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1 **An Environmental-friendly Enzyme-based Nanofibrous Membrane for**
2 **3,3',5,5'-Tetrabromobisphenol Removal**

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11

1 Abstract

2 Chitosan(CS)/poly(vinyl alcohol)(PVA) nanofibrous membranes have inherently poor
3 mechanical strength. To improve the mechanical strength of these membranes,
4 nanocrystalline cellulose (NCC) prepared by a simplified method was added to the
5 former system. Results showed that the tensile strength of membrane with 5% NCC
6 addition was 370% higher than that of the membrane without NCC. Horseradish
7 peroxidase (HRP) was immobilized on the membrane through covalent binding with
8 HRP previously activated with 1,1'-carbonyldiimidazole, and the maximum enzyme
9 loading was approximately 384 mg/g. The physical, chemical properties of
10 immobilized HRP and its application in 3,3',5,5'-tetrabromobisphenol (TBBPA)
11 removal were examined. Results showed that HRP immobilized on CS/PVA-NCC
12 membranes showed greater stabilities and reusability than free HRP and membrane
13 without NCC. The former also exhibited an effective performance (95.9% removal, 3
14 h) for TBBPA removal under the optimum conditions (pH 7, 35 °C). Results showed
15 that HRP immobilized on NCC-incorporated CS/PVA membranes could be used to
16 remove brominated flame-retardants, especially TBBPA from wastewater. Thus, these
17 membranes have potential industrial applications.

18 **Keywords:** 3,3',5,5'-tetrabromobisphenol A, nanocrystalline cellulose, horseradish
19 peroxidase immobilization, nanofibrous membrane, electrospinning

20

1 1. Introduction

2 3,3',5,5'-Tetrabromobisphenol A (TBBPA) is a typical representative of brominated
3 flame-retardants, which are widely used in consumer products ranging from fabrics to
4 plastics and electronics.¹ The three-dimensional structure of TBBPA is illustrated in
5 Fig. S1. Previous studies have confirmed that TBBPA is a bioaccumulative, toxic, and
6 persistent compound, which could act as an endocrine disruptor, an immunotoxicity
7 mediator, and a neurotoxicity effector after a long exposure.²⁻⁴ Given the extensive
8 global use of TBBPA, it has been detected from a variety of samples, including water
9 and wastewater, indoor air, soil, and even in biological matrices. The measured
10 concentration reached an alarming line.⁵ Therefore, developing an appropriate way to
11 remove TBBPA from contaminated water is necessary.

12 Several methods, such as ozonation, adsorption, anaerobic degradation, as well as
13 oxidation, have been applied to remove TBBPA from water.⁶⁻⁹ TBBPA can be
14 degraded using biological methods with a mean half-life of approximately two
15 months.¹⁰ The degradation rate could be higher for nonbiological ones. For example,
16 96% TBBPA could be removed using multiwalled carbon as adsorbents after 60 min.⁸

17 In another case, the removal efficiency of TBBPA was as high as 99.3% through the
18 use of ozonation.¹¹ However, these methods are still greatly limited because of
19 long-cycle length, secondary pollution, high equipment costs, and difficult operations.
20 By contrast, enzyme immobilization on nanofibrous membrane is considered to be a
21 promising technique for the removal of pollutants because of its capability for high
22 efficiency (including both adsorption of nanomaterials and degradation of enzymes),

1 environmental-friendly, and reusability. Our previous studies have found that enzyme
2 immobilized on electrospun nanofibrous membranes (EFMs) have high efficiency on
3 the removal of PPCPs. For example, the removal efficiency for 2,4-dichlorophenol
4 using laccase-CS/PVA nanofibrous membranes could reach as high as 87.6% after 6
5 h.¹² In another case, immobilized HRP showed a great removal efficiency (83.5%) for
6 paracetamol and exhibited excellent reusability.¹³ Enzyme immobilization was
7 considered to be applicable for TBBPA removal because of its structural similarity
8 with the pollutants above. HRP, which has the inherent advantages of relatively wide
9 ranges of pH, temperature, contaminant concentration, and salinity, was extensively
10 studied as an efficient catalyst for the removal of phenols, bisphenols, anilines, and
11 enzidines.¹⁴ Therefore, in the present work, we choose HRP as the model enzyme for
12 the removal of TBBPA.

13 EFMs generally have the inherent disadvantage of low mechanical strength, which
14 restricts their application in industrial applications. Studies have found that the
15 incorporation of nanoscale particles with robust mechanical properties, such as
16 nanotubes, graphites, nanoclays, and inorganic nanoparticles, would increase the
17 mechanical strength of materials.¹⁵ As one of the strongest and stiffest natural
18 materials available, nanocrystalline cellulose (NCC) exhibits remarkable properties,
19 such as high-specific strength, low density, and large surface area, which lead to a
20 distinguished enhancement feature in various matrices.¹⁶ Therefore, we attempted to
21 introduce NCC into our matrix to achieve improved mechanical properties.

