



**Environmental  
Science  
Nano**

**Emerging investigator series: Protein Adsorption and Transformation on Catalytic and Food-Grade TiO<sub>2</sub> Nanoparticles in the Presence of Dissolved Organic Carbon**

Journal:	<i>Environmental Science: Nano</i>
Manuscript ID	EN-ART-01-2019-000130.R3
Article Type:	Paper
Date Submitted by the Author:	03-May-2019
Complete List of Authors:	Kim, Junyeol; University of Notre Dame, Civil and Environmental Engineering and Earth Sciences Doudrick, Kyle; University of Notre Dame, Civil and Environmental Engineering and Earth Sciences

**SCHOLARONE™**  
Manuscripts

## Environmental Significance

The protein corona that forms on engineered nanomaterials (ENMs) in biological systems contributes significantly to the bioactivity of ENMs. While the inherent physicochemical properties of ENMs are known to govern formation of the protein corona, there remains a significant knowledge gap on how exposure of ENMs to environments rich in organic matter (e.g., rivers and wastewater treatment plants) prior to biological uptake affects the adsorption of proteins onto ENMs. This work uses time-resolved, *in-situ* vibrational spectroscopy to investigate the protein sorption dynamics on ENMs in the presence of dissolved organic carbon. Outcomes enhance our understanding of how proteins interact with the surface of ENMs that have been exposed to environmentally relevant scenarios.

1  
2  
3  
4 **Emerging investigator series: Protein Adsorption and Transformation on Catalytic and**  
5  
6 **Food-Grade TiO<sub>2</sub> Nanoparticles in the Presence of Dissolved Organic Carbon**  
7

8  
9 Junyeol Kim<sup>a</sup> and Kyle Doudrick<sup>a\*</sup>  
10

11 <sup>a</sup>Department of Civil and Environmental Engineering and Earth Sciences, University of Notre  
12  
13  
14 Dame, Notre Dame, Indiana, USA, 46637  
15  
16

17  
18  
19 **\*Corresponding Authors and Address:**  
20

21  
22 Kyle Doudrick\*, Department of Civil and Environmental Engineering and Earth Sciences  
23  
24 , University of Notre Dame, Notre Dame, Indiana, USA, 46637  
25  
26  
27 e-mail: [kdoudrick@nd.edu](mailto:kdoudrick@nd.edu)  
28  
29  
30  
31  
32  
33

34  
35 **Journal:** Environmental Science: Nano  
36  
37  
38  
39

40 **Date:** April 29, 2019  
41  
42  
43  
44

45 **Keywords:** Nanoparticle, Protein, Bovine Serum Albumin, Oxalate, Dissolved Organic  
46  
47  
48 Carbon, ATR-FTIR  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

## Abstract

The inherent physicochemical properties of engineered nanomaterials (ENMs) are known to control the sorption of proteins, but knowledge on how the release of ENMs to the environment prior to protein exposure affects this reaction is limited. In this study, time-resolved, *in-situ* infrared spectroscopy was used to investigate the sorption of a model protein, bovine serum albumin (BSA), onto two different types of titanium dioxide (TiO<sub>2</sub>) ENMs (catalytic-grade P90 and food-grade E171) in the presence and absence of a simple dissolved organic carbon molecule, oxalate. Infrared spectroscopy results showed that oxalate adsorbed to P90 through chemisorption interactions, but it adsorbed to E171 through physisorption interactions due to the presence of inherent surface-bound phosphates. Secondary structure and two-dimensional correlation spectroscopy analyses showed that BSA interacted with and unfolded on the surface of P90, but not E171, presumably due to the repulsive forces from the negatively charged phosphates on E171. When oxalate was pre-adsorbed to either P90 or E171, the unfolding of BSA occurred, but along different pathways. This suggests both the “outer” surface chemistry (e.g., oxalate layers) and the mechanism by which this layer is bound to the ENM play a significant role in the adsorption of proteins. Collectively, the results indicate the exposure of ENMs to natural and engineered environments prior to biological uptake affects the resulting protein corona formation, and thus the transport and bioactivity of ENMs.

## INTRODUCTION

Understanding the interaction of engineered nanomaterials (ENMs) with biological

1  
2  
3  
4 systems is important when considering both its applications (e.g., drug delivery) and  
5  
6 implications (e.g., toxicity). Upon contact with biological environments, ENMs can become  
7  
8 coated with proteins that form a “corona” around the ENM.<sup>1</sup> While the inherent  
9  
10 physicochemical properties of the ENM (e.g., size, surface charge) control the protein  
11  
12 sorption behavior, the resulting corona contributes more significantly to the behavior of the  
13  
14 particles in biosystems.<sup>1-26</sup> Due to their widespread use, ENMs have a high probability of  
15  
16 release into the environment throughout their life-cycle.<sup>27, 28</sup> To date, most studies on protein  
17  
18 corona formation have been conducted using pristine or functionalized surfaces, but there is  
19  
20 limited knowledge how ENM exposure to natural and engineered environments *prior* to  
21  
22 biological uptake will affect the protein corona formation.  
23  
24  
25  
26  
27  
28  
29

30 While the general behavior of the formation of protein coronas on ENMs is known,  
31  
32 outcomes are likely to change in realistic scenarios where other macromolecules are present.  
33  
34 A critical knowledge gap exists on how the presence of pre-adsorbed molecules on ENMs  
35  
36 will impact formation of the protein corona. Observations from polymer studies have shown  
37  
38 the type and size of pre-adsorbed polymers on ENMs affects the kinetics and composition of  
39  
40 the protein corona.<sup>29, 30</sup> In the environment, inorganic and organic compounds will attach to  
41  
42 the surface of the ENM, changing the surface chemistry. For example, the presence of  
43  
44 phosphate on titanium dioxide (TiO<sub>2</sub>) was shown to inhibit the sorption of bovine serum  
45  
46 albumin (BSA),<sup>31</sup> and organic matter was found to complex with BSA and compete for TiO<sub>2</sub>  
47  
48 surface sites, altering the protein corona.<sup>32</sup>  
49  
50  
51  
52  
53  
54  
55

56 When ENMs enter natural or engineered environments they may interact with  
57  
58 ubiquitous dissolved organic carbon (DOC), a complex, heterogeneous mixture of organic  
59  
60

1  
2  
3  
4 compounds in aquatic and terrestrial environments formed through natural degradation  
5  
6 processes or exuded from organisms. The physicochemical properties of DOC are source  
7  
8 dependent, but generally humic, fulvic, and transphilic acids are the dominant fractions of  
9  
10 DOC.<sup>33, 34</sup> Humic and fulvic acids are more aromatic, hydrophobic, and have a lower charge  
11  
12 density than transphilic acids, and all contain mostly carboxylic and phenolic functional  
13  
14 groups.<sup>35</sup> These weak acids can adsorb to the ENM surface, effectively changing the transport  
15  
16 and reactivity of ENMs in the environment.<sup>36-42</sup> Further, complexation between proteins and  
17  
18 DOC is expected to be a significant contributor to the formation of the protein corona.<sup>32, 43</sup>  
19  
20 Thus, the alteration of the surface chemistry of the ENM by adsorbed DOC is expected to  
21  
22 change the protein corona's formation behavior.  
23  
24  
25  
26  
27  
28  
29

30 In this study, real-time, *in-situ* vibrational spectroscopy was used to investigate the  
31  
32 adsorption behavior of a model protein, BSA, onto TiO<sub>2</sub> ENMs that were exposed to oxalate,  
33  
34 a dicarboxylic acid representing a structurally simple DOC. *In-situ* spectroscopy has been  
35  
36 successfully used for investigating protein adsorption onto ENMs,<sup>44, 45</sup> and it is advantageous  
37  
38 for understanding the real-time interfacial behavior.<sup>46, 47</sup> To determine the impact of the  
39  
40 surface chemistry of the ENM on formation and transformation of the adsorbed BSA, two  
41  
42 TiO<sub>2</sub> ENMs were investigated: (i) a catalyst-grade TiO<sub>2</sub> (P90) and (ii) a food-grade TiO<sub>2</sub>  
43  
44 (E171) inherently coated with phosphates. This is the first study to use real-time, *in-situ*  
45  
46 spectroscopy to elucidate the formation dynamics and equilibrium of BSA adsorption to TiO<sub>2</sub>  
47  
48 ENMs in the presence of DOC. Specifically, this study (i) provides binding information for  
49  
50 oxalate and BSA onto P90 and E171 and (ii) knowledge on the effect of pre-adsorbed oxalate  
51  
52 on the adsorption behavior of BSA onto P90 and E171.  
53  
54  
55  
56  
57  
58  
59  
60

## MATERIALS AND METHODS

### Materials

(Photo)catalytic-grade TiO<sub>2</sub> (P90) was obtained from Evonik (Essen, Germany). The average primary particle size of the mixed-phase P90 has been previously reported by our group as 12 nm for anatase (86%) and 18 nm for rutile (14%).<sup>48</sup> Food-grade TiO<sub>2</sub> (E171) was purchased from Minerals-Water Ltd. (Purfleet, UK). BSA (AMRESCO Inc., Solon, Ohio) was used as a model protein in this study (Fig. S1). Sodium oxalate (Na<sub>2</sub>C<sub>2</sub>O<sub>4</sub>, 99.5+%) purchased from Alfa Aesar (Ward Hill, MA) was used as a simple DOC surrogate (Fig. S1). In the environment, oxalate is present from the incomplete oxidation of carbohydrates in plants, and thus serves as a representative simple DOC. All experimental and solution preparations used ultrapure water (Thermo-Barnstead Waltham, MA, 18.2 MΩ·cm).

### Materials Characterization

Transmission electron microscopy (TEM; FEI Titan 80-300, Hillsboro, Oregon) was used to determine the primary particle size and morphology of the ENMs in the presence and absence of BSA and oxalate. The ENMs were mixed with BSA and/or oxalate for 30 min, and then a drop was pipetted onto a carbon-lacey copper TEM grid. Before drying completely, a drop of ammonium molybdate tetrahydrate (2% w/v solution adjusted to pH = 7 with NaOH) was added to provide contrast between the ENM and adsorbed organic layer.

The zeta potential ( $\zeta$ ) and point of zero charge (pH<sub>zpc</sub>) of P90 and E171 were measured using phase analysis light scattering (PALS, NanoBrook Omni, Brookhaven instrument corporation, Holtsville, NY). To measure the pH<sub>zpc</sub>, a single stock solution of each

1  
2  
3  
4 TiO<sub>2</sub> (0.119 mg/L) was prepared in 1 mM potassium nitrate (KNO<sub>3</sub>, Sigma Aldrich ACS  
5  
6 Reagent Grade). The stock solution was then split into two samples and the pH adjusted (with  
7  
8 HNO<sub>3</sub> or NaOH) as needed to measure the  $\zeta$  at pH values below and above that of the stock  
9  
10 solution. After each pH adjustment, the solution was pipetted into a cuvette for subsequent  
11  
12 zeta potential analysis. Thirty measurement cycles were conducted at a 15° scattering angle  
13  
14 and 25 °C in triplicate for each sample, and the average and standard errors were reported.  
15  
16  
17  
18  
19 The zeta potential was calculated from electrophoretic mobilities using the Smoluchowski  
20  
21 model (i.e.,  $\kappa a > 1$ ). To analyze the effect of BSA and oxalate adsorption, the zeta potential  
22  
23 was measured after mixing (t = 15 min) with respective concentrations of 3.3 nM and 3.3  
24  
25  $\mu$ M.  
26  
27  
28  
29

30  
31 The hydrodynamic size (HDS) and stability of the ENMs was measured in the  
32  
33 presence and absence of BSA and/or oxalate using dynamic light scattering (DLS,  
34  
35 NanoBrook Omni, Brookhaven Instrument Corporation, Holtsville, NY). Briefly, 5.95 mg/L  
36  
37 TiO<sub>2</sub> was added to either ultrapure water, 0.17  $\mu$ M of BSA, or 0.17 mM of sodium oxalate,  
38  
39 mixed for 15 minutes, and then 3 mL transferred to a cuvette for the DLS measurement. All  
40  
41 measurements were conducted at 25 °C with a 10 s equilibration time. The HDS was  
42  
43 measured in triplicate at a 90° scattering angle, and the average and standard error of the  
44  
45 effective diameters were reported.  
46  
47  
48  
49

50  
51 Diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) and X-ray  
52  
53 photoelectron spectroscopy (XPS) were used to characterize the surface composition of P90  
54  
55 and E171. DRIFTS analyses were conducted using an EasiDiff accessory (PIKE  
56  
57 Technologies, Madison, WI) on a Bruker Vertex 70 FTIR. A 6 mm aperture and 10 kHz  
58  
59  
60



1  
2  
3  
4 scanner velocity were used with a KBr beamsplitter. For each sample, 64 background scans  
5  
6 were conducted followed by 64 sample scans of 4 cm<sup>-1</sup> resolution from 400 to 4000 cm<sup>-1</sup>. All  
7  
8 spectra were analyzed with OPUS 6.5 software (Bruker Corp., Billerica, MA). XPS (PHI,  
9  
10 VersaProbe II) spectra were normalized to the C 1s emission (284.6 eV) and fit using PHI  
11  
12 MultiPak software. All samples were analyzed in a window of 124 to 138 eV with a 0.2 eV  
13  
14  
15  
16  
17 step size.

### 21 22 **Infrared Spectroscopy Experiments**

23  
24  
25 Adsorption mechanisms of BSA and oxalate onto the TiO<sub>2</sub> samples were measured  
26  
27 using time-resolved, *in-situ* attenuated total reflectance Fourier transform infrared (TRIS-  
28  
29 ATR-FTIR) spectroscopy. TRIS-ATR-FTIR experiments were conducted using a Vertex 70  
30  
31 (Bruker Corp., Billerica, MA) equipped with an MCT detector (liquid N<sub>2</sub> cooled MCT,  
32  
33 InfraRed Associates, Inc., Stuart, FL). The ATR flow-through cell (Pike Technologies,  
34  
35 Madison, WI) was outfitted with an AMTIR internal reflection element (IRE) crystal. Spectra  
36  
37 were corrected and complemented using functions such as atmospheric compensation and  
38  
39 baseline correction available in the OPUS 7.2 software package (Bruker Corp., Billerica,  
40  
41 MA).

