Environmental Science: Water Research & Technology



Environmental Science Water Research & Technology

## Emerging investigator series: The Role of Vegetation in Bioretention for Stormwater Treatment in the Built Environment: Pollutant Removal, Hydrologic Function, and Ancillary Benefits

Journal:	Environmental Science: Water Research & Technology	
Manuscript ID	EW-CRV-11-2017-000511.R2	
Article Type:	Critical Review	
Date Submitted by the Author:	23-Mar-2018	
Complete List of Authors:	Muerdter, Claire; University of Iowa, Civil and Environmental Engineering Wong, Carol; Center for Watershed Protection LeFevre, Gregory; University of Iowa, Civil and Environmental Engineering	

SCHOLARONE<sup>™</sup> Manuscripts

1	<b>Emerging investigator series: The Role of</b>
2	Vegetation in Bioretention for Stormwater
3	<b>Treatment in the Built Environment: Pollutant</b>
4	Removal, Hydrologic Function, and Ancillary
5	Benefits
6 7	Claire P. Muerdter, <sup>1,2</sup> Carol K. Wong, <sup>3,4</sup> Gregory H. LeFevre <sup>1,2*</sup>
8 9	<sup>1</sup> Department of Civil and Environmental Engineering, University of Iowa, Iowa City IA, United
10	States; <sup>2</sup> IIHR—Hydroscience and Engineering, University of Iowa, Iowa City IA, United States;
11	<sup>3</sup> Department of Civil and Environmental Engineering, Stanford University, Stanford, CA 94305,
12	United States; <sup>4</sup> Center for Watershed Protection, Ellicott City MD, United States (Current
13	Affiliation)
14 15 16 17 18 19	*Corresponding Author: Email: gregory-lefevre@uiowa.edu; Phone: 319-335-5655; 4105 Seamans Center, Iowa City, Iowa, 52242, United States

#### 20 Table of Contents Entry:



- 22 23 24 Bioretention vegetation contributes to both the hydrologic and pollutant removal functions of bioretention.
- 25 26

#### 27 ABSTRACT

28 Vegetation influences both the hydrologic and pollutant-removal performance of bioretention 29 cells for green infrastructure stormwater management in the built environment. Vegetation can 30 intercept rainfall, lessen erosive sheetflow, ameliorate bioretention soil media clogging to 31 maintain infiltration capacity, and decrease total stormwater volume through transpiration. Plants 32 influence multiple pollutant removal processes, including phytoextraction, in-planta 33 phytotransformation, and alteration of the rhizosphere and associated microbial community. We 34 present the current state of knowledge of vegetative influence on pollutant-removal performance 35 and mechanisms, including for total suspended solids, nitrogen, phosphorus, toxic metals, 36 hydrocarbons, pathogens, and emerging contaminants in urban stormwater. Additional benefits 37 and opportunities for vegetation in bioretention include improved aesthetics of stormwater 38 infrastructure, lessened irrigation / fertilizer demand, provision of urban micro-habitats, thermal 39 attenuation, public education, increased resilience for climate change adaptation, and the 40 potential for air quality improvement as well as biomass and / or food production. We describe 41 plant traits and species that improve pollutant removal and hydrologic function, such as plant 42 biomass and growth rate. We identify key areas of future research need, including a focus on 43 transferrable findings / mechanistic studies, a better understanding of root system / rhizosphere 44 impacts, quantification of the impact of plant shoot harvesting, and further study of emerging 45 organic contaminants and metals. We conclude that vegetation in bioretention systems produces 46 measurable water quality and hydrologic performance benefits, but that plant processes could be 47 substantially further researched and developed to improve stormwater systems.

48

# 50 WATER IMPACT STATEMENT

51 Stormwater runoff is a major source of pollution worldwide. Bioretention can mitigate 52 stormwater flows and pollution. Current knowledge concerning vegetation influence on 53 hydrologic and pollutant removal mechanisms and performance in bioretention is addressed in 54 this review. Analysis of plant traits and specific plants that maximize bioretention function are 55 discussed, with recommendations for further research.