22 This study aimed to develop environmental-friendly CS/PVA-NCC electrospun

1 nanofibrous membranes with high mechanical strength. The materials were applied
2 for the first time to the process of HRP immobilization for the removal of TBBPA
3 from aquatic environments. Results showed that the performance of immobilized HRP
4 for the TBBPA removal was distinguished, thus this method is promising for the
5 removal of chlorophenols from water.

6 **2. Materials and methods**

7 2.1. Materials

8 Low molecular-weight CS ($M_w=20000$), PVA, TBPPA (97%), guaiacol (98%),
9 coomassie brilliant blue (G250), and 1,1'-carbonyldiimidazole (CDI; 97%),
10 fluorescein isothiocyanate (FITC) isomer were purchased from Sigma-Aldrich. Acetic
11 acid, tetraethyl orthosilicate (TEOS), microcrystalline cellulose (MCC), ethanol,
12 disodium hydrogen phosphate, citric acid, sulfuric acid, and
13 2,2'-azinobis-(3-ethylbenzthiazoline-6-sulfonate) were obtained from Sinopharm
14 Chemical Reagent Co., Ltd. Horse radish peroxidase (HRP) was obtained from
15 Sinopharm Chemical Reagent Co., Ltd. Deionized water was used in all experiments.
16 All chemicals used were of analytical grade.

17 NCC was obtained from MCC through acid hydrolysis according to the following
18 procedures.¹⁷ Five grams of MCC was mixed with 50 mL of deionized water, the
19 water/MCC-suspension was placed in an ice bath, and then 87 g of concentrated
20 sulfuric acid was added dropwise at 44 °C with gentle stirring for 2 h. Afterwards,
21 1500 mL of deionized water was quickly added to the mixture to terminate the
22 reaction. The mixture was left to stand for 3 days, during which the supernatant was

1 removed from the sediment and replaced with equal and new deionized water. The
2 suspension obtained was then centrifuged five times, each time involving repeated 5
3 min centrifuge cycles at 10000 rpm. The final centrifugate was subjected to dialysis
4 with deionized water until the wash water maintained a constant pH. The samples
5 were then sonicated in an ice bath for a certain time and naturally dried into powders.

6 2.2. Preparation of CS/PVA-NCC EFMs through electrospinning

7 CS-NCC and PVA-NCC solution were prepared by adding 1 wt% to 8 wt% NCC (dry
8 weight relative to that of dry matrix) to the two solutions containing 3wt% CS/acetic
9 acid (1 mol/L) and 10 wt% PVA, and homogenized under vigorous magnetic stirring
10 at room temperature for 4 h, followed by ultrasonication in a water bath for 2 min.
11 Approximately 2 g of TEOS was dissolved in 3 g of 70 wt% acetic acid solution in a
12 round-bottom plastic bottle. Then, 5 g of CS-NCC and 5 g of PVA-NCC were added
13 to the above solution, followed by vigorous stirring for 45 min in a water bath at
14 60 °C to form a homogeneous CS/PVA sol-gel loaded with NCC.

15 The CS/PVA sol-gel with different NCC loading was added to a plastic syringe with a
16 stainless-steel needle bearing an inner diameter of 0.8 mm. A syringe pump was set to
17 inject the emulsion at a flow rate of 1.2 mL/h. A copper pin connected to a
18 high-voltage generator was placed in the solution. Electrospinning was conducted at
19 22 kV with a tip-to-target distance of 20 cm. CS/PVA-NCC EFMs (CPN EFMs) were
20 then collected on a flat glass covered with aluminum foil for 12 h and then dried for
21 10 h at 40 °C in a vacuum to obtain non-woven fabrics. NCC-free EFMs were also
22 prepared according to our previous study to serve as the control experiment.¹⁸

1 2.3. Characterization of CPN EFMs

2 Electronic-tensile testing machine UT-2060 from U-CAN company was used to
3 examine the mechanical performance of CPN EFMs according to the determination of
4 tensile properties (GB/T 1030.3-2006). The surface morphology of the CPN EFMs
5 was characterized using scanning electron microscopy (SEM), which was conducted
6 on a field emission XL-30 SEM at 30 kV. The average diameter and diameter
7 distribution were determined by choosing 50 fibers at random SEM images and
8 analyzing them using image analysis software Adobe Photoshop CS6, developed by
9 Adobe Systems Inc. HRP labeled with FITC (HRP-FITC) was used to see whether
10 HRP has been successfully immobilized on to the nanofibers using laser confocal
11 scanning microscopy (LSCM; Leica TCS-SP5, Germany). The HRP-labeling
12 procedure was according to the paper published previously.¹⁹ The functional groups of
13 original and enzyme-immobilized nanofibers were determined using a Fourier
14 transform infrared attenuated total reflectance (FTIR/ATR, Bruker-Vector 22)
15 spectrometer that was equipped with a germanium crystal. The background spectra
16 were recorded in air. The residual concentration and activity of HRP was measured
17 through Bradford's method b using an UV-1700 spectrophotometer from Shimadzu.