42  
43  
44 To create the particulate films on the IRE crystal, TiO<sub>2</sub> solutions (1 g/L) were  
45  
46 prepared in ultrapure water by brief mixing followed by bath sonication for 20 minutes  
47  
48 (Branson M3800, 40 kHz). 1 mL of the TiO<sub>2</sub> suspension was evenly distributed on the  
49  
50 surface of the AMTIR crystal using a pipette and allowed to dry. The film was then exposed  
51  
52 to experimental solutions for analysis. The concentration of TiO<sub>2</sub> was chosen based on  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 outcomes from previous literature<sup>49, 50</sup> and on their acceptable infrared (IR) absorbance  
5  
6 intensity observed by us.  
7

8  
9 TRIS-ATR-FTIR experiments were conducted in either batch or flow-through mode.  
10  
11 Batch mode was used to determine the adsorption mechanisms of either oxalate or BSA  
12 alone, while flow-through mode was used to simulate changing environments and  
13 competitive adsorption of oxalate and BSA. All IR spectra were obtained from 4000 to 800  
14  $\text{cm}^{-1}$  at a  $4 \text{ cm}^{-1}$  resolution averaged over 64 scans. To begin, the film was first exposed to  
15 water for 20 min to saturate the  $\text{TiO}_2$  film with water molecules. This allowed for background  
16 water peaks to be removed. In batch mode, the  $\text{TiO}_2$  film was exposed to either a sodium  
17 oxalate (10 mM) or BSA (10  $\mu\text{M}$ ) solution for 20 min, and the final time point was reported.  
18  
19 For flow-through mode ( $\sim 6.8 \text{ mL/min}$ ), either oxalate or BSA was first flowed across the film  
20 for 20 min, followed by flowing the other compound for 20 min (i.e., oxalate followed by  
21 BSA). Spectra was recorded approximately every  $\sim 8.5 \text{ sec}$ . After the second compound,  
22 water was flowed through the system for 20 min to observe the desorption behavior of the  
23 previously adsorbed layers (i.e., oxalate or BSA). All TRIS-ATR-FTIR experiments were  
24 conducted one time, and due to natural heterogeneity resulting from the film preparation,  
25 some variability is expected and thus the results presented are only semi-quantitative.  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50

### 51 **Computational Analysis of IR Data**

52  
53 For the analysis of the BSA secondary structure, the amide I band ( $1600\text{--}1700 \text{ cm}^{-1}$ )  
54 was extracted and curve-fitted using Fityk v.0.9.8.<sup>51</sup> Before curve fitting, the background  
55 spectra were subtracted from the amide I band followed by atmospheric compensation and  
56  
57  
58  
59  
60

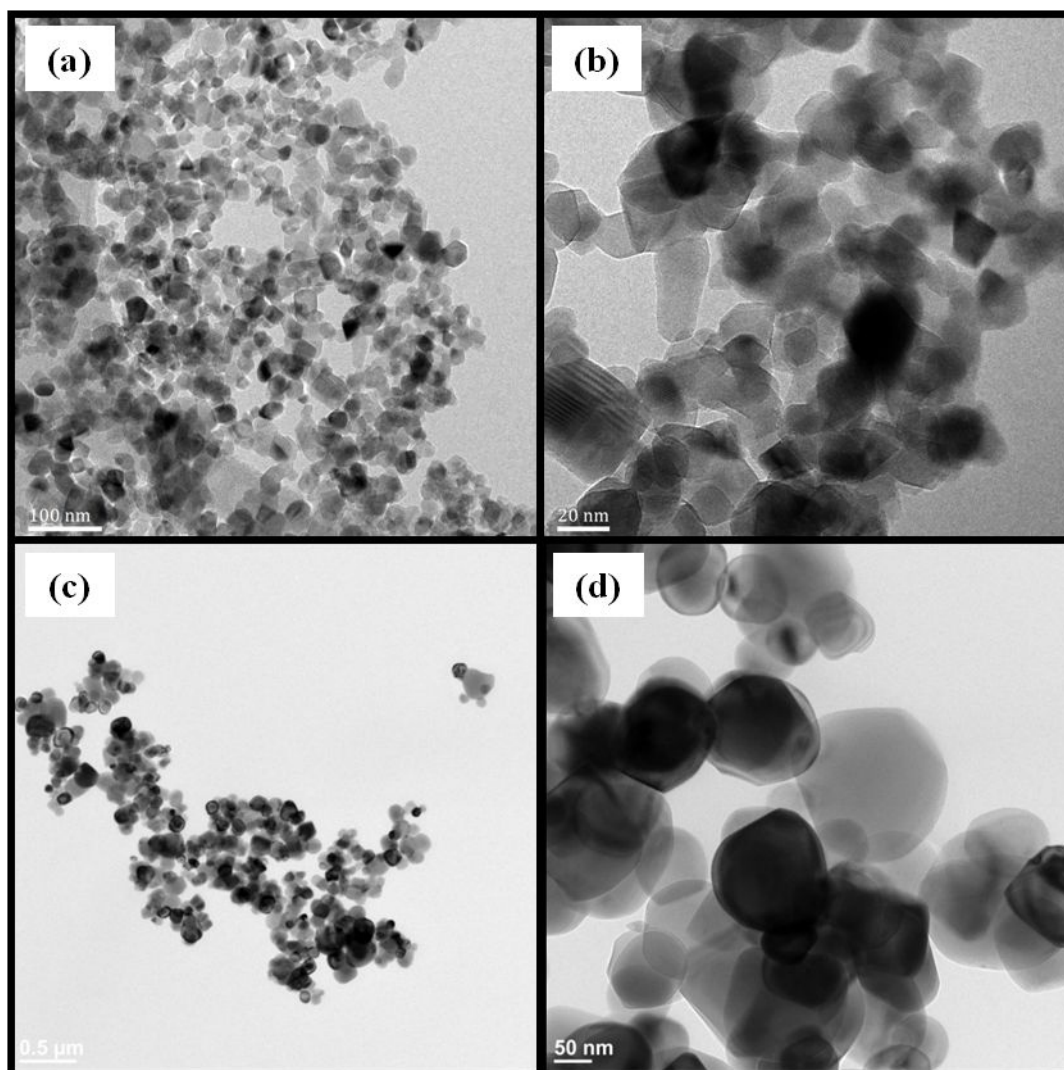
1  
2  
3  
4 baseline correction using OPUS. Goodness of fit was analyzed in-software using  $\chi^2$ ;  
5  
6 numerous fits were conducted, and the mean and standard deviation values are reported for  
7  
8 acceptable fits ( $<0.0002$ ). Generalized two-dimensional correlation spectroscopy (2DCOS)  
9  
10 was used to understand the adsorption dynamics of BSA on  $\text{TiO}_2$  in the presence and absence  
11  
12 of pre-adsorbed oxalate. 2DCOS is a mathematical tool that can interpret overlapping peaks  
13  
14 of IR spectra by spreading the peaks over a two-dimensional array.<sup>52</sup> For example, it allows  
15  
16 the use of spectral intensity to examine the changes in sequential order in the denaturation  
17  
18 process of proteins. The major bands and secondary structure of BSA were analyzed using an  
19  
20 opensource 2DCOS software, 2D-Shige (Shigeaki Morita, Kwansei-Gakuin University,  
21  
22 2004-2005). In this study, two scenarios were analyzed with 2DCOS for P90 and E171: (1)  
23  
24 the BSA spectra with no pre-adsorbed oxalate, and (2) the BSA spectra with pre-adsorbed  
25  
26 oxalate. All spectra data were analyzed following the 2DCOS methods discussed  
27  
28 previously.<sup>52</sup>  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39

## 40 **RESULTS AND DISCUSSION**

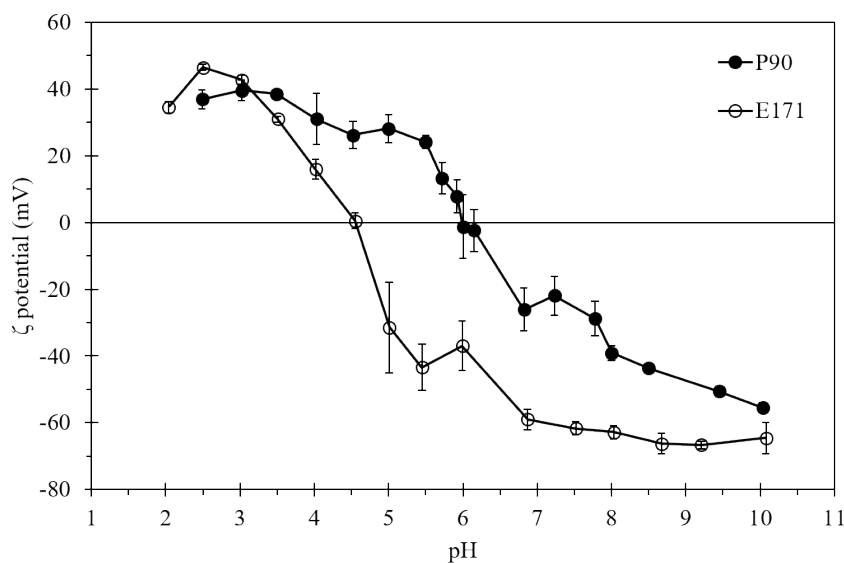
### 41 **TiO<sub>2</sub> Material Characterization**

42  
43  
44  
45 The average primary particle size of E171 was previously reported to range from  
46  
47 106–132 nm with a standard deviation of 38–56 nm.<sup>53</sup> TEM images of the E171 sample used  
48  
49 in this study showed primary particle sizes within this range (Figure 1). DRIFTS and XPS  
50  
51 analyses indicated the presence of phosphate groups on the surface of the E171 (~2.9% at  
52  
53 wt.), while no additional surface moieties were observed for P90 (Fig. S2 and Fig. S3,  
54  
55 respectively). The  $\text{pH}_{\text{zpc}}$  of P90 and E171 was approximately 6.0 and 4.5, respectively  
56  
57  
58  
59  
60

(Figure 2), which is in good agreement with our previous work on the characterization of different E171 samples.<sup>48, 53, 54</sup> The lower  $\text{pH}_{\text{zpc}}$  of E171 is attributed to the negatively charged phosphate groups sorbed to the surface of  $\text{TiO}_2$ . As a result, the zeta potential of E171 remained negative at the experimental pH range used in the study (5.1–6.7), while P90 was mostly positive.



**Figure 1. TEM images of P90 (a, b) and E171 (c, d). Images (b) and (d) are higher magnifications of (a) and (c), respectively.**



**Figure 2. Average  $\zeta$ -potentials of P90 (closed circle) and E171 (open circle) as a function of pH. Error bars represent one standard deviation.**

### Adsorption Behavior of Oxalate and BSA onto P90 and E171

IR spectra were collected for oxalate and BSA separately in the absence and presence of P90 or E171 (Figure 3). The observed IR peak positions for oxalate and BSA are summarized in Table 1. Compared to the solution-phase spectra (i.e., no  $\text{TiO}_2$ ), new peak development and peak shifting upon adsorption may indicate inner-sphere complexation or hydrogen bonding between the adsorbate and adsorbent. The solution-phase peaks for oxalate (Figure 3a) were assigned to the following vibrations:  $1308\text{ cm}^{-1}$  is CO symmetric stretching [ $\nu_s(\text{C-O})$ ],  $1566\text{ cm}^{-1}$  is CO asymmetric stretching [ $\nu_{as}(\text{C-O})$ ], and  $1695\text{ cm}^{-1}$  is CO asymmetric stretching [ $\nu_{as}(\text{C=O})$ ]. The major peaks of the solution-phase IR spectra for BSA (Figure 3b) were attributed to the amide I, II, and III bands ( $1651$ ,  $1547$ , and  $1252\text{ cm}^{-1}$ , respectively). The amide I is predominantly CO stretching with contributions from out-of-

1  
2  
3  
4 phase CN stretching, in-plane NH bending, and CCN deformation.<sup>55</sup> The amide II is  
5  
6 predominantly out-of-phase in-plane NH bending and out-of-phase CN stretching with  
7  
8 contributions from CO bending, CC stretching, and NC stretching.<sup>55</sup> The amide III is  
9  
10 predominantly an in-phase combination of CN stretching and NH bending.<sup>55</sup>  
11  
12  
13

14 For oxalate, the mechanism of adsorption differed for P90 and E171 (Figure 3). For  
15  
16 P90, the peak at 1695 cm<sup>-1</sup> red-shifted -2 cm<sup>-1</sup> while 1566 cm<sup>-1</sup> blue-shifted +6 cm<sup>-1</sup>. These  
17  
18 shifts were accompanied by a reduction in the peak intensity for 1566 cm<sup>-1</sup> and 1308 cm<sup>-1</sup>, an  
19  
20 increase in the peak intensity for 1695 cm<sup>-1</sup>, and the formation of two new peaks at 1425 cm<sup>-1</sup>  
21  
22 [ν(C–O) + ν(C–C)] and 1279 cm<sup>-1</sup> [ν<sub>s</sub>(C–O) + δ(O–C=O)]. For E171, a shift was only  
23  
24 observed for the 1695 cm<sup>-1</sup> peak, with a blue-shift of +4 cm<sup>-1</sup>. Similar to the P90 results, a  
25  
26 new peak was formed at 1425 cm<sup>-1</sup>.  
27  
28  
29  
30  
31