#### 57 **1. INTRODUCTION**

Stormwater runoff generated from impervious surface areas in the built environment causes 58 59 substantial deleterious environmental impacts to surface water quality and disrupts the native hydrologic regime. Consequences of stormwater runoff include degraded aquatic ecosystems.<sup>1</sup> 60 pollution of drinking water sources,<sup>2</sup> human exposure to pathogens,<sup>3</sup> erosion of streambanks, and 61 economic impacts on aquatic recreation through beach closures.<sup>4</sup> Stormwater can accumulate and 62 63 transport pollutants such as nutrients, toxic metals, oil and grease, trace organic contaminants, and pathogens into waterways.<sup>5</sup> A suite of strategies has emerged to mitigate stormwater 64 pollution. Although terminology differs by location (i.e., low-impact development.<sup>6</sup> water 65 sensitive urban design,<sup>7</sup> the sponge city plan,<sup>8</sup> etc.), the strategies all consist of engineered 66 67 stormwater management systems that are based on nature (e.g., soil, plants, etc.) to treat 68 stormwater onsite. These engineered systems are integrated into built landscapes to mitigate 69 changes in hydrology and increased pollution caused by runoff from land development.

70 One technology within the framework of stormwater low-impact development is bioretention cells, sometimes called "rain gardens," "bioinfiltration," or "biofilters." Bioretention cells (Fig. 71 72 1) are engineered infiltration facilities that contain high-permeability bioretention soil media (hereafter: "media") and vegetation to maximize infiltration and remove pollutants from 73 stormwater.<sup>3</sup> The surface of the media is often mulched. An underdrain is sometimes used to 74 75 collect and remove water that infiltrates through the media, especially in situations when the native surrounding soils have a low infiltration rate.<sup>9</sup> Bioretention can aid in restoring pre-76 development hydrology, delaying peak flow and reducing total volume, and is being integrated 77 into some locations for combined sewer overflow prevention.<sup>10–12</sup> Bioretention is also employed 78 79 for pollutant removal of total suspended solids, nitrogen, phosphorus, metals, hydrocarbons, and

pathogens, as well as for temperature mitigation. Bioretention is often applied as a stormwater
best management practice to meet water quality requirements such as total maximum daily
loads.<sup>9</sup> The media is an important component for all of these functions, and vegetation also plays
a significant—if underappreciated—role.

84



**Fig. 1:** Typical bioretention: a. cross-section (Image: Muerdter), b. A vegetated bioretention cell in St. Paul, Minnesota, USA (Photo: LeFevre)

85

86 Despite the importance of vegetation in bioretention design, substantial knowledge gaps exist 87 in areas where plant processes contribute to improved stormwater outcomes. Plants are often 88 selected only for aesthetics, survivorship, or being native to the region, with the vegetative 89 contribution to bioretention pollutant removal and hydrology being overlooked. Because native 90 vegetation is often used for site- and climate-specific resiliency, translating specific vegetation 91 studies to different locations can be difficult. Thus, an understanding of mechanisms rather than 92 mere 'black-box results' is critical in generating transferrable research findings and knowledge. 93 This review examines current research findings on the role of vegetation in bioretention, makes 94 recommendations on the role of plant processes in engineered natural treatment systems such as 95 bioretention, provides context from current practice guidance, and suggests areas of future research need. 96

97

### 98 2. VEGETATION FUNCTIONS IN BIORETENTION

99 **2.1 Hydrologic Processes** 

100 Vegetation contributes to bioretention hydrologic function above, at, and below the media

101 surface, through plant interception of rainwater, surface flow regulation, water infiltration

102 modification, and plant transpiration.

2.1.1 Plant Interception. Above-ground portions of vegetation intercept and store rainwater,
 or channel rainwater to the ground along stems.<sup>13,14</sup> Interception reduces both the total volume of
 stormwater runoff and erosive forces by protecting the soil surface from direct rainfall.<sup>15</sup>
 Interception storage can be substantial; for example, the average per-tree interception of 20 tree
 species was the equivalent of a 0.86 mm storm.<sup>16</sup>

Field studies of interception storage in bioretention are lacking in the literature. Nevertheless, the amount of rainfall intercepted by vegetation can be estimated from various models<sup>17</sup> and previous studies of particular plant species. Plant species create different amounts of interception based on attributes such as surface area and leaf smoothness.<sup>14</sup> For example, generally conifers store more water on plant surfaces than broadleaf trees.<sup>16,18</sup> Seasonality also greatly impacts interception by deciduous plant species.

