2+} in the solution.



54 As an alternative source of Fe^{2+} , ZVI has been employed as a persulfate activator with

55 promising results in the degradation of many organic contaminants.⁸⁻¹⁰ The slow-releasing of
56 Fe^{2+} and recycling of Fe^{3+} at the ZVI surface can prevent the accumulation of excess Fe^{2+} in
57 solution.¹¹ It has been reported that persulfate activation using ZVI for degradation of organic
58 compounds was more effective than Fe^{2+} .¹²⁻¹⁵ Li et al.^{16, 17} demonstrated that particle size of
59 ZVI and solution pH value could influence the degradation performance of acid orange 7
60 (AO7) and the iron corrosion coating, respectively. Oh et al.¹³ suggested that the most likely
61 mechanism for persulfate activation by ZVI did not involve aqueous Fe^{2+} , but direct electron
62 transfer from ZVI or surface-bound Fe^{2+} . Ahn et al.¹⁸ believed that the near-surface Fe^{2+} on a
63 bilayered-structured passive film was contributed to the generation of $\text{SO}_4^{\cdot-}$, rather than direct
64 contact between persulfate and ZVI or Fe^{2+} in the solution.

65 According to TCA, a naturally recalcitrant hydrocarbon with one single bond, has been
66 confirmed to be effectively removed by thermal and UV activation of persulfate.^{19, 20} Xu et
67 al.²¹ reported that $\text{SO}_4^{\cdot-}$, $\cdot\text{OH}$, and $\text{O}_2^{\cdot-}$ were all generated in the thermally activated
68 persulfate system, while $\cdot\text{OH}$ appeared to be the predominant radical species for TCA
69 removal. However, there is less information regarding TCA degradation performance by
70 persulfate activation with transition metals. In this study, a persulfate-ZVI (ZVI without
71 pretreatment) system was applied to remove TCA in aqueous solution, and it was interesting
72 to find that TCA degraded gradually even when persulfate was exhausted. Therefore, other
73 reaction chemistry without persulfate is supposed to be responsible for TCA removal, while
74 most studies using persulfate-ZVI method focused on persulfate oxidation involving $\text{SO}_4^{\cdot-}$ or
75 $\cdot\text{OH}$, and direction reduction action by ZVI was less reported. Hence, the objective of this

76 study was (1) to investigate the degradation performance of TCA in the persulfate-ZVI
77 system, (2) to evaluate the role of persulfate in the system, and (3) to examine the potential
78 oxidation and reduction processes responsible for TCA degradation. TCA degradation
79 performance and the trends of dissolved Fe^{2+} and persulfate were investigated, and then the
80 radical scavenger test and chemical probe method were conducted to examine the responsible
81 processes for TCA removal. Furthermore, the intermediates during TCA degradation and the
82 chloride mass balance were determined.

83 **2. Materials and methods**

84 **2.1. Materials**

85 1,1,1-Trichloroethane (TCA, 99.0%), ZVI powder (99%, 150 μm), isopropanol (IPA,
86 99.7%), *tert*-butyl alcohol (TBA, 99.0%), methyl *tert*-butyl ether (MTBE, 99.9%), sodium
87 bicarbonate (99.5%), and potassium iodide (99.0%) were purchased from Shanghai Jingchun
88 Reagent Co., Ltd. (Shanghai, China). Carbon tetrachloride (CT, 99.5%), 1,10-phenanthroline
89 (99.0%), and *n*-hexane (97%) were purchased from Shanghai Lingfeng Chemical Reagent
90 Co., Ltd. Persulfate (98.0%) was purchased from the Sinopharm Chemical Reagent Co., Ltd.
91 (Shanghai, China). Ultra-pure water from a Milli-Q water process (Classic DI, ELGA) was
92 used for preparing aqueous solutions.

93 **2.2. Experimental procedures**

94 All reactions were conducted in 24 mL volatile organic analysis (VOA) vials fitted with
95 polytetrafluoroethylene (PTFE) lined caps. Stock solution of TCA was prepared by allowing

106 the pure nonaqueous-phase liquid TCA to equilibrate with Milli-Q water overnight with
107 gentle stirring in the dark and later diluted to 20 mg L^{-1} (0.15 mM). When required dosage of
108 ZVI was added to a series of reaction vials, a predetermined amount of persulfate was added
109 to the TCA-containing solution, and later distributed by fully filling the reaction vials. All
110 sample vials were placed in a head-to-bottom rotation drum (with a speed of 5 rpm) to
111 enhance the movement of ZVI powder (Fig. S1).²² The rotation drum was located in the
112 constant temperature chamber to keep the temperature at $20 \pm 1 \text{ }^\circ\text{C}$. The reaction vials were
113 sacrificed at different intervals for immediate analyses. The initial pH in all experiments was
114 unadjusted except in the test for investigating the influence of Fe^{2+} on ZVI reduction. All
115 experiments were conducted in triplicate and the mean values reported.

106 2.3. Analytical methods

107 The concentration of TCA and CT was quantified after extraction with hexane by a gas
108 chromatograph (GC, Agilent 7890A, Palo Alto, CA) equipped with an autosampler (Agilent
109 7693). The method detection limit (MDL) for TCA is $5 \text{ } \mu\text{g L}^{-1}$. The volatile organic
110 intermediates formed in TCA degradation were identified by aqueous samples using an
111 automatic purge and trap (Tekmar Atomx, Mason, OH) coupled to a GC/MS (Agilent
112 7890/5975). The MDL for the intermediates is $0.5 \text{ } \mu\text{g L}^{-1}$. The carboxylic acid intermediates
113 were identified using a GC/MS (Shimadzu GC/MS-QP 2010, Kyoto, Japan) after
114 derivatization with acidic methanol. Details of the analytical methods for TCA and the
115 intermediates are shown in SI Text S1. The concentration of $\text{S}_2\text{O}_8^{2-}$ was determined by a
116 spectrophotometric method using potassium iodide.²³ Ferrous ion and total iron were

117 quantified using 1,10-phenanthroline at a wavelength of 512 nm by a Hach DR 6000
118 spectrophotometer (Loveland, CO).²⁴ The pH was measured with a pH meter (Mettler-Toledo
119 DELTA 320, Greifensee, Switzerland). The concentration of chloride ions was detected by
120 anion chromatograph (Dionex ICS-I000, Sunnyvale, CA).

