



**Highly efficient bacterial removal and disinfection by  
magnetic barium phosphate nanoflakes with embedded iron  
oxide nanoparticles**

Journal:	<i>Environmental Science: Nano</i>
Manuscript ID	EN-ART-04-2018-000403
Article Type:	Paper
Date Submitted by the Author:	12-Apr-2018
Complete List of Authors:	<p>song, junyi; National University of Defense Technology, College of Science; University of California Santa Barbara, Institute for Collaborative Biotechnologies</p> <p>Zhang, Fan; Nanjing Agricultural University, College of Sciences; University of California Santa Barbara, Bren School of Environmental Science and Management</p> <p>HUANG, Yuxiong; University of California, Santa Barbara, Bren School of Environmental Science and Management</p> <p>Keller, Arturo; University of California, Santa Barbara, Bren School of Environmental Science and Management</p> <p>Tang, Xiaoxiu; Nanjing Agricultural University</p> <p>Zhang, wanning; Shanghai Jiao Tong University</p> <p>Jia, weibin; Nanjing Agricultural University</p> <p>Santos, Jerome; University of California, Santa Barbara</p>

### **Environmental Significance Statement**

Contamination of drinking water with bacteria is a persistent public health problem around the world. We developed a novel magnetic barium phosphate nanoflake material (FBP) for efficient bacterial removal and disinfection of drinking water. FBP exhibits much higher removal efficiency than comparable materials (nano and conventional). Importantly, it has potential for microbe control even in high salinity conditions (e.g. treated effluent). FBP can maintain a high bacterial removal efficiency after being reused for five cycles. Several types of forces and mechanisms were also discussed in detail. This work represents a key advance towards the facile synthesis and viable application of magnetic nanocomposites in treating water with high microbial contamination.

1 **Highly efficient bacterial removal and disinfection by magnetic**  
2 **barium phosphate nanoflakes with embedded iron oxide**  
3 **nanoparticles**

4 *Junyi Song*<sup>a, b, ‡</sup>, *Fan Zhang*<sup>c, d, ‡, \*</sup>, *Yuxiong Huang*<sup>d</sup>, *Arturo A. Keller*<sup>d</sup>, *Xiaoxiu Tang*<sup>c</sup>,  
5 *Wanning Zhang*<sup>e</sup>, *Weibin Jia*<sup>f</sup>, *Jerome Santos*<sup>b</sup>

6  
7 a College of Science, National University of Defense Technology, Changsha 410073,  
8 China

9 b Institute for Collaborative Biotechnologies, University of California, Santa Barbara, CA  
10 93106, USA.

11 c College of Sciences, Nanjing Agricultural University, Nanjing 210095, PR China.

12 d Bren School of Environmental Science and Management, University of California,  
13 Santa Barbara, CA 93106, USA.

14 e School of Chemistry and Chemical Engineering, Shanghai Jiao Tong University,  
15 Shanghai 200240, PR China.

16 f College of Life Sciences, Nanjing Agricultural University, Nanjing 210095, PR China.

17 ‡ These authors contributed equally.

18 \* Corresponding Author (Fan Zhang): Tel & Fax: 86-25-84396098; E-mail:  
19 [zhangfan0128@njau.edu.cn](mailto:zhangfan0128@njau.edu.cn)

1  
2  
3  
4 22       **Abstract** Magnetic barium phosphate nanoflakes with embedded iron oxide  
5  
6 23       nanoparticles, Fe<sub>3</sub>O<sub>4</sub>@Ba<sub>3</sub>(PO<sub>4</sub>)<sub>2</sub> (denominated FBP), were prepared through a facile and  
7  
8 24       inexpensive two-step process. FBP was used to purify water heavily contaminated with *E.*  
9  
10 25       *coli* (initial concentration of 5×10<sup>8</sup> CFU/mL). FBP exhibited high removal efficiency  
11  
12 26       (97%) within 30 min at 25 °C and pH 6. We investigated the effects of factors such as pH,  
13  
14 27       ionic strength, co-existing anions, temperature, contact time, material dosage, and initial  
15  
16 28       concentration of bacterial suspension, and developed optimized treatment conditions. The  
17  
18 29       negligible effect of solution ionic strength on bacterial removal efficiency of FBP  
19  
20 30       indicates its potential for microbial control even in high salinity water. Importantly, FBP  
21  
22 31       can maintain a high bacterial removal efficiency of 87% after being reused for five cycles.  
23  
24 32       FBP's magnetic properties allow an easy recovery from water. Several types of forces  
25  
26 33       and mechanisms are thought to be involved in the bacterial removal process by FBP:  
27  
28 34       electrostatic interactions, adhesion to FBP's planar surface, flocculation by polyvalent  
29  
30 35       cations on FBP's surface, oxidation sterilization from Fe<sub>3</sub>O<sub>4</sub> in FBP, irreversible cell  
31  
32 36       structural damage by FBP's edges and corners, and magnetic aggregation under a  
33  
34 37       magnetic field. Thus, FBP is a promising material for effectively treating water with high  
35  
36 38       microbial contamination.

37  
38  
39  
40 39  
41 40       **Keywords:** magnetic nanoflake; *E. coli* removal; highly salinity water; flocculation;  
42  
43 41       disinfection; recycling

## 1. Introduction

Microbial contamination of drinking water supplies is a major emerging environmental issue around the world, and even low doses of pathogenic bacteria can pose severe threats to public health, with potentially widespread impacts.<sup>1, 2</sup> As an indicator of fecal contamination, *Escherichia coli* (*E. coli*) is commonly used to evaluate and monitor water quality.<sup>3</sup> Hence, it is common to evaluate the effectiveness of disinfection using the removal efficiency of *E. coli*. There is a need to develop novel biocides, since conventional biocides can become less and less efficient and effective due to the emergence of antibiotic-resistant strains.<sup>4</sup> Recently, a number of new materials or methods have been proposed to remove bacteria from water, for example, polyoxometalate supported ionic liquid phases,<sup>5</sup> single cell imprinting on Ag-ZnO bimetallic nanoparticle modified graphene oxide sheets,<sup>6</sup> silver-modified zeolite,<sup>7</sup> cellulose nanofibers and activated carbon membranes,<sup>8</sup> pompon-like ZnO-PANI heterostructures,<sup>9</sup> electrocoagulation and electro-Fenton<sup>10</sup>. These materials showed removal and/or bactericidal properties. However, developing more sustainable materials for bacterial disinfection is still of great urgency and significance, especially with low cost and easy separation.

It has been proven that Fe<sub>3</sub>O<sub>4</sub> nanoparticles (denominated FNPs) generate reactive oxygen species (ROS) when interacting with bacteria, leading to protein oxidation and DNA damage, and finally resulting in cell death<sup>11, 12</sup>. In addition, FNPs can be easily recovered from treated water in the post-treatment using a magnetic field.<sup>1, 11, 13</sup> However, if the FNPs do not possess strong bactericidal property, the aggregates of FNPs that are

1  
2  
3  
4 67 not removed could then conceivably serve as anchoring points for further microbial  
5  
6 68 colonization, which would bring up an undesired secondary contamination. Therefore,  
7  
8  
9 69 there has been an increasing number of studies to develop advanced functionalized  
10  
11 70 magnetic nanocomposites for improving removal and disinfection of *E. coli* from  
12  
13 71 contaminated water.<sup>13-16</sup> Magnetic graphene-carbon nanotube iron nanocomposites,<sup>13</sup>  
14  
15 72 magnetic chitosan-graphene oxide composites<sup>12</sup>, anti-fimbrial modified magnetic reduced  
16  
17 73 graphene oxide nanoheaters<sup>17</sup>, and Ag-CoFe<sub>2</sub>O<sub>4</sub>-GO nano-composites<sup>18</sup> have been  
18  
19 74 reported to be effective adsorbents and antibacterial agents. However, these materials  
20  
21 75 were generally used to purify *E. coli* contaminated water with an initial concentration  
22  
23 76 range of 10<sup>2</sup>-10<sup>7</sup> colony forming unit (CFU) per mL, and their removal capacities are  
24  
25 77 limited to < 10<sup>7</sup> CFU/mg, which constrains their application range to moderately  
26  
27 78 contaminated water. Furthermore, disinfection of *E. coli* from high-salinity wastewater  
28  
29 79 using these magnetic nanomaterials is limited.

