



Metallomics

Identification and Physiological Comparison of Plant Species that Show Positive or Negative Co-Occurrence with Selenium Hyperaccumulators

Journal:	<i>Metallomics</i>
Manuscript ID	MT-ART-09-2019-000217.R1
Article Type:	Paper
Date Submitted by the Author:	30-Oct-2019
Complete List of Authors:	Reynolds, Ray Jason; Colorado State University, Biology Jones, Rachel; Colorado State University, Biology Stonehouse, Gavin; Colorado State University, Biology El Mehdawi, Ali; Colorado State University, Biology Lima, Leonardo; Colorado State University, Biology Fakra, Sirine; Lawrence Berkeley National Laboratory Pilon-Smits, Elizabeth; Colorado State University, Biology

SCHOLARONE™
Manuscripts

ARTICLE

Identification and Physiological Comparison of Plant Species that Show Positive or Negative Co-Occurrence with Selenium Hyperaccumulators

Received 00th January 20xx,
Accepted 00th January 20xx

DOI: 10.1039/x0xx00000x

R. Jason B. Reynolds^a, Rachel R. Jones^a, Gavin C. Stonehouse^a, Ali F. El Mehdawi^a, Leonardo W. Lima^a, Sirine C. Fakra^b and Elizabeth A.H. Pilon-Smits^{a*}

In these studies we identified and compared the properties of plant species that showed positive or negative co-occurrence with selenium (Se) hyperaccumulators in their natural habitat. The main questions addressed were: which species are most abundant directly adjacent to hyperaccumulators, and which are absent? How do Se accumulation and tolerance compare in species found to positively or negatively co-occur with hyperaccumulators? Approaches included field surveys, x-ray microprobe analysis of field samples, and a lab Se tolerance and accumulation study. When 54 hyperaccumulators across two naturally seleniferous sites were surveyed for their five nearest neighboring species, and the relative abundance of these species around hyperaccumulators compared to that in the overall vegetation, some species were identified to positively or negatively co-occur with hyperaccumulators. Several positively co-occurring species showed high Se accumulation capability (up to 900 mg Se/kg dry weight), which may reflect Se tolerance. Leaf X-ray microprobe analysis found relatively more organic forms of Se in two positively co-occurring species than in a negatively co-occurring one. There were elevated soil Se levels around Se hyperaccumulators, and neighbors of Se hyperaccumulators had a higher tissue Se concentration as compared to when the same species grew elsewhere in the area. The elevated soil Se levels around Se hyperaccumulators –likely resulting from litter deposition– may significantly affect the local plant community, facilitating Se-tolerant plant community members but lowering the fitness of Se-sensitive members.

Introduction

The element selenium (Se) is a micronutrient for many species, but becomes toxic at elevated levels, with a narrow concentration range between sufficiency and toxicity¹. The natural Se distribution in the earth's crust is variable, and its primary bioavailable form in soils is SeO_4^{2-} (selenate)².

While Se is not essential for plants, it can be taken up and assimilated due to the chemical similarity between Se and the plant macronutrient sulfur (S)³. One reason why Se is toxic is the non-specific incorporation of selenocysteine into proteins, which can disrupt protein function; the other mechanism by which Se becomes toxic to plants is due to oxidative stress from inorganic selenate and selenite⁴.

Interestingly, Se can benefit plant growth at low tissue levels, typically <5 mg Se/kg DW (dry weight), by providing physiological benefits related to induction of oxidative stress resistance mechanisms⁵. Moreover, at higher levels, plants can enjoy ecological benefits from Se accumulation, owing to

protection from fungal infections or herbivory⁶. The tissue concentration where plants start to experience toxicity varies by species: some are quite Se-sensitive (toxicity <100 mg Se/kg DW), while others are moderately Se tolerant (toxicity >500 mg Se/kg DW). Remarkably, some plants show extreme Se tolerance and also hyperaccumulate Se to >1,000 mg Se/kg DW while growing in naturally seleniferous habitats⁷. Examples are *Astragalus bisulcatus* (Fabaceae) which may contain 14,000 mg Se/kg DW and *Stanleya pinnata* (Brassicaceae), with 4,000 mg Se/kg DW⁸. These hyperaccumulators also differ from other species in that they have higher Se to S ratios, accumulate relatively more organic Se (particularly methyl-selenocysteine), and sequester Se in specific tissues (leaf epidermis and margins) and organs (reproductive organs)^{7,9}.

Likely, Se hyperaccumulation serves as an elemental defense, protecting plants from generalist herbivores and some fungal pathogens¹⁰⁻¹⁴. Furthermore, concentration of Se around hyperaccumulators may result in elemental allelopathy, inhibiting the growth of Se-sensitive other plant species¹⁵. Supporting evidence for Se-mediated allelopathy is that soil collected adjacent to hyperaccumulators was found to be Se-enriched (~10 fold), and inhibited germination and growth of Se-sensitive species *Arabidopsis thaliana*¹⁵. This Se enrichment is thought to be due to deposition of hyperaccumulator litter, and perhaps root exudation, all in the form of highly bioavailable organic selenocompounds^{16,17}.

^a Colorado State University, Biology Department, Fort Collins, CO 80523, USA.

^b Advanced Light Source, Lawrence Berkeley National Laboratory, Berkeley, CA 94720, USA

* Corresponding author: epsmits@colostate.edu Tel.: +1-970-491-4991

1
2
3 Interestingly, there is also evidence that some native plants
4 growing around Se hyperaccumulators are positively affected
5 by the concentrated Se around hyperaccumulators¹⁸. Their
6 proximity to Se hyperaccumulators was associated with higher
7 Se accumulation, higher biomass and less herbivory damage.
8 Thus, if neighbors of hyperaccumulators are able to tolerate the
9 associated high levels of Se, positive allelopathy may result,
10 which may be a combination of physiological benefits
11 (enhanced growth) as well as protection from herbivory^{18,19}.
12 However, if they cannot tolerate the Se, they may suffer
13 toxicity¹⁵.

14
15 The hypothesis that may be derived from these earlier
16 observations is that through these allelopathic processes,
17 hyperaccumulators may affect the fitness of other plant species,
18 favoring Se-tolerant individuals and exerting a negative effect
19 on Se-sensitive individuals. This may over time lead to
20 differences in vegetation patterns relative to communities
21 without hyperaccumulators. In an earlier study it was found
22 that sparser vegetation cover was observed directly around
23 hyperaccumulators, and there were some differences in species
24 composition¹⁷. The studies described here focused on
25 identifying plant species in the community that showed the
26 clearest positive or negative co-occurrence with
27 hyperaccumulators, and characterizing their Se-related
28 properties. The specific questions addressed were: 1) How does
29 soil Se concentration relate to distance from Se
30 hyperaccumulators? 2) Is Se concentration in a
31 hyperaccumulator correlated with that in its neighboring
32 plants? 3) Which species are found in the near vicinity of
33 hyperaccumulators (<50 cm, under the canopy), and do they
34 occur more frequently there, as compared to the overall
35 landscape? 4) How do Se tolerance and accumulation
36 properties compare between species that positively or
37 negatively co-occur with hyperaccumulators, and what are the
38 patterns of Se tissue distribution and chemical speciation in
39 these species?
40

41 First, a field survey approach was used across two naturally
42 seleniferous sites, to pinpoint species of interest. The shale soils
43 at these sites support several Se hyperaccumulator plant
44 species, as described in earlier studies^{8,15}. The survey was
45 followed by a controlled laboratory Se tolerance and
46 accumulation experiment using selected species. Furthermore,
47 x-ray microprobe analysis was done to image the distribution of
48 Se and other elements in intact frozen field-collected samples
49 from species of interest, as well as forms of Se accumulated (Se
50 speciation).
51

52 Results

53
54 Fifty-four hyperaccumulator individuals from the species *S.*
55 *pinnata* or *A. bisulcatus* were selected across two field sites
56 (Pine Ridge and Coyote Ridge natural areas, Fort Collins, CO,
57 USA). The five species growing nearest to these
58 hyperaccumulators were recorded and counted (Table 1, "Near
59 5 counts"). The relative species abundance (RSA) data
60

calculated from these counts, as a fraction of the overall RSA of
the same species were used to calculate the ratios shown in
Figure 1. Plant species found relatively more frequently near
hyperaccumulators than expected based on Daubenmire
surveys have ratios >1, and species that were found less
frequently next to hyperaccumulators compared to overall
vegetation have ratios <1 (Fig. 1, left side vs. right side).