121 **3. Results and discussion**

122 **3.1. Effects of persulfate concentration and ZVI dosage on TCA degradation in the** 123 **persulfate-ZVI systems**

124 Experiments were conducted to determined TCA degradation performance by a
125 combination of persulfate and ZVI. The control experiments revealed that less than 5% loss
126 of TCA due to volatilization and TCA was stable after addition of persulfate or Fe²⁺ alone at
127 20 °C (Fig. S2). Our previous study showed that TCA could be effectively removed in a soil
128 slurry system by means of pretreated ZVI powder.²⁵ In this study, the control test was
129 conducted using the same ZVI (0.05 g, 150 μm) but without pretreatment in the absence of
130 persulfate, and the result showed no observable TCA removal over 12 h, representing that
131 direct reduction of ZVI was not sufficient for TCA degradation as the ZVI powder had been
132 passivated under the experimental conditions (Fig. 1a). However, TCA removal did occur
133 when persulfate at various concentrations was applied together with ZVI, indicating that the
134 combination of ZVI with persulfate was efficient in promoting the degradation of TCA,
135 which is also considered as a competitive advantage for this method applied in field as ZVI
136 could be used without pretreatment. When persulfate concentration increased from 1.5 mM to

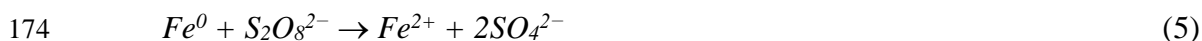
137 9.0 mM, the TCA degradation efficiency increased from 29% to 97%. Further increase of
138 persulfate concentration to 12.0 mM almost did not increase TCA removal, indicating that
139 there was an optimum persulfate concentration at fixed dosage of ZVI in the persulfate-ZVI
140 system. The results are also consistent with the conclusions of some other researchers.^{14, 26, 27}
141 Therefore, persulfate concentration in subsequent experiments was set at 9.0 mM.

142 The effect of ZVI dosage on TCA removal in the persulfate-ZVI system was evaluated
143 by changing ZVI loading while keeping persulfate concentration at 9.0 mM (Fig. 1b). The
144 dosages of ZVI 0.05 g, 0.10 g, 0.15 g, and 0.20 g (corresponding to 2.08~8.33 g/L) were
145 applied. It can be seen that the degradation efficiency of TCA decreased with the increase of
146 ZVI dosage in the persulfate-ZVI system. When increasing ZVI amount from 0.05 g to 0.20 g,
147 TCA removal declined from 97% to 70%. However, Li et al.¹⁶ reported that the degradation
148 efficiency of AO7 increased with higher ZVI dosage when mili-ZVI (1 mm) and micro-ZVI
149 (150 μm) were used in the persulfate system, and similar results were found by some other
150 researchers as well.^{13, 28, 29} In addition, other studies demonstrated an optimum dosage of ZVI
151 in the persulfate-ZVI systems, and contaminants removal decreased when ZVI loading was
152 increased over the optimum dosage.^{12, 15, 27, 30} It was assumed that different characteristics of
153 the contaminants and mechanisms of degradation might be responsible for the various results
154 of the effect of ZVI dosage, and interesting results of persulfate decomposition and
155 mechanism of TCA degradation would be discussed in the following sections. It should be
156 noted that ZVI dosage in the following experiments was set at 0.05 g.

157 **3.2. Trends of dissolved iron species change and persulfate decomposition in the**

158 **persulfate-ZVI systems**

159 To further confirm the roles of persulfate and ZVI in the combination system, the trends
160 of total dissolved iron and ferrous iron in the solution and the consumption of persulfate
161 during the reaction process were investigated. Two possibilities for the generation of Fe²⁺ in
162 the persulfate-ZVI system were reported: one was due to the corrosion of ZVI under both
163 aerobic and anaerobic conditions (eqs 3-4), and the other one was due to direct oxidation by
164 persulfate (eq 5).^{11, 29, 30} Furthermore, Fe³⁺ generated upon Fe²⁺ oxidation would initiate the
165 release of Fe²⁺ at the ZVI surface as presented in eq 6. As shown in Fig. 2, the concentration
166 of Fe²⁺ increased with the decomposition of persulfate and reached 502.5 mg/L at 2 h when
167 persulfate was almost exhausted, and then remained approximately constant. The same trend
168 of total dissolved iron was observed during the first 2 h. Therefore, the corrosion of persulfate
169 was mainly responsible for the production of dissolved iron in this study. Li et al.¹⁶ also
170 reported that the increase of Fe²⁺ and the total dissolved iron release rate was accompanied by
171 a gradual persulfate decomposition.



176 Since Fe²⁺ released from ZVI was believed to play a critical role in the activation of
177 persulfate by ZVI in most existing studies, a parallel test was conducted under the same
178 conditions, except that ZVI was replaced with Fe²⁺, to compare the degradation efficiency of

179 TCA by persulfate activated with ZVI and Fe^{2+} directly. The initial concentration of Fe^{2+} was
180 set at 500 mg L^{-1} (8.9 mM) corresponding to the final released Fe^{2+} amount in the
181 persulfate-ZVI system. The results were showed in Fig. S3. TCA removal was 7.4% when
182 Fe^{2+} was applied for persulfate activation, and TCA degradation only occurred within the first
183 10 minute and then stalled. This limited removal for TCA can be explain by the following
184 two reasons: (a) the destruction of $\text{SO}_4^{\bullet-}$ might occurred in the presence of excess Fe^{2+} . (b)
185 the fast reaction between Fe^{2+} and persulfate^{11,31}.

186 As illustrated in Fig. 2, persulfate decomposed gradually and almost completely
187 exhausted during the first 2 h, which corresponded well with the trends of the total dissolved
188 iron and Fe^{2+} in the solution. Li et al.¹⁶ also found a gradual and nearly complete
189 decomposition of persulfate in 3 h and the contaminant (AO7) was completely degraded after
190 2 h meanwhile when a micro-ZVI/persulfate system was applied. Moreover, some studies
191 focused on persulfate-ZVI systems demonstrated that a rapid and complete removal of
192 contaminants was achieved within a few minutes when persulfate was exhausted as well.^{11,16,}
193^{32,33} However, TCA removal was only 40% at 2 h and kept increasing when persulfate was
194 depleted in this study. Therefore, the degradation performance of TCA after 2 h was supposed
195 to be independent of persulfate chemistry. Thus, we hypothesized that a two-stage process
196 involving persulfate oxidation and ZVI reduction was contributed to TCA degradation in this
197 persulfate-ZVI system: (1) in the 1st-stage (0~2 h), TCA was removed by both persulfate
198 oxidation and ZVI reduction processes. (2) in the 2nd-stage (2~12 h), an enhanced ZVI
199 reduction was responsible for TCA degradation. To confirm the proposed stages during TCA

200 degradation, the radical scavenger tests, the chemical probe method, and the formation of
201 intermediates during TCA degradation were evaluated in the persulfate-ZVI systems.