18 2.4. Immobilization of HRP

19 CPN EFMs were abundant in hydroxyl groups, which could be activated using CDI to
20 form imidazole carbamate following a reaction with HRP solution. In a typical
21 immobilization experiment, 200 mg of dry composite nanofibers were immersed in
22 200 mL of anhydrous THF containing 30 mg/mL CDI. This mixture was then placed

1 in a 500 mL conical flask, which was fixed on horizontal shaker at 100 r/min and
2 25 °C for 6 h. Activated fibrous membranes were washed three times with THF to
3 remove excess CDI and reaction by-products. After the final wash, the fibrous
4 membranes were removed from solvent and washed with distilled water several times.
5 Then, 10 mg activated nanofibers was placed in 10 mL of HRP solution (2 mg/mL in
6 citric acid-disodium hydrogen phosphate buffer solution) for immobilization at 25 °C
7 under mild shaking (150 r/min). The effects of pH (4.0, 5.0, 6.0, 6.5, 7.0, 7.5, and 8.0)
8 and time (1, 2, 4, 6, 8, and 10 h) on HRP immobilization were analyzed. After enzyme
9 immobilization, the membranes were fetched out from the solution and rinsed with
10 CPBS until no soluble protein was detected in the washings.

11 2.5. Determination of free and immobilized enzymes

12 A colorimetric assay was used to determine the activity of peroxidase with reference
13 to the method of Nicell.²⁰ In a typical procedure, a suitable amount of free or
14 immobilized HRP sample (supernatant from centrifugation at 4000 r/min) was added
15 to a cuvette containing 3 mL of CPBS (0.1 M, pH 6.0) and 0.05 mL of guaiacol (20
16 mM), followed by the addition of 1 µL of 30% H₂O₂ to initiate the reaction.
17 Absorbance at 436 nm was continuously recorded with a UV-1700 spectrometer. One
18 unit of HRP activity was defined as the amount of enzyme that produced 1 µmol
19 tetraguaiacol/min under assay conditions. For data analysis, the activity results were
20 converted to relative activities-equal to the percentage of measured activity out of the
21 maximum activity.

22 To determine the effect of immobilization on the activity of HRP, the activity assays

1 were repeated using a series of H₂O₂ substrates with a concentration ranging from 0.2
2 mM to 10 mM to achieve corresponding catalytic rates, which were used to calculate
3 V_{\max} and K_m using the Lineweaver-Burk plot.²¹ Similar assays were also performed
4 with free enzymes as the control experiments.

5 2.6. Stabilities of the free and immobilized HRP

6 Thermal stabilities of the catalases were evaluated in terms of different temperature
7 effects on the activity of HRP. The activities of both immobilized and free HRP were
8 measured at pH 6 and different temperatures (4, 20, 30, 40, 50, 60, and 70 °C). The
9 pH stabilities of the catalases were examined by evaluating enzyme activity at 30 °C
10 from pH 3.0 to 10.0 for batch experiments. The operational stability associated with
11 the reusability of immobilized HRP was determined through 10 times of
12 measurements within one day under the optimum conditions. After each reaction, the
13 immobilized HRP was washed with CPBS (pH 6) to remove any residual substrate.
14 Storage stabilities of free and immobilized HRP were determined at 4 °C in CPBS
15 (pH 7) for 30 days. Residual activity was calculated every 3 days.

16 2.7. Removal of TBBPA

17 Effects of various pH values (3.0, 4.0, 5.0, 6.0, 7.0, and 8.0), temperatures (15, 20, 25,
18 30, 35, 40, 45, 50, and 55 °C), and reaction times (10, 20, 30, 60, and 120 min) on the
19 removal of 3 mg/L of TBBPA were studied. CPN EFMs, as well as free and
20 immobilized HRP, were used to treat TBPPA solution. A series of 50 mL conical
21 flasks with tightly closed screw caps were placed on a horizontal shaker at a speed of
22 150 r/min and were used in all TBBPA treatments. TBBPA concentration in the

1 reaction mixture was analyzed using HPLC (Agilent 1200) as described in the next
2 section. The amount of TBBPA that was degraded using immobilized laccase was
3 calculated with the following Eq. (1):

$$4 \quad q_D = q_0 - q_S - q_A \quad (1)$$

5 where q_D is the amount of TBBPA degraded using HRP-EFMs; q_0 is the initial
6 amount of TBBPA in solution, q_S is the amount of TBBPA retained in the solution,
7 and q_A is the amount of TBBPA that the EFMs absorbed. All treatments were
8 repeated three times, and the average values were obtained.

9 2.8. TBBPA and its analysis with HPLC

10 An HPLC instrument with two pumps, an auto sampler, and a photodiode array
11 detector was used for the analysis. TBBPA and its oxidation products were extracted
12 from the flasks using a syringe with a filter cap, and the treated solution was analyzed
13 using HPLC. Standard solutions of TBBPA were prepared to obtain a standard curve.
14 The mobile phase was composed of acetonitrile (75%) and 60 mM acetic acid
15 solution (30%), and eluted at 1.0 mL/min. UV spectra of TBBPA and reaction
16 products were obtained at 207 nm.