Species found more frequently growing near
hyperaccumulators at all three sites are *Alyssum simplex*,
Oenothera suffrutescens and *Tragopogon dubius*, and two
species were found to positively co-occur in two of the three
sites: *Bromus tectorum* and *Nasella viridula*. The species with
the highest Near 5/Daubenmire ratio was *Ericameria nauseosa*
in the Pine Ridge 2 site. In regards to species less abundant near
hyperaccumulators than expected, there were three that stood
out: *Hesperostipa comata*, *Euphorbia brachycera* and *Rosa*
woodsii; each of these species were at least four-fold less
abundant near hyperaccumulators than expected based on
Daubenmire data. It should be noted that only species that
occurred in both the Daubenmire plots and as one of the
nearest five species are included in Figure 1.

The average leaf Se concentration in those plants found as one
of the five nearest species to the hyperaccumulator ranged
from 4.3 mg Se/kg dry weight for *Bouteloua dactyloides* to 2,974
mg Se/kg DW for *Symphyotrichum ericoides* (Table 1). The
hyperaccumulator *S. pinnata* averaged 1,459 ± 196 mg Se/kg
dry weight for Pine Ridge site 1 and 1,107 ± 28 mg Se/kg dry
weight for Coyote Ridge; there was no *S. pinnata* located at Pine
Ridge site 2. The hyperaccumulator *A. bisulcatus* averaged
8,284 ± 992 mg Se/kg dry weight for Pine Ridge site 1 and 13,927
± 253 mg Se/kg dry weight for Pine Ridge site 2; there was no *A.*
bisulcatus located at the Coyote Ridge site. After *S. ericoides*
(hyperaccumulator co-occurrence pattern: +), the species next
highest in Se were *Gutierrezia sarothrae* (+), *Townsendia*
hookeri (+), *Physaria bellii* (+), and *Bromus inermis* (-/+,
depending on the site).

To investigate the effect of the hyperaccumulator's proximity in
the field, when possible the leaf Se concentration was also
determined of plants of the same species, collected at a
minimum of 50 m from a hyperaccumulator. The Se
concentration was lower at this longer distance for all species
surveyed (Table 1B, near vs. far columns). The effect of the
hyperaccumulator's proximity on soil Se concentration was also
investigated: Se concentration in soil was analyzed at 0.5 meter
successive intervals from the stem of five hyperaccumulator
plants (0 m) to 2 m from the plant across all directions. The soil
Se concentration significantly decreased with distance from the
hyperaccumulator (Fig. 2).

Table 1. Plant species occurrence as influenced by the vicinity of hyperaccumulators at Pine Ridge locations 1 (A, lower in Se) and 2 (B, higher in Se) and Coyote Ridge (C). "Species code": USDA abbreviations of species. "Near 5 counts": number of times species was observed as one of the nearest 5 species to a hyperaccumulator. "Near 5 RSA": relative species abundance, i.e. fraction of the total of all 5 nearest species to all hyperaccumulators (5 x 54 = 270 counts). "Daub RSA": relative species abundance for that species based on Daubenmire plot canopy cover data. Last column in A and second to last column in B show the average Se concentration of all collected leaves for plants found as one of the 5 nearest to hyperaccumulators \pm SEM. The last column in B ("far") shows the plant Se concentration of the same species found growing on the same site but at least 50 m from any hyperaccumulator. ND: no data. If there are no values for hyperaccumulators ASB12 and STPI, they were not encountered as one of the nearest five to the other hyperaccumulator or they were not surveyed in the Daubenmire plots.

A

Species code	Species name	Near 5 Counts	Near 5 RSA	Daub % Canopy	Daub RSA	Avg. mg Se/kg DW
ALSI	<i>Alyssum simplex</i>	4	9.1	3.1	5.8	10 \pm 1
AMPS	<i>Ambrosia psilostachya</i>	1	2.3	0.1	0.2	18
ASB12	<i>Astragalus bisulcatus</i>	0	0	0.6	1.1	335 \pm 40
BOCU	<i>Bouteloua curtipendula</i>	3	6.8	6.9	13.1	3 \pm 1
BOD2	<i>Bouteloua dactyloides</i>	5	11.4	8.2	15.4	4 \pm 1
BOGR2	<i>Bouteloua gracilis</i>	1	2.3	0	0	12
BRIN2	<i>Bromus inermis</i>	1	2.3	0.2	0.4	5
COAR4	<i>Convolvulus arvensis</i>	7	15.9	2.4	4.5	10 \pm 1
MEOF	<i>Melilotus officinalis</i>	6	13.6	2.4	4.5	17 \pm 3
NAVI	<i>Nassella viridula</i>	2	4.6	0.3	0.6	10 \pm 0.3
OESU3	<i>Oenothera suffrutescens</i>	2	4.6	0.6	1.2	9 \pm 5
PASM	<i>Pascopyrum smithii</i>	8	18.2	11.8	22.2	9 \pm 1
TAOF	<i>Taraxacum officinale</i>	1	2.3	0	0	ND
TRDU	<i>Tragopogon dubius</i>	3	6.8	0.3	0.6	20 \pm 10

B

Species code	Species name	Near 5 Counts	Near 5 RSA	Daub % Canopy	Daub RSA	avg mg Se/kg DW	Near	Far
ALSI	<i>Alyssum simplex</i>	3	4.7	4.1	2.6	337 \pm 189	0	
ARLU	<i>Artemisia ludoviciana</i>	6	3.5	2.9	2.3	183 \pm 28	0	
ASB12	<i>Astragalus bisulcatus</i>	0	0	1.5	2.4	8284 \pm 992	n/a	
ASPU	<i>Asclepias pumila</i>	1	0.4	4.0	6.3	39 \pm 0	ND	
BRTE	<i>Bromus tectorum</i>	1	4.7	2.0	1.5	452 \pm 0	ND	
BRIN2	<i>Bromus inermis</i>	3	1.6	22.9	18.1	369 \pm 73	ND	
ERNA10	<i>Ericameria nauseosa</i>	1	0.8	1.7	1.5	223 \pm 0	ND	
HEAN3	<i>Helianthus annuus</i>	2	3.1	6.6	4.4	219 \pm 106	3.4	
OESU3	<i>Oenothera suffrutescens</i>	1	2.0	1.0	0.7	397 \pm 0	ND	
OPPO	<i>Opuntia polyacantha</i>	1	0.4	0.2	0.3	95 \pm 0	ND	
PASM	<i>Pascopyrum smithii</i>	4	6.7	18.0	10.8	503 \pm 78	13.8	
PSSPS	<i>Pseudoroegneria spicata</i>	2	6.7	3.0	3.3	54 \pm 19	ND	
RHTR	<i>Rhus trilobata</i>	1	1.2	2.5	1.7	41 \pm 0	6.7	
SYER	<i>Symphotrichum ericoides</i>	3	1.6	3.0	2.0	3838 \pm 309	0.4	
STPI	<i>Stanleya pinnata</i>	0	0	0	0	1459 \pm 196	n/a	
TRDU	<i>Tragopogon dubius</i>	3	1.6	1.9	1.5	300 \pm 70	64.7	
YUGL	<i>Yucca glauca</i>	1	6.7	2.0	1.2	39 \pm 0	ND	