32 The [ν<sub>s</sub>(C–O)] and [ν<sub>as</sub>(C–O)] peaks at 1566 cm<sup>-1</sup> and 1308 cm<sup>-1</sup>, respectively, are  
33  
34 considered to be weakly adsorbed outer-sphere complexes driven by hydrogen bonds  
35  
36 between COO<sup>-</sup> of oxalates and O–H<sup>δ+</sup> and/or water molecules anchored on the surface of  
37  
38 TiO<sub>2</sub>.<sup>56-58</sup> The peaks at 1693–1699 cm<sup>-1</sup>, 1425 cm<sup>-1</sup>, and 1279 cm<sup>-1</sup> are associated with inner-  
39  
40 sphere complexes between COO<sup>-</sup> and Ti ions of TiO<sub>2</sub>.<sup>58-60</sup> Thus, P90 conclusively forms an  
41  
42 inner-sphere complex with oxalate because of the significant decrease in peak intensity for  
43  
44 1566 cm<sup>-1</sup> and 1308 cm<sup>-1</sup>, the red-shift of 1566 cm<sup>-1</sup>, the increased intensity for 1693 cm<sup>-1</sup>,  
45  
46 and the formation of new peaks at 1425 cm<sup>-1</sup> and 1279 cm<sup>-1</sup>. For E171, oxalate presumably  
47  
48 formed a mixture of outer- and inner-sphere complexes because the physisorption related  
49  
50 peaks at 1566 cm<sup>-1</sup> and 1308 cm<sup>-1</sup> were still present and the chemisorption related peak at  
51  
52 1425 cm<sup>-1</sup> was formed. E171 is coated with phosphate groups resulting in a negatively  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 charged surface at the experimental pH (i.e., Figure 2), and oxalate is also negatively charged  
5  
6 at this pH (i.e.,  $\text{HC}_2\text{O}_4^-/\text{C}_2\text{O}_4^{2-}$ ). This results in repulsive forces that reduce the transport of  
7  
8 oxalate to the surface of  $\text{TiO}_2$ , inhibiting the formation of inner-sphere complexes. However,  
9  
10 IR evidence of chemisorption suggests phosphate does not completely inhibit oxalate from  
11  
12 entering the inner Stern layer, which is supported by previous reports on competitive sorption  
13  
14 between oxalate and phosphate.<sup>61-65</sup> Thus, either the phosphate does not completely cover the  
15  
16 surface or oxalate is able to displace the bound phosphate. The latter is presumably the  
17  
18 dominate mechanism as phosphate and oxalate have similar binding constants for  $\text{TiO}_2$ <sup>65</sup>, and  
19  
20 thus exchange is possible.  
21  
22  
23  
24  
25

26  
27 For BSA, sorption onto P90 and E171 was apparent as wavenumber shifts were  
28  
29 observed for the amide I and III bands and the absorption intensity the amide I/II ratio  
30  
31 increased compared to the solution-phase spectra (Figure 3; Table 1). This suggests  
32  
33 conformation changes of the secondary structure occurred upon adsorption.<sup>66-70</sup> The  
34  
35 wavenumber shift for the amide I band was identical for P90 and E171, despite their  
36  
37 difference in surface chemistry. The blue-shift of the amide I from 1651 to 1657  $\text{cm}^{-1}$  was  
38  
39 presumably from the reduction of intramolecular hydrogen bonding and transition dipole  
40  
41 coupling of BSA,<sup>69, 71, 72</sup> resulting from protein unfolding upon adsorption to the  $\text{TiO}_2$ .  
42  
43  
44  
45  
46  
47

48  
49 Though the peak shift of the amide I was identical for P90 and E171, the difference in  
50  
51 the amide I shape and the amide I/II ratio suggests the conformational changes upon  
52  
53 adsorption were different. Protein adsorption in aqueous solution follows a general three-step  
54  
55 process: (1) transport from the bulk to the surface film, (2) attachment to the surface, and (3)  
56  
57 rearrangement on the surface, which will change the surface potential and expose different  
58  
59  
60

1  
2  
3  
4 portions of the protein that facilitate further adsorption and structural changes. In an unmixed  
5  
6 system, transport is driven by diffusion and is typically the controlling kinetic process.<sup>73</sup>  
7  
8 Attachment occurs through long- and short-range forces, but long-range forces (e.g.,  
9  
10 electrostatics) are critical to guiding the protein to favorable sorption sites, whereas  
11  
12 immobilization only occurs through strong interactions with the surface.<sup>74</sup> In our system, the  
13  
14 zeta potential of P90 and E171 at the experimental pH of 5.1 was approximately +27.5 and  
15  
16 -35 mV, respectively. Thus, at this pH the surface of P90 was composed mostly of  $\text{Ti-OH}_2^+$   
17  
18 and  $\text{Ti-OH}$  groups (i.e.,  $\text{pH}_{\text{zpc}} = 6.0$ ), and E171 was dominated by  $\text{H}_2\text{PO}_4^-$  groups (i.e.,  $\text{pK}_{\text{a1}} =$   
19  
20 2.15,  $\text{pK}_{\text{a2}} = 7.20$ ). The  $\text{pH}_{\text{zpc}}$  of BSA is approximately 4.8<sup>75</sup> and thus it would be slightly  
21  
22 deprotonated.  
23  
24  
25  
26  
27  
28  
29

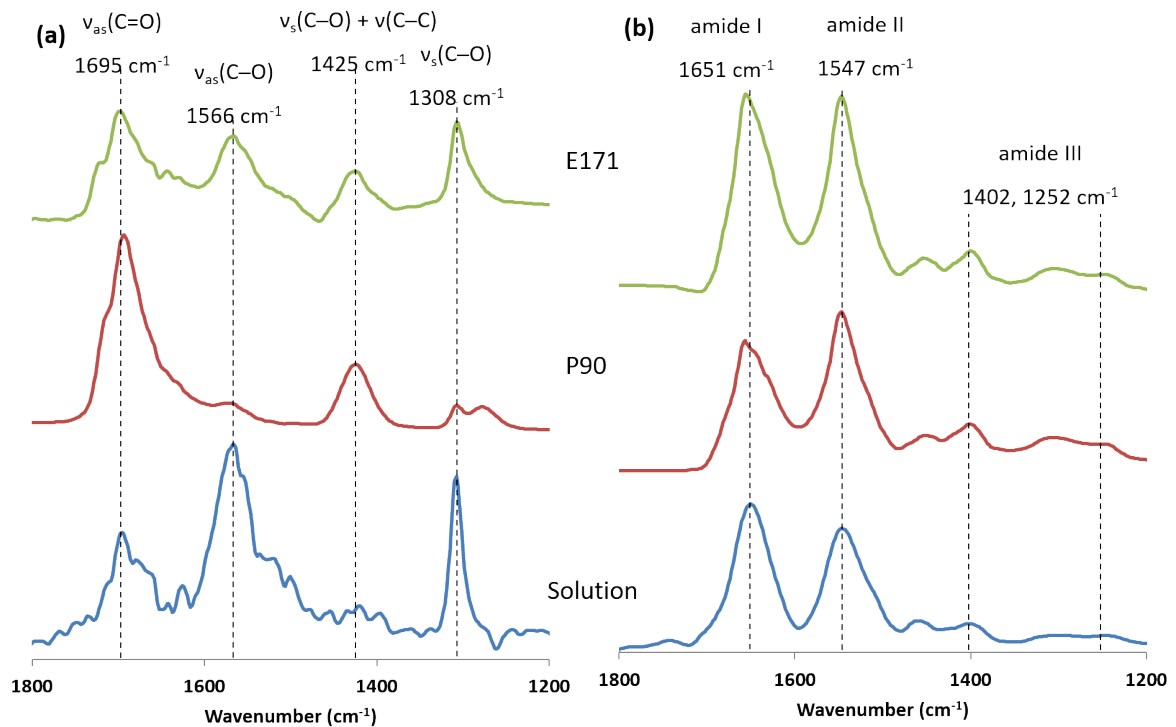
30 For BSA sorption to P90, electrostatic attraction between  $\text{Ti-OH}_2^+$  groups and  
31  
32 deprotonated moieties on BSA was likely the initial driving force, followed by strong  
33  
34 interactions and structural rearrangement upon entering the inner Stern layer. For E171,  
35  
36 electrostatic repulsion between negatively charged phosphate groups and deprotonated  
37  
38 moieties on BSA suggests transport to the inner layer and chemisorption are not favorable;  
39  
40 however, the  $\text{pH}_{\text{zpc}}$  represents the global charge, but BSA consists of a mixture negatively  
41  
42 charged (i.e., carboxyl), positively charged (i.e., amino), polar, and nonpolar segments. This  
43  
44 results in a complex sorption mechanism that can be spontaneous even in unfavorable  
45  
46 electrostatic conditions.<sup>76</sup> Two modes of attraction are possible: (1) the amino groups were  
47  
48 attracted to the negatively charged phosphate groups or (2) non-electrostatic forces between  
49  
50  $\text{TiO}_2$  and BSA overcame the weak repulsive forces, allowing BSA to attach to the  $\text{TiO}_2$   
51  
52 surface and then rearrange. Both or one of these mechanisms are possible and have been  
53  
54  
55  
56  
57  
58  
59  
60



discussed at length previously for protein adsorption to surfaces,<sup>76-83</sup> however, previous studies have shown phosphate can either complex with BSA<sup>31, 84</sup> or block the adsorption of BSA,<sup>85</sup> which conflicts with either of the proposed attachment hypotheses. Kinetic and secondary structure analyses were conducted to further elucidate the adsorption mechanism of BSA onto P90 and E171 under varying scenarios.

**Table 1. Summary of major IR bands for BSA and oxalate in the absence and presence of P90 or E171. Oxalate<sup>58, 60, 86, 87</sup> and BSA<sup>60, 88</sup> bands were previously identified.**

BSA			Oxalate			Assignment
Solution Phase	P90	E171	Solution Phase	P90	E171	
			1695	1693	1699	$\nu_{as}(C=O)$
1651	1657	1657				Amide I
			1566	1572	1566	$\nu_{as}(C-O)$
1547	1547	1547				Amide II
1402, 1252	1400, 1304, 1248	1400, 1304, 1248				Amide III
				1425	1425	$\nu_s(C-O) + \nu(C-C)$
			1308	1308	1308	$\nu_s(C-O)$
				1279		$\nu_s(C-O) + \delta(O-C=O)$

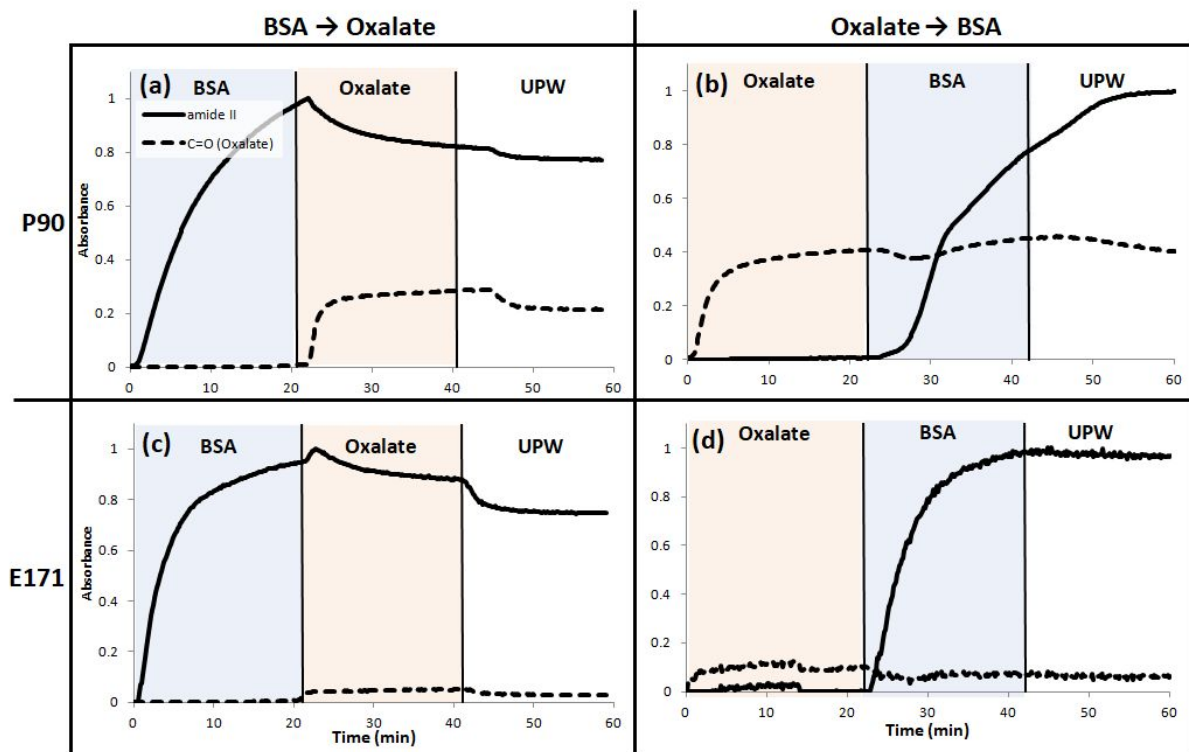


**Figure 3. Batch ATR-FTIR spectra (64 scans, 4 cm<sup>-1</sup>) of oxalate and BSA in the presence of P90 and E171 films. (a) From bottom to top: aqueous oxalate, oxalate and P90, oxalate and E171. (b) From bottom to top: aqueous BSA, BSA and P90, and BSA and E171. The unadjusted pH of samples was 6.7 for oxalate and 5.1 for BSA.**

### Sorption Kinetics of BSA

The dynamic sorption behavior of BSA in the presence of oxalate was investigated for P90 and E171 under two experimental scenarios: (1) a BSA solution was flowed through the system, followed by an oxalate solution (Figure 4 a,c); (2) an oxalate solution was flowed through the system, followed by a BSA solution (Figure 4 b,d). The purpose of the first scenario was to evaluate the effect of DOC (i.e., oxalate) on adsorbed BSA that was already formed on TiO<sub>2</sub> (e.g., a protein coated ENM entering the aquatic environment). The purpose

of the second scenario was to investigate how pre-adsorbed DOC affected the adsorbed BSA (e.g., a DOC coated ENM entering the bloodstream). The flow-through experiments are useful since this presents a more realistic scenario (e.g., a lake that has a pseudo-constant high concentration of DOC compared to the protein). For both scenarios, an additional phase was conducted at the end to investigate desorption behavior by water. The C=O and amide II vibrations were selected as representative bands for oxalate and BSA, respectively. The amide II band is good for monitoring and quantifying protein adsorption because it is less sensitive to structural changes and has less interferences from the absorbance of the water band at  $\sim 1640\text{ cm}^{-1}$ .<sup>30, 49, 50</sup>



**Figure 4.** Flow-through ATR-FTIR spectra (64 scans,  $4\text{ cm}^{-1}$ ) for BSA in the presence and absence of pre-formed oxalate layers on P90 (a, b) and E171 (c, d) films. All

1  
2  
3  
4 **absorbances were normalized to the maximum value recorded. The amide II (1547 cm<sup>-1</sup>)**  
5  
6 **and C=O (1693 cm<sup>-1</sup> and 1699 cm<sup>-1</sup> for P90 and E171, respectively) peaks were selected**  
7  
8 **for BSA and oxalate, respectively. Each dataset has three sequential phases: flow**  
9  
10 **BSA/oxalate solution → flow oxalate/BSA solution → flow water. The vertical lines**  
11  
12 **represent the time point at which each phase transitioned (e.g., 20 min).**  
13  
14  
15  
16  
17  
18  
19