202 **3.3. Effect of solution pH on TCA degradation performance**

203 Because the pH of groundwater is nearly neutral, TCA degradation performance under
204 various initial pH values (from pH 6 to 8) was investigated. The solutions were unbuffered
205 (adjusted with 0.1 M sulfuric acid and 0.1 M sodium hydroxide) and phosphate (0.1 M)
206 buffered, respectively. In the pH-unadjusted tests, the initial pH value of the reaction solution
207 was 3.0, which then dropped to 2.8 at 2 h (Fig. S4) and readily increased to 5.9 at 12 h. In this
208 study, persulfate was the limiting reagent and hence ZVI remained after persulfate had been
209 exhausted within 2 h. The solution pH firstly decreased due to the formation of bisulfate
210 (HSO_4^-) byproduct of persulfate and the acid intermediate products of TCA,³⁴ and then
211 readily increased after 2 h due to the hypothetical ZVI reduction process. The trend of
212 solution pH value was consistent with the results of Hussain et al.²⁹ and Liang and Lai.³²

213 As shown in Fig. 3, the influence of the initial solution pH in the unbuffered systems
214 (pH 6~8) was found to be negligible at the tested pH values. Similar with the unadjusted test,
215 the solution pH in the unbuffered systems dropped to 3.1~3.2 at 2 h, and then increased to
216 around 6.0 after 2 h. Therefore, there was no significant difference in the degradation
217 efficiency of TCA. In the buffered solutions, the pH variation was within 0.2 unit during the
218 course of tests, and the TCA degradation efficiency decreased with increasing pH. The
219 maximum removal of TCA (54.4% after 12 h) occurred at pH 6, whereas no obvious TCA
220 removal was observed at pH 8. In the pH-unadjusted and -unbuffered tests, iron corrosion and

221 Fe^{2+} generation were promoted in acid condition and hence enhanced TCA removal, which
222 would be further explained in Section 3.4. However, in the buffered solutions at neutral and
223 alkaline pHs, iron corrosion was inhibited.²⁶ In addition, the precipitation of iron species
224 occurred in the alkaline condition, and consequently inhibited TCA removal.²⁹ It should be
225 noted that the reactive oxygen species (e.g., $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$) reacted slower with phosphate
226 anions than with TCA, and the impact of phosphate on the TCA degradation performance in
227 this study is deduced to be minimal.²⁰

228 3.4. Radical scavenger tests in the persulfate-ZVI systems

229 As mentioned in Introduction, the persulfate oxidation process involving $\text{SO}_4^{\cdot-}$ or $\cdot\text{OH}$ is
230 believed to play an important role for the destruction of organic pollutants in the
231 persulfate-ZVI systems. Therefore, radical scavenger tests were conducted to identify the
232 existence of $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ during different stages in this study. Both $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ were
233 scavenged in the system using isopropanol (IPA, $k_{\text{SO}_4^{\cdot-}} = 6 \times 10^7 \text{ M}^{-1}\text{s}^{-1}$ and $k_{\cdot\text{OH}} = 2.8 \times 10^9$
234 $\text{M}^{-1}\text{s}^{-1}$), and $\cdot\text{OH}$ was scavenged by *tert*-butyl alcohol (TBA, $k_{\cdot\text{OH}} = (3.8\sim 7.6) \times 10^8 \text{ M}^{-1}\text{s}^{-1}$),
235 which is unreactive with $\text{SO}_4^{\cdot-}$ ($k_{\text{SO}_4^{\cdot-}} = (4.0\sim 9.1) \times 10^5 \text{ M}^{-1}\text{s}^{-1}$).³⁵ Firstly, IPA and TBA (300
236 mM, 2000 times more than initial TCA concentration) was introduced before reaction
237 initiation respectively. As shown in Fig. 4a, both IPA and TBA had scavenging effects for
238 TCA degradation, and IPA had a higher inhibition effect compared to TBA. The results
239 revealed that the oxidation action including both $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ was partly responsible for
240 TCA degradation. Zhao et al.³⁰ demonstrated that the reaction in persulfate-ZVI system was
241 completely quenched by adding methanol as $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ scavengers, and moderately

242 inhibited by adding TBA as $\cdot\text{OH}$ scavenger. They suggested that $\text{SO}_4^{\cdot-}$ was the predominant
243 radical species responsible for 4-chlorophenol degradation, and similar results of scavenger
244 tests were obtained by Hussain et al.²⁷ However, in this study, 83.0% of the TCA removal was
245 still achieved after 2h with the addition of IPA, and 90.1% in the presence of TBA. It was
246 speculated that other processes besides persulfate oxidation were important to induce TCA
247 degradation.

248 In order to further investigated the role of $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ in TCA degradation, the
249 scavenger tests were carried out during the hypothetical 2nd-stage. A volume of 1 mL of
250 supernatant was removed from each reaction vial and replaced with 1 mL of IPA or TBA after
251 2 h of TCA degradation in the persulfate-ZVI system (corresponding to 545 mM and 436 mM
252 for IPA and TBA, respectively), and 1 mL of water was injected into the vial for the
253 replacement of alcohols in the control test. As shown in Fig. 4b, the influences of IPA and
254 TBA were found to be negligible for TCA degradation when the alcohols added after 2 h,
255 implying that there was no $\text{SO}_4^{\cdot-}$ or $\cdot\text{OH}$ existing in the 2nd-stage when persulfate was
256 exhausted. The results further confirmed that TCA degradation during the 2nd-stage was
257 independent of persulfate oxidation process.

258 **3.5. Enhanced reduction performance in persulfate-ZVI systems**

259 To confirm the occurrence of reduction process in the persulfate-ZVI system, the
260 chemical probe method was introduced in the persulfate-ZVI system. Carbon tetrachloride
261 (CT) was often used as a probe compound for the reduction action, because CT was reported
262 to be highly resistant to both $\cdot\text{OH}$ and $\text{SO}_4^{\cdot-}$ oxidation.^{36, 37} Therefore, CT was chosen as a

263 probe compound to identify the reduction process in this study. As shown in Fig. 5a, in the
264 control test with ZVI alone, 13.4% CT was lost likely due to volatilization and/or ZVI
265 reduction. However, CT was completely removed after 12 h in the persulfate-ZVI system,
266 indicating the presence of reduction process in the system.