C

Species code	Species	OBS Counts	OBS RSA	EXP % Canopy	EXP RSA	Avg. (mg Se/kg DW)
ACHY	<i>Achnatherum hymenoides</i>	6	0.024	1.36	0.010	28 \pm 6
ALSI	<i>Alyssum simplex</i>	12	0.047	4.10	0.026	4 \pm 1
AMPS	<i>Ambrosia psilostachya</i>	2	0.008	0.10	0.002	10
ARDR4	<i>Artemisia dracunculul</i>	2	0.011	0	0	30 \pm 0.3
ARFR4	<i>Artemisia frigida</i>	1	0.004	0.37	0.002	50
ARLU	<i>Artemisia ludoviciana</i>	9	0.035	2.85	0.023	20 \pm 9
ARPU9	<i>Aristida purpurea</i>	2	0.008	2.69	0.027	8 \pm 4
ASB12	<i>Astragalus bisulcatus</i>	0	0	0.14	0.35	1458 \pm 94
ASMO7	<i>Astragalus mollissimus</i>	1	0.006	0	0	219
ASTE5	<i>Astragalus tenellus</i>	4	0.016	0.74	0.018	45 \pm 12
BOCU	<i>Bouteloua curtipendula</i>	10	0.039	7.33	0.046	57 \pm 10
BRAR5	<i>Bromus arvensis</i>	12	0.047	2.00	0.015	43 \pm 4
COUM	<i>Comandra umbellata</i>	2	0.008	1.01	0.008	213 \pm 12
ELEL5	<i>Elymus elymoides</i>	1	0.004	0.52	0.005	71
ERNA10	<i>Ericameria nauseosa</i>	2	0.008	1.65	0.015	6
EUBR	<i>Euphorbia brachycera</i>	1	0.004	0.69	0.017	68
EVNU	<i>Evolvulus nuttallianus</i>	3	0.012	1.75	0.012	152 \pm 66
GUSA2	<i>Gutierrezia sarothrae</i>	3	0.012	1.86	0.019	597 \pm 222
HEAN3	<i>Helianthus annuus</i>	8	0.031	6.64	0.044	28 \pm 6
HECO26	<i>Hesperostipa comata</i>	20	0.079	22.84	0.163	59 \pm 4
LASE	<i>Lactuca serriola</i>	9	0.035	0.05	0.001	89 \pm 12
LIDA	<i>Linaria dalmatica</i>	2	0.008	1.26	0.012	188 \pm 101
LILE3	<i>Linum lewisii</i>	5	0.020	2.09	0.015	46 \pm 15
MEDE	<i>Mentzelia decapetala</i>	1	0.004	0.31	0.008	27
NAVI	<i>Nassella viridula</i>	7	0.028	4.73	0.026	48 \pm 10
OESU3	<i>Oenothera suffrutescens</i>	5	0.020	1.03	0.007	31 \pm 10
OEVI	<i>Oenothera villosa</i>	2	0.008	0.84	0.007	96 \pm 64
PASM	<i>Pascopyrum smithii</i>	17	0.067	18.00	0.108	70 \pm 23
PHBE2	<i>Physaria bellii</i>	5	0.020	0.17	0.004	375 \pm 69
POPR	<i>Poa pratensis</i>	2	0.008	0.24	0.004	156 \pm 108
PSSPS	<i>Pseudoroegneria spicata</i>	17	0.067	3.03	0.033	62 \pm 5
PSTE5	<i>Psoraleidium tenuiflorum</i>	3	0.012	2.53	0.017	8 \pm 1
RACO3	<i>Ratibida columnifera</i>	1	0.004	0.31	0.008	16
RHTR	<i>Rhus trilobata</i>	4	0.016	3.01	0.020	5 \pm 2
ROWO	<i>Rosa woodsii</i>	1	0.004	1.15	0.012	6
STPI	<i>Stanleya pinnata</i>	1	0.004	0.57	0.014	1107 \pm 28
SYAS3	<i>Symphotrichum ascendens</i>	3	0.012	0.95	0.023	17 \pm 8
SYER	<i>Symphotrichum ericoides</i>	4	0.016	1.88	0.015	382
SYFE	<i>Symphotrichum fendleri</i>	3	0.017	0	0	1129 \pm 126
THME	<i>Thelesperma megapotamicum</i>	2	0.008	0.33	0.003	150 \pm 78
TOHO	<i>Townsendia hookeri</i>	3	0.012	0.24	0.006	394 \pm 85
TRDU	<i>Tragopogon dubius</i>	17	0.067	2.01	0.012	53 \pm 7
TRRA5	<i>Tragia ramosa</i>	5	0.020	2.79	0.033	31 \pm 6
VETH	<i>Verbascum thapsus</i>	2	0.008	0.05	0.000	19 \pm 7

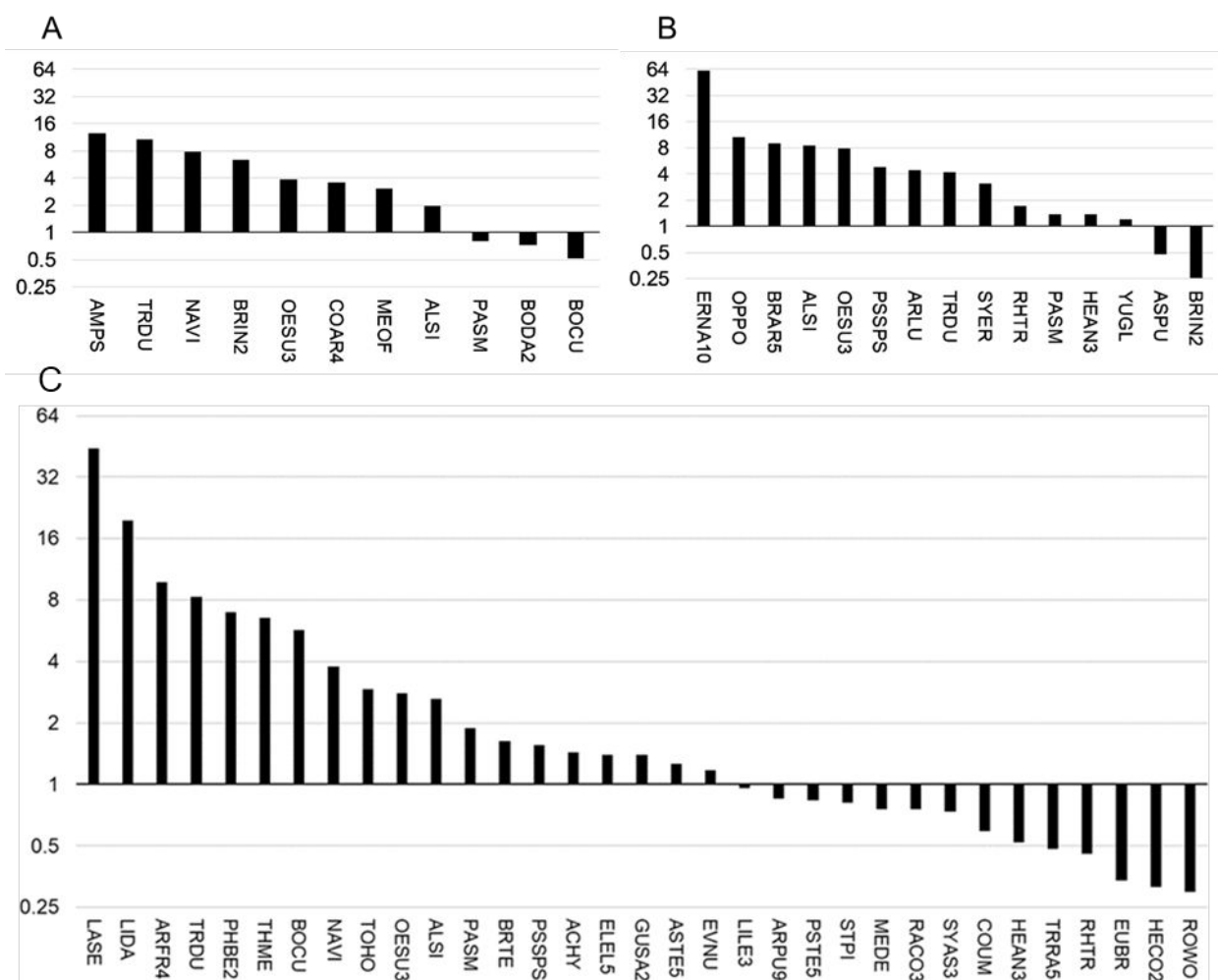


Figure 1. Positive and negative co-occurrence of different plant species with Se hyperaccumulator species at Pine Ridge, locations 1 (A, lower in Se) and 2 (B, higher in Se) and Coyote Ridge (C). See Table 1 for explanation of species abbreviations, and the observation data used to calculate the ratios shown here, which represent the fold difference in the frequency the species as one of the nearest five to hyperaccumulators relative to the frequency of the same species in the overall area (from Daubenmire plot data).

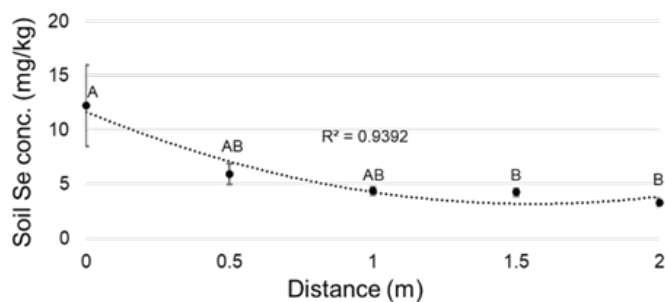


Figure 2. Soil Se concentration measured at half meter successive intervals from hyperaccumulators (n=5). Shown data represent mean and Standard Error of the Mean. Letters above graph indicate which means were significantly different (ANOVA and Tukey Kramer, $p < 0.01$).