20  
21 The IR absorbance values of BSA and oxalate adsorption onto P90 were  
22  
23 approximately an order of magnitude greater than E171 (Fig. S4) because of the greater  
24  
25 available surface area per mass for P90 and more favorable sorption mechanism (i.e.,  
26  
27 chemisorption). To equally compare the kinetics between samples, the absorbance values  
28  
29 were normalized to the maximum observed value for each phase (e.g., for the BSA phase in  
30  
31 Figure 4a, all values were normalized to 0.301).  
32  
33  
34  
35

36 The Elovich, pseudo zero-order (PZO), pseudo first-order (PFO), pseudo second-  
37  
38 order (PSO), and parabolic diffusion models are commonly used to empirically fit sorption  
39  
40 data.<sup>89-92</sup> None of these models were found to accurately describe any of the BSA adsorption  
41  
42 (i.e., to P90 and E171 in the presence and absence of pre-adsorbed oxalate) datasets across all  
43  
44 time points, indicating the presence of different sorption mechanisms. Heterogeneous  
45  
46 sorption reactions are not easily described by a single kinetic model, and often sorption  
47  
48 reactions are modeled without consideration of early, diffusion driven timepoints.<sup>91, 93</sup>  
49  
50 Schmidt et al.<sup>49</sup> found the simplified Elovich and PSO models to fit the adsorption of BSA to  
51  
52 montmorillonite, but early time points in the Elovich model were not predicted well,  
53  
54 suggesting diffusional control in this region. In our study, we found a mixture of PZO and  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 PSO models at early and late times, respectively, described the data well ( $R^2 > 0.99$ , data not  
5 shown). This is presumably due to the mixture of sorption mechanisms involved with protein  
6 adsorption including physisorption, chemisorption, and structural rearrangement. These  
7 empirical kinetic models generally lead to “good fits” because early and late times do not  
8 skew the correlation significantly. Because of the complexity of the sorption mechanisms, the  
9 inconsistency in the models used to obtain rate constants, and the variability from a lack of  
10 experimental repetitions, we instead report the initial slopes of the sorption kinetic curves to  
11 gain a general understanding of the behavior. The normalized absorbances were used to  
12 calculate the initial rates of adsorption for BSA and oxalate, which corresponded to the initial  
13 linear slope of the datasets.  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29

30 For the first scenario, the initial rates of adsorption of BSA onto P90 (Figure 4a) and  
31 E171 (Figure 4c) were 0.106 and 0.235 a.u.  $s^{-1}$ , respectively. The faster rate of adsorption  
32 onto E171 suggests the sorption behavior was dominated by physisorption compared to P90,  
33 which was likely a mixture of physi- and chemi-sorption. For P90 and E171, the addition of  
34 oxalate after BSA sorption caused BSA to partially desorb, which continued into the  
35 subsequent water phase. This indicates a fraction of BSA was weakly or reversibly adsorbed.  
36 The desorption was more apparent for E171, which confirms the sorption of BSA was  
37 dominated by physisorption mechanisms.  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50

51 For the second scenario, with oxalate pre-adsorbed to the surface, the sorption kinetics  
52 and behavior of BSA changed for both P90 and E171 (Fig. 4; Fig. S5a). The initial rate for  
53 P90 (Fig. 4b) was 0.011 a.u.  $s^{-1}$ ; however, the kinetics for P90 with pre-adsorbed oxalate were  
54 not straightforward and the use of a single initial rate is misleading. Instead, four distinct  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 kinetic regions were observed (Fig. S5b): (1) BSA adsorption was initially slow ( $0.011 \text{ a.u. s}^{-1}$ )  
5  
6  
7  $^1$ ); (2) the rate increased slightly ( $0.019 \text{ a.u. s}^{-1}$ ); (3) the rate increased ( $0.098 \text{ a.u. s}^{-1}$ ) to a  
8  
9 similar rate as BSA adsorption on bare P90 ( $0.106 \text{ a.u. s}^{-1}$ ); (4) the rate slowed ( $0.037 \text{ a.u. s}^{-1}$ ),  
10  
11 and then sorption continued into the water phase until reaching equilibrium. In the first two  
12  
13 regions, BSA was presumably competing with oxalate for sorption sites, and oxalate was  
14  
15 being displaced or rearranged as indicated by the decrease in the oxalate absorption intensity.  
16  
17 In the third region, BSA adsorption preceded at a similar rate as to when no oxalate was  
18  
19 present, suggesting oxalate had been partially displaced in favor of BSA. However, the  
20  
21 absorption intensity of oxalate began increasing back to its original value precisely in this  
22  
23 region even though there was no oxalate being added. This suggests oxalate was only being  
24  
25 temporarily displaced or rearranged initially and then re-adsorbed simultaneously with BSA.  
26  
27 In the last region, BSA sorption slowed as it approached saturation conditions. This suggests  
28  
29 oxalate partially hindered the rate of adsorption of BSA but did not completely inhibit it.  
30  
31 During the water phase, BSA was not desorbed as observed in the first scenario, rather it  
32  
33 continued to adsorb, further providing evidence oxalate slows the adsorption kinetics of BSA  
34  
35 with continued rearrangement and sorption. The adsorption behavior of BSA and oxalate in  
36  
37 both scenarios is conceptualized in the schematic presented in Figure 5.  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47

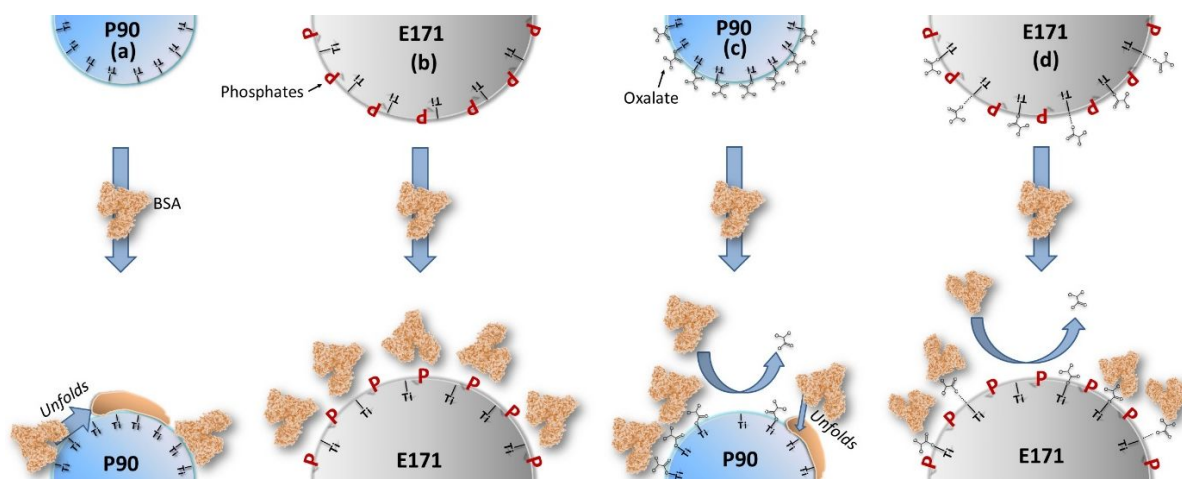
48 For the adsorption of BSA onto E171 with pre-adsorbed oxalate (Fig. 4d), the initial  
49  
50 rate was slower ( $0.141 \text{ a.u. s}^{-1}$ ) than in the absence of pre-adsorbed oxalate (i.e., Fig. 4b;  
51  
52  $0.235 \text{ a.u. s}^{-1}$ ). No desorption of BSA from E171 occurred during the water phase, contrary to  
53  
54 the first scenario, suggesting there is some degree of interaction between oxalate and BSA  
55  
56 resulting in irreversible sorption of BSA. Similar dynamic interactions between natural  
57  
58  
59  
60

1  
2  
3  
4 organic matter (NOM) and BSA were observed for TiO<sub>2</sub>, where the adsorption behavior was  
5  
6 dependent on the “history” of exposure.<sup>32</sup> When NOM and BSA were introduced  
7  
8 simultaneously, an NOM-BSA complexed species adsorbed to TiO<sub>2</sub> and multilayer sorption  
9  
10 was inhibited. When the species were introduced sequentially (i.e., NOM and then BSA), the  
11  
12 NOM dominated inner-layer adsorption with BSA populating the outer-layers of the  
13  
14 macromolecular coating. Similar results have been reported regarding the competitive  
15  
16 adsorption between humic acid and BSA onto TiO<sub>2</sub>.<sup>94</sup> Humic acid adsorbed to TiO<sub>2</sub> without  
17  
18 displacing BSA due to non-specific hydrophobic interactions between adsorbate species.  
19  
20  
21  
22  
23  
24  
25

26 The association of DOC (e.g., oxalate) with proteins has been contributed to  $\pi$ - $\pi$   
27  
28 interactions, dipole-dipole bonding, electrostatic attraction, and hydrophobic interactions.<sup>43,</sup>  
29  
30 <sup>95-98</sup> However, because both oxalate ( $pK_{a2} = 4.26$ ) and BSA ( $pH_{zpc} = 4.8$ ) are negatively  
31  
32 charged at the experimental pH range (5.1–6.7), electrostatic attraction is not expected to be a  
33  
34 major contributor. Additionally, the hydrophilic nature of oxalate will limit hydrophobic  
35  
36 interactions. Therefore, dipole-dipole interactions (e.g., hydrogen bonds) and/or  $\pi$ - $\pi$   
37  
38 interactions of C atoms of BSA and oxalate are presumably the main complexation  
39  
40 mechanisms that overcome the electrostatic repulsion between BSA and oxalate, resulting in  
41  
42 irreversible sorption of BSA.  
43  
44  
45  
46  
47  
48

49 For all scenarios, the sorption of BSA onto E171 was faster compared to P90 (Fig.  
50  
51 S5), presumably due to the physisorption-dominated mechanisms due to the phosphate bound  
52  
53 to the surface of the TiO<sub>2</sub>. For both P90 and E171, the addition of oxalate to the system prior  
54  
55 to BSA resulted in slower sorption kinetics of BSA, and its post addition resulted in the  
56  
57 desorption and/or rearrangement of surface-adsorbed BSA. The rebound in oxalate  
58  
59  
60

adsorption after desorption suggests its displacement was only temporary. These results collectively indicate the “history” of the nanoparticle exposure can affect the kinetic and equilibrium conditions, and thus the resulting transport and bio-activity behavior of the nanoparticle. Consequently, accurate modeling of the protein corona formation will not be straightforward, and simply knowing the pristine or manufactured state of the nanoparticle or the environment it is exposed to (e.g., blood) will not be sufficient. Knowledge of the surface chemistry just prior to the protein corona formation will be critical.



**Figure 5.** A representative schematic of the adsorption behavior of BSA onto (left to right): bare P90, bare E171, P90 with pre-adsorbed oxalate, E171 with pre-adsorbed oxalate. The “P” represents phosphate groups. Figure is not to scale.

The adsorption behavior of BSA and oxalate onto the TiO<sub>2</sub> ENMs was confirmed by analyzing the stability, HDS, and zeta potential (Table S2 and Fig. S6), and by imaging with TEM (Fig. S7). In water, P90 is stable but aggregated with an HDS of  $279 \pm 3.4$  nm. In the presence of BSA and oxalate, the HDS of P90 increased to  $1116 \pm 75.6$  nm and  $2112 \pm 167$  nm, respectively. A corresponding decrease in the zeta potential from positive to negative

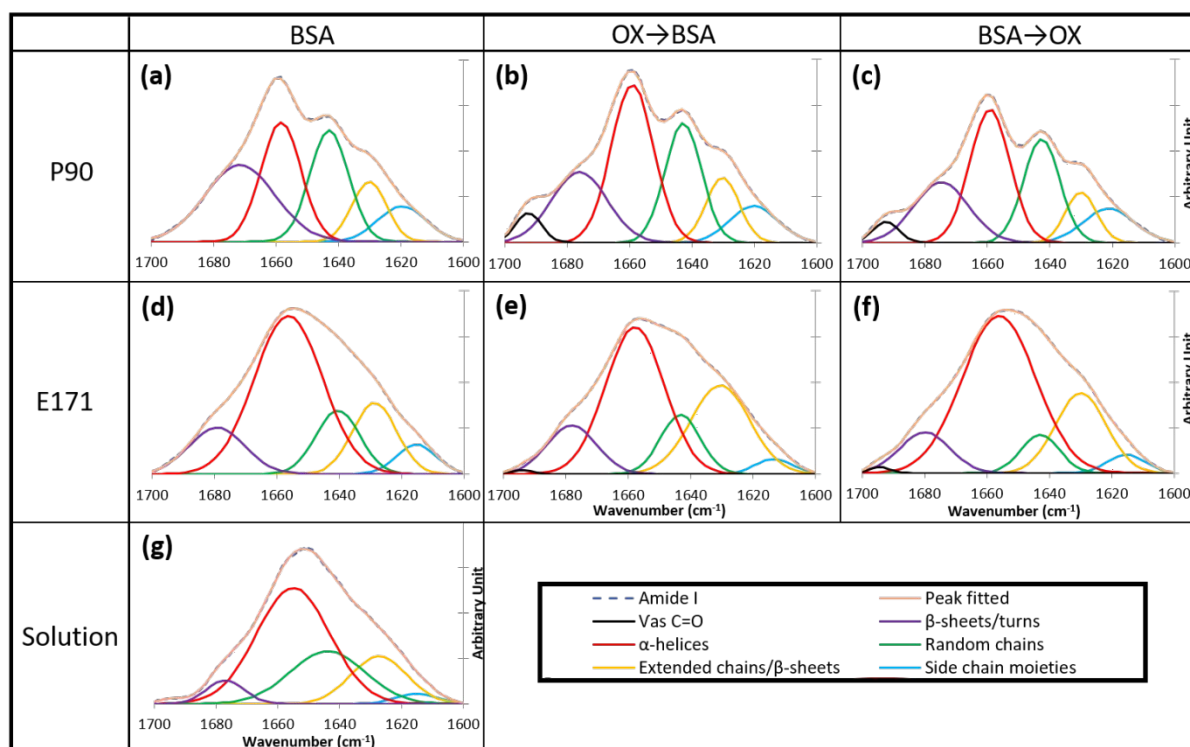


1  
2  
3  
4 values was observed for each. In the presence of both BSA and oxalate, the HDS did not  
5  
6 change significantly for P90 ( $302 \pm 4.2$  nm), but the negative magnitude of the zeta potential  
7  
8 increased. This suggests the BSA and oxalate complex adsorbs to P90 and stabilizes it  
9  
10 through increased electrostatic repulsion. For E171, only the addition of BSA significantly  
11  
12 altered the HDS ( $315 \pm 9.9$  nm to  $464 \pm 0.3$  nm) and zeta potential ( $-27.3 \pm 2.9$  mV to  $-0.2 \pm$   
13  
14  $0.8$  mV). Results of the settling (i.e., stability) experiments in Fig. S6 showed only P90 with  
15  
16 oxalate became unstable within 2 hrs, but both P90 and E171 with BSA became unstable  
17  
18 within 24 hrs. The mixture of oxalate and BSA did not result in instability within 24 hrs,  
19  
20 confirming the complexed BSA-oxalate species exhibits a different adsorption behavior than  
21  
22 BSA or oxalate alone. This agrees with previous outcomes on the competitive sorption  
23  
24 between NOM and BSA.<sup>32</sup> TEM images confirmed the presence of BSA adsorbed to P90 and  
25  
26 E171 (Fig. S7). These outcomes collectively support the molecular-scale observations that  
27  
28 protein adsorption, which alters the surface chemistry of the ENMs, is significantly affected  
29  
30 by both the inherent surface chemistry of the ENM and the exposure to DOC.  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42