**2.1.2 Surface Flow.** The capability for vegetation to slow overland flow and reduce erosion has been quantified in other settings, but has yet to be quantified in bioretention.<sup>19–21</sup> The impact on overland flow can vary greatly between vegetation types. For example, the Manning's Roughness Coefficient value<sup>13</sup> for "woods with dense underbrush" (0.80) is >five-fold the value for "short grass" (0.15). Slowing surface flow with vegetation presence can decrease erosion, preventing the movement of bioretention mulch and media that otherwise would be scoured off 120 of the inlet area of the cell and redistributed to other parts of the bioretention cell. Mulch is 121 important for the removal of metals and hydrocarbons, thus an evenly distributed layer of mulch 122 throughout the bioretention cell is desired.<sup>9</sup>

123 2.1.3 Stormwater Infiltration. Media clogging due to sediment influx is the main cause of 124 failure in bioretention.<sup>22</sup> In a clogged system, partially treated or untreated water can pond for 125 longer than desired, permitting mosquito development. Clogging can also cause water to 126 overflow the bioretention cell, bypassing treatment and creating flooding.<sup>23</sup> Many bioretention 127 design manuals specify a maximum allowable ponding time, for example, 48 hours.<sup>24</sup>

128 The roots of bioretention vegetation create macropores and root channels that enhance media 129 hydraulic conductivity and prevent clogging. Specifically, more extensive, thick roots and 130 vigorous vegetation growth rates increase infiltration over time and are recommended for 131 clogging prevention. For example, under low flow rates, a shrub (Buxus sinica) facilitated faster bioretention infiltration than turf grass, which has a shallow root system.<sup>25</sup> Similarly, *Melaleuca* 132 133 ericafolia, a thick-rooted Australian native shrub/tree, increased hydraulic conductivity (155 mm hr<sup>-1</sup> to 295 mm hr<sup>-1</sup> after 56 weeks) in bioretention columns over time.<sup>26</sup> Hydraulic conductivity 134 135 decreased in unplanted controls and treatments with other vegetation. Vegetation growth during 136 the study period was not reported; thus the causation of differential hydraulic conductivity by 137 plant roots must be presumed from treatment design. A field study in Australia, however, did 138 document a correlation between vigorous vegetation growth and significant increases in infiltration.<sup>27</sup> Larger root biomass also correlated to greater increases in infiltration than smaller 139 root biomass in Oregon, USA.<sup>28</sup> Similarly, a field study in France found two to four-times higher 140 141 hydraulic conductivity in parts of an infiltration basin with actively growing plants vis-à-vis bare areas or vegetated areas during seasons of plant rest.<sup>29</sup> Thus, seasonality and the extent of growth 142

of a root network over time can impact infiltration rates. It should be noted that in all of these studies hydraulic conductivity measurements were not decoupled from the impact of evaporation and transpiration.

146 The ratio of root depth to media depth should be considered in the bioretention design 147 process. Root depth will vary depending on plant species, climate (typically deeper roots are 148 found in dry climates), and the presence of an internal water storage layer in the bioretention design (which creates a saturated layer, discouraging root growth).<sup>30</sup> Deeper root systems 149 150 facilitate enhanced water infiltration into the media through root channels and macropores. Very 151 aggressively growing roots may be able to penetrate and clog a bioretention underdrain. 152 Additionally, denser plantings with increased infiltration and roots that reach the bottom of the 153 mesocosm have been linked with lessened nitrate removal from stormwater, in comparison with less-dense plantings,<sup>31</sup> presumably due to the formation of preferential flow paths. Thus, less-154 155 effective pollution removal performance may sometimes be a tradeoff of the increased 156 infiltration and clogging prevention created through root density. The depth of the mature plant 157 root system should be considered in the initial design, not just the root depth of the initial planted 158 material. Measurements of root depth in bioretention research include an average longest root of 29.1 cm for three forb species in Maryland, USA bioretention<sup>32</sup> and the majority of roots for two 159 Australian species, a sedge and a woody species, to be above 63 cm.<sup>26</sup> Media depth will vary 160 161 depending on available space, budget, and climate. Deeper media maximizes outflow volume reduction,<sup>9</sup> and thus will be preferable in climates that receive high-volume precipitation events, 162 163 whether those events are frequent (e.g., temperate or tropical climates) or infrequent (e.g., arid).