Selected species that differed in their co-occurrence properties (Table 1, Fig. 1) were further investigated for their Se speciation and localization, as well as their Se tolerance. *Bromus inermis* was selected as a negatively co-occurring species, and *Alyssum simplex* and *Artemisia frigida* as positively co-occurring species. Leaves from these three species were collected in the field and analyzed by x-ray microprobe analysis for Se distribution (micro x-ray fluorescence, μ XRF) and chemical speciation (micro x-ray absorption near-edge structure, μ XANES). Selenium was distributed throughout the leaf of *B. inermis* (Fig. 3A, B), with a concentration in a pattern of straight parallel lines, likely the vasculature. The chemical speciation of Se was determined in the *B. inermis* leaf locations indicated (Fig. 3B) via Se K-edge XANES. Around half of the Se appeared to be organic and the other half inorganic (Fig. 3C). The organic forms made up half of the Se and were fitted as 40% C-Se-C compounds (possibly selenomethionine or methyl-selenocysteine) and 9% selenodiglutathione (SeGSH₂, C-S-Se-S-C). The other half consisted of inorganic Se and was fitted as elemental Se (26% Se⁰), selenite (17% Se^{IV}) and selenate (5% Se^{VI}).

The second species tested, *A. simplex*, showed diffuse Se distribution throughout the leaf, with increased concentration in the central vein and slight concentration in the other vasculature and in the stellate trichomes (Fig. 4A-C). Calcium was very concentrated in the trichomes (much more so than

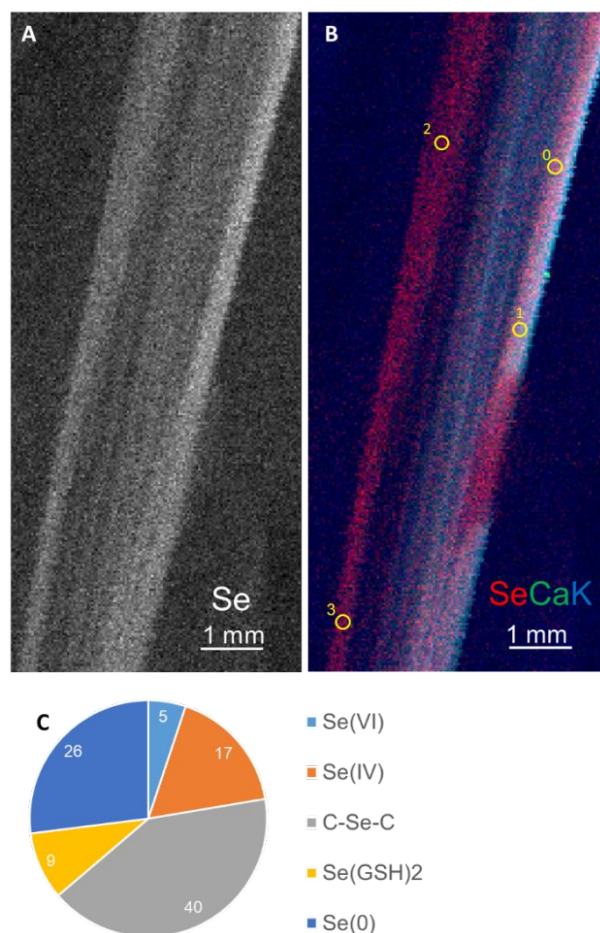


Figure 3. Selenium localization and speciation in *Bromus inermis* leaf collected at Pine Ridge. (A) μ XRF distribution map of Se. (B) Tricolor-coded μ XRF map of Se (red), Ca (green) and K (blue). Yellow circles indicate locations of Se K-edge XANES spots. (C) Results of least-square linear combination fitting of Se XANES spectra.

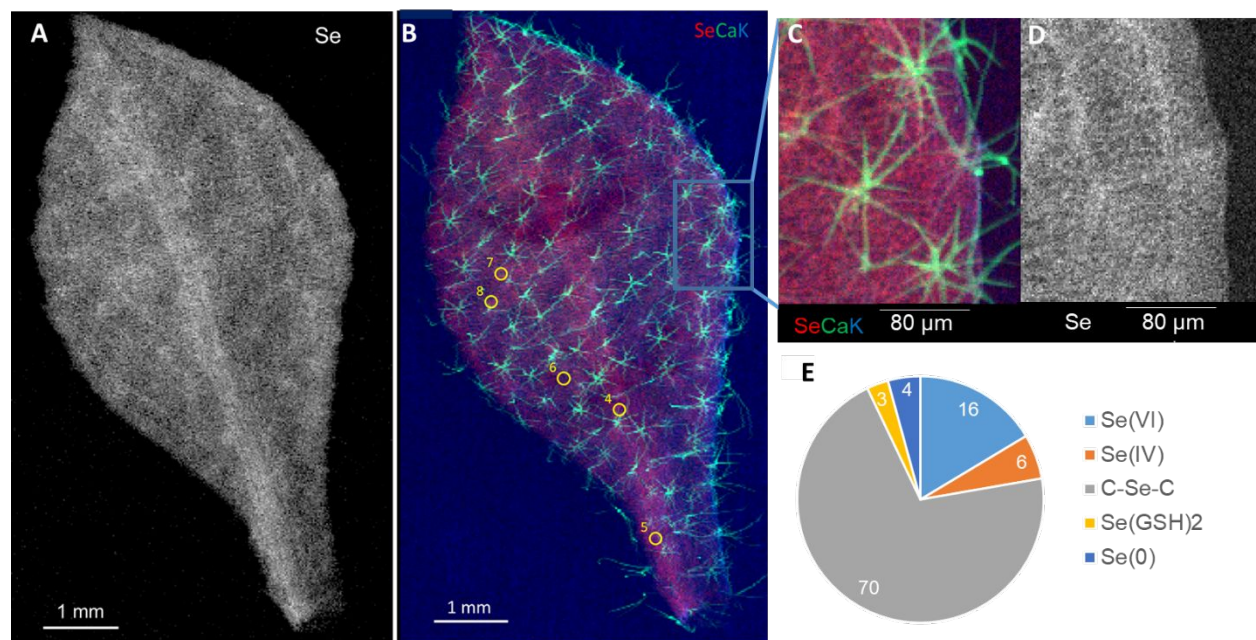
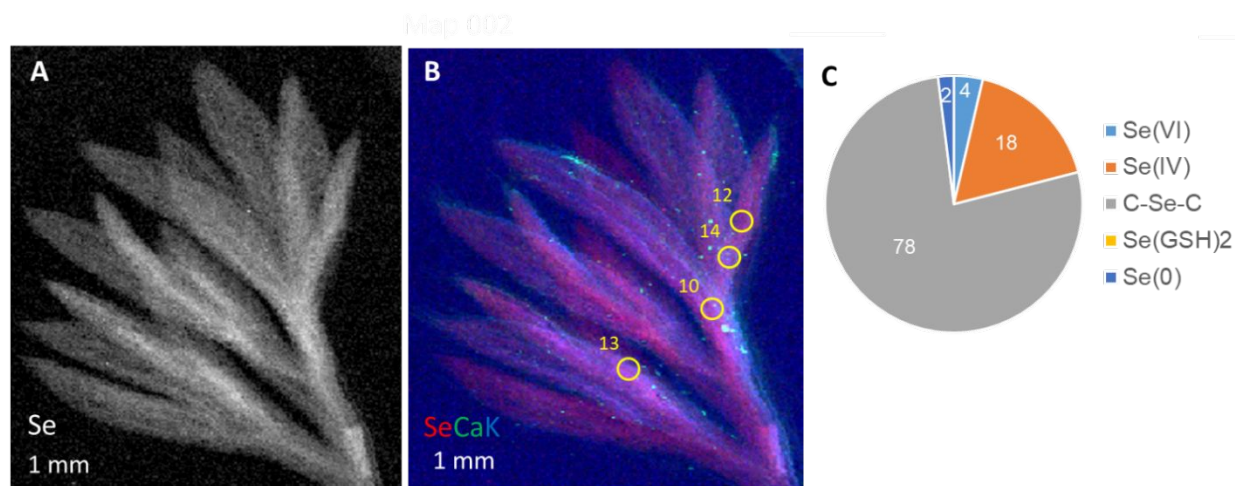


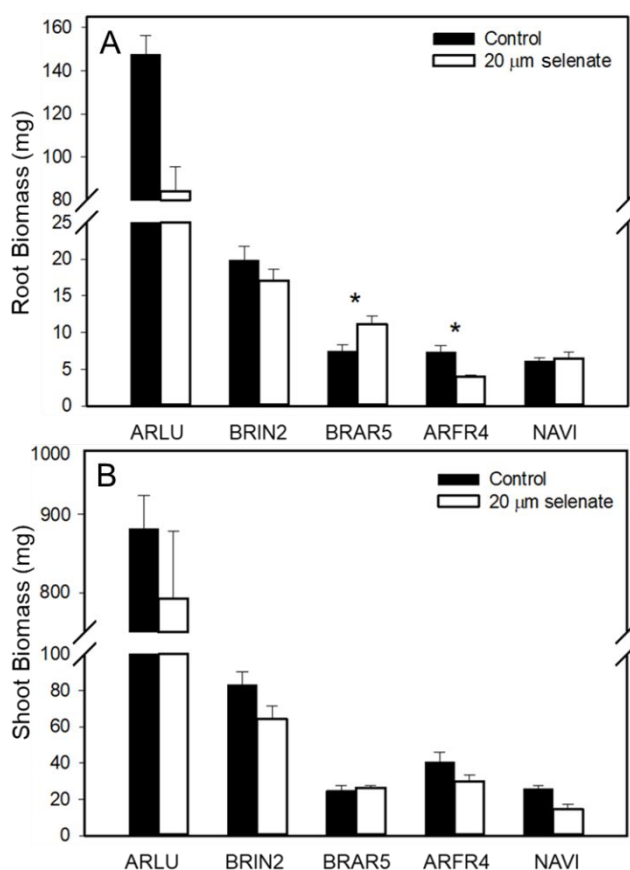
Figure 4. Selenium localization and speciation in *Alyssum simplex* leaf collected at Pine Ridge. (A, D) μ XRF distribution maps of Se. (B, C) Tricolor-coded μ XRF maps of Se (red), Ca (green) and K (blue). Yellow circles indicate locations of Se K-edge XANES spots. (E) Results of least-square linear combination fitting of Se XANES spectra.