### 43 **Analysis of the Secondary Structure of BSA**

44  
45 The secondary structure of BSA was evaluated for the two scenarios described in the  
46  
47 previous section to determine conformational changes occurring upon adsorption to  $\text{TiO}_2$  in  
48  
49 the presence and absence of DOC (i.e., oxalate). The secondary structure of BSA was  
50  
51 investigated using the amide I band ( $\sim 1600\text{--}1700$   $\text{cm}^{-1}$ ), which is sensitive to changes in  
52  
53 secondary structure.<sup>68, 69</sup> The amide I band of BSA was deconvoluted into  $\beta$ -sheets/turns,  $\alpha$ -  
54  
55 helices, random chains, extended chains/ $\beta$ -sheets, and side chain moieties (Fig. 6).  
56  
57  
58  
59  
60

Wavenumber ranges for each component of the BSA secondary structure are summarized in Table S1.<sup>44, 55, 68, 99-101</sup> The  $\nu_{\text{as}}(\text{C}=\text{O})$  peak for oxalate was also fitted for relevant experimental results. The deconvoluted peak area percentages were determined as a function of the total amide I area to evaluate specific component changes with varying conditions (Table 2). The  $\alpha$ -helix portion was used to evaluate perturbations and unfolding of the BSA structure.<sup>68, 102</sup>



**Figure 6.** The amide I band and secondary structure deconvoluted peaks of BSA for the different experimental scenarios involving (a-f) P90 and E171 and (g) the solution-phase with no  $\text{TiO}_2$  or oxalate present.

The  $\alpha$ -helix content of BSA on the bare IRE was 53.6% (i.e., solution-phase). Upon adsorption to P90, the  $\alpha$ -helix content decreased to 26.3%, which is in agreement with previous studies showing when proteins adsorb to surfaces they lose their helical structure due to strong interactions between the protein and surface (e.g., chemisorption, hydrogen bonding).<sup>70, 76, 103, 104</sup> In the presence of pre-adsorbed oxalate on P90, the  $\alpha$ -helix content was

1  
2  
3  
4 comparatively higher at 34.0%, indicating a lower degree of unfolding. Likewise, even when  
5  
6 BSA was adsorbed first, the subsequent addition of oxalate to the system resulted in a higher  
7  
8  $\alpha$ -helix content (32.9%) compared to the solution-phase BSA. These results suggest the  
9  
10 displacement of BSA by oxalate allowed BSA to fold back into its  $\alpha$ -helix structure, and this  
11  
12 agrees with the kinetic data (i.e., Figure 4) showing the desorption and/or rearrangement of  
13  
14 BSA when oxalate was introduced. The unfolding of BSA  $\alpha$ -helix has been suggested to  
15  
16 proceed through  $\beta$ -sheets/chains and then to  $\beta$ -sheet/turns.<sup>49, 68</sup> The total  $\beta$ -content (i.e.,  
17  
18 extended and turns) for P90, P90 with pre-adsorbed oxalate, and P90 with adsorbed BSA and  
19  
20 subsequent oxalate exposure was 40.9%, 30.9%, and 30.1%, respectively. This increase in  $\beta$ -  
21  
22 content agrees with the decreasing trend observed for the  $\alpha$ -helix content. The addition of a  
23  
24 final water desorption phase for the latter two experiments did not significantly alter (i.e.  
25  
26 <5%) the  $\alpha$ -helix content (not shown, 34.5 and 35%, respectively). Thus, even though the  
27  
28 desorption behavior differed between the two scenarios (i.e., Figure 4 a, b), water did not  
29  
30 have a measurable effect on conformational changes. Collectively, these results suggest the  
31  
32 presence of oxalate, whether it is pre-adsorbed or introduced after BSA adsorption, affects  
33  
34 the conformational state of BSA on P90.

35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45 The  $\alpha$ -helix and total  $\beta$ -content of BSA on E171 in the absence of oxalate was 54.9%  
46  
47 and 31.5%, respectively. These were similar to the solution-phase BSA contents (53.6%,  
48  
49 21.4%, respectively), suggesting BSA was weakly sorbed to E171, and the presence of  
50  
51 phosphate groups on the TiO<sub>2</sub> surface deterred conformational changes. The presence of pre-  
52  
53 adsorbed oxalate on E171 reduced the  $\alpha$ -helix content to 44.7%, again indicating the ability  
54  
55 of oxalate to induce conformational changes, as seen with P90. Interestingly, the  $\alpha$ -helix  
56  
57  
58  
59  
60

content of BSA on both E171 and P90 pre-adsorbed with oxalate were markedly different (44.7% and 34%, respectively). This suggests the just the presence of oxalate alone does not determine the conformation of BSA, but rather the binding mechanism of oxalate to TiO<sub>2</sub> affects how BSA conforms on the surface. The energetics of oxalate would be different since, as discussed previously, oxalate is strongly bound in the inner Stern layer to P90, but weakly bound in the outer Stern layer to E171 due to its phosphate coating. Thus, the phosphate coating appears to play a significant role in altering the sorption of oxalate and in turn the conformational changes of BSA upon adsorption.

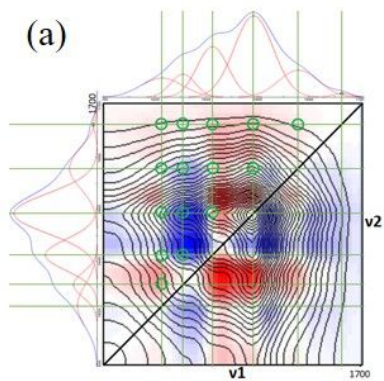
**Table 2. Relative percent contribution of each secondary structure of BSA adsorbed on P90 and E171 ENMs. Goodness of fit was analyzed using  $\chi^2$ ; numerous fits were conducted, and the mean and standard deviation values are reported for acceptable fits (<0.0002). OX = oxalate.**

		Side chain moieties	Extended chains/ $\beta$ -sheets	Random chains	$\alpha$ -helices	$\beta$ - sheets/turns	V <sub>as</sub> C=O
	BSA solution	2.9 ± 0.3	15.6 ± 2.8	22.1 ± 4.0	53.6 ± 4.9	5.8 ± 0.3	-
<b>P90</b>	BSA	9.3 ± 0.1	11.4 ± 0.1	23.6 ± 0.0	26.3 ± 0.0	29.5 ± 0.0	-
	BSA → OX	10.6 ± 0.0	9.6 ± 0.0	23.0 ± 0.0	32.9 ± 0.1	20.5 ± 0.1	3.3 ± 0.0
	OX → BSA	9.1 ± 0.1	10.9 ± 0.2	22.2 ± 0.1	34.0 ± 0.1	20.0 ± 0.1	3.0 ± 0.0
<b>E171</b>	BSA	4.9 ± 0.6	19.4 ± 2.8	8.7 ± 3.3	54.9 ± 1.4	12.1 ± 0.3	-
	BSA → OX	3.5 ± 0.0	20.3 ± 0.0	7.3 ± 0.1	58.1 ± 0.1	10.2 ± 0.0	0.6 ± 0.0
	OX → BSA	2.8 ± 0.1	27.4 ± 0.3	12.1 ± 0.3	44.7 ± 0.3	12.5 ± 0.1	0.4 ± 0.0

### Analysis of the Sequential Changes of the Secondary Structure using 2DCOS

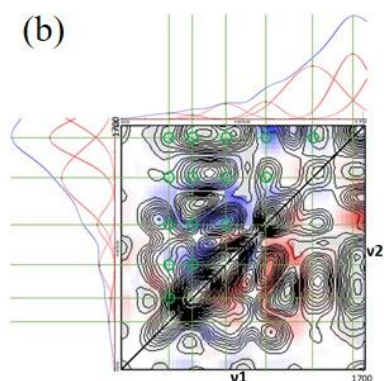
1  
2  
3  
4 Changes in the protein secondary structure upon adsorption are often rapid and  
5  
6 simultaneous, making it difficult to determine the unfolding sequence. 2DCOS is a useful  
7  
8 technique that can be used to deconvolute overlapping one-dimensional IR peaks that occur  
9  
10 due to an external perturbation (i.e., time).<sup>49, 52, 105, 106</sup> It was used herein to determine the  
11  
12 sequence of secondary structural changes that occurred during the first minute of adsorption.  
13  
14 Figure 7 shows synchronous (real components) and asynchronous (imaginary components)  
15  
16 2DCOS plots and the spectrum analysis for the amide I secondary structures of BSA upon  
17  
18 adsorption onto P90 and E171 in the absence and presence of oxalate. The sequential changes  
19  
20 of the secondary structure of BSA upon adsorption was determined by the location and signs  
21  
22 of cross peaks ( $\nu_1$ ,  $\nu_2$ ) in the synchronous and asynchronous spectral plots. When the sign of  
23  
24 cross peaks is positive in the synchronous plot, both  $\nu_1$  and  $\nu_2$  increase or decrease  
25  
26 simultaneously as adsorption progresses. In contrast, when the sign of cross peaks is negative,  
27  
28 an opposing trend occurs, and  $\nu_1$  increases while  $\nu_2$  decreases or vice versa. For the  
29  
30 asynchronous plot, the cross peak shows a sequential change of spectral intensities that  
31  
32 occurs as the external perturbation (time) is applied to the two waves,  $\nu_1$  and  $\nu_2$ . In other  
33  
34 words, variations in the dynamic changes observed for  $\nu_1$  and  $\nu_2$  can be seen from the  
35  
36 asynchronous spectrum, where associated signs are expressed in red (+) and blue (-) (Fig. 7  
37  
38 and Fig. S8). For the asynchronous cross peaks, there are four scenarios that can occur in the  
39  
40 adsorption event and these are separated into the upper left and lower right regions of the  
41  
42 spectrum plot (e.g., Fig. S8): (1) if the cross peaks are located in the upper left (i.e.,  $\nu_1 < \nu_2$ )  
43  
44 and are positive (i.e., red), then the lower wavenumber,  $\nu_1$ , occurs before the higher  
45  
46 wavenumber,  $\nu_2$ ; (2) if cross peaks are located in the lower right ( $\nu_1 > \nu_2$ ) and are positive  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 (red), then the higher wavenumber,  $\nu_1$ , occurs before the lower wavenumber,  $\nu_2$ ; (3, 4) if the  
5  
6  
7 sign of the cross peaks is negative (blue), then the order is reversed for those regions.  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60



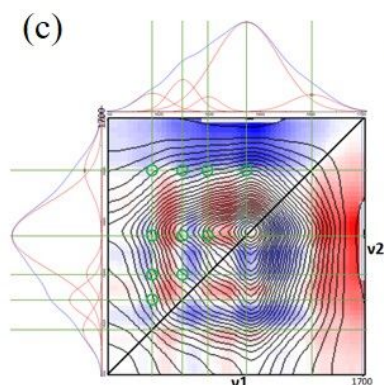
BSA (P90) Asynchronous signal	v1 v2	1623	1631	1642	1658	1676	1693
$\beta$ -sheets	1693	+	+	+	+	+	
$\beta$ -sheets/turns	1676	-	-	+	+		-
$\alpha$ -helix	1658	-	-	+		-	-
Random chains	1642	-	-		-	-	-
$\beta$ -sheets	1631	+		+	+	+	-
Side chain moieties	1623		-	+	+	+	-

Sequential order: (First) 1642  $\rightarrow$  1658  $\rightarrow$  1676  $\rightarrow$  1623  $\rightarrow$  1631  $\rightarrow$  1693  $\text{cm}^{-1}$  (Last)  
 (First) Random chains  $\rightarrow$   $\alpha$ -helix  $\rightarrow$   $\beta$ -sheets/turns  $\rightarrow$  Side chain moieties  $\rightarrow$   $\beta$ -sheets  $\rightarrow$   $\beta$ -sheets (Last)



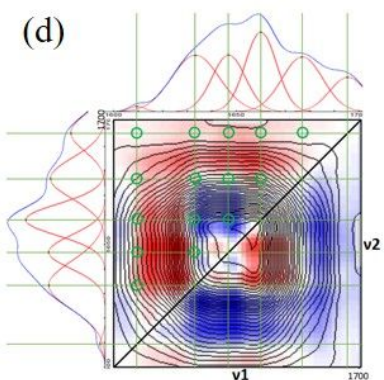
OX $\rightarrow$ BSA (P90) Asynchronous signal	v1 v2	1619	1629	1643	1659	1679	1695
$\beta$ -sheets	1695	(-, -)	(-, -)	(-, -)	(-, -)	(-, -)	
$\beta$ -sheets/turns	1679	-	-	-	+		(-, +)
$\alpha$ -helix	1659	-	-	-		-	(-, +)
Random chains	1643	-	-		+	+	(-, +)
$\beta$ -sheets	1629	+		+	+	+	(-, +)
Side chain moieties	1619		-	+	+	+	(-, +)