164 2.1.4 Transpiration. Transpiration is the process by which water is taken up by the plant
 165 roots, transported through the plant tissue, and evaporated from leaf surfaces. Transpiration of

166 water by vegetation helps maximize the volume of stormwater treated by the bioretention cell by 167 decreasing the total water exported to the underdrain / surrounding soil. Lessening total water 168 export may also lower the transport of soluble pollutants out of bioretention cells. 169 Evapotranspiration, a more inclusive term than transpiration, consists of abiotic evaporation as 170 well as transpiration. In seasonal climates, evapotranspiration can vary substantially throughout the year as the weather changes.<sup>33</sup> Work on evapotranspiration in bioretention is growing (e.g., 171 references<sup>34–36</sup>), although vegetation differences are not examined in most studies. Bioretention 172 vegetation type was linked to varying evaporation rates in one Wisconsin, USA study.<sup>37</sup> 173 174 Vegetation differences caused four-fold evapotranspiration variation. The shrub treatment had the highest average evapotranspiration rate (9.2 mm day<sup>-1</sup>), which was not significantly 175 different than the prairie treatment (7.9 mm day<sup>-1</sup>). The turfgrass treatment evapotranspiration 176 averaged 5.9 mm day<sup>-1</sup>, and the bare soil control averaged 2.1 mm day<sup>-1</sup>. Although 177 178 transpiration and evaporation were not explicitly decoupled in this study, higher transpiration 179 in the shrub and prairie treatments than the turfgrass evapotranspiration is also likely.

180 Transpiration data alone, decoupled from evapotranspiration, is very limited in bioretention. 181 In one study in Utah, total annual transpiration by bioretention cell vegetation was 7% (=5,600 liters) of the inflow volume during the growing season.<sup>38</sup> Different plant species can transpire at 182 widely varying rates (e.g., 3–25 Mg/yr among five tree species),<sup>39</sup> depleting soil moisture and 183 184 thus regenerating the hydrologic storage capacity of the media between events. For example, 185 prior to storm events, bioretention mesocosms planted with prairie and shrub vegetation had 186 significantly lower soil volumetric water content at depths of 0-0.15 and 0.30-0.45 m compared to turfgrass.<sup>40</sup> Specific studies of tree evapotranspiration and transpiration rates in 187 188 bioretention are needed in addition to forb, grass, and shrub data. As reviewed in Berland et 189 al.<sup>14</sup>, tree evapotranspiration rates in urban forests can have high in inter- and intraspecies 190 variation, but can be substantial (*e.g.*,  $\sim 2.5 \times 10^4$  kg yr<sup>-1</sup> for *Gleditsia triacanthos*, 191 honeylocust).<sup>39</sup> The effect of planting density on transpiration should also be considered. In a 192 non-bioretention pot study, densely planted trees transpired at lower rates than those planted 193 farther apart.<sup>41</sup>

194 Crop coefficients, developed in agriculture to predict evapotranspiration rates, could be a 195 useful tool in bioretention modeling, while recognizing the different conditions between agriculture and bioretention.<sup>42</sup> The rate of plant transpiration could be estimated from the 196 197 evapotranspiration rate, using the ratio of transpiration to evapotranspiration for the specific plant (*e.g.*, reference  $^{43}$ ). Crop coefficient evapotranspiration calculations also account for water 198 199 stress. When plants are water-stressed, *i.e.*,  $\leq 2x$  the wilting point, transpiration rates are substantially lowered.<sup>44</sup> Water stress on the vegetation in bioretention cells between precipitation 200 201 events will occur in many climates, because the media is designed to drain rapidly. Saturated 202 zones, a continually damp area of the media created by upturned underdrain elbows, can provide 203 a source of water for vegetation between natural rainfall events to minimize water stress.