Se), while potassium was present throughout the leaf (Fig. 4B). The form of Se was investigated in the *A. simplex* leaf at the positions indicated (Fig. 4B), and found to be predominantly organic (73%), consisting of 70% C-Se-C compounds and 3% SeGSH2 (Fig. 4D). The inorganic Se fraction was comprised of three forms: 16% Se(VI), 6% Se(IV) and 4% Se(0).

The third species tested, *A. frigida*, also showed diffuse Se distribution in the leaf with clearly higher levels in what appears to be the vasculature (Fig. 5A, B). XANES revealed that the Se in the *A. frigida* leaf at the locations indicated (Fig. 5B) consisted primarily of organic Se with C-Se-C structure (78%); the remainder was inorganic: 18% Se(IV), 4% Se(VI) and 2% Se(0)

Figure 5. Selenium localization and speciation in *Artemisia frigida* leaf collected at Pine Ridge. (A) μ XRF distribution map of Se. (B) Tricolor-coded μ XRF map of Se (red), Ca (green) and K (blue). Yellow circles indicate locations of Se K-edge XANES spots. (C) Results of least-square linear combination fitting of Se XANES spectra.



(Fig. 5C). Thus, Se localization was similar across the three species, showing diffuse distribution throughout the leaf with concentration in the vasculature. *Alyssum simplex* additionally appeared to store some Se in its trichomes. The Se speciation differed between negatively and positively co-occurring species in that the positively co-occurring species had relatively more organic Se (three-quarters vs. half).

A follow-up lab Se tolerance study was carried out, using seeds from species that differed in co-occurrence (+, - or neutral) with hyperaccumulators, collected from the areas where the field study was performed. Five species had sufficient germination to provide meaningful results: *Artemisia ludoviciana* (+)(ARLU), *Bromus inermis* (-) (BRIN2), *Bromus tectorum* (+) (BRTE), *Artemisia frigida* (+) (ARFR4) and *Nasella viridula* (~0) (NAVI) (Fig. 1). The average biomass for plants of these five species was compared between a 20 μ m selenate treatment and a control treatment (Fig. 6). For *A. ludoviciana*, *B. inermis* and *A. frigida*, the root and shoot biomass were on average lower for plants treated with selenate than for the control treatment, but this was only statistically significant for roots of *A. frigida* (Fig. 6). For *N. viridula* root biomass was unaffected, but shoot biomass less for plants treated with selenate (Fig. 6). Exceptionally, for

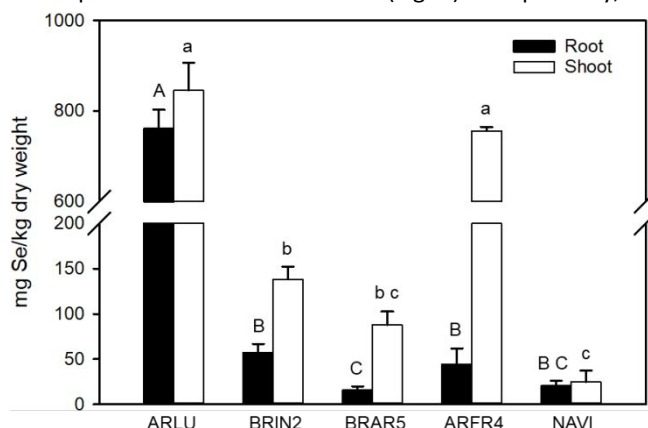


Figure 7. Selenium concentration in the roots and shoots of species treated with or without 20 μ m sodium selenate in controlled lab conditions. Full names of the species can be found in Table 1. In cases where Se was not detectable in one or more of the replicates, half of the detection limit was used, and the resulting values thus are estimates; this is the case for BRAR5, ARFR4 and NAVI root as well as NAVI shoot.

B. tectorum the root biomass was significantly larger with selenate than without; shoot biomass was unaffected.

When the selenate-supplied plants were analyzed for their tissue Se concentration, the two *Artemisia* species had by far the highest levels (Fig 7). In *A. ludoviciana*, shoot Se was ~800 mg/kg DW and *A. frigida* contained ~700 mg Se/kg DW. When compared to the next highest shoot concentration in *B. inermis*, *A. ludoviciana* was around five-fold higher. For roots, *A. ludoviciana* had ~700 mg Se/kg DW while *B. inermis* (the next highest) had a 10-fold lower level. The Se concentration was generally higher in shoot than in root for all species, but this was much less pronounced for *A. ludoviciana* and *N. viridula*.

Discussion

This study addressed a significant question in regard to the ecology of hyperaccumulators: Does the presence of Se hyperaccumulators correspond with differences in their local plant communities, particularly with respect to the presence or absence of other plant species near hyperaccumulators? The hypothesis was that some species would be disproportionately more or less abundant around hyperaccumulators (*A. bisulcatus* and *S. pinnata*), depending on their Se sensitivity: less sensitive species showing positive co-occurrence, and more Se-sensitive species negative co-occurrence. Candidate species, found to occur more or less frequently near hyperaccumulators than in the overall landscape, were then further characterized by lab uptake and tolerance studies and x-ray microprobe analysis.

Twenty-two plant species were found at least three-fold more frequently near hyperaccumulators than could be expected based on their overall local abundance, and thus can be said to show positive co-occurrence with Se hyperaccumulators (Fig. 1). The species of hyperaccumulator, *A. bisulcatus* or *S. pinnata*, did not matter in this respect, so the effect may be more related to a higher Se content experienced by neighbors rather than any other species-specific effect like nitrogen fixing capability (data not shown). The average Se concentrations for all of these positively co-occurring individuals exceeded levels that begin to show toxicity for most plant species^{20,21} and in some cases were at or near Se hyperaccumulator levels (>1,000 mg Se/kg DW)¹⁰. The Se levels in plants growing near hyperaccumulators were higher than in plants of the same species growing further from hyperaccumulators. Moreover, the soil Se concentration under the canopy of hyperaccumulators was found to be elevated. Thus, it appears that hyperaccumulators are surrounded by a Se "hot spot", and plants growing in this area experience elevated tissue Se levels. Some species that were found to be relatively more abundant in zones surrounding hyperaccumulators may benefit from these elevated Se levels while other species may experience a negative effect from growing there, perhaps due to Se toxicity, as hypothesized. Positive effects of Se on plants may be physiological or ecological. Selenium at tissue levels <5 mg Se/kg DW can benefit plant physiology by increasing antioxidant capacity²². At higher levels (>5 mg Se/kg DW), Se increasingly protects plants from herbivory¹¹⁻¹⁴. Of course, to

enjoy these positive ecological effects, plants have to tolerate the tissue Se levels they experience.

Tragopogon dubius was found in all three sites to be at least four-fold over-abundant near hyperaccumulators. *Tragopogon dubius* is an annual or biennial that is often found on disturbed ground (<http://swbiodiversity.org>). Its overabundance may be related to its ability to take advantage of bare ground, which was found to be more prevalent near hyperaccumulators in our studies; a similar trend was found in earlier study¹⁵. The average Se levels in *T. dubius* were 91 ± 30 mg/kg DW (Table 1) across all three sites, with a high of 488 mg Se/kg DW at Pine Ridge 2. Thus, it appears to be fairly Se-tolerant. Interesting to note in this respect is that in the lab Se tolerance study, the single *T. dubius* plant obtained accumulated 882 mg Se/kg DW in its shoot when fed 20 μ M Se, with no apparent toxicity symptoms (results not shown due to lack of replication).