Sequential order: (First) 1659  $\rightarrow$  1679  $\rightarrow$  1643  $\rightarrow$  1619  $\rightarrow$  1629  $\rightarrow$  1695  $\text{cm}^{-1}$  (Last)  
 (First)  $\alpha$ -helix  $\rightarrow$   $\beta$ -sheets/turns  $\rightarrow$  Random chains  $\rightarrow$  Side chain moieties  $\rightarrow$   $\beta$ -sheets  $\rightarrow$   $\beta$ -sheets (Last)



BSA (E171) Asynchronous signal	v1 v2	1617	1629	1639	1655	1680
$\beta$ -sheets/turns	1680	-	-	-	-	
$\alpha$ -helix	1655	-	+	+		+
Random chains	1639	-	-		-	+
$\beta$ -sheets	1629	-		+	-	+
Side chain moieties	1617		+	+	+	+

Sequential order: (First) 1680  $\rightarrow$  1639  $\rightarrow$  1629  $\rightarrow$  1655  $\rightarrow$  1617  $\text{cm}^{-1}$  (Last)  
 (First)  $\beta$ -sheets/turns  $\rightarrow$  Random chains  $\rightarrow$   $\beta$ -sheets  $\rightarrow$   $\alpha$ -helix  $\rightarrow$  Side chain moieties (Last)



OX $\rightarrow$ BSA (E171) Asynchronous signal	v1 v2	1609	1633	1646	1659	1676	1695
$\beta$ -sheets	1695	+	+	+	+	+	
$\beta$ -sheets/turns	1676	+	+	+	+		-
$\alpha$ -helix	1659	+	-	-		-	-
Random chains	1646	+	-		+	-	-
$\beta$ -sheets	1633	+		+	+	-	-
Side chain moieties	1609		-	-	-	-	-

Sequential order: (First) 1609  $\rightarrow$  1659  $\rightarrow$  1646  $\rightarrow$  1633  $\rightarrow$  1676  $\rightarrow$  1695  $\text{cm}^{-1}$  (Last)  
 (First) Side chain moieties  $\rightarrow$   $\alpha$ -helix  $\rightarrow$  Random chains  $\rightarrow$   $\beta$ -sheets  $\rightarrow$   $\beta$ -sheets/turns  $\rightarrow$   $\beta$ -sheets (Last)

1  
2  
3  
4 **Figure 7. 2DCOS synchronous and asynchronous spectra and the sequential order of**  
5 **the BSA secondary structures upon adsorption to P90 and E171. In the plots, the**  
6 **contour lines (black) and the colored regions (blue, red) represent the absorption**  
7 **intensities from synchronous and asynchronous spectrum, respectively. The IR peaks**  
8 **are the secondary structures of BSA. In the tables, the horizontal row and vertical**  
9 **column wavenumbers are  $\nu_1$  and  $\nu_2$ , respectively. Except for (b), all synchronous**  
10 **secondary structure peaks were +, so the + or – shown in the tables are for the**  
11 **asynchronous spectra peaks. For (b), the parentheses represent the synchronous and**  
12 **asynchronous spectra, respectively, e.g., (-,-).**  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29

30 At early time points (<1 min), an additional secondary structure peak was observed  
31 between 1690–1700  $\text{cm}^{-1}$ , which has been identified as an additional  $\beta$ -sheets peak.<sup>55, 100, 101</sup>  
32 This  $\beta$ -sheets peak was distinguishable from the adjacent  $\beta$ -sheets/turns peak at early time  
33 points when the intensity of the  $\beta$ -sheets/turns peak was weak. As time progressed, the  
34 intensity the  $\beta$ -sheets/turns peak increased, while the  $\beta$ -sheets peak did not, and it became  
35 negligible in comparison. This additional  $\beta$ -sheets peak was not observed in the secondary  
36 structure peaks of BSA adsorbed onto E171.  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47

48 For BSA adsorption to P90, the sequence of secondary structures observed  
49 immediately upon adsorption was random chains  $\rightarrow$   $\alpha$ -helix  $\rightarrow$   $\beta$ -sheets/turns  $\rightarrow$  side chain  
50 moieties  $\rightarrow$   $\beta$ -sheets (1631  $\text{cm}^{-1}$ )  $\rightarrow$   $\beta$ -sheets (1693  $\text{cm}^{-1}$ ). In the presence of pre-adsorbed  
51 oxalate, the adsorption sequence of the BSA secondary structure was:  $\alpha$ -helix  $\rightarrow$   $\beta$ -  
52 sheets/turns  $\rightarrow$  random chains  $\rightarrow$  side chain moieties  $\rightarrow$   $\beta$ -sheets (1629  $\text{cm}^{-1}$ )  $\rightarrow$   $\beta$ -sheets  
53  
54  
55  
56  
57  
58  
59  
60



1  
2  
3  
4 (1695  $\text{cm}^{-1}$ ). Upon adsorption, the sequential changes of the secondary structure in the  
5  
6 absence and presence of pre-adsorbed oxalate were similar with a general trend indicating the  
7  
8 unfolding of BSA from  $\alpha$ -helix to  $\beta$ -sheets.  
9

10  
11 For BSA adsorption to E171 in the absence of oxalate, the sequence of secondary  
12  
13 structures observed upon adsorption was:  $\beta$ -sheets/turns  $\rightarrow$  random chains  $\rightarrow$   $\beta$ -sheets  $\rightarrow$   $\alpha$ -  
14  
15 helix  $\rightarrow$  side chain moieties. In the presence of pre-adsorbed oxalate, the adsorption sequence  
16  
17 of the BSA secondary structure was side chain moieties  $\rightarrow$   $\alpha$ -helix  $\rightarrow$  random chains  $\rightarrow$   $\beta$ -  
18  
19 sheets (1633  $\text{cm}^{-1}$ )  $\rightarrow$   $\beta$ -sheets/turns  $\rightarrow$   $\beta$ -sheets (1695  $\text{cm}^{-1}$ ). In the absence of oxalate, the  
20  
21 adsorption of BSA onto E171 was random and no clear unfolding trend was observed. As  
22  
23 discussed previously, this suggests phosphate inhibits the binding of BSA to  $\text{TiO}_2$ , and no  
24  
25 mode of attraction (e.g., amino group attraction to  $\text{Ti-OH}_2^+$ ) allows BSA to approach the  
26  
27 surface. This also agrees with the secondary structure analysis that showed minimal  $\alpha$ -helix  
28  
29 loss. In the presence of oxalate, the resulting sequence suggests unfolding occurred from  $\alpha$ -  
30  
31 helix to  $\beta$ -sheets/turns and this behavior was comparable to P90 in the presence of oxalate.  
32  
33 The similar unfolding of BSA on P90 and E171 with pre-adsorbed oxalate indicates oxalate  
34  
35 serves a specific role in the adsorption of BSA.  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47

## 48 **Conclusion**

49  
50 The inherent surface chemistry of ENMs is known to control the sorption of proteins,  
51  
52 but little evidence has been provided on how the release of ENMs to the environment prior to  
53  
54 biological uptake affects the protein corona formation. In this study, we used TRIS-ATR-  
55  
56 FTIR to investigate the sorption of a model protein, BSA, onto two different types of  $\text{TiO}_2$   
57  
58  
59  
60

1  
2  
3  
4 ENMs (catalytic-grade P90 and food-grade E171) in the presence and absence of a simple  
5  
6 DOC molecule, oxalate. TiO<sub>2</sub> is a commonly used model ENM in fate, transport, and toxicity  
7  
8 studies because of its stability, widespread use, and low cost, but historically these studies  
9  
10 have been conducted with catalytic grade TiO<sub>2</sub> (e.g., P90) instead of the more prevalent food-  
11  
12 grade TiO<sub>2</sub>, E171. Though a single protein (i.e., BSA) was used in this study, outcomes can  
13  
14 be collectively used to understand protein corona formation. The inherent presence of  
15  
16 phosphates on the surface of E171 will affect the sorption behavior of DOC and proteins, the  
17  
18 presence of pre-adsorbed DOC on TiO<sub>2</sub> will significantly change the adsorption behavior of  
19  
20 proteins. The mechanism by which DOC is bound to TiO<sub>2</sub> will impact the adsorption  
21  
22 behavior of proteins significantly more than considering just the presence of DOC on the  
23  
24 TiO<sub>2</sub> surface.

25  
26  
27  
28  
29  
30  
31  
32 Beyond medical applications, ENMs are unlikely to be taken up by organisms without  
33  
34 first being exposed to natural (e.g., streams) or engineered systems (e.g., wastewater  
35  
36 treatment plant). Thus, when considering the impact of the protein corona on the bioactivity  
37  
38 of ENMs, not only must the “history” of the ENM be considered to understand how sorbed  
39  
40 compounds affect its surface chemistry, but also the mechanism by which these compounds  
41  
42 are bound to the ENM (e.g., chemisorbed), as this will impact the ensuing sorption kinetics  
43  
44 and equilibrium behavior of proteins. Moving forward, using the outcomes of the simplistic  
45  
46 systems used in this study as baseline knowledge, future studies should consider increasingly  
47  
48 complex systems that have protein mixtures, more complex DOCs (e.g., humic acid), and the  
49  
50 presence of other potentially complexing constituents (e.g., calcium), all while keeping in  
51  
52 mind that the inherent ENM properties can affect these outcomes.

## Acknowledgements

This work was made possible through financial support provided by the University of Notre Dame and the National Science Foundation (CBET 1705770). The authors further thank the Center for Environmental Science and Technology at Notre Dame (CEST) and the Notre Dame Integrated Imaging Facility (NDIIF) for access to select instrumentation.

## REFERENCES

1. T. Cedervall, I. Lynch, S. Lindman, T. Berggard, E. Thulin, H. Nilsson, K. A. Dawson and S. Linse, Understanding the nanoparticle-protein corona using methods to quantify exchange rates and affinities of proteins for nanoparticles, *Proc. Natl. Acad. Sci. U. S. A.*, 2007, **104**, 2050-2055.
2. M. Lundqvist, J. Stigler, G. Elia, I. Lynch, T. Cedervall and K. A. Dawson, Nanoparticle size and surface properties determine the protein corona with possible implications for biological impacts, *Proc. Natl. Acad. Sci. U. S. A.*, 2008, **105**, 14265-14270.
3. A. E. Nel, L. Madler, D. Velegol, T. Xia, E. M. V. Hoek, P. Somasundaran, F. Klaessig, V. Castranova and M. Thompson, Understanding biophysicochemical interactions at the nano-bio interface, *Nat. Mater.*, 2009, **8**, 543-557.
4. D. Walczyk, F. B. Bombelli, M. P. Monopoli, I. Lynch and K. A. Dawson, What the Cell "Sees" in Bionanoscience, *J. Am. Chem. Soc.*, 2010, **132**, 5761-5768.
5. C. C. Ge, J. F. Du, L. N. Zhao, L. M. Wang, Y. Liu, D. H. Li, Y. L. Yang, R. H. Zhou, Y. L. Zhao, Z. F. Chai and C. Y. Chen, Binding of blood proteins to carbon nanotubes reduces cytotoxicity, *Proc. Natl. Acad. Sci. U. S. A.*, 2011, **108**, 16968-16973.
6. W. B. Hu, C. Peng, M. Lv, X. M. Li, Y. J. Zhang, N. Chen, C. H. Fan and Q. Huang, Protein Corona-Mediated Mitigation of Cytotoxicity of Graphene Oxide, *ACS Nano*, 2011, **5**, 3693-3700.
7. M. Lundqvist, J. Stigler, T. Cedervall, T. Berggard, M. B. Flanagan, I. Lynch, G. Elia and K. Dawson, The Evolution of the Protein Corona around Nanoparticles: A Test Study, *ACS Nano*, 2011, **5**, 7503-7509.
8. M. P. Monopoli, D. Walczyk, A. Campbell, G. Elia, I. Lynch, F. B. Bombelli and K. A. Dawson, Physical-Chemical Aspects of Protein Corona: Relevance to in Vitro and in Vivo Biological Impacts of Nanoparticles, *J. Am. Chem. Soc.*, 2011, **133**, 2525-2534.
9. A. Lesniak, F. Fenaroli, M. R. Monopoli, C. Aberg, K. A. Dawson and A. Salvati, Effects of the Presence or Absence of a Protein Corona on Silica Nanoparticle Uptake and Impact on Cells, *ACS Nano*, 2012, **6**, 5845-5857.
10. M. P. Monopoli, C. Aberg, A. Salvati and K. A. Dawson, Biomolecular coronas provide the biological identity of nanosized materials, *Nat. Nanotechnol.*, 2012, **7**, 779-786.
11. C. D. Walkey and W. C. W. Chan, Understanding and controlling the interaction of nanomaterials with proteins in a physiological environment, *Chem. Soc. Rev.*, 2012, **41**, 2780-2799.
12. A. Lesniak, A. Salvati, M. J. Santos-Martinez, M. W. Radomski, K. A. Dawson and C. Aberg, Nanoparticle Adhesion to the Cell Membrane and Its Effect on Nanoparticle Uptake Efficiency, *J. Am. Chem. Soc.*, 2013, **135**, 1438-1444.
13. S. R. Saptarshi, A. Duschl and A. L. Lopata, Interaction of nanoparticles with proteins: relation to bio-reactivity of the nanoparticle, *Journal of Nanobiotechnology*, 2013, **11**, 12.
14. S. Tenzer, D. Docter, J. Kuharev, A. Musyanovych, V. Fetz, R. Hecht, F. Schlenk, D. Fischer, K. Kiouptsi, C. Reinhardt, K. Landfester, H. Schild, M. Maskos, S. K. Knauer and R. H. Stauber, Rapid formation of plasma protein corona critically affects nanoparticle pathophysiology, *Nat. Nanotechnol.*, 2013, **8**, 772-U1000.
15. P. Del Pino, B. Pelaz, Q. Zhang, P. Maffre, G. U. Nienhaus and W. J. Parak, Protein corona formation around nanoparticles - from the past to the future, *Materials Horizons*, 2014, **1**, 301-313.
16. D. Docter, C. Bantz, D. Westmeier, H. J. Galla, Q. B. Wang, J. C. Kirkpatrick, P. Nielsen, M. Maskos and R. H. Stauber, The protein corona protects against size- and dose-dependent toxicity of amorphous silica nanoparticles, *Beilstein Journal of Nanotechnology*, 2014, **5**, 1380-1392.