30  
31  
32  
33  
34  
35 80 In a previous study we synthesized several types of magnetic phosphate  
36  
37 81 nano-composites by a simple and inexpensive two-step method.<sup>19-22</sup> These materials  
38  
39 82 exhibited high removal efficiency for several dyes (e.g. malachite green, Congo red,  
40  
41 83 methyl blue) from water by hydrogen bonding, ionic interaction, and flocculation. Thus,  
42  
43 84 there is a potential application for removing bacteria by magnetic phosphate  
44  
45 85 nano-composites through similar interactions with different groups outside the cell  
46  
47 86 membrane.

48  
49  
50  
51 87 In this work, we applied a novel synthesis approach to prepare magnetic barium  
52  
53 88 phosphate nanoflakes with embedded iron oxide nanoparticles, Fe<sub>3</sub>O<sub>4</sub>@Ba<sub>3</sub>(PO<sub>4</sub>)<sub>2</sub>

1  
2  
3  
4 89 (denominated FBP), which possess enhanced saturated magnetization ( $M_s$ ) and surface  
5  
6 90 area compared with our previous work<sup>19</sup>. The removal efficiency of *E. coli* by FBP  
7  
8 91 nanoflakes was evaluated and compared with bare FNPs, in various environmental  
9  
10 92 conditions, including pH, temperature, reaction time, FBP dosage, and initial bacterial  
11  
12 93 concentration, as well as the effects of ionic strength and competitive anions. In addition,  
13  
14 94 the reusability of the FBP nanoflakes was examined through five consecutive bacterial  
15  
16 95 removal cycles. The interactions between *E. coli* and magnetic materials were also  
17  
18 96 investigated in detail by applying several characterization techniques.  
19  
20  
21  
22

## 23 97 **2. Experimental**

### 24 98 **2.1. Materials and reagents**

25  
26  
27 99 Ferric chloride ( $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$ ), sodium acetate (NaAc), polyethylene glycol 2000  
28  
29 100 (PEG 2000), ethylene glycol (EG), barium nitrate, cetyltrimethyl ammonium bromide  
30  
31 101 (CTAB), sodium phosphate, sodium nitrate, sodium sulfate, sodium chloride, nitric acid,  
32  
33 102 and sodium hydroxide were all purchased from Shanghai Sinopharm Chemical Reagent  
34  
35 103 Co., Ltd. (China). All chemicals and reagents were of analytical grade and used as  
36  
37 104 received without any further purification.  
38  
39  
40  
41

### 42 105 **2.2. Magnetic barium phosphate synthesis**

43  
44 106 In the first step, bare black FNPs were prepared using a solvothermal method: 1.35 g  
45  
46 107  $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$ , 1 g PEG 2000, and 3.6 g NaAc were mixed into 40 mL EG by ultrasonic  
47  
48 108 dispersion for 30 min at 30 °C. The mixed solution was transferred into two 50 mL  
49  
50 109 Teflon bottles held in a stainless-steel autoclave, sealed, and maintained at 200 °C for 5 h.  
51  
52  
53 110 After the autoclave cooled to room temperature, the powder was washed with deionized  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 111 (DI) water and then with ethanol for five times each. Then, samples were dried in a  
5  
6 112 vacuum oven at 60 °C for 3 h. In the second step, an olive magnetic FBP powder was  
7  
8 113 obtained with the similar process: First, 0.06 g FNPs, 0.01 g CTAB, and 7.5 mmol  
9  
10 114 barium nitrate were ultrasonically dispersed into 40 mL of DI water for 15 min at 30 °C.  
11  
12 115 5 mmol sodium phosphate was dissolved in 10 mL of DI water and then delivered  
13  
14 116 drop-by-drop slowly into the FNPs/CTAB/Ba(NO<sub>3</sub>)<sub>2</sub> suspension. CTAB and sodium  
15  
16 117 phosphate were used as versatile soft template and precipitant, respectively. After  
17  
18 118 mechanical stirring for 30 min, the mixed suspension was transferred into two 50 mL  
19  
20 119 Teflon bottles, sealed, heated, cooled, washed, and dried with the same procedure as for  
21  
22 120 the FNPs. At the end of the synthesis, around 1.3 g FBP were obtained at a cost of about  
23  
24 121 \$0.08/g, which indicates the economic efficiency and practicality of this material.  
25  
26  
27  
28  
29

### 30 122 **2.3. Characterization**

31  
32 123 Scanning electron microscopy (SEM) studies were performed on a FEI XL40 Sirion  
33  
34 124 FEG Digital Scanning Microscope with an Oxford energy dispersive spectroscopy (EDS)  
35  
36 125 analysis system. Transmission electron microscope (TEM) image was obtained via a FEI  
37  
38 126 Tecnai G2 Sphera 200 kV Cryo. X-ray diffraction (XRD) patterns were obtained on a  
39  
40 127 Bruker (model AXS D8, Germany) advanced XRD. Specific surface area was measured  
41  
42 128 with a N<sub>2</sub> adsorption isotherm by using an ASAP 2020 M Micromeritics instrument at  
43  
44 129 100 °C and via the Brunauer–Emmett–Teller (BET) method. The corresponding zeta  
45  
46 130 potential (ZP) values of the materials were determined via a ZP analyzer (Malvern,  
47  
48 131 Model Nano ZS, England). The magnetic performance of the materials was investigated  
49  
50 132 by a physical property measurement system (PPMS, Model-9, Quantum Design, USA).  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60



1  
2  
3  
4 133 The released concentrations of  $\text{Fe}^{3+}$  and  $\text{PO}_4^{3-}$  into the solution from materials during the  
5  
6 134 removal process were analyzed with an inductively coupled plasma-atomic emission  
7  
8 135 spectrometry (ICP-AES, Optima 2100, PerkinElmer, USA) and an ion chromatograph  
9  
10  
11 136 (Thermo, model ICS-5000), respectively.

#### 137 **2.4. Bacteria suspension preparation**

138 *E. coli* was cultivated in 100 mL of Luria Broth (LB) growth medium, consisting of 10  
139 g/L tryptone, 5 g/L bacto-yeast extract, and 10 g/L NaCl. The strain was shaken at 220  
140 rpm and 37 °C in a thermostatic incubator until the logarithmic growth late phase was  
141 reached. Cells were then separated from the growth medium by centrifugation (6000 rpm  
142 for 5 min at 15 °C). The bacterial pellets were washed three times with sterilized DI water  
143 to remove the residual growth medium and were then re-suspended in sterilized DI water  
144 to obtain bacterial stock solutions with a cell density of approximately  $5 \times 10^8$  CFU/mL.  
145 The concentration of the bacteria suspension was established using the optical density at  
146 600 nm ( $\text{OD}_{600}$ ) measured with a Shimadzu (Japan, model UV1700) ultraviolet  
147 spectrophotometer.