Oenothera suffeutescens was also found on all three sites to positively co-occur with hyperaccumulators. The Se levels found in this species in the field indicate a capacity to accumulate relatively high levels of Se (up to 397 mg Se/kg DW), coupled with Se tolerance. Similar to *T. dubius*, this species is often found in disturbed areas and has low water needs (<http://swbiodiversity.org>).

The third species found to positively co-occur with hyperaccumulators on all three sites was *Alyssum simplex*. It had high Se levels growing in the field (up to 603 mg Se/kg DW), indicating high Se tolerance. Indeed, the two seedlings obtained in the lab study each accumulated above 1,000 mg Se/kg DW in the shoot after feeding with 20 μ M selenate, without toxicity symptoms (results not shown for lack of sufficient replication). Similar to the two species discussed above, *A. simplex* is common in disturbed areas, and is an annual (<http://swbiodiversity.org>). As a shallow rooting winter annual, it may indeed be well suited to life near a hyperaccumulator: Se levels in the soil near hyperaccumulators have been found to be lowest in spring and highest in fall, likely due to leaching¹⁶; competition for light is minimal, as the hyperaccumulator has not yet produced canopy; Se at the surface has been found to be lower relative to deeper horizons¹⁵. This seasonal and spatial Se difference may have given the opportunity to develop the tolerance shown by the field and lab plants sampled. *Bromus tectorum* (field brome) was found to positively co-occur with hyperaccumulators on two of the three sites. It is a winter annual grass species²³ that inhabits disturbed sites. *Bromus tectorum* had similar field levels of Se (up to 452 mg/kg DW) to the above species. When grown from seeds collected from the Pine Ridge 2 site, and treated with or without Se, the root biomass of the Se-treated plants was actually higher compared to the control treatment and the shoot biomass was nearly equal. The shoot Se levels in lab conditions reached nearly 150 mg Se/kg DW, indicating tolerance to Se and perhaps a positive growth effect. In a previous study, *S. ericoides* and *A. ludoviciana* were found to be facilitated by growing next to hyperaccumulators, showing higher Se levels and reduced

herbivory¹⁸. In the current lab study there were no herbivores, so the positive growth response must have been entirely due to physiological factors.

The list of species found less often near hyperaccumulators than expected is much smaller, because many species found in the Daubenmire plots were not at all found as one of the top five species near hyperaccumulators, and thus not counted. Only the species found in both plots could be compared and, of course, those species that negatively co-occur, by definition, should be absent more often as one of the nearest five species. Nevertheless, a few species were found with high relative abundance in the landscape overall, but infrequently near hyperaccumulators, perhaps because they are sensitive to Se. The two that stand out in this regard are *B. inermis* at the Pine Ridge 2 site and *Hesperostipa comata* at Coyote Ridge; both were ~23% of the Daubenmire canopy cover and ~4-fold less abundant near hyperaccumulators.

If Se hyperaccumulators, via the Se in their tissues and surrounding soil, affect neighboring species positively or negatively, is this effect Se concentration-dependent, and how large is their sphere of influence? When looking at correlation of Se in the hyperaccumulator and the Se content in the surrounding plants, a significant correlation was found for quite a few positively co-occurring species (with enough replication), both as individual species and as a group (notably, *T. dubius*, *A. simplex*, *O. suffrutescens* and *B. tectorum*), suggesting the Se enhancement effect in neighbors is proportional to the hyperaccumulator's Se concentration. However, a correlation between surface soil Se levels at the base of the hyperaccumulator and the leaf levels in the nearest five plants was not found; this may be due to fluctuating soil Se levels and forms, the limited differences in distance (most were under the canopy) and variation in other properties in the soil such as sulfur forms and levels (S inhibits Se uptake). In addition, no correlation was found between the surface soil Se concentration beneath the hyperaccumulator directly next to the stem and the leaf Se levels in the hyperaccumulator. Still, neighboring plant species contained higher Se levels when growing next to hyperaccumulators than when growing far away (Table 1B).

There was a pronounced local increase in Se concentration in soil around hyperaccumulators close to the plant, but beyond the canopy edge (>50 cm), soil Se decreased substantially (Fig. 2) Therefore, the sphere of influence of hyperaccumulators on surrounding vegetation due to higher local soil Se concentration is likely highest under their canopy. The reason for the higher Se under the canopy is likely leaf litter deposition and subsequent decomposition, but may also be root exudation^{16,17}. The forms of Se in hyperaccumulator species has been found to be organic, in the form of non-proteinogenic amino acids. This is hypothesized to be a possible tolerance mechanism, as it prevents oxidative stress from inorganic Se forms as well as toxic effects of non-specific incorporation of seleno-amino acids into proteins^{9,17,24}. Thus, hyperaccumulator litter

decomposition and root deposition processes locally release organic Se into the soil, which is more readily accumulated by plants than inorganic Se²⁵. Therefore, even slight increases in local soil Se concentration, in organic forms, likely already elevate Se levels in neighboring vegetation, and the Se accumulated might be expected to be enriched in organic forms. In this context, it is interesting to note that the negatively co-occurring species *B. inermis* had relatively more inorganic and less organic Se (Fig. 3), compared to the two positively co-occurring species *A. simplex* (Fig. 4) and *A. frigida* (Fig. 5).

The observed differences in the fraction of organic Se between positive and negative co-occurring species is also interesting because it may reflect Se tolerance differences. Species with lower Se tolerance typically accumulate relatively more inorganic Se; the negative co-occurrence of *B. inermis* may be related to lower Se tolerance to the elevated soil Se levels around hyperaccumulators. In the lab Se tolerance study, *B. inermis* did not stand out for being more Se sensitive than positively co-occurring species. However, it is noteworthy that its tissue Se levels were much lower than that of the positively co-occurring species in the lab experiment. Also interesting to note is that the field leaf Se levels for *B. inermis* in the area where it showed negative co-occurrence (Pine Ridge area 2) were over 2.5-fold higher than those obtained in the controlled study (370 vs. 140 mg Se/kg DW, respectively). Thus, it may well be that *B. inermis* would show more negative effects at higher tissue Se levels. Interestingly, in Pine Ridge area 1, where overall Se levels in the soils and vegetation were much lower (only 5 mg/kg for *B. inermis*), *B. inermis* actually showed positive co-occurrence with hyperaccumulators. Thus, it is possible that the mode of interaction between a certain neighboring plant species and a hyperaccumulator can be both positive and negative, depending on the experienced Se concentration in the neighbor. As stated above, low tissue Se levels often confer positive growth effects and higher anti-oxidant production. The threshold tissue Se levels between Se benefit and Se toxicity will be species-dependent.

In the lab Se tolerance and accumulation study, the two positively co-occurring species *A. ludoviciana* and *A. frigida* accumulated much higher Se levels than *B. inermis* and the other species tested (Fig. 7). In the field they also contained relatively high Se levels compared to other species, but not as high as they did in the lab study. The levels obtained in the lab (~800 mg/kg DW) appeared to be toxic, considering the reduction in biomass, albeit only significant for ARFR root. However, at the Se levels experienced in the field (183 and 50 mg/kg DW), they may have experienced benefits, considering their positive co-occurrence with hyperaccumulators. Leaves from *A. frigida* and *A. simplex* collected in the field had a relative large fraction (~75%) of organic Se, relative to *B. inermis* (50%), suggestive of higher Se tolerance. Incidentally, in *B. inermis* a large fraction of elemental Se was found; this may be due to endophytes or soil microbes: many bacterial and fungal endophytes from hyperaccumulators growing at the Pine Ridge site have been found to produce elemental Se²⁶⁻²⁸. In this

context it is also interesting to note that there is evidence that Se hyperaccumulator rhizobiomes are more similar to each other than to related non-hyperaccumulator plant species on the same site²⁹. Some microbial taxa were more prevalent around hyperaccumulator roots; it is possible that these exert an effect on growth or Se accumulation in neighboring plant species. Of further interest to note is that *A. simplex* showed some Se accumulation in its trichomes, similar to the Se hyperaccumulator *A. bisulcatus*⁹. Sequestration of Se in trichomes may provide Se tolerance and also may be effective in protection from herbivory.