17. C. Gunawan, M. Lim, C. P. Marquis and R. Amal, Nanoparticle-protein corona complexes govern the biological fates and functions of nanoparticles, *Journal of Materials Chemistry B*, 2014, **2**, 2060-2083.
18. L. Treuel, S. Brandholt, P. Maffre, S. Wiegele, L. Shang and G. U. Nienhaus, Impact of Protein Modification on the Protein Corona on Nanoparticles and Nanoparticle-Cell Interactions, *ACS Nano*, 2014, **8**, 503-513.
19. C. D. Walkey, J. B. Olsen, F. Y. Song, R. Liu, H. B. Guo, D. W. H. Olsen, Y. Cohen, A. Emili and W. C. W. Chan, Protein Corona Fingerprinting Predicts the Cellular Interaction of Gold and Silver Nanoparticles, *ACS Nano*, 2014, **8**, 2439-2455.
20. G. Caracciolo, S. Palchetti, V. Colapicchioni, L. Digiacomo, D. Pozzi, A. L. Capriotti, G. La Barbera and A. Lagana, Stealth Effect of Biomolecular Corona on Nanoparticle Uptake by Immune Cells, *Langmuir*, 2015, **31**, 10764-10773.
21. X. J. Cheng, X. Tian, A. Q. Wu, J. X. Li, J. Tian, Y. Chong, Z. F. Chai, Y. L. Zhao, C. Y. Chen and C. Ge, Protein Corona Influences Cellular Uptake of Gold Nanoparticles by Phagocytic and Nonphagocytic Cells in a Size-Dependent Manner, *ACS Appl. Mater. Interfaces*, 2015, **7**, 20568-20575.
22. D. Docter, D. Westmeier, M. Markiewicz, S. Stolte, S. K. Knauer and R. H. Stauber, The nanoparticle biomolecule corona: lessons learned - challenge accepted?, *Chem. Soc. Rev.*, 2015, **44**, 6094-6121.
23. P. Foroozandeh and A. A. Aziz, Merging Worlds of Nanomaterials and Biological Environment: Factors Governing Protein Corona Formation on Nanoparticles and Its Biological Consequences, *Nanoscale Res. Lett.*, 2015, **10**, 12.
24. C. C. Ge, J. Tian, Y. L. Zhao, C. Y. Chen, R. H. Zhou and Z. F. Chai, Towards understanding of nanoparticle-protein corona, *Archives of Toxicology*, 2015, **89**, 519-539.
25. R. Liu, W. Jiang, C. D. Walkey, W. C. W. Chan and Y. Cohen, Prediction of nanoparticles-cell association based on corona proteins and physicochemical properties, *Nanoscale*, 2015, **7**, 9664-9675.
26. D. Pozzi, G. Caracciolo, L. Digiacomo, V. Colapicchioni, S. Palchetti, A. L. Capriotti, C. Cavaliere, R. Z. Chiozzi, A. Puglisi and A. Lagana, The biomolecular corona of nanoparticles in circulating biological media, *Nanoscale*, 2015, **7**, 13958-13966.
27. F. Gottschalk, T. Sonderer, R. W. Scholz and B. Nowack, Possibilities and limitations of modeling environmental exposure to engineered nanomaterials by probabilistic material flow analysis, *Environmental Toxicology and Chemistry*, 2010, **29**, 1036-1048.
28. A. A. Keller, S. McFerran, A. Lazareva and S. Suh, Global life cycle releases of engineered nanomaterials, *Journal of Nanoparticle Research*, 2013, **15**, 1692.
29. V. M. Gun'ko, L. I. Mikhalovska, P. E. Tomlins and S. V. Mikhalovsky, Competitive adsorption of macromolecules and real-time dynamics of Vroman-like effects, *Physical Chemistry Chemical Physics*, 2011, **13**, 4476-4485.
30. D.-H. Tsai, M. Davila-Morris, F. W. DelRio, S. Guha, M. R. Zachariah and V. A. Hackley, Quantitative Determination of Competitive Molecular Adsorption on Gold Nanoparticles Using Attenuated Total Reflectance-Fourier Transform Infrared Spectroscopy, *Langmuir*, 2011, **27**, 9302-9313.
31. Z. Xu and V. H. Grassian, Bovine Serum Albumin Adsorption on TiO<sub>2</sub> Nanoparticle Surfaces: Effects of pH and Coadsorption of Phosphate on Protein-Surface Interactions and Protein Structure, *The Journal of Physical Chemistry C*, 2017, **121**, 21763-21771.
32. S. Shakiba, A. Hakimian, L. R. Barco and S. M. Louie, Dynamic Intermolecular Interactions Control Adsorption from Mixtures of Natural Organic Matter and Protein onto Titanium Dioxide

- Nanoparticles, *Environmental Science & Technology*, 2018, **52**, 14158-14168.
33. M. Kitis, J. E. Kilduff and T. Karanfil, Isolation of dissolved organic matter (dom) from surface waters using reverse osmosis and its impact on the reactivity of dom to formation and speciation of disinfection by-products, *Water Research*, 2001, **35**, 2225-2234.
34. J. S. Gaffney, N. A. Marley and S. B. Clark, *Humic and Fulvic Acids and Organic Colloidal Materials in the Environment*, American Chemical Society, 1996.
35. H. K. Tan, *Humic Matter in Soil and the Environment: Principles and Controversies*, CRC Press, 2nd edn., 2014.
36. C. Jiang, G. R. Aiken and H. Hsu-Kim, Effects of Natural Organic Matter Properties on the Dissolution Kinetics of Zinc Oxide Nanoparticles, *Environmental Science & Technology*, 2015, **49**, 11476-11484.
37. R. Grillo, A. H. Rosa and L. F. Fraceto, Engineered nanoparticles and organic matter: A review of the state-of-the-art, *Chemosphere*, 2015, **119**, 608-619.
38. M.-H. Shen, Y.-G. Yin, A. Booth and J.-F. Liu, Effects of molecular weight-dependent physicochemical heterogeneity of natural organic matter on the aggregation of fullerene nanoparticles in mono- and di-valent electrolyte solutions, *Water Research*, 2015, **71**, 11-20.
39. B. Gu, J. Schmitt, Z. Chen, L. Liang and J. F. McCarthy, Adsorption and desorption of different organic matter fractions on iron oxide, *Geochimica et Cosmochimica Acta*, 1995, **59**, 219-229.
40. A. M. Vindedahl, M. S. Stemig, W. A. Arnold and R. L. Penn, Character of Humic Substances as a Predictor for Goethite Nanoparticle Reactivity and Aggregation, *Environmental Science & Technology*, 2016, **50**, 1200-1208.
41. A. M. Vindedahl, J. H. Strehlau, W. A. Arnold and R. L. Penn, Organic matter and iron oxide nanoparticles: aggregation, interactions, and reactivity, *Environmental Science: Nano*, 2016, **3**, 494-505.
42. E. M. Hotze, S. M. Louie, S. Lin, M. R. Wiesner and G. V. Lowry, Nanoparticle core properties affect attachment of macromolecule-coated nanoparticles to silica surfaces, *Environmental Chemistry*, 2014, **11**, 257-267.
43. J. E. Tomaszewski, M. Madliger, J. A. Pedersen, R. P. Schwarzenbach and M. Sander, Adsorption of Insecticidal Cry1Ab Protein to Humic Substances. 2. Influence of Humic and Fulvic Acid Charge and Polarity Characteristics, *Environmental Science & Technology*, 2012, **46**, 9932-9940.
44. B. E. Givens, Z. Xu, J. Fiegel and V. H. Grassian, Bovine serum albumin adsorption on SiO<sub>2</sub> and TiO<sub>2</sub> nanoparticle surfaces at circumneutral and acidic pH: A tale of two nano-bio surface interactions, *Journal of Colloid and Interface Science*, 2017, **493**, 334-341.
45. A. Jedlovsky-Hajdú, F. B. Bombelli, M. P. Monopoli, E. Tombácz and K. A. Dawson, Surface Coatings Shape the Protein Corona of SPIONs with Relevance to Their Application in Vivo, *Langmuir*, 2012, **28**, 14983-14991.
46. I. A. Mudunkotuwa and V. H. Grassian, Biological and environmental media control oxide nanoparticle surface composition: the roles of biological components (proteins and amino acids), inorganic oxyanions and humic acid, *Environmental Science: Nano*, 2015, **2**, 429-439.
47. I. A. Mudunkotuwa, A. A. Minshid and V. H. Grassian, ATR-FTIR spectroscopy as a tool to probe surface adsorption on nanoparticles at the liquid–solid interface in environmentally and biologically relevant media, *Analyst*, 2014, **139**, 870-881.
48. K. Doudrick, O. Monzón, A. Mangonon, K. Hristovski and P. Westerhoff, Nitrate Reduction in Water Using Commercial Titanium Dioxide Photocatalysts (P25, P90, and Hombikat UV100), *Journal of Environmental Engineering*, 2012, **138**, 852-861.

- 1
- 2
- 3
- 4 49. M. P. Schmidt and C. E. Martínez, Kinetic and Conformational Insights of Protein Adsorption onto
- 5 Montmorillonite Revealed Using in Situ ATR-FTIR/2D-COS, *Langmuir*, 2016, **32**, 7719-7729.
- 6 50. D.-H. Tsai, F. W. DelRio, A. M. Keene, K. M. Tyner, R. I. MacCuspie, T. J. Cho, M. R. Zachariah and
- 7 V. A. Hackley, Adsorption and Conformation of Serum Albumin Protein on Gold Nanoparticles
- 8 Investigated Using Dimensional Measurements and in Situ Spectroscopic Methods, *Langmuir*, 2011,
- 9 **27**, 2464-2477.
- 10 51. M. Wojdyr, Fityk: a general-purpose peak fitting program, *Journal of Applied Crystallography*, 2010,
- 11 **43**, 1126-1128.
- 12 52. I. Noda and Y. Ozaki, *Two-Dimensional Correlation Spectroscopy: Applications in Vibrational and*
- 13 *Optical Spectroscopy*, WILEY, 2004.
- 14 53. Y. Yang, K. Doudrick, X. Bi, K. Hristovski, P. Herckes, P. Westerhoff and R. Kaegi, Characterization
- 15 of Food-Grade Titanium Dioxide: The Presence of Nanosized Particles, *Environmental Science &*
- 16 *Technology*, 2014, **48**, 6391-6400.
- 17 54. J. J. Faust, K. Doudrick, Y. Yang, P. Westerhoff and D. G. Capco, Food grade titanium dioxide
- 18 disrupts intestinal brush border microvilli in vitro independent of sedimentation, *Cell Biology and*
- 19 *Toxicology*, 2014, **30**, 169-188.
- 20 55. A. Barth, Infrared spectroscopy of proteins, *Biochimica et Biophysica Acta (BBA) - Bioenergetics*,
- 21 2007, **1767**, 1073-1101.
- 22 56. S. Benkoula, O. Sublemontier, M. Patanen, C. Nicolas, F. Sirotti, A. Naitabdi, F. Gaie-Levrel, E.
- 23 Antonsson, D. Aureau, F.-X. Ouf, S.-I. Wada, A. Etcheberry, K. Ueda and C. Miron, Water adsorption
- 24 on TiO<sub>2</sub> surfaces probed by soft X-ray spectroscopies: bulk materials vs. isolated nanoparticles,
- 25 *Scientific Reports*, 2015, **5**, 15088.
- 26 57. P. A. Connor, K. D. Dobson and A. J. McQuillan, Infrared Spectroscopy of the TiO<sub>2</sub>/Aqueous Solution
- 27 Interface, *Langmuir*, 1999, **15**, 2402-2408.
- 28 58. T. H. Yoon, S. B. Johnson, C. B. Musgrave and G. E. Brown, Adsorption of organic matter at
- 29 mineral/water interfaces: I. ATR-FTIR spectroscopic and quantum chemical study of oxalate adsorbed
- 30 at boehmite/water and corundum/water interfaces, *Geochimica et Cosmochimica Acta*, 2004, **68**, 4505-
- 31 4518.
- 32 59. S. J. Hug and D. Bahnemann, Infrared spectra of oxalate, malonate and succinate adsorbed on the
- 33 aqueous surface of rutile, anatase and lepidocrocite measured with in situ ATR-FTIR, *Journal of*
- 34 *Electron Spectroscopy and Related Phenomena*, 2006, **150**, 208-219.
- 35 60. A. D. Weisz, L. García Rodenas, P. J. Morando, A. E. Regazzoni and M. A. Blesa, FTIR study of the
- 36 adsorption of single pollutants and mixtures of pollutants onto titanium dioxide in water: oxalic and
- 37 salicylic acids, *Catalysis Today*, 2002, **76**, 103-112.
- 38 61. M. A. Elsheikh, Z. Abidin, N. Matsue and T. Henmi, Competitive Adsorption of Oxalate and
- 39 Phosphate on Allophane at Low Concentration, *Clay Science*, 2008, **13**, 181-188.
- 40 62. C. N. Guppy, N. W. Menzies, P. W. Moody and F. P. C. Blamey, Competitive sorption reactions
- 41 between phosphorus and organic matter in soil: a review, *Soil Research*, 2005, **43**, 189-202.
- 42 63. S. Nagarajah, A. M. Posner and J. P. Quirk, Competitive Adsorption of Phosphate with
- 43 Polygalacturonate and other Organic Anions on Kaolinite and Oxide Surfaces, *Nature*, 1970, **228**, 83.
- 44 64. M. Long, J. Brame, F. Qin, J. Bao, Q. Li and P. J. J. Alvarez, Phosphate Changes Effect of Humic
- 45 Acids on TiO<sub>2</sub> Photocatalysis: From Inhibition to Mitigation of Electron–Hole Recombination,
- 46 *Environmental Science & Technology*, 2017, **51**, 514-521.
- 47 65. P. A. Connor and A. J. McQuillan, Phosphate Adsorption onto TiO<sub>2</sub> from Aqueous Solutions: An in
- 48
- 49
- 50
- 51
- 52
- 53
- 54
- 55
- 56
- 57
- 58
- 59
- 60