267 To distinguish the reduction actions between the 1st- and 2nd-stage, TCA and CT
268 degradation was investigated in the 2nd-stage separately. Firstly, the experiment started with
269 persulfate and ZVI but no TCA or CT, and then after 2 h when persulfate was nearly
270 exhausted, a volume of 1 mL of supernatant was removed from each vial and replaced with 1
271 mL of TCA or CT stock solution to generate the desired initial concentration (0.15 mM). It
272 can be seen from Fig. 5b that nearly complete CT and TCA removal was achieved after 14 h
273 when the degradation was initiated from 2 h, further revealing the presence of reduction
274 action in the 2nd-stage and the important role played by ZVI reduction in TCA degradation.

275 As discussed above, a two-stage process involving persulfate oxidation and ZVI
276 reduction was hypothesized for TCA removal. In the 1st-stage, persulfate was responsible for
277 the generation of both SO_4^- and $\cdot\text{OH}$ for the degradation of TCA. On the other hand, Fe^{2+}
278 formed by persulfate corrosion was speculated to enhance TCA reduction by ZVI. Therefore,
279 parallel tests were conducted with the addition of Fe^{2+} but without persulfate in ZVI system
280 to investigate the effect of Fe^{2+} on the reduction performance of TCA (Fig. S5). The initial
281 concentration of Fe^{2+} was set at 500 mg L^{-1} (8.9 mM) corresponding to the final released Fe^{2+}
282 amount in the persulfate-ZVI system. In pH-unadjusted test, the solution pH increased from
283 5.3 to 6.9 in the Fe^{2+} -ZVI system (data not shown), and 17% TCA was removed after 12 h.

284 Moreover, the initial solution pH was adjusted to 3.0 (with 0.1 M H₂SO₄) to simulate TCA
285 degradation in the persulfate-ZVI system, and resulting 38% of the TCA removal after 12 h,
286 confirming that the presence of Fe²⁺ in the solution could improve TCA removal.

287 The enhanced TCA degradation in Fe²⁺-ZVI system indicated that some surface
288 reactions proceeded by aqueous Fe²⁺. Li et al.¹⁷ classified the iron compounds in
289 persulfate-ZVI system into three groups as Fe metal (Fe⁰), Fe₃O₄/FeO (Fe²⁺), and
290 Fe₂O₃/FeOOH (Fe³⁺), and a two-layer structure on ZVI surface in the persulfate-ZVI system
291 was proposed, where the inner layer was goethite (α -FeOOH) and magnetite (Fe₃O₄), and the
292 outer layer was mainly consisted by hematite (α -Fe₂O₃). With the adsorption and
293 incorporation of aqueous Fe²⁺ into the hematite lattice, the hematite would convert to
294 magnetite, enhancing the conductivity of surface layer and allowing electrons to be
295 transferred from ZVI to TCA.³⁸ Therefore, the presence of Fe²⁺ in solution could enhance
296 TCA reduction by ZVI. Similar enhancement of TCE degradation was observed when Fe²⁺
297 was simultaneous with un-pretreated ZVI in aqueous solution.³⁹

298 In addition, the relationship between Fe²⁺ and ZVI could also explain the effect of ZVI
299 amount on TCA removal in Section 3.1. Since persulfate was the limiting reagent in the
300 system, the final concentration of Fe²⁺ produced from persulfate corrosion could be
301 considered as constant. With the increase in ZVI dosage, the amount of Fe²⁺ converted
302 precipitates to magnetite form was increased, and hence, the amount of Fe²⁺ adsorbed into the
303 outer layer on ZVI surface was decreased, resulting in the decrease of surface conductivity
304 and the efficiency of electron transformation. Therefore, the degradation performance of TCA

305 decreased with the increase of ZVI dosage in the persulfate-ZVI system.

306 Hence, although the persulfate chemistry involving $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ had no effect on TCA
307 removal in the 2nd-stage, persulfate was considered to enhance TCA reduction by ZVI in
308 playing roles in the solution acidification, the Fe^{2+} generation, and the formation of
309 conductive iron oxides in the persulfate-ZVI system.

310 **3.6. Mechanism of TCA degradation in the persulfate-ZVI systems**

311 The formation of intermediates, including the volatile chlorinated hydrocarbons and the
312 carboxylic acid, during TCA degradation was further investigated through GC/MS analysis
313 (Fig. 6). Since the adsorption of TCA by iron precipitates under the experimental conditions
314 was negligible (Fig. S2), a chloride mass balance in the persulfate-ZVI system was illustrated
315 in Fig. 7 as well. The theoretical value of Cl in parent TCA was used as the total amount of Cl
316 in the system (100%), which was divided into 4 parts: (1) and (2) Cl in TCA and DCA (1,1-
317 dichloroethane), i.e. the calculated percentage of Cl in TCA and DCA at the given time. (3)
318 Cl released in the solution, i.e. the percentage of Cl^- measured in the solution. (4) unknown
319 Cl, i.e. the percentage of Cl in the undetectable chlorinated intermediates or loss by TCA
320 volatilization. In our previous studies, various intermediates were confirmed from $\text{SO}_4^{\cdot-}$ - and
321 $\cdot\text{OH}$ -induced TCA degradation in the UV/persulfate and VUV/persulfate processes.^{20, 40}
322 However, as shown in Fig. 6, DCA, a confirmed byproduct of TCA by ZVI reduction,⁴¹ was
323 the only chlorinated compound identified by GC/MS in the persulfate-ZVI system. During
324 the 1st-stage, none of volatile or carboxylic acid intermediates were detected when persulfate
325 was simultaneous with ZVI. However, the unknown Cl was 9.7% at 2 h, indicating the

326 generation of other undetectable chlorinated intermediates during the first 2 h. In the
327 2nd-stage, the occurrence and accumulation of DCA was observed during TCA degradation.
328 At the end of the reaction, the concentration of DCA was 0.058 mM, and the percentages of
329 Cl released into the solution and Cl in DCA gradually increased to 51.5% and 25.3%,
330 respectively. The formation of DCA after 2 h proved the existence of ZVI reduction during
331 the 2nd-stage, but TCA dechlorination was not complete as the percentage of Cl in the
332 undetected intermediates was 21.2% in the end. It is supposed that the undetected products
333 are nonvolatile chlorinated compounds other than carboxylic acid intermediates.