#### 148 **2.5. *E. coli* removal**

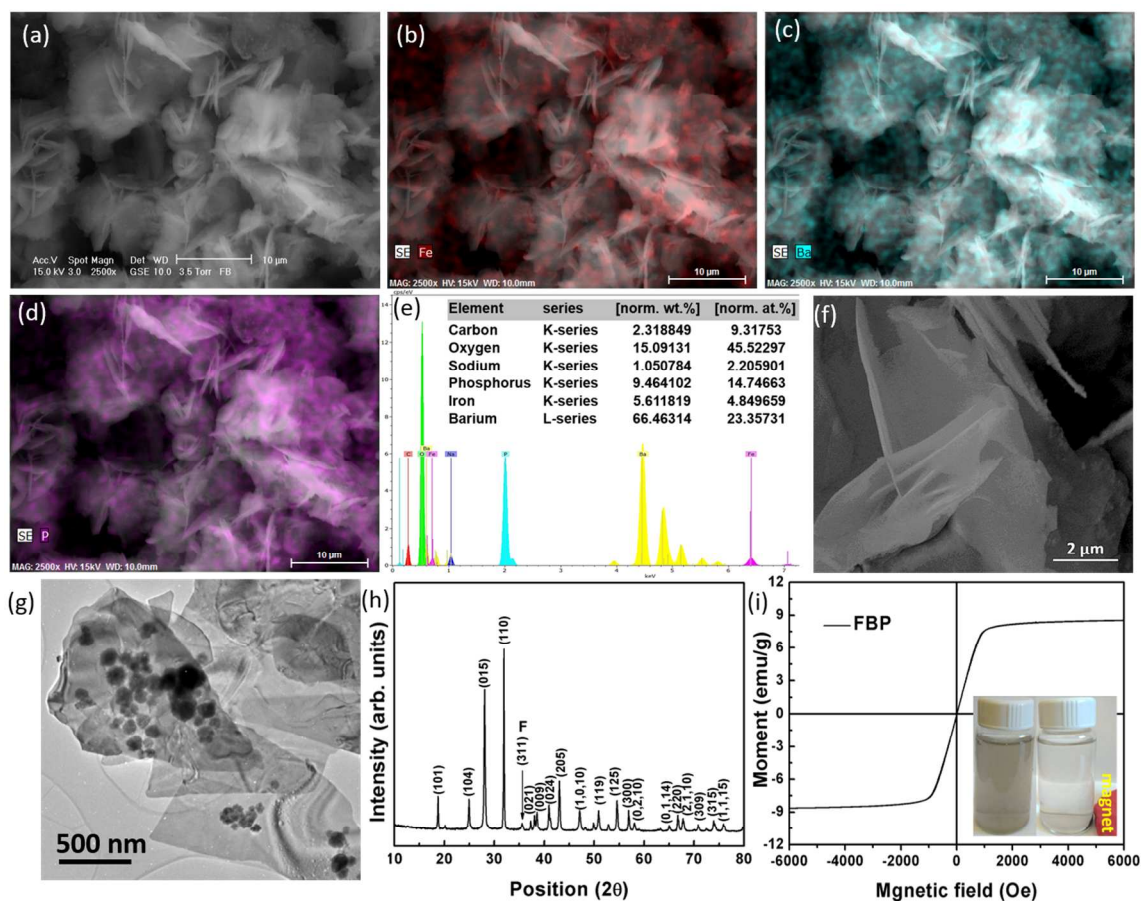
149 Generally, 0.1 g of either FBP or FNPs was added into 50 mL of *E. coli* suspension  
150 ( $5 \times 10^8$  CFU/mL) at 25 °C and pH 6.0 with mechanical stirring for 30 min to investigate  
151 the removal efficiency. The pH value was adjusted from 5 to 9 to study its effect on *E.*  
152 *coli* removal efficiency. To investigate the influence of temperature and dosage of  
153 material, the studies were done at different temperatures (10, 15, 20, 25, 30, 35, 40 °C)  
154 and dosage (0.02 g, 0.05 g, 0.08 g, 0.10 g, 0.15 g). A kinetic study was conducted by

1  
2  
3  
4 155 changing the interaction time (2, 5, 10, 20, 30, 40, 50 min). The maximum removal  
5  
6 156 capacity of the magnetic materials (FBP and FNPs) was evaluated via adjusting the initial  
7  
8 157 concentration of *E. coli* ranging from  $2.5 \times 10^8$  to  $12.5 \times 10^8$  CFU/mL. A neodymium  
9  
10 158 magnet (diameter of 50 mm, thickness of 5 mm) was used to separate the bacteria-loaded  
11  
12 159 material from aqueous solution, and the residual concentration of *E. coli* was calculated  
13  
14 160 via  $OD_{600}$ . To evaluate the reusability of FBP, the bacteria-loaded FBP particles from the  
15  
16 161 first cycle were directly added into the same bacterial suspension to start the next  
17  
18 162 treatment at the same experimental conditions. This was repeated for 5 cycles. All  
19  
20 163 experiment were repeated three times and the average and standard deviation were  
21  
22 164 reported. The interactions between *E. coli* and the magnetic materials were analyzed  
23  
24 165 using optical microscope, SEM, Confocal fluorescent images, and dilution plate count.  
25  
26 166 The related experimental methods are described in the Supporting Information.  
27  
28  
29  
30  
31

### 32 167 **3. Results and discussion**

#### 33 168 **3.1. Characterization of materials**

34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60



169

170 **Figure 1.** (a) SEM image of FBP; corresponding element mapping of (b) Fe, (c) Ba, and  
 171 (d) P in FBP; (e) EDS results, (f) SEM close-up of FBP, (g) TEM image of FBP, (h)  
 172 XRD pattern and (i) magnetic properties of FBP.

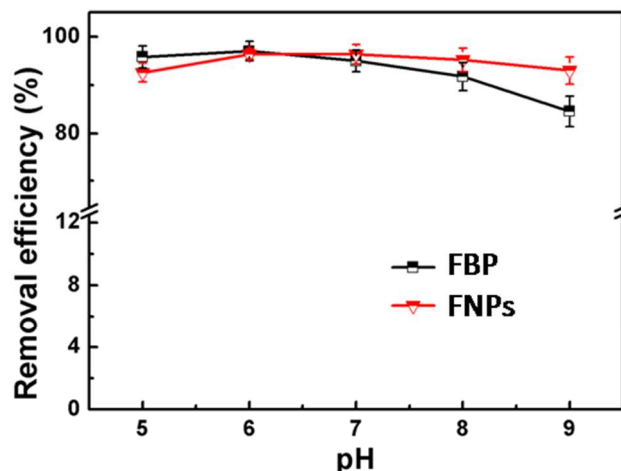
173 As seen from SEM images of FBP (Figures 1a and 1f), FBP exhibits a flake-like  
 174 structure with the sharp edges and corners. The thickness is around 40 nm and the length  
 175 is around 7  $\mu\text{m}$ , which is larger than the length of an *E. coli* cell (700 nm-1.2  $\mu\text{m}$ ).<sup>18</sup> The  
 176 plane and large flakes of FBP improve the adhesion of *E. coli* to its surface. EDS results  
 177 (Figure 1e) demonstrate the presence of Fe, with a content of 4.85% in the FBP  
 178 nanoflakes. The element mapping results (Figure 1b-d) confirm the uniform distribution  
 179 of Fe, Ba, and P in the composites. From the EDS X-ray scan, which can go to a depth of

1  
2  
3  
4 180 up to 2  $\mu\text{m}$  under the surface, the presence of Fe was uniformly distributed inside the  
5  
6 181 nanoflakes rather than simply adhering physically to the surface or blending into the  
7  
8 182 barium phosphate. This may reduce atmospheric air oxidation and acid corrosion of the  
9  
10 183 embedded FNPs within FBP.<sup>18</sup> The morphology of the FBP nanoflakes was further  
11  
12 184 examined using TEM. Figure 1g shows a thin FBP nanoflake with sharp edges and  
13  
14 185 corners, with embedded FNPs. Pure FNPs possess a spherical shape with an average size  
15  
16 186 of  $\sim 100$  nm (Figure S1). Hence, FNPs are much smaller than an *E. coli* cell, and they  
17  
18 187 may attach to the *E. coli* cell surface.  
19  
20  
21  
22

23 188 On the basis of XRD data (Figure 1h), the main characteristic peaks are attributed to  
24  
25 189 the diffraction of rhombohedral crystal planes of  $\text{Ba}_3(\text{PO}_4)_2$  (JCPDS No. 25-0028).<sup>23</sup> The  
26  
27 190 small peak marked at  $35.8^\circ$  confirmed the presence of Fe with a relatively lower  
28  
29 191 concentration in the composites. The sharp peaks revealed the fine crystallinity of the  
30  
31 192 FBP nano-flakes.<sup>23</sup> As shown in Figure 1i, the magnetic saturation ( $M_s$ ) of FBP is around  
32  
33 193 8.7 emu/g, which is higher than similar magnetic bactericides<sup>18</sup> and FBP from a previous  
34  
35 194 study with larger iron oxide particles<sup>19</sup>. A strong magnetic response results in easier and  
36  
37 195 faster post-treatment separation, as shown in the embedded picture in Figure 1i. More  
38  
39 196 than 94% of FBP can be separated from water with a small neodymium magnet within 1  
40  
41 197 min. The surface area of FBP is  $48 \text{ m}^2/\text{g}$ , which is three times higher than the surface area  
42  
43 198 of barium phosphate blended iron oxide submicron particles ( $14.8 \text{ m}^2/\text{g}$ ) reported  
44  
45 199 previously<sup>19</sup>. This can improve *E. coli* removal because FBP's larger surface area  
46  
47 200 provides more exposed active sites for adhesion.  
48  
49  
50  
51  
52  
53