204

### 2.2 Stormwater Quality Benefits

Multiple plant-related mechanisms impact pollutant removal in bioretention. After a brief introduction to the mechanisms (Figure 2), the plant impacts on pollutant processing are discussed in the context of specific pollutants. Typical stormwater concentrations and sources of pollutants are available in the literature and other sources (*e.g.*, references<sup>45-47</sup>). Design choices for specific sites should consider the pollutants of highest concern for that location.



211

212 Fig. 2: Pollutant removal mechanisms that can occur in vegetated bioretention systems.

- 213 (Illustration: Wong).
- 214
- 215

## 2.2.1 Mechanisms of Plant-Related Bioretention Pollution Removal

216 **2.2.1.1 Phytoextraction and Phytodegradation Mechanisms.** Phytoextraction is the 217 process of direct pollutant uptake from soil and its translocation into plant tissues, either above or 218 below ground.<sup>48</sup> Phytoextraction moves the pollutant into the plant tissue without chemical 219 modification, for example, the uptake of lead into plant shoots and roots from contaminated soil. 220 The lead remains in the same form as in the soil, *i.e.*, it is not mineralized or altered to a different 221 form. Phytoextraction can be an advantage when metals of commercial value are taken up into 222 plants because the metals can be removed from the plant tissue and recovered.<sup>48</sup> After phytoextraction, pollutants are often transported to the plant vacuole for sequestration and to prevent harm to active plant metabolic processes.<sup>49</sup> Phytoextraction depends on a number of factors such as temperature, plant phenology (*i.e.*, seasonality), and media components.<sup>50</sup> In contrast to phytoextraction, phytodegradation chemically alters the pollutant, ideally lowering of pollutant toxicity. For example, some pollutants form conjugates with sugars or amino acids after entering plant tissue, and can thus escape detection by methods that only measure the parent pollutant and not the conjugated form.<sup>51–54</sup>

230 **2.2.1.2 Rhizosphere Mechanisms.** The rhizosphere, *i.e.*, the zone adjacent to and 231 influenced by plant roots, has very distinct abiotic and biotic characteristics from the surrounding 232 soil, thus impacting pollutant fate.<sup>55</sup> These characteristics include redox conditions, pH, and the 233 microbial community. For example, field bioretention studies show higher bacterial abundance 234 in planted bioretention cell areas than unplanted,<sup>56,57</sup> and higher bacterial abundance in areas 235 with deeply rooted plants vis-à-vis turfgrass.<sup>56</sup>

236 Multiple factors contribute to the rhizosphere effect, notably oxygen introduction from plant roots<sup>58</sup> and root exudates. Soil oxygen levels impact redox conditions. For example, aerobic 237 conditions in soil oxidize ferrous iron and increase the P sorption capacity.<sup>59</sup> Oxygen levels also 238 239 impact rhizosphere microbial community structure and function; for example, creating 240 significantly greater aerobic nitrifying bacterial populations in the rhizosphere than the bulk soil during plant growth seasons.<sup>60</sup> In addition to oxygen, root exudates influence the rhizosphere 241 242 microbial community. Root exudates are a complex mixture of sugars, organic acids, and 243 secondary plant metabolite compounds that are released through plant roots. Simple carbohydrates in exudates, which can represent 30% of a plant's net fixed carbon,<sup>61</sup> stimulate 244 245 microbial growth in the rhizosphere and can increase cometabolic pollutant degradation. In

bioretention, runoff supplemented with dissolved organic carbon increased microbial populations
 and degradation of trace organic contaminants such as atrazine and fipronil.<sup>62</sup> Thus,
 carbohydrates in root exudates may perform a similar function.