17 2.9. Data analysis

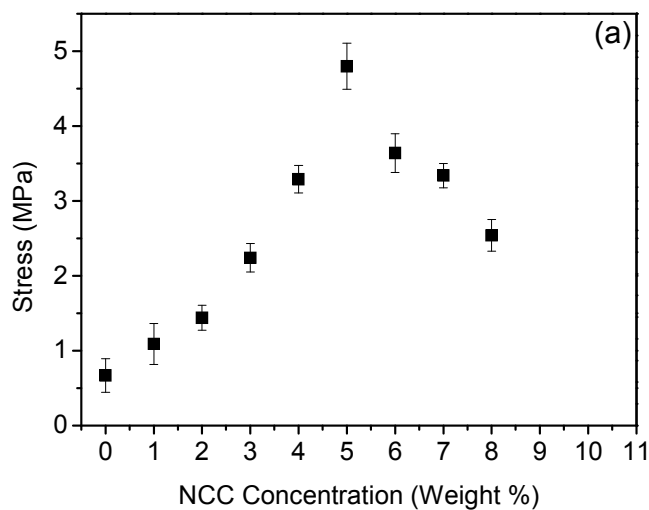
18 One-way ANOVA was applied to measure the statistical significance of the various
19 conditions (pH, temperature) for the HRP removal of TBBPA. Turkey's procedure
20 was used to evaluate differences among carriers, as well as free and immobilized HRP,
21 at a family error rate of 5%. Data were considered as significantly different from one
22 another if $p < 0.05$. Design Expert 7.0.0 was used throughout the statistical analysis.

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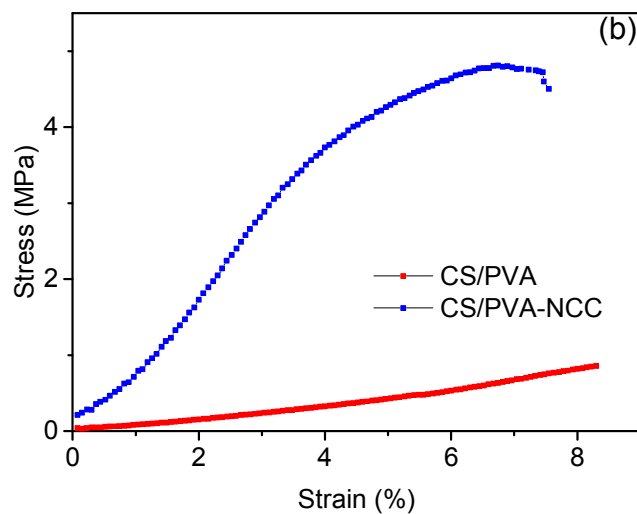
2 **3. Results and discussion**

3 3.1. Mechanical properties of CPN EFMs

4 The effect of different NCC loadings ranging from 1% to 8% (w/w) of the final dry
5 weight on the tensile strength of CPN EFMs was studied (Fig.1a). The reinforced
6 nanofibers added with 1% NCC to 5% NCC accordingly improved tensile-strength
7 values. The optimum amount of NCC incorporated for best tensile strength of CPN
8 EFMs was 4% to 5%, consistent with the previous studies.^{22,23} A typical example of
9 the stress-strain behavior of CS/PVA loading with 5% (w/w) of NCC electrospun
10 nanofibers and neat CS/PVA EFMs (CP EFMs) is shown in Fig. 1b. The average
11 tensile strength for NCC incorporated nanofibers reached 4.85 MPa compared with
12 0.85 MPa of neat CS/PVA nanofibers. The 370% tensile-strength increment of 5%
13 NCC-incorporated film compared with non-NCC film can be attributed the
14 reinforcing effect of NCC and was compatible with the former results. This finding
15 could be attributed to the strong intra- and intermolecular forces among NCC, CS, and
16 PVA, as well as the high aspect ratio of NCC itself.²⁴ The formation of the networked
17 structure above the percolation threshold, which was a result of hydrogen bonding,
18 may also contribute to the strong reinforcing effect of NCC.



1



2

3 **Fig. 1**

4 3.2. Characterization of CPN EFMs before and after HRP immobilization

5 CPN EFMs were activated through CDI, following by bio-conjugation with HRP.

6 Specifically, the free hydroxyl groups on EFMs surfaces were activated with CDI,

7 after which the amino groups of the enzymes conducted the condensation reaction

8 with the imidazolyl carbamate groups of the activated EFMs (Fig. S2). This process is

9 available when attaching HRP on the CPN EFMs, wherein the immobilized HRP

1 exhibited good stability through reutilization.

2 The morphologies of NCC and CPN EFMs with different NCC loadings were
3 characterized using SEM, the images were shown in Fig. S3 and Fig. S4, respectively.

4 As is shown in Fig. S3, NCC showed a rod-like structure and distributed unevenly,
5 this might be due to the agglomeration of the nanoparticles. We could see from Fig.
6 S4 that nanofibers with NCC loading from 4% to 5% showed best surface
7 morphology. A typical example of 5% NCC-incorporated nanofibers was shown in
8 Fig. S5. The image demonstrated that the EFMs possessed the feature of being
9 randomly arrayed and bead-free with an average diameter of 138.4 ± 20 nm (Fig. S5a).