### 54 201 **3.2. Removal of *E. coli* using FBP and FNPs**

55  
56  
57  
58  
59  
60



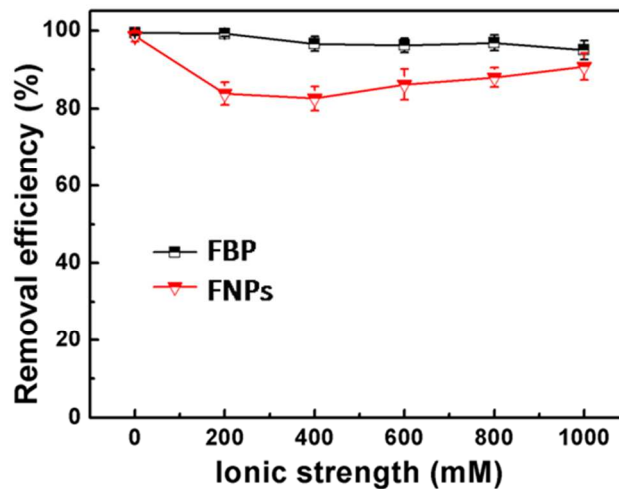
202

203 **Figure 2.** Removal efficiency of *E. coli* (initial concentration,  $C_0 = 5 \times 10^8$  CFU/mL,  
204 25 °C, 30 min) using 2.0 mg/mL of FBP or FNPs at different pH ranging from 5 to 9.

205 A series of experiments were conducted to investigate the influence of different  
206 factors such as temperature, time interval, dosage, initial concentration, and pH on *E. coli*  
207 removal using FBP. There was no substantial temperature effect on *E. coli* removal  
208 efficiency from 10 °C to 40 °C (Figure S2). Improved removal efficiency was obtained  
209 when the FBP dosage was increased from 0.02 g to 0.10 g, while increased initial  
210 concentration of *E. coli* led to decreased removal efficiency (Figure S3 and S4). On the  
211 basis of optimized treatment conditions, the removal capacity of the FBP nanoflakes  
212 ( $2.43 \times 10^8$  CFU/mg) is higher than many previous studies (Table S1), demonstrating high  
213 potential of FBP nanoflakes for treating water heavily contaminated with bacteria.

214 Figure 2 showed the effect of pH on the removal of *E. coli* by FBP and FNPs.  
215 Bacterial cells possess a negative surface charge within a pH range of 5.0–9.0,<sup>1</sup> while  
216 FBP exhibits a positive ZP at pH < 7 and the isoelectric point of FNPs is slightly above 6  
217 (Figure S5). Thus, the slightly decreased removal efficiency of FBP from 96% to 84%  
218 may be due to the change in the particle's surface charge (Figure S5), which implies that

1  
2  
3  
4 219 electrostatic interaction is the driving force in the removal process. A similar  
5  
6 220 phenomenon was also reported in a previous study with amine-modified magnetic  
7  
8 221 nanoparticles.<sup>1</sup> The negligible effect of pH on the removal efficiency of FNPs at the  
9  
10 222 tested pH range implies that the electrostatic forces between FNPs and *E. coli* were weak.

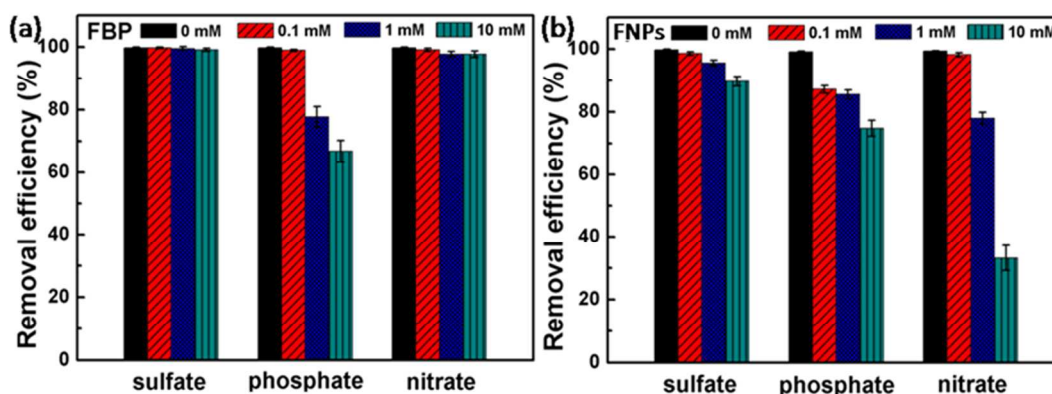


223  
224 **Figure 3.** Effect of ionic strength (0-1000 mM NaCl) on *E. coli* removal efficiency  
225 ( $C_0=5\times 10^8$  CFU/mL, 25 °C, pH 6, 30 min) with a dose of 2.0 mg/mL of FBP and FNPs.

226 The ionic strength of the suspension also influences the surface charge of bacteria  
227 and particles, and thus affects the electrostatic interaction between bacterial cells and  
228 particle surfaces.<sup>4</sup> As shown in Figure 3, more than 95% of *E. coli* cells could be  
229 removed by FBP over the entire range of ionic strength tested (0-1000 mM NaCl),  
230 indicating no significant effect was induced by changes in ionic strength. This can be  
231 explained by the positive surface charge of FBP over the entire ionic strength range  
232 examined (Figure S6) which is beneficial to the attachment of negatively charged *E. coli*  
233 onto FBP's surface via electrostatic interaction. Therefore, FBP can be applied to  
234 separate bacteria from high-salinity water.

235 However, the increase in ionic strength from 0 to 400 mM led to the significant

1  
2  
3  
4 236 reduction on removal efficiency of FNPs. This is due to a transition of the ZP of FNPs  
5  
6 237 from positive to negative values as ionic strength increases (Figure S6). The enhanced  
7  
8 238 repulsion between *E. coli* and FNPs can result in a decrease in removal efficiency. The  
9  
10 239 gradual increase in the removal efficiency of FNPs at an ionic strength higher than 600  
11  
12 240 mM (Figure S6) could be ascribed to the decrease in electrostatic repulsion between *E.*  
13  
14 241 *coli* and FNPs. A similar phenomenon was also reported for interactions between  
15  
16 242 biphasic calcium phosphate ceramic and protein.<sup>24</sup>



243  
244 **Figure 4.** Effect of anions on the removal efficiency of *E. coli* ( $C_0=5\times 10^8$  CFU/mL,  
245 25 °C, pH 6, 30 min) with a dose of 2.0 mg/mL of (a) FBP and (b) FNPs.

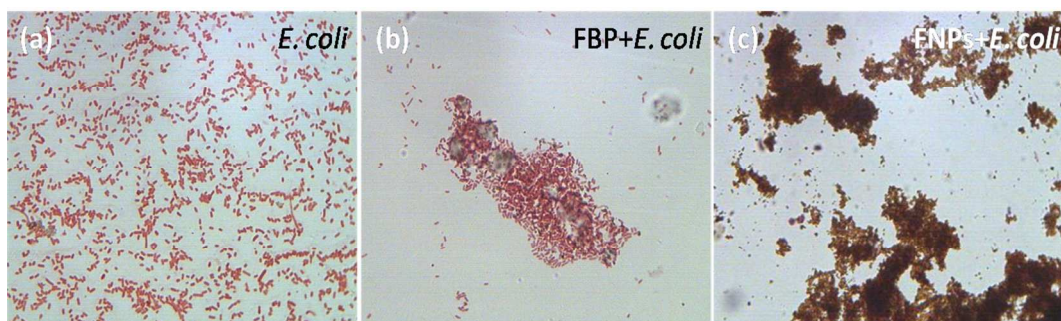
246 Anions, including sulfate, phosphate, and nitrate are commonly present in surface  
247 water and groundwater.<sup>4</sup> These coexisting anions might adsorb onto the surface of the  
248 nanomaterials and affect the bacterial attachment process.<sup>25</sup> The typical concentrations of  
249 sulfate, phosphate, and nitrate in natural water are usually lower than 10 mM.<sup>26</sup> Therefore,  
250 the effects of these anions at a concentration range of 0-10 mM on the bacterial capture  
251 efficiency by FBP and FNPs was investigated in this work. As shown in Figure 4, the  
252 presence of sulfate and nitrate did not have substantial effects on the removal efficiency  
253 of *E. coli* by FBP, while increasing the concentration of phosphate from 0 to 10 mM