249

## 2.2.2 Impact of Plant-Related Mechanisms on Specific Bioretention Pollutants

2.2.2.1 Total Suspended Solids (TSS). TSS removal rates in bioretention are typically 250 high.<sup>9</sup> The main mechanisms of total suspended solids (TSS) removal in bioretention are settling 251 / sedimentation and filtration by the mulch and media.<sup>9</sup> High (>80%) TSS removal has been 252 documented in unvegetated bioretention systems.<sup>63</sup> Nevertheless, improved TSS removal in field 253 254 bioretention cells after planting (vis-à-vis unvegetated bioretention cells) is attributed to media 255 stabilization and vegetation presence minimizing mulch and media movement in the bioretention cell.<sup>64</sup> Vegetation may contribute to maximizing sedimentation by slowing stormwater flow, 256 which allows more even distribution of solids throughout the bioretention cell.<sup>9</sup> Long-term, 257 258 vegetation's main function in bioretention TSS removal is to prevent media clogging by TSS deposition in the mulch and media. This is accomplished by root growth maintenance of 259 stormwater infiltration rates.<sup>26–29</sup> 260

An additional benefit of TSS capture is the concurrent removal of many other particleassociated pollutants, including several metals, P, and hydrophobic organic contaminants such as PCBs and dioxins.<sup>65–67</sup> Thus, stormwater regulations on total suspended solids levels simultaneously control other pollutants.

265 **2.2.2.2** Nitrogen. Reported nitrogen removal rates in bioretention cells vary widely 266 (from net export to 99% removal<sup>21</sup>), with plant presence usually facilitating increased nitrogen 267 uptake compared to unplanted conditions. Multiple studies document higher total nitrogen (TN) 268 removal<sup>59,68–72</sup> total dissolved N (TDN),<sup>37,70</sup> ammonium (NH<sub>4</sub><sup>+</sup>) removal,<sup>70</sup> and nitrate (NO<sub>3</sub><sup>-</sup>) /

 $NO_x$  ( $NO_3^- + NO_2^-$ ) removal,<sup>68,70,73–75</sup> in planted bioretention compared to unplanted systems. 269 270 Even with salt-containing influent (present in cold climates where deicing salt is used), vegetation presence improves TN, TDN, and NO<sub>x</sub> removal in bioretention.<sup>76</sup> In some cases, plant 271 presence and/or type did not yield significant N impacts.<sup>70,71,74,75,77,78</sup> The lack of difference in 272 273 these cases is likely due to inherent variation among plant species and/or the non-plant 274 components of the studies (e.g., media type, saturation conditions). Significant differences have been documented among vegetation types for the removal efficacy of TN and/or TDN,<sup>70,73,75–77,79</sup> 275 nitrate or NO<sub>x</sub>,<sup>68,73–76</sup> ammonium,<sup>70,75,76</sup> and dissolved organic nitrogen.<sup>75</sup> Indeed, plant selection 276 can represent the difference between N export and N removal.<sup>37,73</sup> The strongest performing 277 278 plant species for N removal are listed in Table 2.

Nitrogen-processing mechanisms in bioretention influenced by plants can be organized into biological mechanisms and hydrological mechanisms. All of these mechanisms can potentially be influenced by plant age.<sup>1</sup> Reported literature values may be lower than would occur in wellestablished bioretention sites, because many studies are conducted immediately after planting. Further research in this area is warranted.

284 Biological mechanisms include the direct plant uptake of N and the rhizosphere influence on 285 the media microbial community. First, direct plant uptake will occur because N is essential for plant growth.<sup>80</sup> Plants are typically 2–5% N by dry weight.<sup>81</sup> Therefore, bioretention plants will 286 assimilate N from the media and stormwater.  $NO_3^-$  and  $NH_4^+$  are the two major forms of N taken 287 up by plants.<sup>81–83</sup> As an anion,  $NO_3^-$  is water-soluble and plant-accessible.  $NH_4^+$  can be captured 288 289 in the soil via sorption or ion exchange and subsequently assimilated by plants. Some plant species can also take up organic N compounds,<sup>84–86</sup> which is relevant in bioretention because 290 291 organic N is typically a component of incoming stormwater. Ideal plants for bioretention should

have high water-use efficiency, *i.e.*, a high conversion of available water to biomass, which includes N. Water-use efficiency can vary between plant species, e.g., between ~16 mg N (L  $H_2O$ )<sup>-1</sup> and 93 mg N (L  $H_2O$ )<sup>-1</sup> in a study of eight plant species, <sup>87</sup> and as a plant ages.<sup>88</sup>