The specific questions for this study can now be re-addressed: 1) How does soil Se concentration relate to distance from Se hyperaccumulators? It was found that soil Se concentration is higher adjacent to hyperaccumulators, but quickly diminishes after 50 cm, around the average canopy diameter (56 cm) we found for hyperaccumulators in this study. This supports the hypothesis that hyperaccumulators enrich their surrounding soil with Se; they likely influence the soil Se speciation as well, into more organic forms¹⁸. The average distance of the nearest five species to hyperaccumulators was ~20 cm, so well within the area of influence where hyperaccumulators affect Se form and concentration. The concentration in the hyperaccumulator correlated with the Se concentration in multiple species that showed positive co-occurrence. Therefore, it is possible that Se hyperaccumulators are providing selection pressures to neighboring plants by enriching them with Se, which may be positive or negative depending on the resulting Se concentration in the neighbour, and its tolerance. How this local influence of hyperaccumulators is projected into the greater landscape and to higher and lower trophic levels, and the possible cumulative effects over time will be interesting questions for further study.

Experimental

Field sampling and surveys

Field data collection started late June in Pine Ridge Natural Area Fort Collins, CO (latitude: 40.545496, longitude: -105.133213) and Coyote Ridge Natural Area, Fort Collins, CO (latitude: 40.480898, longitude: -105.125547). Both are naturally seleniferous sites located on shale soil, and harbor hyperaccumulator populations as described before^{8,15}. A total of 54 hyperaccumulator plants (*A. bisulcatus* and *S. pinnata*) were located, and the five plant species growing nearest to the hyperaccumulator were identified and recorded. In addition, a soil sample was taken at the base of the stem for each hyperaccumulator as well as at the base of each of the five nearest plant species. Soil samples were taken by brushing aside all organic matter at the surface and sampling the top layer (0-5 cm depth). In addition to soil samples, the youngest mature leaf was collected from each hyperaccumulator and each of the five nearest species. For the reference a 50 m baseline was used and six 25 m transects were located along it, three East and three West, alternating. Along each of these east and west

transects a Daubenmire quadrat was used at each meter alternating sides (north and south) each time²⁹. The data from the Daubenmire plots were then used to estimate canopy cover³⁰.

Lab studies

A follow-up lab study was performed using seeds collected in the areas described above. A total of 15 species were sown on Pro-Mix brand potting soil in the lab. Of these 15 species, five species, *Tragopogon dubius*, *Bromus inermis*, *Bromus tectorum*, *Artemisia ludoviciana*, and *Artemisia frigida*, produced at least six seedlings. The seedlings from these five species were split into two groups, with at least 3 replicates each: a control group, given only ¼ strength Hoagland's solution³¹, and a Se treatment group given the same solution with 20 µM sodium selenate added. All species were treated for five weeks at a 14:10 Light : Dark photoperiod. They were then harvested, washed, roots and shoots separated and dried in a 50°C drying oven (24 hours). After drying, they were weighed and total biomass recorded for root and shoot separately. The Se accumulated in the root and shoot material was analyzed as described below.

Elemental analysis

Soil samples were dried at 50°C for 2 days. Each sample was homogenized and sieved through a 1 mm screen. Approximately 400 mg of each sample was weighed and placed in a 25 mm x 200 mm glass acid digestion tube. Two ml of ultra-trace grade concentrated (70%) nitric acid was added to each sample. After addition of acid, each sample was heated to 60°C for two hours and then 125°C for 6 hours on a digestion block. After cooling, each sample was diluted to 10 ml with distilled water. This digest was subjected to elemental analysis via inductively coupled plasma mass spectrometry (ICP-MS) using a Perkin-Elmer Elan DRCII instrument with a detection limit for Se of approximately 0.01 ppb in the digest. Appropriate standards and controls were included in each analysis.

After drying, leaf samples were crushed to homogenize, and approximately 100 mg was weighed, placed in a 25 mm x 200 mm glass acid digestion tube and 1 ml of ultra-trace grade 70% nitric acid was added to each. The tubes were then placed on a heating block and heated to 60°C for 2 hours and then 125°C for 6 hours. After cooling, the samples were diluted to 10 ml with distilled water. All leaf samples were then analyzed using Inductively Coupled Plasma Optical Emission Spectroscopy (ICP-OES), model Perkin-Elmer 7300 DV, according to the manufacturer's instructions, including appropriate standards and controls.

X-ray microprobe analyses

Selenium, calcium (Ca) and potassium (K) distribution and Se speciation were analyzed using X-ray microprobe analysis (XRF mapping and Se K-edge XANES). Analyses were performed at beamline 10.3.2 (X-ray Fluorescence Microprobe) of the Advanced Light Source (ALS), at Lawrence Berkeley National Lab (Berkeley, CA, USA) using a Peltier cooling stage (-25 °C). Localization of Se, Ca and K was determined on intact leaves collected in the field, frozen immediately and kept frozen during

shipping and analysis. Micro-focused X-ray fluorescence (μ XRF) maps were recorded at 13 keV incident energy, using $20\ \mu\text{m} \times 20\ \mu\text{m}$ pixel size, a beam spot size of $7\ \mu\text{m} \times 7\ \mu\text{m}$, using 50 ms dwell time. Maps were then deadtime-corrected and decontaminated. Selenium K-edge micro X-ray absorption near-edge structure (μ XANES) spectroscopy (in the range 12,500–13,070 eV) was used to determine Se speciation on various spots as indicated on the μ XRF maps, close to areas showing high Se concentration. Spectra were energy calibrated using a red amorphous Se standard, with the main peak set at 12,660 eV. Least-square linear combination fitting of the μ XANES data was performed in the range of 12,630 to 12,850 eV using a library of 52 standard selenocompounds following procedures described in Fakra et al.³². All data were recorded in fluorescence mode using a 7-elements Ge solid state detector (Canberra, ON) and processed using custom LabVIEW programs available at the beamline.

Statistical analyses

Statistical analyses were done using R (ver. X64 3.32). T-tests were performed comparing the root and shoot biomass for the lab study ($\alpha=0.05$). One-way Anova was used to compare all of the different species in the lab experiment ($\alpha=0.05$).

Conclusions

Overall, the results from these studies support the hypothesis that hyperaccumulators affect local Se distribution and, with that, the composition and properties of surrounding vegetation. Plant species composition around hyperaccumulators is different. Some plant species show positive, co-occurrence with hyperaccumulators, and others negative co-occurrence with hyperaccumulators. Positive or negative co-occurrence may depend on Se tolerance, which is related to the forms of Se accumulator. Neighbors of Se hyperaccumulators had a higher tissue Se concentration as compared to when the same species grew elsewhere in the area, which was related with elevated soil Se levels around Se hyperaccumulators. These “Se hot spots” may derive from hyperaccumulator concentration and deposition of Se. By concentrating Se in and around them in highly bioavailable organic forms, hyperaccumulators may have a large effect on their local plant community, facilitating Se-tolerant plant community members but lowering the fitness of Se-sensitive members. These effects on the local vegetation is in turn expected to affect other trophic levels, as well as overall Se cycling in the local ecosystem.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

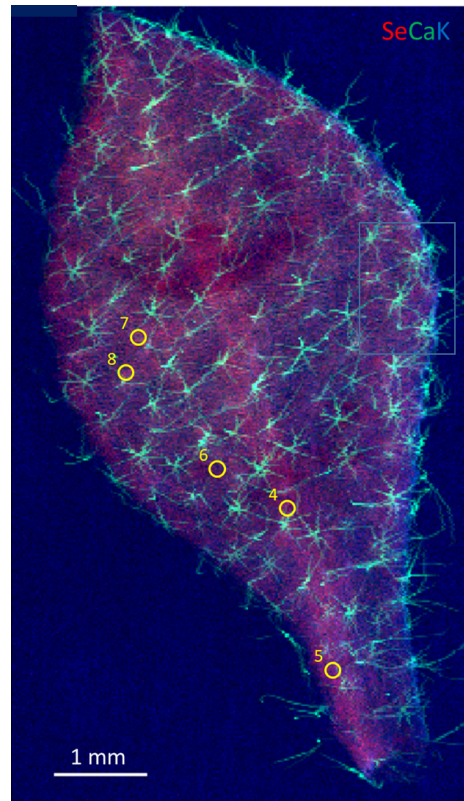
This research was funded by National Science Foundation, grant number IOS-1456361 to E.A.H.P-S. This research used resources

of the Advanced Light Source, which is a DOE Office of Science User Facility under contract no. DE-AC02-05CH11231.