10 The diameter (137 ± 26 nm) and structure of EFMs had no substantial change after
11 binding with HRP. Comparing Fig. S5b with Fig. S3a, we can find that the surface of
12 the CPN EFMs transformed from smooth to coarse, and many HRP molecules became
13 evenly dispersed on the nanofibers after immobilization. Fig. S3c showed that the
14 HRP-FITC had been successfully immobilized onto the nanofibers. The strong
15 fluorescence emitted by the nanofibers should be resulted from the combined effect of
16 covalently binding and physical adsorption of HRP-FITC to the nanofibers.

17 FTIR spectra of CPN EFMs before and after CDI activation were obtained to
18 characterize the EFMs (Fig. S4). Compared with the original CPN EFMs, a new peak
19 at 1653 cm^{-1} was found in the spectra of activated CPN EFMs, which was attributed to
20 C=O stretching vibrations. This change suggested that EFMs had been successfully
21 activated through CDI and could be a favorable carrier for covalent binding.

22 3.3. Effect of reaction time and pH on HRP activity and enzyme loading on CPN

1 EFMs

2 The reaction time and pH ($p < 0.05$) significantly influenced the covalent binding

3 efficiency of the enzyme and EFMs. Fig. 2 showed that the optimum pH for HRP

4 immobilization was 7.5 with a maximum HRP loading of 368 mg/g, whereas pH

5 lower or higher than 7.5 caused a distinct decrease in the loading efficiency, which

6 might be attributed to changes in the diffusion rate of the enzyme and variations in the

7 microenvironment.²⁵ We could also conclude the effect of time on the immobilization

8 efficiency based on Fig. 2. The immobilization efficiency almost leveled off after 8 h

9 of incubation, which might be because of the stability of imidazole carbamate groups

10 in hydrolysis aqueous buffer that made them react and couple with amines slowly.²⁶

11 According to Fig. 2, the loading of HRP reached a maximum of 384 mg/g after 8 h at

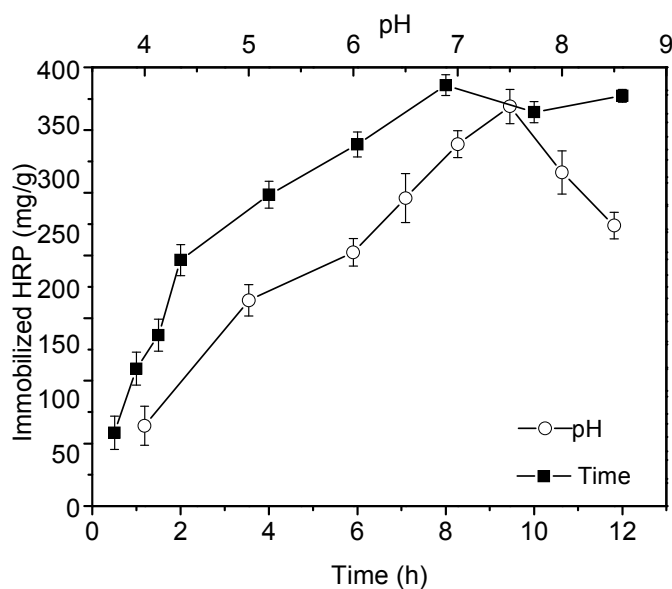
12 pH 7 and 35 °C, which was higher than other supports.^{27,28} This result may be

13 attributed to the abundant –OH groups on the backbone of the CPN EFMs, which

14 could be activated efficiently through CDI. Thus, sufficient bonding sites are provided

15 for the amino groups of HRP immobilization.

16



1

2 **Fig. 2**

3 3.4. Effects of immobilization on kinetic parameters

4 Kinetic parameters, namely, the Michaelis constant K_m and V_{max} , were measured using
 5 H_2O_2 as a substrate with Lineweaver–Burk plots. The kinetic parameters for free,
 6 HRP-CP EFMs and HRP-CPN EFMs are shown in Table 1.

7 According to Table 1, the specific activities of the HRP-CP EFMs and HRP-CPN
 8 EFMs was 81% and 69.3% of its free form, respectively. The specific activities of the
 9 HRP-CP EFMs and HRP-CPN EFMs were higher than other supports.^{29,30} These
 10 differences can be ascribed to the biocompatible characteristics of the polymers we
 11 used, which increased the accessibility between enzyme and carrier. Various methods
 12 of immobilization or different sources of HRP can also cause these differences. In
 13 contrast with HRP-CP EFMs, the higher specific activities of HRP-CPN EFMs might
 14 be resulted from the incorporation of NCC, which increased the biocompatibility of
 15 the composite membrane. Compared with free HRP, the immobilized HRP showed a
 16 significantly lower V_{max} , whereas the K_m value was significantly higher. The higher K_m

1 of immobilized HRP in this study suggested that the immobilized HRP had a lower
 2 affinity for H₂O₂ than their free form because K_m embodies the affinity of an enzyme
 3 with its substrate.³¹ This result may be attributed to the low accessibility of the
 4 substrate to the active site of the immobilized enzyme from the increased diffusion
 5 limitation. The low possibility to form a substrate-enzyme complex that arised from
 6 enzyme-conformational changes may also be an explanation for this phenomenon. As
 7 shown in Table 1, the V_{max} of the enzymes demonstrate a significant decrease upon
 8 immobilization. The reason for this low V_{max} could be explained with the steric effects,
 9 as well as bulk and diffusional effects.