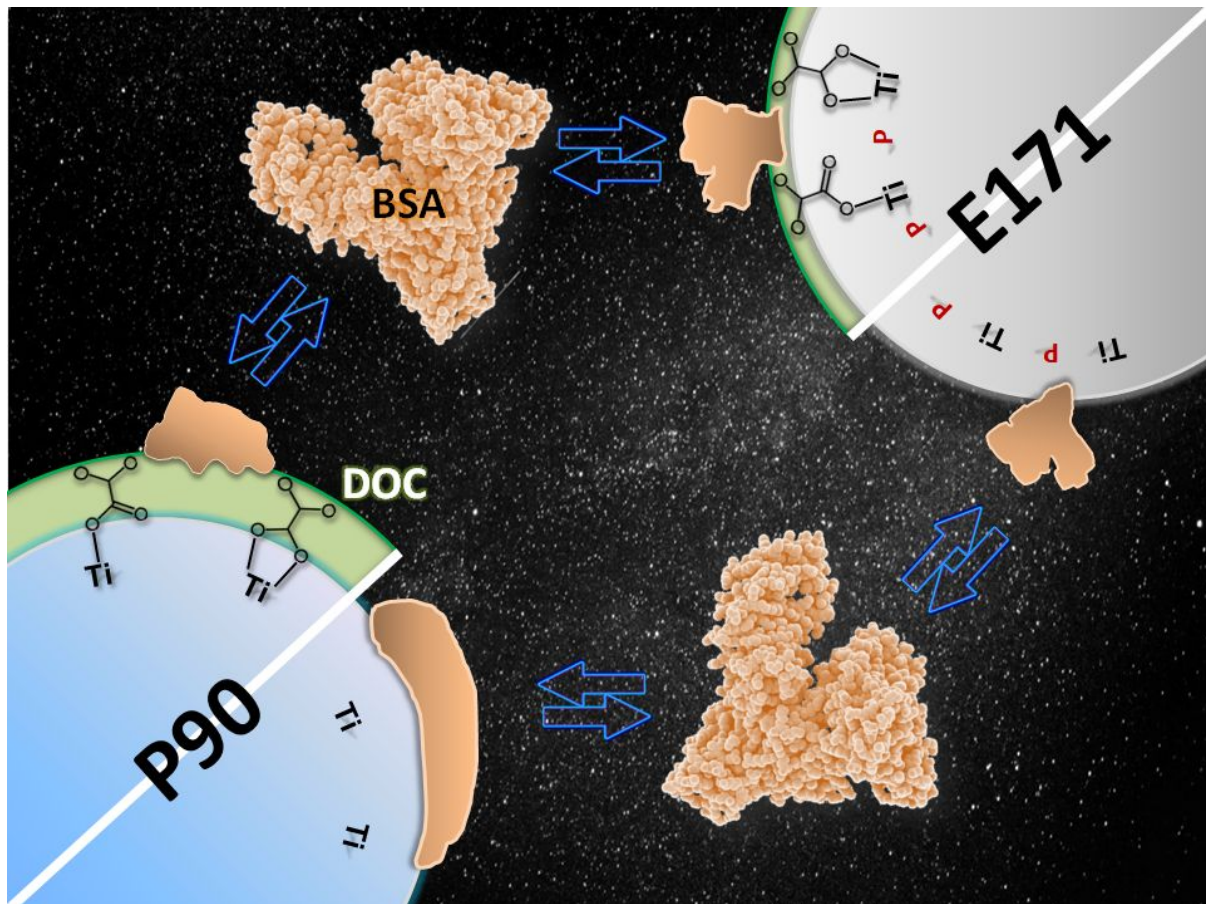
- 1  
2  
3 Situ Internal Reflection Infrared Spectroscopic Study, *Langmuir*, 1999, **15**, 2916-2921.
- 4  
5 66. L. A. Buchanan and A. El-Ghannam, Effect of bioactive glass crystallization on the conformation and  
6 bioactivity of adsorbed proteins, *Journal of Biomedical Materials Research Part A*, 2009, **93A**, 537-  
7 546.
- 8  
9 67. S. J. McClellan and E. I. Franses, Adsorption of bovine serum albumin at solid/aqueous interfaces,  
10 *Colloids and Surfaces A: Physicochemical and Engineering Aspects*, 2005, **260**, 265-275.
- 11  
12 68. P. Roach, D. Farrar and C. C. Perry, Interpretation of Protein Adsorption: Surface-Induced  
13 Conformational Changes, *Journal of the American Chemical Society*, 2005, **127**, 8168-8173.
- 14  
15 69. A. Barth and C. Zscherp, What vibrations tell about proteins, *Quarterly Reviews of Biophysics*, 2002,  
16 **35**, 369-430.
- 17  
18 70. L. Song, K. Yang, W. Jiang, P. Du and B. Xing, Adsorption of bovine serum albumin on nano and bulk  
19 oxide particles in deionized water, *Colloids and Surfaces B: Biointerfaces*, 2012, **94**, 341-346.
- 20  
21 71. B. Liedberg, B. Ivarsson and I. Lundström, Fourier transform infrared reflection absorption  
22 spectroscopy (FT-IRAS) of fibrinogen adsorbed on metal and metal oxide surfaces, *Journal of  
23 Biochemical and Biophysical Methods*, 1984, **9**, 233-243.
- 24  
25 72. J. Lipkowski, *Chapter One - Biomimetic Membrane Supported at a Metal Electrode Surface: A  
26 Molecular View*, Academic Press, 2014.
- 27  
28 73. N. G. Hoogeveen, M. A. C. Stuart and G. J. Fler, Polyelectrolyte Adsorption on Oxides: I. Kinetics  
29 and Adsorbed Amounts, *Journal of Colloid and Interface Science*, 1996, **182**, 133-145.
- 30  
31 74. K. Kubiak-Ossowska and P. A. Mulheran, What Governs Protein Adsorption and Immobilization at a  
32 Charged Solid Surface?, *Langmuir*, 2010, **26**, 7690-7694.
- 33  
34 75. A. Márquez, T. Berger, A. Feinle, N. Hüsing, M. Himly, A. Duschl and O. Diwald, Bovine Serum  
35 Albumin Adsorption on TiO<sub>2</sub> Colloids: The Effect of Particle Agglomeration and Surface  
36 Composition, *Langmuir*, 2017, **33**, 2551-2558.
- 37  
38 76. W. Norde and J. Lyklema, The adsorption of human plasma albumin and bovine pancreas ribonuclease  
39 at negatively charged polystyrene surfaces: I. Adsorption isotherms. Effects of charge, ionic strength,  
40 and temperature, *Journal of Colloid and Interface Science*, 1978, **66**, 257-265.
- 41  
42 77. W. Norde, Driving forces for protein adsorption at solid surfaces, *Macromolecular Symposia*, 1996,  
43 **103**, 5-18.
- 44  
45 78. W. Norde and J. Lyklema, The adsorption of human plasma albumin and bovine pancreas ribonuclease  
46 at negatively charged polystyrene surfaces: II. Hydrogen ion titrations, *Journal of Colloid and Interface  
47 Science*, 1978, **66**, 266-276.
- 48  
49 79. W. Norde and J. Lyklema, The adsorption of human plasma albumin and bovine pancreas ribonuclease  
50 at negatively charged polystyrene surfaces: III. Electrophoresis, *Journal of Colloid and Interface  
51 Science*, 1978, **66**, 277-284.
- 52  
53 80. W. Norde and J. Lyklema, The adsorption of human plasma albumin and bovine pancreas ribonuclease  
54 at negatively charged polystyrene surfaces: IV. The charge distribution in the adsorbed state, *Journal of  
55 Colloid and Interface Science*, 1978, **66**, 285-294.
- 56  
57 81. W. Norde and J. Lyklema, The adsorption of human plasma albumin and bovine pancreas ribonuclease  
58 at negatively charged polystyrene surfaces: V. Microcalorimetry, *Journal of Colloid and Interface  
59 Science*, 1978, **66**, 295-302.
- 60  
61 82. W. Norde and J. Lyklema, Thermodynamics of protein adsorption. Theory with special reference to the  
62 adsorption of human plasma albumin and bovine pancreas ribonuclease at polystyrene surfaces,  
63 *Journal of Colloid and Interface Science*, 1979, **71**, 350-366.



- 1
- 2
- 3
- 4 83. K. Wang, C. Zhou, Y. Hong and X. Zhang, A review of protein adsorption on bioceramics, *Interface Focus*, 2012, **2**, 259.
- 5
- 6 84. Y. L. Jeyachandran, E. Mielczarski, B. Rai and J. A. Mielczarski, Quantitative and Qualitative
- 7 Evaluation of Adsorption/Desorption of Bovine Serum Albumin on Hydrophilic and Hydrophobic
- 8 Surfaces, *Langmuir*, 2009, **25**, 11614-11620.
- 9
- 10 85. Z. Yang and C. Zhang, Adsorption/desorption behavior of protein on nanosized hydroxyapatite
- 11 coatings: A quartz crystal microbalance study, *Applied Surface Science*, 2009, **255**, 4569-4574.
- 12
- 13 86. K. D. Dobson and A. J. McQuillan, In situ infrared spectroscopic analysis of the adsorption of aliphatic
- 14 carboxylic acids to TiO<sub>2</sub>, ZrO<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub>, and Ta<sub>2</sub>O<sub>5</sub> from aqueous solutions, *Spectrochimica Acta Part*
- 15 *A: Molecular and Biomolecular Spectroscopy*, 1999, **55**, 1395-1405.
- 16
- 17 87. H. B. Stuart, *Infrared Spectroscopy: Fundamentals and Applications*, WILEY, 2005.
- 18
- 19 88. A. Baral, L. Satish, D. P. Das, H. Sahoo and M. K. Ghosh, Construing the interactions between MnO<sub>2</sub>
- 20 nanoparticle and bovine serum albumin: insight into the structure and stability of a protein-
- 21 nanoparticle complex, *New Journal of Chemistry*, 2017, **41**, 8130-8139.
- 22
- 23 89. C. Aharoni and F. C. Tompkins, *Kinetics of Adsorption and Desorption and the Elovich Equation*,
- 24 Academic Press, 1970.
- 25
- 26 90. Y. S. Ho and G. McKay, Pseudo-second order model for sorption processes, *Process Biochemistry*,
- 27 1999, **34**, 451-465.
- 28
- 29 91. L. Largette and R. Pasquier, A review of the kinetics adsorption models and their application to the
- 30 adsorption of lead by an activated carbon, *Chemical Engineering Research and Design*, 2016, **109**,
- 31 495-504.
- 32
- 33 92. Y. Xiao, J. Azaiez and J. M. Hill, Erroneous Application of Pseudo-Second-Order Adsorption Kinetics
- 34 Model: Ignored Assumptions and Spurious Correlations, *Industrial & Engineering Chemistry*
- 35 *Research*, 2018, **57**, 2705-2709.
- 36
- 37 93. A. Pavlatou and N. A. Polyzopoulos, The role of diffusion in the kinetics of phosphate desorption: the
- 38 relevance of the Elovich equation, *Journal of Soil Science*, 1988, **39**, 425-436.
- 39
- 40 94. H. Wu, N. I. Gonzalez-Pech and V. H. Grassian, Displacement reactions between environmentally and
- 41 biologically relevant ligands on TiO<sub>2</sub> nanoparticles: insights into the aging of nanoparticles in the
- 42 environment, *Environmental Science: Nano*, 2019, **6**, 489-504.
- 43
- 44 95. M. Bayrakci, O. Gezici, S. Z. Bas, M. Ozmen and E. Maltas, Novel humic acid-bonded magnetite
- 45 nanoparticles for protein immobilization, *Materials Science and Engineering: C*, 2014, **42**, 546-552.
- 46
- 47 96. Y. Li, W. Tan, L. K. Koopal, M. Wang, F. Liu and W. Norde, Influence of Soil Humic and Fulvic Acid
- 48 on the Activity and Stability of Lysozyme and Urease, *Environmental Science & Technology*, 2013, **47**,
- 49 5050-5056.
- 50
- 51 97. W. F. Tan, L. K. Koopal and W. Norde, Interaction between Humic Acid and Lysozyme, Studied by
- 52 Dynamic Light Scattering and Isothermal Titration Calorimetry, *Environmental Science & Technology*,
- 53 2009, **43**, 591-596.
- 54
- 55 98. J. E. Tomaszewski, R. P. Schwarzenbach and M. Sander, Protein Encapsulation by Humic Substances,
- 56 *Environmental Science & Technology*, 2011, **45**, 6003-6010.
- 57
- 58 99. P. Roach, D. Farrar and C. C. Perry, Surface Tailoring for Controlled Protein Adsorption: Effect of
- 59 Topography at the Nanometer Scale and Chemistry, *Journal of the American Chemical Society*, 2006,
- 60 **128**, 3939-3945.
100. Rustem I. Litvinov, Dzhigangir A. Faizullin, Yuriy F. Zuev and John W. Weisel, The  $\alpha$ -Helix to  $\beta$ -
- Sheet Transition in Stretched and Compressed Hydrated Fibrin Clots, *Biophysical Journal*, 2012, **103**,

- 1  
2  
3 1020-1027.
- 4 101. A. Adamiano, G. Lesci Isidoro, D. Fabbri and N. Roveri, Adsorption of bovine serum albumin onto  
5 synthetic Fe-doped geomimetic chrysotile, *Journal of The Royal Society Interface*, 2015, **12**, 20150186.
- 6 102. B. E. Givens, N. D. Diklich, J. Fiegel and V. H. Grassian, Adsorption of bovine serum albumin on  
7 silicon dioxide nanoparticles: Impact of pH on nanoparticle–protein interactions, *Biointerphases*, 2017,  
8 **12**, 02D404.
- 9 103. H. Pan, M. Qin, W. Meng, Y. Cao and W. Wang, How Do Proteins Unfold upon Adsorption on  
10 Nanoparticle Surfaces?, *Langmuir*, 2012, **28**, 12779-12787.
- 11 104. A. Sethuraman, G. Vedantham, T. Imoto, T. Przybycien and G. Belfort, Protein unfolding at interfaces:  
12 Slow dynamics of  $\alpha$ -helix to  $\beta$ -sheet transition, *Proteins: Structure, Function, and Bioinformatics*,  
13 2004, **56**, 669-678.
- 14 105. I. Noda, Vibrational two-dimensional correlation spectroscopy (2DCOS) study of proteins,  
15 *Spectrochimica Acta Part A: Molecular and Biomolecular Spectroscopy*, 2017, **187**, 119-129.
- 16 106. V. A. Shashilov and I. K. Lednev, Two-dimensional correlation Raman spectroscopy for characterizing  
17 protein structure and dynamics, *Journal of Raman Spectroscopy*, 2009, **40**, 1749-1758.
- 18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

## TOC/Abstract Art



The adsorption and unfolding behavior of bovine serum albumin onto catalytic- and food-grade titanium dioxide nanoparticles is dependent on the surface chemistry of the nanoparticles and their environmental exposure history.