References

- 1 F.M. Fordyce, Selenium deficiency and toxicity in the environment, In: O. Selinus, B. Alloway, J.A. Centeno, et al., eds., *Essentials of medical geology*, revised edn., Dordrecht, Springer, 2013, pp. 375–416.
- 2 A. Kabata-Pendias, *Trace elements in soils and plants*, Boca Raton, FL, USA, CRC Press/Taylor & Francis Group (2010), pp. 548.
- 3 T. Sors, D. Ellis, and D. Salt, Selenium uptake, translocation, assimilation and metabolic fate in plants, *Photosynth. Res.*, 2005, **86**, 373.
- 4 D. van Hoewyk, A tale of two toxicities: Malformed selenoproteins and oxidative stress both contribute to selenium stress in plants. *Ann. Bot.*, 2013, **112**, 965.
- 5 R. Feng, C. Wei, and S. Tu, The roles of selenium in protecting plants against abiotic stresses. *Environ. Exp. Bot.*, 2013, **87**, 58.
- 6 M. Schiavon, and E.A.H. Pilon-Smits, The fascinating facets of plant selenium accumulation – biochemistry, physiology, evolution and ecology, *New Phytol.*, 2017, **213**, 1582.
- 7 P.J. White, Selenium accumulation by plants, *Ann. Bot.*, 2016, **117**, 217.
- 8 M.L. Galeas, L.H. Zhang, J.L. Freeman, M. Wegner, and E.A.H. Pilon-Smits, Seasonal fluctuations of selenium and sulfur accumulation in selenium hyperaccumulators and related nonaccumulators, *New Phytol.*, 2007, **173**, 517.
- 9 J.L. Freeman, L.H. Zhang, M.A. Marcus, S.C. Fakra, S.P. McGrath, and E.A.H. Pilon-Smits, Spatial imaging, speciation, and quantification of selenium in the hyperaccumulator plants *Astragalus bisulcatus* and *Stanleya pinnata*, *Plant Physiol.*, 2006, **142**, 124.
- 10 R.S. Boyd, and S.N. Martens, The raison d'être for metal hyperaccumulation by plants, In: A.J.M. Baker, J. Proctor, R.D. Reeves (eds.), *The Vegetation of Ultra-mafic (Serpentine) Soils*, Intercept Limited, Andover, 1992, 279–289.
- 11 B. Hanson, G.F. Garifullina, S.D. Lindblom, A.L. Wangeline, A. Ackley, K. Kramer, A.P. Norton, C.B. Lawrence, and E.A.H. Pilon-Smits, Selenium accumulation protects *Brassica juncea* from invertebrate herbivory and fungal infection, *New Phytol.*, 2003, **159**: 461.
- 12 J.L. Freeman, S.D. Lindblom, C.F. Quinn, S.C. Fakra, M.A. Marcus, and E.A.H. Pilon-Smits, Selenium accumulation protects plants from herbivory by Orthoptera via toxicity and deterrence. *New Phytol.*, 2007, **175**, 490.
- 13 J.L. Freeman, C.F. Quinn, S.D. Lindblom, E.M. Klumper, and E.A.H. Pilon-Smits, Selenium protects the hyperaccumulator *Stanleya pinnata* against black-tailed prairie dog herbivory in native seleniferous habitats. *Am. J. Bot.*, 2009, **96**, 1075.
- 14 C.F. Quinn, J.L. Freeman, R.J.B. Reynolds, J.J. Cappa, S.C. Fakra, M.A. Marcus, S.D. Lindblom, E.K. Quinn, L.E. Bennett, and E.A.H. Pilon-Smits, Selenium hyperaccumulation offers protection from cell disruptor herbivores. *BMC Ecol.*, 2010, **10**, 19.
- 15 A.F. El Mehdawi, C.F. Quinn, and E.A.H. Pilon-Smits, Effects of selenium hyperaccumulation on plant-plant interactions: Evidence for elemental allelopathy? *New Phytol.*, 2011, **191**, 120.
- 16 C.F. Quinn, K.A. Wyant, A.L. Wangeline, J. Shulman, M.L. Galeas, J.R. Valdez, J.R. Self, M.W. Paschke, and E.A.H. Pilon-Smits, Enhanced decomposition of selenium hyperaccumulator litter in a seleniferous habitat - evidence for specialist decomposers? *Plant Soil*, 2010, **341**, 51.
- 17 A.F. El Mehdawi, S.D. Lindblom, J.J. Cappa, S.C. Fakra, and E.A.H. Pilon-Smits, Do selenium hyperaccumulators affect

- selenium speciation in neighboring plants and soil? An X-ray microprobe analysis. *Int. J. Phytorem.*, 2015, **17**, 753.
- 18 A.F. El Mehdawi, C.F. Quinn, and E.A.H. Pilon-Smits, Selenium hyperaccumulators facilitate selenium-tolerant neighbors via phytoenrichment and reduced herbivory. *Curr. Biol.* 2011, **21**, 1440.
- 19 A.F. El Mehdawi, R.J.B. Reynolds, C.N. Prins, S.D. Lindblom, J.J. Cappa, S.C. Fakra, and E.A.H. Pilon-Smits, Analysis of selenium accumulation, speciation and tolerance of potential selenium hyperaccumulator *Symphyotrichum ericoides*. *Physiol. Plant.*, 2014, **152**, 70.
- 20 T.A. Brown, and A. Shrift, Selenium: Toxicity and tolerance in higher plants. *Biol. Rev.*, 1982, **57**, 59.
- 21 S. Kaur, N. Kaur, K.H.M. Siddique, and H. Nayyar, Beneficial elements for agricultural crops and their functional relevance in defence against stresses, *Arch. Agron. Soil Sci.*, 2016, **62**, 905.
- 22 N. Rani, K.S. Dhillon, and S.K. Dhillon, Critical levels of selenium in different crops grown in an alkaline silty loam soil treated with selenite-Se, *Plant Soil*, 2005, **277**, 367.
- 23 K.G. Beck, Downy brome (*Bromus tectorum*) and Japanese brome (*Bromus japonicus*) biology, ecology, and management, 2009, accessed August 26 2019, <http://mining.state.co.us/SiteCollectionDocuments/DownybromeandJapanesebromeliteraturereviewColoradoDRMSDec09.pdf>
- 24 J.L. Freeman, M. Tamaoki, C. Stushnoff, C.F. Quinn, J.J. Cappa, J. Devonshire, S.C. Fakra, M.A. Marcus, S.P. McGrath, D. Van Hoewyk, and E.A.H. Pilon-Smits, Molecular mechanisms of selenium tolerance and hyperaccumulation in *Stanleya pinnata*. *Plant Physiol.*, 2010, **153**, 1630.
- 25 A. Zayed, C.M. Lytle, and N. Terry, Accumulation and volatilization of different chemical species of selenium by plants. *Planta* 1998, **206**, 284.
- 26 S.D. Lindblom, J.R. Valdez-Barillas, S.C. Fakra, M.A. Marcus, A.L. Wangeline, and E.A.H. Pilon-Smits, Influence of microbial associations on selenium localization and speciation in roots of *Astragalus* and *Stanleya* hyperaccumulators. *Environ. Exp. Bot.*, 2013, **88**, 33.
- 27 S.D. Lindblom, A.L. Wangeline, J.R. Valdez-Barillas, B. deVilbiss, S.C. Fakra, and E.A.H. Pilon-Smits, Fungal endophyte *Alternaria tenuissima* can affect growth and selenium accumulation in its hyperaccumulator host *Astragalus bisulcatus*. *Front. Plant Sci.*, 2018, **9**, 1213.
- 28 M. Sura-de Jong, R.J.B. Reynolds, K. Richterova, L. Musilova, L.C. Staicu, I. Chocholata, J.J. Cappa, S. Taghavi, D. van der Lelie, T. Frantik, I. Dolinova, M. Strojcek, A.T. Cochran, P. Lovecka, and E.A.H. Pilon-Smits, Selenium hyperaccumulators harbor a diverse endophytic bacterial community characterized by high selenium resistance and plant growth promoting properties. *Front. Plant Sci.*, 2015, **6**, 113.
- 29 A.T. Cochran, J. Bauer, J.L. Metcalf, P. Lovecka, M. Sura-de Jong, S. Warris, P.J.W. Mooijman, I. van der Meer, R. Knight, and E.A.H. Pilon-Smits, Plant selenium hyperaccumulation affects rhizosphere: Enhanced species richness and altered species composition. *Phytobiomes*, 2018, **2**, 82.
- 30 B. Coulloudon, K. Eshelman, and J. Gianola, Sampling vegetation attributes, BLM Technical Reference. BLM Business center, Denver, CO, 1999.
- 31 D.R. Hoagland, and D.I. Arnon, The water culture method for growing plants without soil, *Circ. Calif. Agr. Exp. St.*, 1938, **52**, 347.
- 32 S.C. Fakra, B. Luef, C.J. Castelle, S.W. Mullin, K.H. Williams, M.A. Marcus, D. Schichnes, J.F. Banfield, Correlative cryogenic spectromicroscopy to investigate selenium bioreduction products, *Environ. Sci. Technol.*, 2018, **52**, 503.



Reynolds and coworkers investigated effects of selenium hyperaccumulator plants on local vegetation. Shown is elemental distribution in *Alyssum simplex*.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41