295 The second biological mechanism of plant influence on N removal in bioretention is 296 rhizosphere interactions with the media microbial community. Ammonium can be nitrified to 297 nitrite by Nitrosomonas spp. bacteria and nitrite can be nitrified to nitrate by Nitrobacter spp. bacteria.<sup>89</sup> Nitrate can be easily leached from bioretention. Due to nitrification, bioretention 298 effluent concentrations can be higher than the input nitrate concentration.<sup>21,68,90–93</sup> Plants can 299 affect this export though both direct nitrate uptake<sup>91</sup> and the influence of the rhizosphere on 300 301 microbial nitrification and denitrification. In a study of microbes present in media, higher levels 302 of four nitrification and denitrification genes occurred in the media samples of densely or 303 moderately vegetated cores than from areas with minimal or moderate vegetation, suggesting greater biotransformation capacity.<sup>57</sup> An additional potential impact on bioretention nitrogen 304 305 cycling is the microbial production of nitrous oxide and methane, both greenhouse gases. One study reported<sup>94</sup> that although nitrous oxide emissions were affected by plant root structure, the 306 307 total amount of incoming nitrogen being converted to greenhouse gases was small (<1.5% of the 308 incoming nitrogen load). Thus, the emission of greenhouse gases from properly functioning 309 bioretention cells should be minimal.

Without design and maintenance management, plant presence in bioretention can facilitate N export due to plant nutritional needs and senescing biomass. Organic matter is usually included in media to stimulate plant growth, often in the form of compost. Compost, however, contributes to N export via leaching, particularly immediately after installation.<sup>95</sup> A minimal amount of compost should therefore be used in order to minimize nutrient export while providing for plant

315 growth. Another consequence of plant presence is the reintroduction of N from decomposing, 316 senesced plant biomass. This biomass can contribute organic N, which can be mineralized into 317  $NO_3^-$  and leach out of the bioretention cell.<sup>92</sup> Shoot harvesting and removal from the bioretention 318 cell permanently removes this N from the bioretention system.

319 Lastly, hydraulic factors, including the presence of a bioretention saturated zone and overall 320 hydraulic conductivity, impact N removal and the plants in bioretention. The use of saturation 321 zones in bioretention continues to be investigated to promote microbial denitrification and 322 attenuate plant water stress, but the exact impact on plant survival has not been quantified. Saturated zones enhanced the plant removal of multiple N species in some studies<sup>70</sup> but not in 323 others.<sup>1</sup> This variation appears to depend on both the individual plant species used and the 324 325 media/study configurations varying between studies. The second hydraulic-related mechanism is 326 the influence of root architecture on hydraulic conductivity. Plants with more extensive root 327 systems are speculated to be the most effective at promoting N removal. For example, in a study in Texas,<sup>68</sup> Big Muhly grass (*Muhlenbergia lindheimeri*), a large bunch grass with a root depth 328 329 of ~460 mm in the mesocosms, removed significantly more NO<sub>x</sub> than Buffalograss 609 (Buchloe 330 *dactyloides*), a turf grass with roots only in the top ~100 mm of the media. Similarly, a *Carex* sp. 331 with a dense root architecture and many fine root hairs was the most successful out of five tested plant species at NO<sub>x</sub> and TN removal in an Australian column study.<sup>73</sup> Nevertheless, excessive 332 333 hydraulic conductivity promoted by high root density and roots reaching the bottom of the media may provide insufficient contact time for maximum removal of nitrate.<sup>31</sup> Therefore, an extensive 334 335 root network that does not penetrate to the bottom of the media appears to be the most favorable 336 architecture for N removal.