10 Table 1. Specific activity, K_m , and V_{max} of the free and immobilized HRP.

HRP	Specific activity (U)	K_m ($\mu\text{mol/mL}$)	V_{max} ($\mu\text{mol}/(\text{mg}\cdot\text{min})$)
Free HRP	237.6	2.6	670.4
HRP-CP EFMs	164.7	4.1	479.2
HRP-CPN EFMs	192.4	3.7	529.3

11

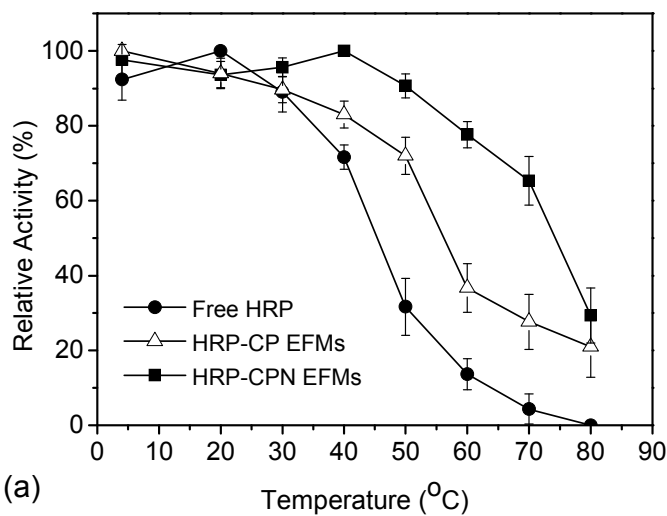
12 3.5. Stabilities of Free and immobilized HRP

13 Stabilities are of vital importance for the potential biotechnological applications of the
 14 immobilized enzymes. To further investigate whether the incorporation of NCC into
 15 CPN EFMs brought any excellent performance for HRP immobilization, comparative
 16 stabilities among free HRP, HRP immobilized on CS/PVA, and CPN EFMs (HRP-CP
 17 EFMs, HRP-CPN EFMs) in terms of thermal, operational, and storage stability were

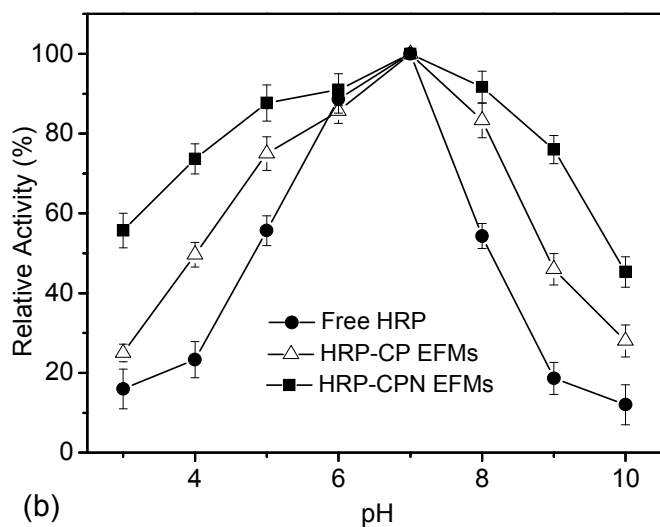
1 conducted.

2 Fig. 3 showed the effect of temperature (a) and pH (b) on the catalytic capabilities of
3 the free and immobilized HRPs. Fig. 3a shows that the relative activity of the
4 immobilized HRP changed significantly slower than its free form with the increase of
5 temperature ($p < 0.05$); the activity of HRP-CPN EFMs declined slower than that of
6 HRP-CP EFMs. Therefore, the thermal stability of three HRP kinds followed the
7 sequence of HRP-CPN EFMs > HRP-CP EFMs > free HRP. The significant
8 improvement of temperature resistance in HRP after immobilization on HRP-CPN
9 EFMs might be a consequence of great mechanical strength resulting from the
10 addition of NCC, as well as a suitable microenvironment from the framework of the
11 carrier.

12 Fig. 3b shows the effects of pH on the relative activity of free and immobilized HRPs.
13 Both free and immobilized HRPs achieved the maximum activity at pH 7, which was
14 similar to that of HRP immobilized on perlite reported using Seyed-Fakhreddin
15 Torabi.³² Within the test pH range, immobilized HRP exhibited higher activity than its
16 free form. A typical example was the case where immobilized HRP retained 75%
17 activity, whereas free HRP retained only 18% at pH 9. Compared with neat CP EFMs,
18 incorporation of NCC increased the mechanical strength and maintained the
19 membrane microstructure, which may explain the high HRP activity.