334 **4. Conclusions**

335 The results in this study showed that TCA could be effectively removed by means of the
336 persulfate-ZVI system, and both persulfate oxidation and ZVI reduction were responsible for
337 TCA degradation. Increasing persulfate concentration from 1.5 mM to 9.0 mM ensured a
338 significant increase in the TCA removal, and lower ZVI dosage resulted in higher TCA
339 degradation efficiency. A two-stage process splitted at 2 h was proposed during TCA
340 degradation. The oxidation process in the 1st-stage involving both $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ was proved
341 by the radical scavenger tests, and an enhanced ZVI reduction action was confirmed in both
342 stages by the results of CT degradation performance. The presence of persulfate was
343 contributed to the generation of $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ and the enhancement of ZVI reduction. The
344 only confirmed intermediate was 1,1-dichloroethane and the chloride mass balance results
345 showed that TCA dechlorination was not complete.

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351 **Supplementary material**

352 Text S1 and Figures S1~S5 were listed in the Supporting Information.

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- 420

Figure Captions

Fig. 1 Degradation performance of TCA under various (a) persulfate concentrations (Conditions: $[TCA]_0 = 0.15$ mM, $[ZVI]_0 = 0.05$ g, 20 °C) and (b) ZVI dosages (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, 20 °C).

Fig. 2 Trends of Fe^{2+} and the total dissolved iron and persulfate decomposition during reaction time (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C).

Fig. 3 Effect of initial pH on TCA removal performance (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C)

Fig. 4 Degradation performance of TCA with the addition of IPA and TBA. IPA and TBA was added (a) before reaction and (b) after 2 h of TCA degradation (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, $[IPA]_{a0} = [TBA]_{a0} = 300$ mM, $[IPA]_{b0} = 545$ mM, $[TBA]_{b0} = 436$ mM, 20 °C).

Fig. 5 Degradation performance of TCA and CT in the persulfate-ZVI system. TCA and CT was added (a) before reaction and (b) after 2 h of the reaction between of persulfate and ZVI (Conditions: $[TCA]_0 = [CT]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C).

Fig. 6 Evolution of volatile organic intermediates during TCA degradation (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C).

Fig. 7 Chloride mass balance during TCA degradation (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C).

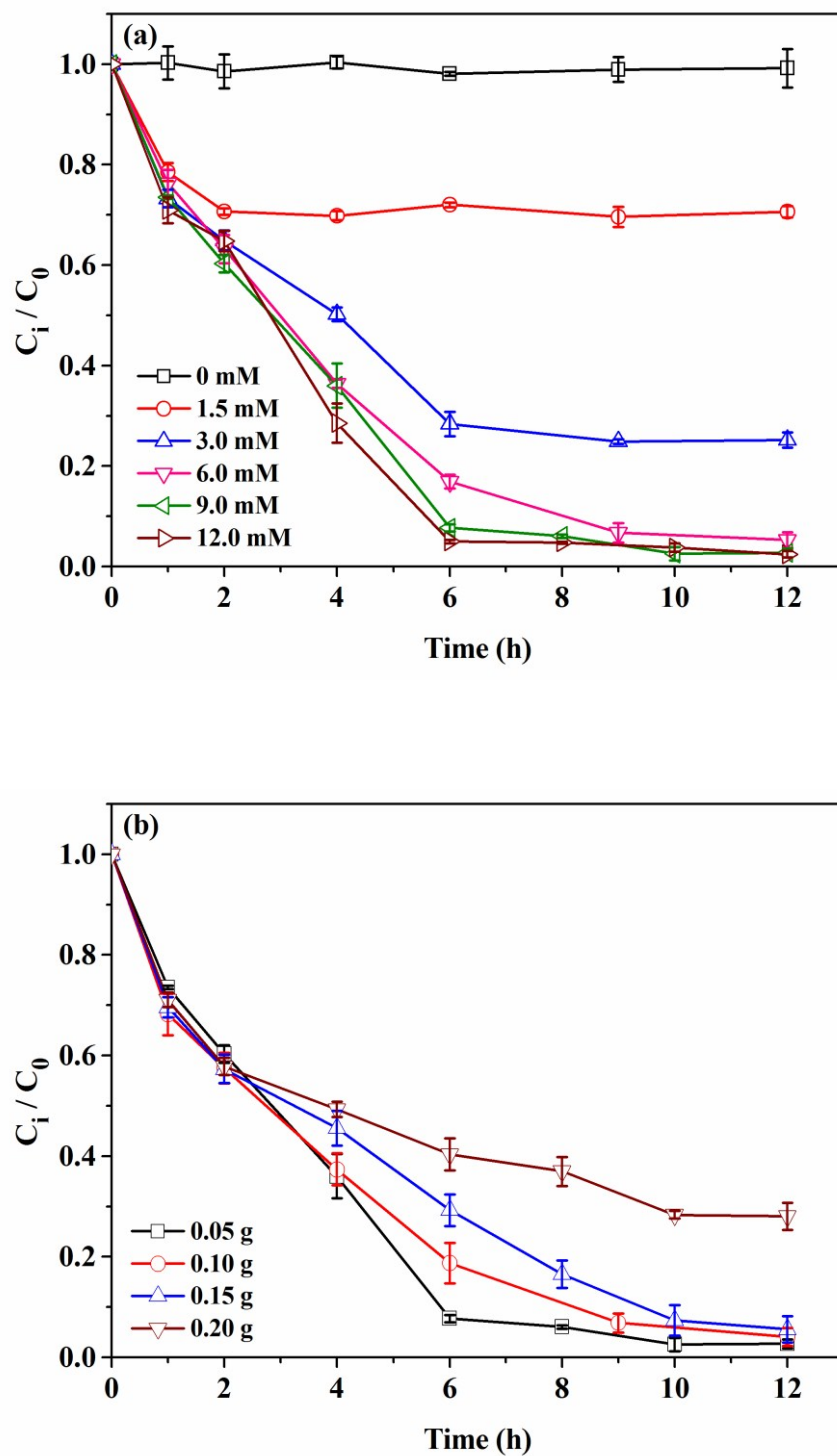


Fig. 1 Degradation performance of TCA under various (a) persulfate concentrations (Conditions: $[TCA]_0 = 0.15$ mM, $[ZVI]_0 = 0.05$ g, 20 °C) and (b) ZVI dosages (Conditions: $[TCA]_0 = 0.15$ mM, $[\text{persulfate}]_0 = 9.0$ mM, 20 °C).