1  
2  
3  
4 254 decreased FBP's *E. coli* removal ability from 98% to 67%. The ZP of bacteria was  
5  
6 255 negative under all examined conditions regardless of the types of anion or concentration  
7  
8 256 levels.<sup>4</sup> Although the ZP of FBP was influenced by the presence of the anions at  
9  
10 257 difference concentrations (Figure S7), the ZP remained positive at almost all  
11  
12 258 concentrations of sulfate or nitrate, which resulted in a consistent bacterial removal  
13  
14 259 efficiency of FBP in the presence of these anions. Given its large negative charge,  
15  
16 260 phosphate can be easily adsorbed onto the positively charged surface of FBP, leading to a  
17  
18 261 decreased number of active sites on FBP's surface. Therefore, the bacterial removal  
19  
20 262 efficiency of FBP decreased as phosphate concentration increased. For FNPs, the  
21  
22 263 increasing concentration of sulfate, phosphate, or nitrate resulted in decreased removal  
23  
24 264 efficiency of *E. coli*, especially with the addition of nitrate. The adsorbed anions on the  
25  
26 265 surface of FNPs may lead to a major shift in their surface ZP from positive to negative  
27  
28 266 (Figure S7) and enhance the repulsive energy barriers for *E. coli* to attach onto the  
29  
30 267 surface of FNPs. However, it should be noted that the typical concentrations of phosphate  
31  
32 268 and nitrate in natural water are usually below 1 mM.<sup>25</sup> Thus, FBP and FNPs would  
33  
34 269 effectively capture bacterial cells from natural water with the typical concentrations of  
35  
36 270 the coexisting anions.<sup>25</sup> In the case of co-existence of phosphate or nitrite at elevated  
37  
38 271 concentrations, high bacterial removal efficiency may still be obtained by increasing the  
39  
40 272 material dosage (Figure S3).

### 41 273 **3.3. Flocculation effect**

42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60





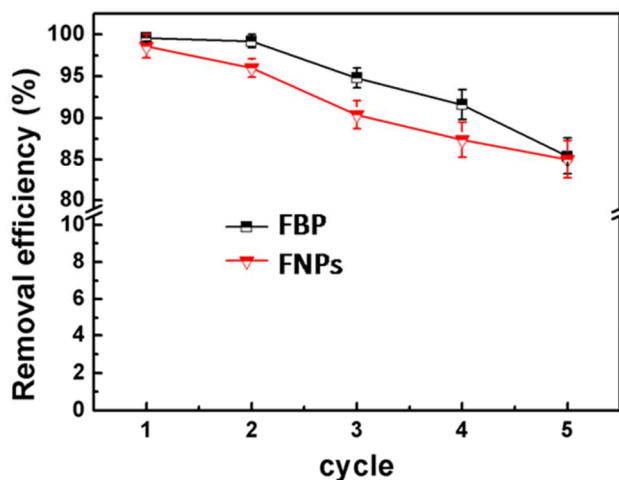
274

275 **Figure 5.** Optical microscope images of (a) pure *E. coli*; *E. coli* adhered to (b) FBP and  
276 (c) FNPs after treatment ( $C_0=5\times 10^8$  CFU/mL, 25 °C, pH 6, material dosage 2.0 mg/mL,  
277 30 min).

278 In order to understand the interaction between *E. coli* and the magnetic materials,  
279 optical microscopy was employed. The images show that pure *E. coli* cells before  
280 treatment are dispersed homogeneously (Figure 5a). After being treated with FBP (Figure  
281 5b) and FNPs (Figure 5 c), the bacterial cells aggregated substantially onto the magnetic  
282 materials. This was evidenced by the large agglomerates of bacteria (red) and the  
283 magnetic materials (black). The surface charge of bacteria is opposite to that of the  
284 magnetic materials at pH 6. Therefore, flocs of polyvalent cations on the surface of the  
285 magnetic material and the negative cell membrane form rapidly due to charge  
286 neutralization, which agrees with previous studies.<sup>4, 27</sup> Importantly, a smaller area of  
287 material and larger area of bacteria can be observed in Figure 5b compared with Figure  
288 5c, which implies that FBP is more effective in flocculating *E. coli* than the FNPs.  
289 Electrostatic interaction may aid substantially in the flocculation due to charge  
290 neutralization.<sup>27</sup> The polyvalent cations on FBP's surface can promote flocculation and  
291 accelerate the removal of bacteria.<sup>27, 28</sup> This may be due to the fact that FBP nanoflakes  
292 possess more polyvalent cations and more positively charged surfaces than FNPs. It

293 should be noted that pure  $\text{Ba}_3(\text{PO}_4)_2$  nanoflakes showed a lower removal efficiency (61%)  
294 for *E. coli* (Figure S8) than FBP though they shared a similar shape and surface area.<sup>23</sup>  
295 Therefore, the flocculation effect from the iron in FBP (e.g. iron nanoparticles, iron ions,  
296 iron oxides, and hydrolysis products) contribute to *E. coli* removal.<sup>27, 28</sup> Furthermore,  
297  $\text{PO}_4^{3-}$  groups associated with FBP can combine with  $\text{H}^+$  to help release  $\text{Fe}^{3+}$  and  $\text{Fe}^{2+}$ ; this  
298 can increase the flocculation effect of FBP as hydrolysis products (e.g.  $[\text{Fe}(\text{OH})]^{2+}$ ,  
299  $[\text{Fe}(\text{OH})_2]^+$ ,  $[\text{Fe}(\text{OH})]^+$ ) can also produce agglomeration of microorganisms<sup>27</sup>.

### 300 3.4. Reuse of FBP and FNPs



301

302 **Figure 6.** Removal efficiency of *E. coli* at different cycles ( $C_0=5 \times 10^8$  CFU/mL, 25 °C,  
303 pH 6, 30 min) with an initial dosage of 2.0 mg/mL FBP and FNPs.

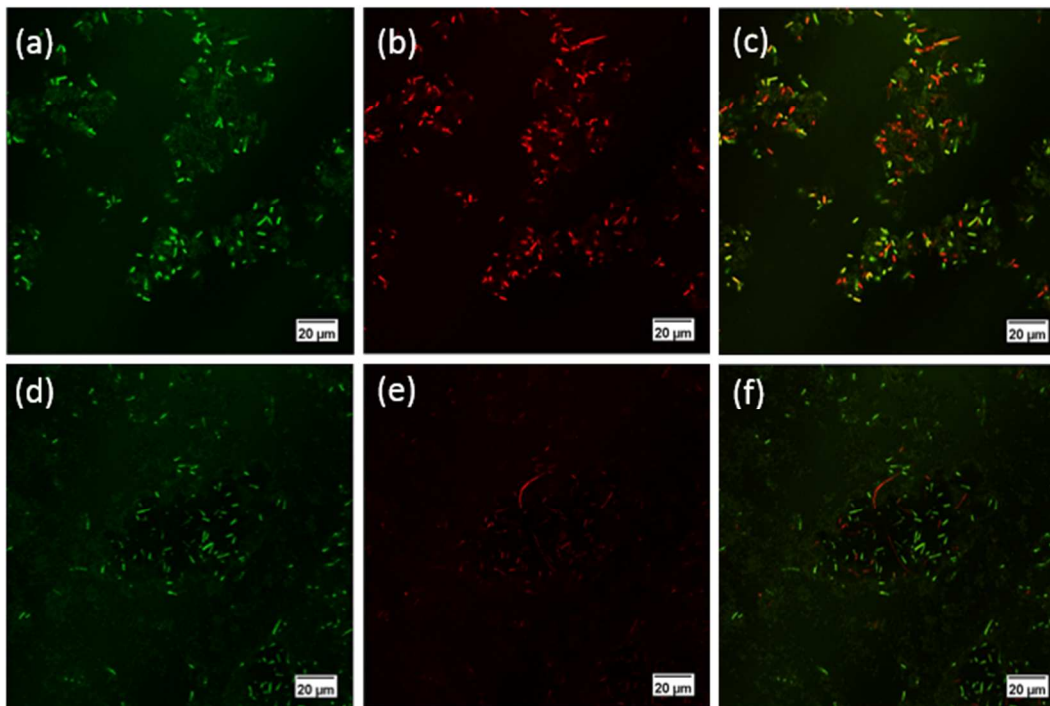
304 The recovery and reuse of FBP and FNPs throughout five consecutive bacteria  
305 removal experiments are illustrated in Figure 6. It can be observed that the *E. coli*  
306 removal efficiencies of both FBP and FNPs were more than 90% in the first cycle, and  
307 then decreased slowly after each cycle. This tendency can be explained by the fact that  
308 the materials were covered by more bacteria after additional cycles, which can result in  
309 decreased surface charge from positive to negative values (Figure S9) and less sites

1  
2  
3  
4 310 available for other bacterial attachment. However, FBP still removes around 87% of *E.*  
5  
6 311 *coli* after five cycles and FNPs remove 86% under the same conditions. Thus, the  
7  
8 312 electrostatic interaction may be not the dominant mechanism for the high removal  
9  
10  
11 313 efficiency during the reuse.