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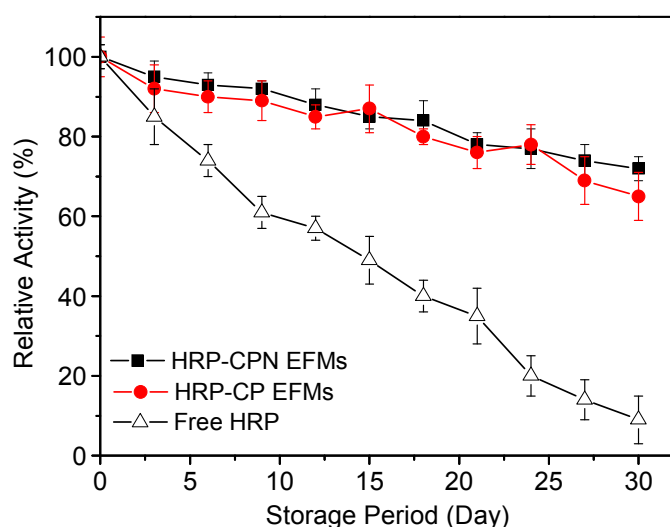


2

3 **Fig. 3**

4 Compared with immobilized enzymes, free enzymes are sensitive to the surrounding
5 environment and may easily become inactive. The superior performance of storage
6 stability is a great preponderance for immobilized enzymes, which can greatly reduce
7 the cost in their biotechnological and industrial applications. Fig. 4 shows the residual
8 activity of the free and immobilized enzymes. As the storage period increased, the
9 HRP-CPN EFMs showed a higher stability over the other two forms of HRP. The

1 relative activity of free HRP declined sharply as time passed. After 30 days, the
 2 immobilized HRP could still retain 72% (CPN EFMs) and 65% (CP EFMs) of its
 3 initial activity, whereas the free HRP only retained 9%. This phenomenon might be
 4 attributed to the limited conformational changes of the HRP, which helped retain its
 5 stability. Accordingly, the support and the methods used in the immobilization
 6 provided a long shelf life than their free counterpart and can be a preferable carrier for
 7 future applications.



8
 9 **Fig. 4**

10 3.6. Removal of TBBPA from water by HRP

11 Fig. 5 shows the influence of pH, temperature, and time on the removal of TBBPA
 12 using HRP-CPN EFMs, free HRP, and neat CP EFMs. The removal efficiency of
 13 HRP-CPNEFMs remained between 56% and 84% at 35 °C with the pH ranging from
 14 4 to 10 and between 52% and 84% at pH 7, and a temperature range of 15–55 °C.

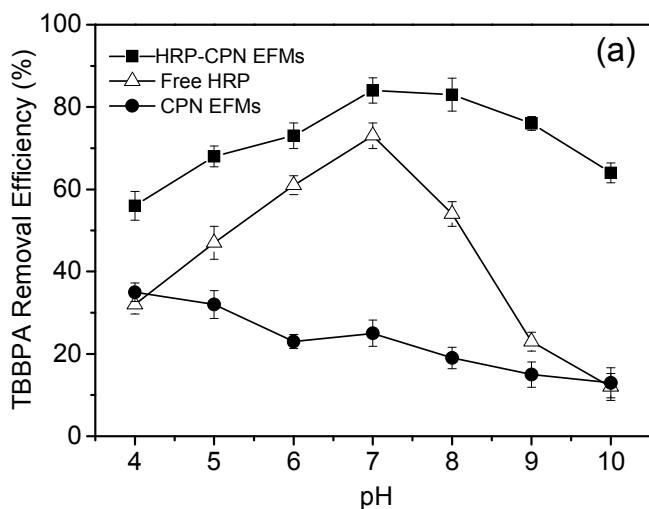
15 Fig. 5a indicates the change of TBBPA removal efficiency at pH values that varied
 16 from 4–10. The optimum pH for free enzymes was 7 with a degradation rate of 73%.

1 A high removal efficiency of TBBPA was obtained using HRP-CPN EFMs at pH 4–
2 10, and the maximum degradation rate was 84% at an optimum pH of 7. This result
3 could definitely demonstrate that HRP-CPNEFMs is preferable for the removal of
4 TBBPA from water.

5 As shown in Fig. 5b, the temperature undoubtedly affected the TBBPA removal
6 efficiency and the optimal temperature for both immobilized and free HRP to remove
7 TBBPA from water was at 35 °C with removal rates of 84% and 73%, respectively. In
8 the temperature range of 35–55 °C, the degradation rate of TBBPA using free HRP
9 decreased sharply as the temperature increased. When temperature was 55 °C, the
10 removal efficiency of TBBPA was only 21% using free HRP, whereas the removal
11 efficiency was 55% using HRP-CPN EFMs. The high thermal stability of
12 immobilized enzymes might be because of the appropriate porous structure and
13 surface characteristics of the membrane, as well as the multipoint complexation of
14 peroxidase with the support.³³