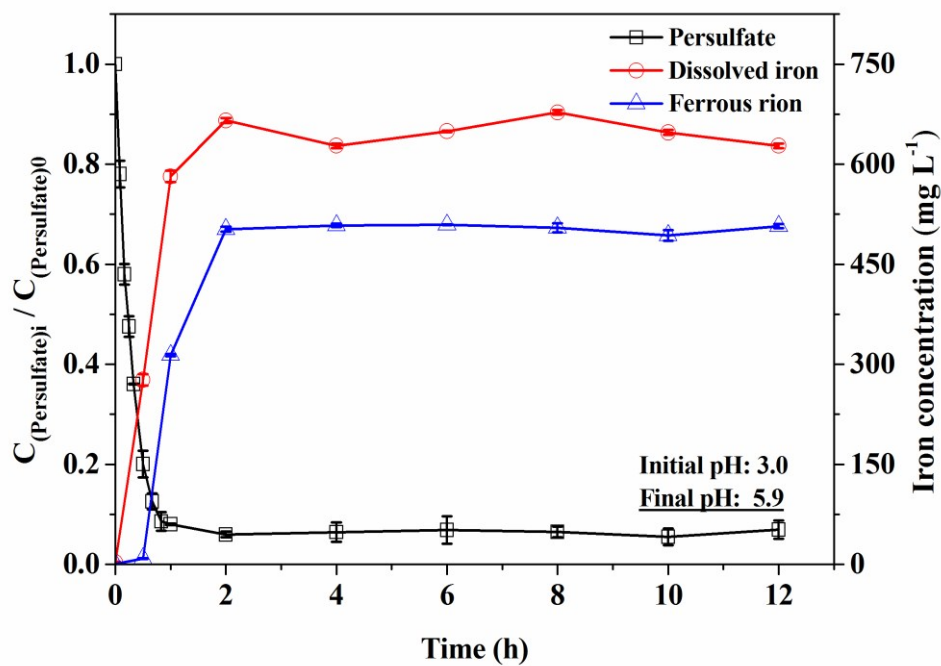


Fig. 2 Trends of Fe^{2+} and the total dissolved iron and persulfate decomposition during reaction time (Conditions: $[\text{TCA}]_0 = 0.15 \text{ mM}$, $[\text{persulfate}]_0 = 9.0 \text{ mM}$, $[\text{ZVI}]_0 = 0.05 \text{ g}$, $20 \text{ }^\circ\text{C}$).

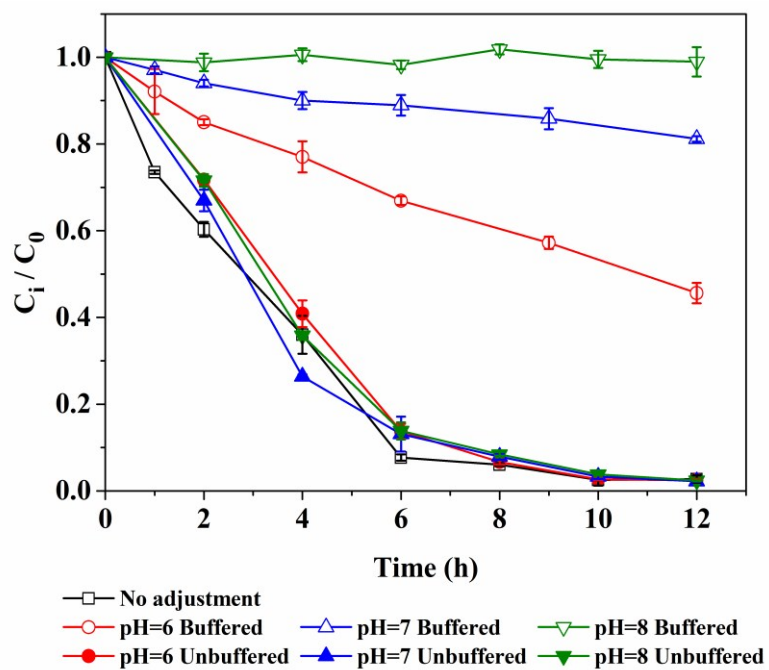


Fig. 3 Effect of initial pH on TCA removal performance (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C)