12  
13 314 The chemical stability of the magnetic materials and the retention of adsorbed  
14  
15 315 bacteria are significant in the reuse process to avoid secondary pollution (from the release  
16  
17 316 of components of the FBP, FNPs or bacteria) to the treated water.<sup>18</sup> On the basis of  
18  
19 317 ICP-AES and ion chromatography analysis, there was no significant release of Fe<sup>3+</sup> or  
20  
21 318 PO<sub>4</sub><sup>3-</sup> to the solution from the magnetic materials during five cycles, which demonstrates  
22  
23 319 the chemical stability of FBP and FNPs in the removal process. After bacterial treatment,  
24  
25 320 the FBP and FNPs were rinsed with 50 mL of sterilized DI water three times to determine  
26  
27 321 whether the captured bacteria on the surface of these materials could be released back  
28  
29 322 into the environment, resulting in secondary pollution. On the basis of dilution plate  
30  
31 323 count results (Figure S10), FBP that had been loaded with bacteria released less live *E.*  
32  
33 324 *coli* (0.48%) into solution than FNPs (0.96%). This may be due to the stronger  
34  
35 325 electrostatic binding and flocculation by FBP or better disinfection effect by FBP.  
36  
37 326 Furthermore, after being rinsed ultrasonically for 5 min, FBP acquired a positive surface  
38  
39 327 charge again, which implies that the captured cells were washed off and the materials  
40  
41 328 could be reused.

### 42 329 **3.5. Disinfection effect**

43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60



330

331 **Figure 7.** Confocal fluorescent images of live and dead bacterial cells treated with FBP  
332 (a–c) and FNPs (d–f) under mechanical stirring and stained with SYTO9 (green) and PI  
333 (red). (c, f) Overlaying images of *E. coli* stained with SYTO9 (live and dead) and PI  
334 (dead).

335 In order to confirm whether the cells were affected by the magnetic materials in a  
336 reversible manner or not, fluorescent-based cell tests of live and dead bacterial cells were  
337 conducted (Figure 7). The strong green fluorescence intensity region (Figure 7a and d)  
338 represents the proportion of live and dead bacteria, and the heavy red fluorescence  
339 intensity region (Figure 7b and e) indicates the dead fraction.<sup>18</sup> It can be clearly observed  
340 from Figure 7b that most of the bacteria adhered to FBP's surface are dead, indicating  
341 irreversibly damaged cell membranes or mass cell death upon exposure to FBP.<sup>3</sup>  
342 However, most of bacteria cells were still alive after being treated with FNPs (Figure 7d  
343 and e).

1  
2  
3  
4 344 Removal kinetics are shown in Figure S11 and S12. Maximum removal was  
5  
6 345 achieved after 30 min, with no substantial improvement with longer times. The kinetics  
7  
8 346 behavior of the disinfection process was determined by considering the disinfection at  
9  
10 347 different time intervals; the confocal fluorescent images were used to evaluate the ratio of  
11  
12 348 live and dead bacterial cells treated with FBP and FNPs. As shown in Figure S13, around  
13  
14 349 30%, 70%, and 90% of bacteria were inactivated after treatment with FBP under shaking  
15  
16 350 for 5 min, 10 min, and 30 min, respectively. However, FNPs showed a lower disinfection  
17  
18 351 ratio (less than 40%) after 30 min (Figure S14) under shaking compared with FBP under  
19  
20 352 shaking (Figure S13).

21  
22  
23  
24  
25 353 FBP under mechanical stirring (Figure 7) or shaking (Figure S13g-i) generated a  
26  
27 354 similar disinfection effect, implying that the stirring paddle itself was not likely the cause  
28  
29 355 of cell death. Furthermore, without any mechanical mixing, FBP resulted in much lower  
30  
31 356 disinfection (less than 20%) (Figure S15) than FBP under mechanical stirring (Figure 7)  
32  
33 357 or shaking (Figure S13). This implies that the mechanical mixing process can improve  
34  
35 358 FBP's physical bactericidal action because it allows more bacteria to contact the sharp  
36  
37 359 FBP nanoflake edges. The pure cations ( $\text{Fe}^{2+}$ ,  $\text{Fe}^{3+}$ , or  $\text{Ba}^{2+}$ ) showed no noticeable  
38  
39 360 toxicity to *E. coli* (Figure S16). Once the cell membrane is broken by FBP, ROS or some  
40  
41 361 cations in solution can react with protein, DNA, or other matters inside the cell, leading  
42  
43 362 to cell death.

44  
45  
46  
47  
48  
49 363 Dilution plate count assay was conducted to measure cell proliferation after treatment  
50  
51 364 with FBP and FNPs (Figure S17). Cell proliferation was weakest for the treatment using  
52  
53 365 FBP under stirring, which is consistent with confocal fluorescent results. To further  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 366 investigate the underlying mechanism and inspect the bacterial morphological change, *E.*  
5  
6 367 *coli* cells adhered with FBP and FNPs were visualized using SEM. Compared with pure  
7  
8  
9 368 *E. coli* (Figure S18a), most of the bacterial cells fused onto the FBP surface with an  
10  
11 369 irregular shape (Figure S18b) which seemed to exhibit a plasmolysis phenomenon<sup>18, 28</sup>.  
12  
13 370 Only a small number of *E. coli* cells were found fixed onto the FBP plane (Figure S18c  
14  
15  
16 371 and d). But some of the cell membranes were deformed from the normal rods to round  
17  
18 372 shapes. Similar results were seen in other antibacterial materials, such as magnetic  
19  
20 373 chitosan-graphene oxide composite and graphene-based antibacterial paper<sup>12, 29</sup>. The  
21  
22  
23 374 leakage of intracellular constituents and change of morphology may be due to altered cell  
24  
25 375 permeability resulting from the interaction between positively charged FBP and the  
26  
27 376 negatively charged membrane of cells<sup>30</sup>. Moreover, iron oxide nanoparticles have been  
28  
29  
30 377 shown to generate ROS when interacting with bacteria, resulting in protein oxidation and  
31  
32 378 cell death<sup>31,32</sup>. Additionally, *E. coli* cells were observed to be divided into two parts after  
33  
34  
35 379 contacting the sharp FBP edge (Figure S18e and f), demonstrating physical damage.

36  
37 380 It was interesting to note that the FNPs aggregated and adhered to some of the *E. coli*  
38  
39 381 cells, forming pits on the bacterial surface and leading to irregular shapes (Figure S18g-i).  
40  
41  
42 382 This implies that adhesion of FNPs onto the surface of *E. coli* may cause destruction of  
43  
44 383 cell integrity, similar to other materials<sup>13,33</sup>. Although the FNPs may lead to disinfection  
45  
46  
47 384 via oxidation, the small spherical structure of FNPs with low positive charge may result  
48  
49 385 in weaker adhesion, piercing, or other physical action to bacteria cells compared with  
50  
51 386 FBP. Although the FNPs may lead to disinfection via oxidation, the small spherical  
52  
53  
54 387 structure of FNPs with low positive charge may result in weaker adhesion, piercing, or

388 other physical action to bacteria cells compared with FBP.