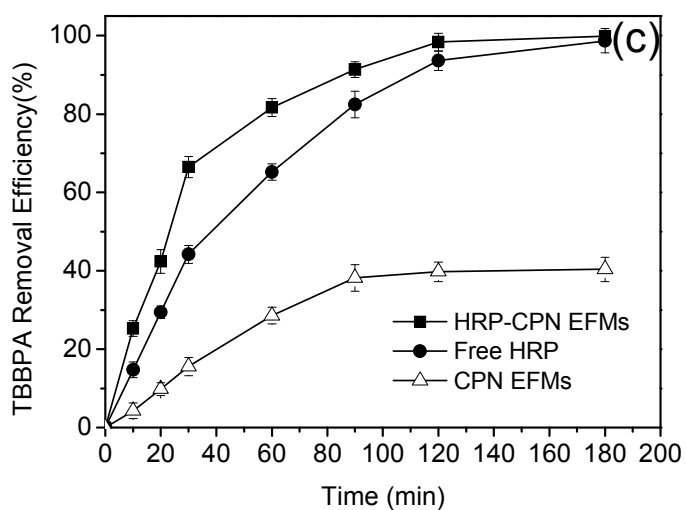
15 As shown in Fig. 5c, the degradation rate of TBBPA using the three forms could be
16 affected with time. When the time was less than 2 h, the removal efficiency increased
17 as the time increased. The degradation rate was almost leveled off after 2 h with a
18 removal rate of 98.34% (HRP-CPN EFMs), 93.66% (free HRP), and 39.8% (neat CP
19 EFMs), which might be because of the decreased concentration of TBBPA, HRP, and
20 H₂O₂ in the reaction system.³⁴ Furthermore, the generated polymer attacked the active
21 center of the enzyme and combined with the said center, thus the enzyme lost its
22 catalytic activity and lead to the decrease in its reaction rate.³⁵ We can also conclude

1 that the HRP-CPN EFMs were the most effective material among the three forms for
2 the removal of TBBPA from water. For example, the TBBPA removal efficiency using
3 HRP-CPN EFM was 66.5% after 30 min, which was significantly higher than that of
4 free HRP (44.23%) and neat CPN EFM (15.6%).



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6 s

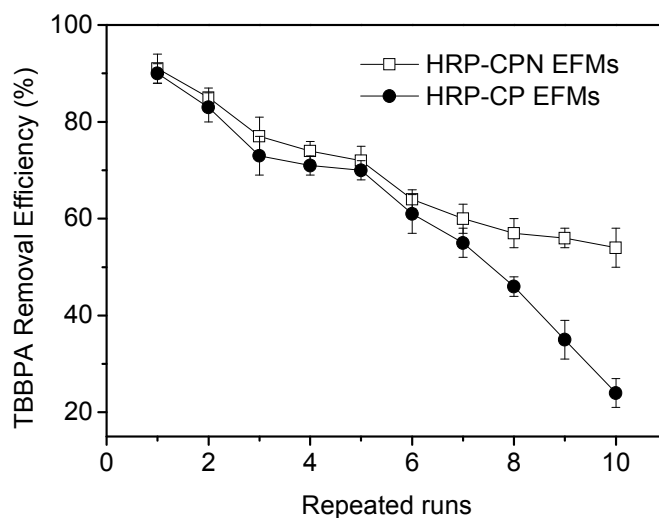


7

8 **Fig. 5**

9 3.7. Reusability of HRP-CPN EFMs

1 The main problems for free enzymes are inactivation, easy bleeding and hard to
2 separate.³⁶ The immobilized enzymes were superior to the free HRPs in terms of
3 reusability. The removal efficiency of TBBPA using HRP-CPN and HRP-CP EFMs
4 during different batch operation runs is shown in Fig. 6. The TBBPA removal
5 efficiency using immobilized enzymes decreased with the increase of reuse numbers.
6 The decrease in TBBPA removal efficiency could be explained through the loss and
7 inactivation of the enzyme, as well as the damage of the membrane. After six repeated
8 runs, approximately 60% of TBBPA could be removed through HRP-CPN and
9 HRP-CP EFMs. However, the HRP-CPN EFMs showed better TBBPA removal
10 efficiency than HRP-CP EFMs as the repeated runs increased. This result could be
11 attributed to the improvement of mechanical strength through introducing NCC into
12 the CS/PVA mixed matrix.



13

14 **Fig. 6**

15 **4. Conclusions**

16 An environmental-friendly nanofibrous membrane was fabricated. Compared with CP

1 EFMs, NCC incorporation increased mechanical strength and enzyme loading. NCC
2 incorporation also effectively maintained the catalytic activity of HRP. Thermal and
3 storage stabilities, as well as pH, were enhanced after immobilization. HRP-CPN
4 EFMs exhibited effective performance (95.9% removal, 3 h) under the optimum
5 conditions. After six repeated runs, removal efficiency remained as high as 60%. Thus,
6 these membranes have potential applications in the field of enzyme immobilization.

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15 **Figure Captions**

16 **Fig. 1** Mechanical properties of nanofibrous membranes: (a) Influence of NCC
17 content on the tensile behaviour of the membrane; (b) Stress-strain behaviour of CP
18 and CPN EFMs

19 **Fig. 2** Effects of pH and time on HRP immobilization efficiency

20 **Fig. 3** Effect of temperature (a) and pH (b) on free and immobilized HRP

21 **Fig. 4** Storage stability of free and immobilized HRP at 4 °C.

22 **Fig. 5** Effect of pH (a), temperature (b) and time (c) on TBBPA efficiency of free and

- 1 immobilized HRP.
- 2 **Fig. 6** Removal efficiency of TBBPA by HRP-CPN and HRP-CP EFMs during
- 3 repeated runs