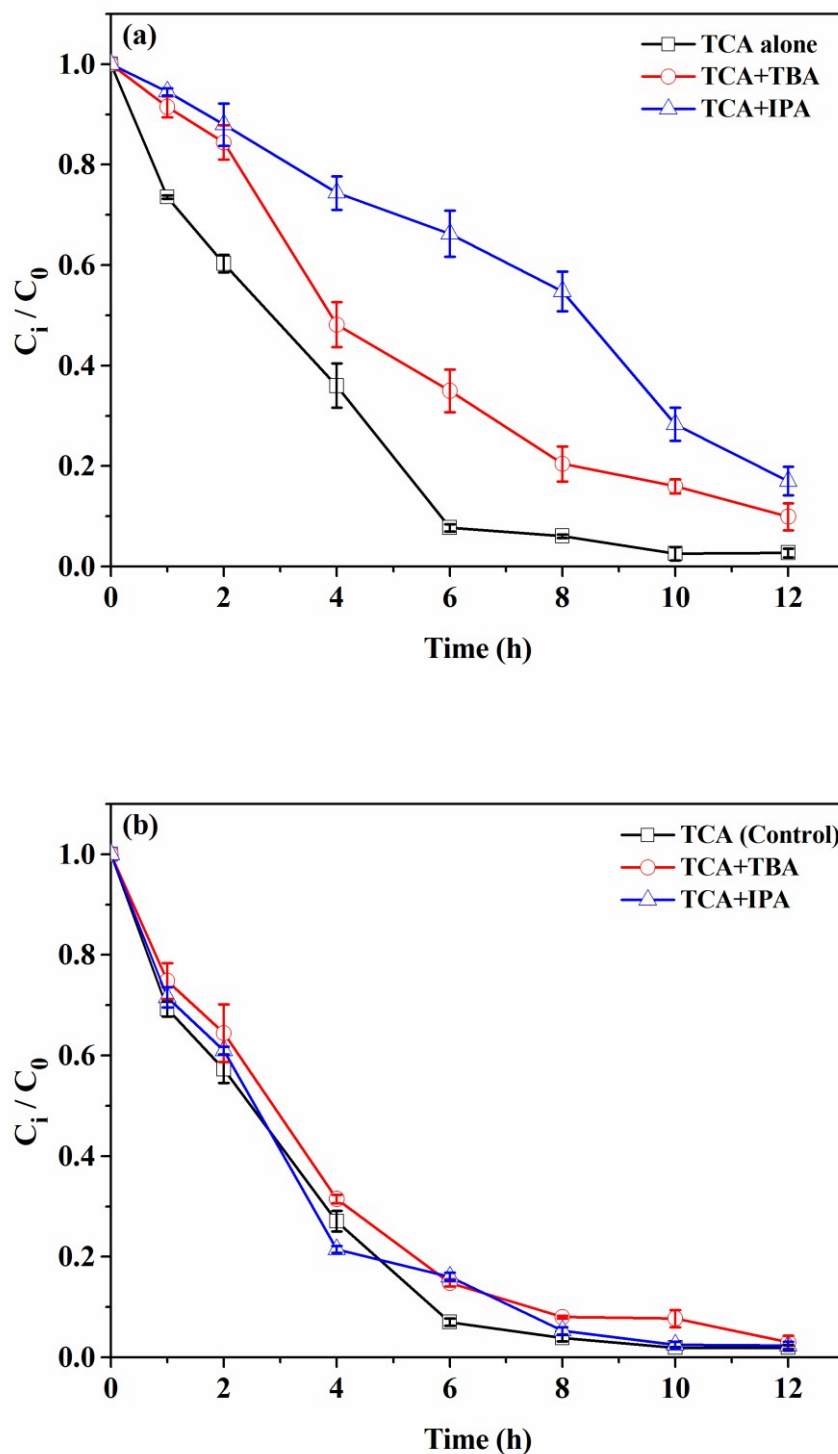


Fig. 4 Degradation performance of TCA with the addition of IPA and TBA. IPA and TBA was added (a) before reaction and (b) after 2 h of TCA degradation (Conditions: $[TCA]_0 = 0.15$ mM, $[\text{persulfate}]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, $[IPA]_{a0} = [TBA]_{a0} = 300$ mM, $[IPA]_{b0} = 545$ mM, $[TBA]_{b0} = 436$ mM, 20°C).

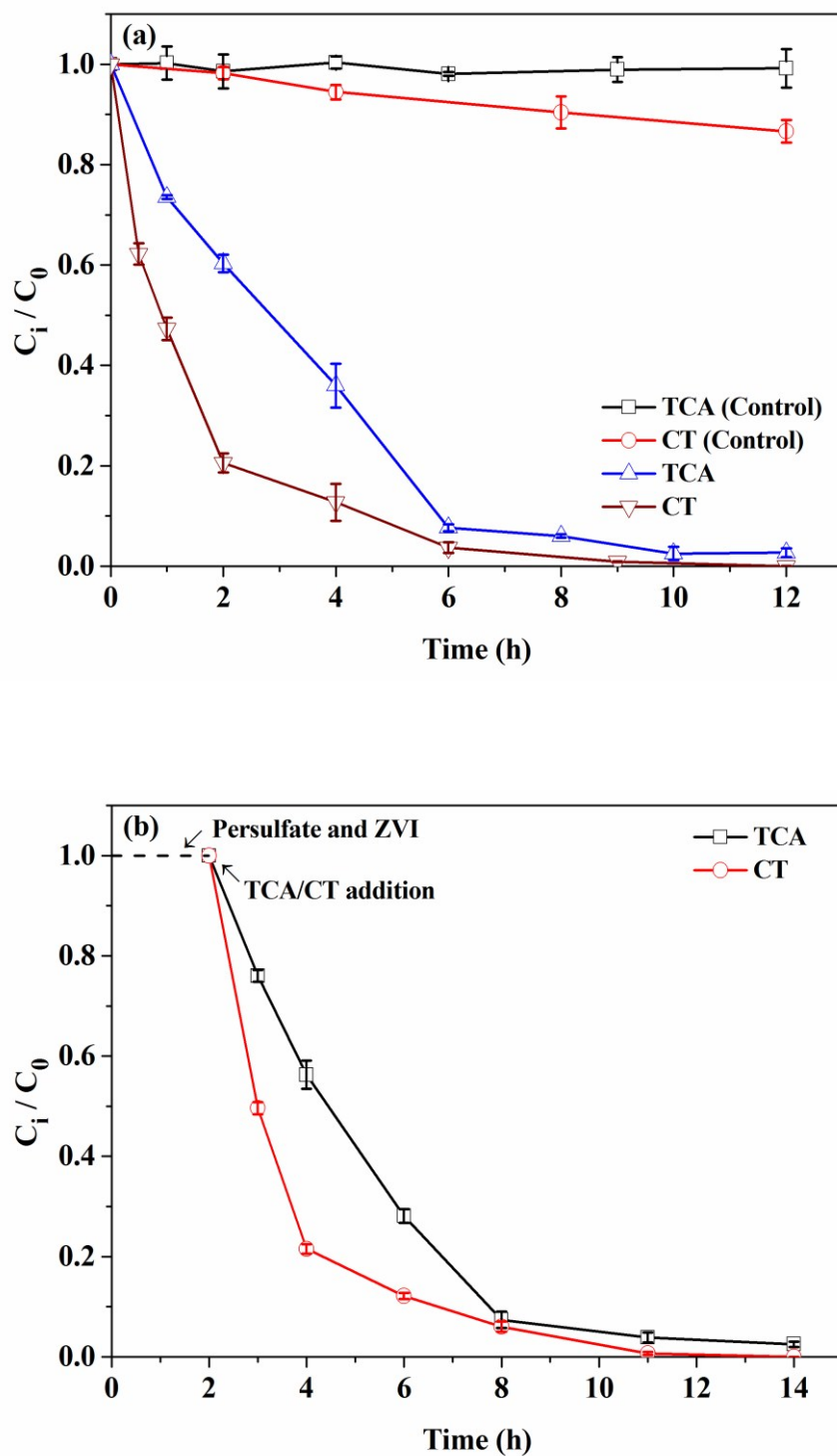


Fig. 5 Degradation performance of TCA and CT in the persulfate-ZVI system. TCA and CT was added (a) before reaction and (b) after 2 h of the reaction between of persulfate and ZVI (Conditions: $[TCA]_0 = [CT]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C).