#### 389 **4. Conclusions**

390 In this study, magnetic FBP nanoflakes were synthesized, to be used for disinfection  
391 and removing bacteria from contaminated waters. The bacterial removal efficiency of  
392 FBP was systematically studied compared with bare FNPs. Results showed that FBP  
393 could remove 97% of *E. coli* at the optimized experimental conditions (initial  
394 concentration of  $5 \times 10^8$  CFU/mL, pH=6, 25 °C, reaction time of 30 min, material dosage  
395 of 2 mg/mL). FBP was still an effective disinfectant even in high salinity waters and in  
396 the presence of common anions at concentrations typical of most water sources.  
397 Moreover, 87% of *E. coli* cells could be removed by FBP even in the fifth reuse cycle,  
398 demonstrating good reusability. FBP performs better as a disinfectant than FNPs due to  
399 its composition and morphology: (1) the higher positive charge of FBP due to the  
400 addition of barium phosphate to FNPs improves electrostatic interactions, altering cell  
401 permeability and inducing plasmolysis; (2) the special two-dimensional flake structure of  
402 FBP with sharp edges and corners can induce the physical disruption of the bacterial cell  
403 membrane, releasing cellular components, and resulting in cell death, similar to magnetic  
404 graphene<sup>34</sup>. The removal and disinfection mechanisms of FBP include: electrostatic  
405 adhesion onto surface, flocculation effect of polyvalent cations, oxidation inactivation by  
406 Fe<sub>3</sub>O<sub>4</sub> in FBP, irreversible cell damage by FBP's sharp structure. Cell damage from the  
407 sharp edges was confirmed the low disinfection when FBP is added without mechanical  
408 mixing or stirring; shaking or stirring of FBP resulted in similarly high disinfection via  
409 cell damage. Generally, taking into account the easy preparation, magnetic separation,

1  
2  
3  
4 410 high removal efficiency, effective disinfection, and good reusability, magnetic FBP  
5  
6 411 nanoflakes are an ideal material for water disinfection.  
7

8  
9 412 **Conflicts of interest**

10  
11 413 There are no conflicts to declare.  
12

13  
14 414 **Acknowledgements**

15  
16 415 This work was financially supported by the Jiangsu Natural Science Funds of China  
17  
18 416 (Grant No. SBK2017020336), the Fundamental Research Funds for the Central  
19  
20 417 Universities (Grant No. KYZ201747), and the National Natural Science Funds of China  
21  
22 418 (Grant No. 51402153). This work was in part supported by the National Science  
23  
24 419 Foundation (NSF) and the U.S. Environmental Protection Agency (EPA) under  
25  
26 420 NSF-EF0830117. Any opinions, findings, conclusions or recommendations expressed in  
27  
28 421 this material are those of the authors do not necessarily reflect the views of the funding  
29  
30 422 agencies.  
31  
32  
33

34  
35 423 **Electronic supplementary information**

36  
37 424 (1) Characterization experiment methods: Optical microscope, Fluorescent-based cell  
38  
39 425 live/dead test, SEM of *E. coli* with material, and Dilution plate count method; (2) Effects  
40  
41 426 of temperature, interaction time, dosage, and initial *E. coli* concentration on bacterial  
42  
43 427 removal; (3) Comparisons of removal capacity for *E. coli* by per mg material in this work  
44  
45 428 with other recent reports; (4) Zeta potentials of FBP and FNPs at different pH, ionic  
46  
47 429 strength, co-existing anions, and cycle; (5) Bacterial removal effect of rare barium  
48  
49 430 phosphate nanoflake; (6) Dilution plating procedure results; (7) Kinetics of removal and  
50  
51 431 disinfection process by FBP and FNPs; (8) Confocal fluorescent images of *E. coli* cells  
52  
53  
54  
55  
56  
57  
58  
59  
60



1  
2  
3  
4 432 treated in solutions containing metal ions leached from FBP; (9) Cell proliferation after

5  
6 433 treatment by FBP and FNPs, and (10) SEM images of *E. coli* with FBP and FNPs

7  
8  
9 434 treatment after magnetic separation.

10  
11 435  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60436 **References**

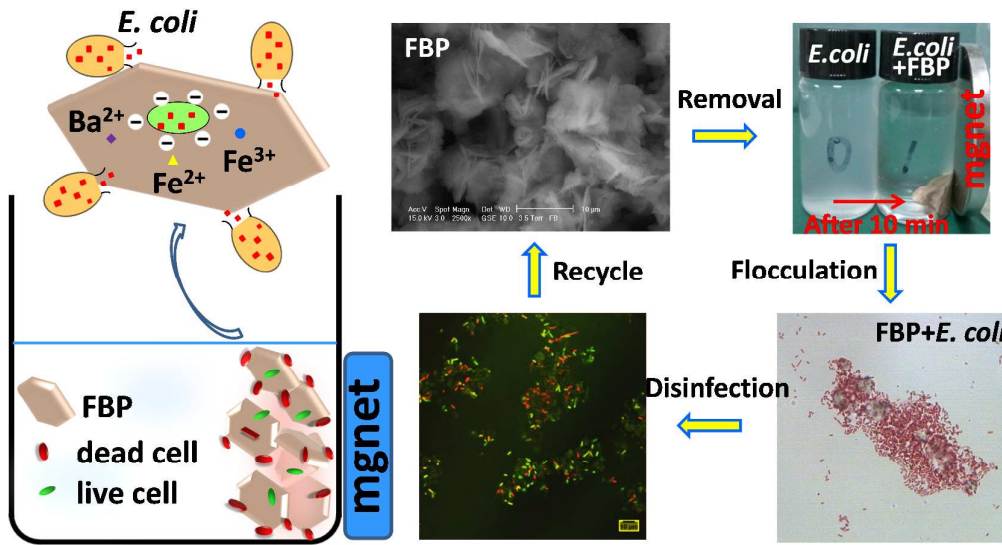
- 437 1. S. Zhan, Y. Yang, Z. Shen, J. Shan, Y. Li, S. Yang and D. Zhu, Efficient removal of pathogenic  
438 bacteria and viruses by multifunctional amine-modified magnetic nanoparticles, *Journal of*  
439 *hazardous materials*, 2014, **274**, 115-123.
- 440 2. S. Singh, K. Barick and D. Bahadur, Inactivation of bacterial pathogens under magnetic  
441 hyperthermia using Fe<sub>3</sub>O<sub>4</sub>-ZnO nanocomposite, *Powder Technology*, 2015, **269**, 513-519.
- 442 3. S. Zhan, D. Zhu, S. Ma, W. Yu, Y. Jia, Y. Li, H. Yu and Z. Shen, Highly efficient removal of  
443 pathogenic bacteria with magnetic graphene composite, *ACS applied materials & interfaces*,  
444 2015, **7**, 4290-4298.
- 445 4. Y. Jin, F. Liu, C. Shan, M. Tong and Y. Hou, Efficient bacterial capture with amino acid modified  
446 magnetic nanoparticles, *Water research*, 2014, **50**, 124-134.
- 447 5. S. Herrmann, L. De Matteis, J. M. de la Fuente, S. G. Mitchell and C. Streb, Removal of Multiple  
448 Contaminants from Water by Polyoxometalate Supported Ionic Liquid Phases (POM-SILPs),  
449 *Angewandte Chemie International Edition*, 2017, **56**, 1667-1670.
- 450 6. E. Roy, S. Patra, A. Tiwari, R. Madhuri and P. K. Sharma, Single cell imprinting on the surface of  
451 Ag-ZnO bimetallic nanoparticle modified graphene oxide sheets for targeted detection, removal  
452 and photothermal killing of E. Coli, *Biosensors and Bioelectronics*, 2017, **89**, 620-626.
- 453 7. L. Akhigbe, S. Ouki and D. Saroj, Disinfection and removal performance for Escherichia coli and  
454 heavy metals by silver-modified zeolite in a fixed bed column, *Chemical Engineering Journal*,  
455 2016, **295**, 92-98.
- 456 8. M. Hassan, R. Abou-Zeid, E. Hassan, L. Berglund, Y. Aitomäki and K. Oksman, Membranes based  
457 on cellulose nanofibers and activated carbon for removal of Escherichia coli bacteria from water,  
458 *Polymers*, 2017, **9**, 335.
- 459 9. T. Zou, C. Wang, R. Tan, W. Song and Y. Cheng, Preparation of pompon-like ZnO-PANI  
460 heterostructure and its applications for the treatment of typical water pollutants under visible  
461 light, *Journal of Hazardous Materials*, 2017, **338**, 276-286.
- 462 10. E. Anfruns-Estrada, C. Bruguera-Casamada, H. Salvadó, E. Brillas, I. Sirés and R. M. Araujo,  
463 Inactivation of microbiota from urban wastewater by single and sequential electrocoagulation  
464 and electro-Fenton treatments, *Water research*, 2017, **126**, 450-459.
- 465 11. S. Singh, K. Barick and D. Bahadur, Fe<sub>3</sub>O<sub>4</sub> embedded ZnO nanocomposites for the removal of  
466 toxic metal ions, organic dyes and bacterial pathogens, *Journal of Materials Chemistry A*, 2013, **1**,  
467 3325-3333.
- 468 12. Y. Jiang, J.-L. Gong, G.-M. Zeng, X.-M. Ou, Y.-N. Chang, C.-H. Deng, J. Zhang, H.-Y. Liu and S.-Y.  
469 Huang, Magnetic chitosan-graphene oxide composite for anti-microbial and dye removal  
470 applications, *International journal of biological macromolecules*, 2016, **82**, 702-710.
- 471 13. V. K. Sharma, T. J. McDonald, H. Kim and V. K. Garg, Magnetic graphene-carbon nanotube iron  
472 nanocomposites as adsorbents and antibacterial agents for water purification, *Advances in*  
473 *colloid and interface science*, 2015, **225**, 229-240.
- 474 14. K. Niemirowicz, I. Swiecicka, A. Z. Wilczewska, K. H. Markiewicz, U. Surel, A. Kułakowska, Z.  
475 Namiot, B. Szyńska, R. Bucki and H. Car, Growth arrest and rapid capture of select pathogens  
476 following magnetic nanoparticle treatment, *Colloids and Surfaces B: Biointerfaces*, 2015, **131**,  
477 29-38.
- 478 15. S. Chella, P. Kollu, E. V. P. Komarala, S. Doshi, M. Saranya, S. Felix, R. Ramachandran, P.

- 1  
2  
3 479 Saravanan, V. L. Koneru and V. Venugopal, Solvothermal synthesis of MnFe<sub>2</sub>O<sub>4</sub>-graphene  
4 480 composite—Investigation of its adsorption and antimicrobial properties, *Applied Surface Science*,  
5 481 2015, **327**, 27-36.
- 6  
7 482 16. J. Chen, B. Duncan, Z. Wang, L.-S. Wang, V. M. Rotello and S. R. Nugen, Bacteriophage-based  
8 483 nanoprobe for rapid bacteria separation, *Nanoscale*, 2015, **7**, 16230-16236.
- 9 484 17. F. Halouane, R. Jijie, D. Meziane, C. Li, S. K. Singh, J. Bouckaert, J. Jurazek, S. Kurungot, A. Barras  
10 485 and M. Li, Selective isolation and eradication of E. coli associated with urinary tract infections  
11 486 using anti-fimbrial modified magnetic reduced graphene oxide nanoheaters, *Journal of Materials*  
12 487 *Chemistry B*, 2017, **5**, 8133-8142.
- 13  
14 488 18. S. Ma, S. Zhan, Y. Jia and Q. Zhou, Highly efficient antibacterial and Pb (II) removal effects of  
15 489 Ag-CoFe<sub>2</sub>O<sub>4</sub>-GO nanocomposite, *ACS applied materials & interfaces*, 2015, **7**, 10576-10586.
- 16 490 19. F. Zhang, X. Yin, J. Lan and W. Zhang, Application of Ba<sub>3</sub>(PO<sub>4</sub>)<sub>2</sub>/Fe<sub>3</sub>O<sub>4</sub> as a novel magnetic  
17 491 adsorbent to remove methyl blue from aqueous solution, *Journal of materials science*, 2016, **51**,  
18 492 3525-3535.
- 19  
20 493 20. F. Zhang, X. Yin and W. Zhang, Development of magnetic Sr<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>(OH)/Fe<sub>3</sub>O<sub>4</sub> nanorod for  
21 494 adsorption of Congo red from solution, *Journal of Alloys and Compounds*, 2016, **657**, 809-817.
- 22 495 21. F. Zhang, B. Ma, X. Jiang and Y. Ji, Dual function magnetic hydroxyapatite nanopowder for  
23 496 removal of malachite green and Congo red from aqueous solution, *Powder Technology*, 2016,  
24 497 **302**, 207-214.
- 25  
26 498 22. F. Zhang, Z. Wei, W. Zhang and H. Cui, Effective adsorption of malachite green using magnetic  
27 499 barium phosphate composite from aqueous solution, *Spectrochimica Acta Part A: Molecular and*  
28 500 *Biomolecular Spectroscopy*, 2017, **182**, 116-122.
- 29 501 23. F. Zhang, Z. Zhao, R. Tan, Y. Guo, L. Cao, L. Chen, J. Li, W. Xu, Y. Yang and W. Song, Selective and  
30 502 effective adsorption of methyl blue by barium phosphate nano-flake, *Journal of colloid and*  
31 503 *interface science*, 2012, **386**, 277-284.
- 32  
33 504 24. X. Zhu, H. Fan, D. Li, Y. Xiao and X. Zhang, Protein adsorption and zeta potentials of a biphasic  
34 505 calcium phosphate ceramic under various conditions, *Journal of Biomedical Materials Research*  
35 506 *Part B: Applied Biomaterials*, 2007, **82**, 65-73.
- 36 507 25. Y. Jin, J. Deng, J. Liang, C. Shan and M. Tong, Efficient bacteria capture and inactivation by  
37 508 cetyltrimethylammonium bromide modified magnetic nanoparticles, *Colloids and Surfaces B:*  
38 509 *Biointerfaces*, 2015, **136**, 659-665.
- 39  
40 510 26. L. C. Roberts, S. J. Hug, T. Ruettimann, M. M. Billah, A. W. Khan and M. T. Rahman, Arsenic  
41 511 removal with iron (II) and iron (III) in waters with high silicate and phosphate concentrations,  
42 512 *Environmental Science & Technology*, 2004, **38**, 307-315.
- 43  
44 513 27. M. W. Tenney and W. Stumm, Chemical flocculation of microorganisms in biological waste  
45 514 treatment, *Journal (Water Pollution Control Federation)*, 1965, **37**, 1370-1388.
- 46 515 28. E. Neyens and J. Baeyens, A review of thermal sludge pre-treatment processes to improve  
47 516 dewaterability, *Journal of hazardous materials*, 2003, **98**, 51-67.
- 48 517 29. W. Hu, C. Peng, W. Luo, X. Li, D. Li, Q. Huang and C. Fan, Graphene-Based Antibacterial Paper,  
49 518 *ACS Nano*, 2010, **4**, 4317-4323.
- 50  
51 519 30. E. I. Rabea, M. E. T. Badawy, C. V. Stevens, G. Smagghe and W. Steurbaut, Chitosan as  
52 520 Antimicrobial Agent: Applications and Mode of Action, *Biomacromolecules*, 2003, **4**, 1457-1465.
- 53 521 31. N. Tran, A. Mir, D. Mallik, A. Sinha, S. Nayar and T. J. Webster, Bactericidal effect of iron oxide  
54 522 nanoparticles on Staphylococcus aureus, *International Journal of Nanomedicine*, 2010, **5**,  
55 523 277-283.

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

- 524 32. B. Stephen Inbaraj, T. Y. Tsai and B. H. Chen, Synthesis, characterization and antibacterial activity  
525 of superparamagnetic nanoparticles modified with glycol chitosan, *Science and Technology of*  
526 *Advanced Materials*, 2012, **13**, 015002.
- 527 33. S. Singh, K. Barick and D. Bahadur, Surface engineered magnetic nanoparticles for removal of  
528 toxic metal ions and bacterial pathogens, *Journal of Hazardous Materials*, 2011, **192**, 1539-1547.
- 529 34. G. Gollavelli, C.-C. Chang and Y.-C. Ling, Facile synthesis of smart magnetic graphene for safe  
530 drinking water: heavy metal removal and disinfection control, *ACS Sustainable Chemistry &*  
531 *Engineering*, 2013, **1**, 462-472.
- 532

Table of Contents



Highly efficient bacterial removal and disinfection of drinking water by recyclable magnetic barium phosphate nanoflakes with embedded iron oxide nanoparticles