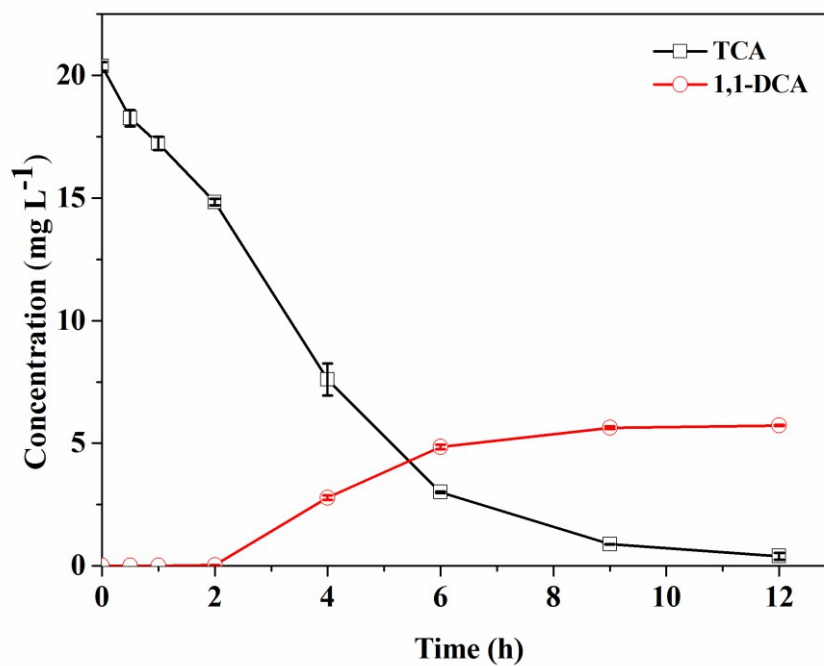


Fig. 6 Evolution of volatile organic intermediates during TCA degradation (Conditions: $[\text{TCA}]_0 = 0.15 \text{ mM}$, $[\text{persulfate}]_0 = 9.0 \text{ mM}$, $[\text{ZVI}]_0 = 0.05 \text{ g}$, $20 \text{ }^\circ\text{C}$).

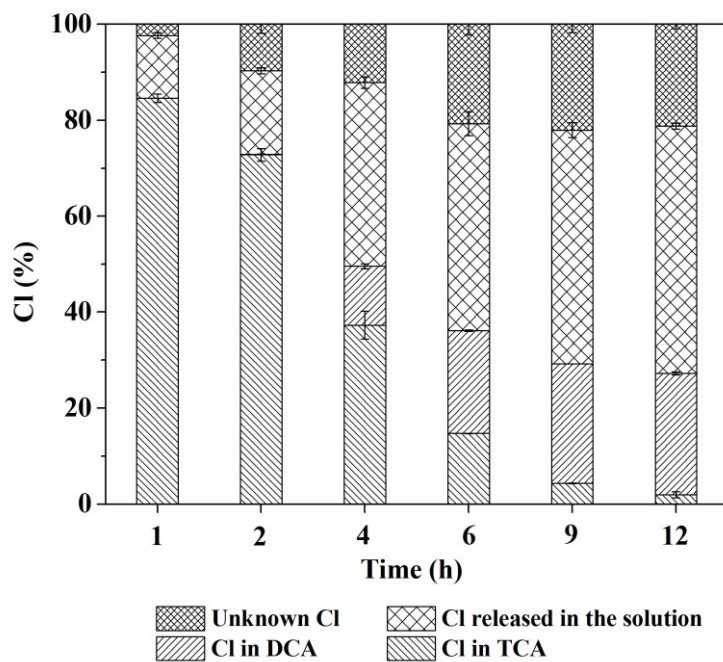
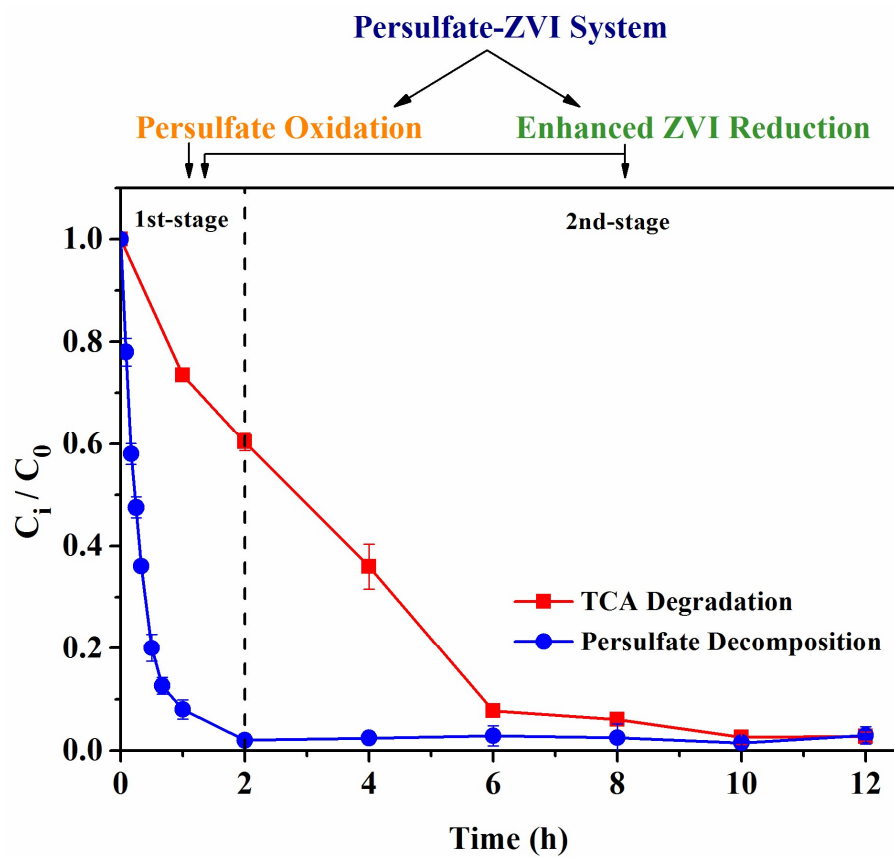


Fig. 7 Chloride mass balance during TCA degradation (Conditions: $[TCA]_0 = 0.15$ mM, $[persulfate]_0 = 9.0$ mM, $[ZVI]_0 = 0.05$ g, 20 °C).

Graphical abstract



TCA degradation performance by both persulfate oxidation and enhanced ZVI reduction in the persulfate-ZVI system