2.2.2.3 Phosphorus. Phosphorus removal rates in bioretention cells vary widely,

ranging from removal to net export.<sup>96</sup> Although P removal can be high (e.g., 81%)<sup>73</sup> without 338 plants,<sup>71</sup> plant presence can create increased P uptake vis-à-vis unplanted treatments, especially 339 for dissolved P, which plants uptake directly.<sup>59,69,73,77</sup> For example, in a study with an influent 340 concentration of 2.5–3.5 mg TP  $L^{-1}$ , >80% of which was dissolved, plant storage in *Carex* 341 342 appressa was the dominant (64% on average) P sink in the system, illustrating the importance of vegetation in treating dissolved P.<sup>97</sup> P removal can differ with vegetation type, in addition to the 343 influence of P type.<sup>73,74,77</sup> For example, in an Australian mesocosm study,<sup>75</sup> only one of twenty 344 345 tested plant species removed significantly more TP than the unplanted control. In contrast, all but 346 one tested species removed more total dissolved P than the unplanted control. Other studies report minimal or no significant difference in P removal among different plant speices.<sup>1,31,68,78</sup> 347 348 These results are likely due to plants that are inherently similar in their P uptake abilities, and/or 349 low dissolved P concentrations in the influent. Of note for United States bioretention is that the 350 majority of previous studies on P and plant uptake occurred outside of the United States, with 351 several species that do not have American counterparts of the same genus.

352 The main mechanisms of P removal in bioretention are media sorption (dissolved P), plant/fungal uptake (dissolved P) and mulch/media filtration (particulate P).<sup>96,97</sup> Phosphorus 353 354 processing mechanisms in bioretention influenced by plants include direct plant and mychorrizal 355 uptake, plant alteration of media, and the introduction of P back to the bioretention cell from 356 senesced plant biomass. Plants directly assimilate P for normal physiological functioning (ATP production, nucleic acids, and phospholipids).<sup>98</sup> Plants take up dissolved inorganic 357 orthophosphate  $(H_2PO_4^{-} \text{ or } HPO_4^{2^{-}})$ , and thus are expected to have a larger impact on 358 359 phosphate than particulate-associated P. The phosphorus fraction in plant tissue can vary widely depending on species, but is typically 0.2–0.5% P by dry weight<sup>81,99</sup>–an order of magnitude less 360

361 than the N content. Nevertheless, plants can concentrate P, with xylem sap P levels 100 to 1,000 times in the soil.<sup>81</sup> Plants with associated mychorrhizal fungi may assimilate P more rapidly; in 362 363 one study, 75% of applied TP was removed from the liquid medium within two hours of 364 application by mychorrizal-innoculated pine (Pinus sylvestris) plants, vis-à-vis >8 hours for nonmychorrizal control pine plants.<sup>100</sup> Additionally, mycorrhizae can store excess P for future plant 365 use.<sup>101</sup> In a field study, plant-mycorrhizal associations were found in 4 out of 11 dominant 366 bioretention plant species from nine bioretention sites.<sup>102</sup> Further work is needed to quantify the 367 368 impacts that such mycorrhizal colonization has on bioretention pollutant removal dynamics.

369 Plants can also influence P in bioretention by altering media infiltration. The gradient created 370 by root removal of P from the soil solution encourages desorption of P from the soil or 371 particulate matter. Plant roots also facilitate oxidization of the media's ferrous iron, increasing media P sorption ability.<sup>59</sup> Between storm events, vegetation appears to help temporarily retain 372 373  $PO_4$ -P, especially in media with the greatest sorption capacity, through a not fully elucidated mechanism.<sup>59</sup> As a negative impact on P removal, P can also leach from compost/other organic 374 matter included in the media to support plant growth.<sup>95,103</sup> Thus, as with N, minimal OM (or OM 375 with very low P content) should be incorporated if phosphorus removal is critical, and the plant 376 377 palette adjusted accordingly. As with N, dead vegetative biomass can also contribute P back to 378 the bioretention cell upon decomposition. P concentration in stormwater has been correlated to the amount of tree canopy over streets, which introduces dead biomass to the stormwater.<sup>104</sup> This 379 380 challenge can be avoided in bioretention through vegetation shoot harvesting.

2.2.2.4 Metals. Metal removal from stormwater influent in bioretention is typically
 high. The most common metals in stormwater are copper, zinc, and lead, although other metals
 can be present.<sup>105</sup> Metals vary in their intrinsic properties and thus in their bioretention behavior.

In a planted 'bioretention box' in Norway, overall mass reduction rates were 90% for zinc, 82% for lead, and 72% for copper.<sup>106</sup> Removal can be high in nonvegetated bioretention: in both planted and unplanted treatments in a greenhouse study,<sup>107</sup> >92% of input metals were removed in the upper 27 cm of soil, with the majority of metal removal occurring in the mulch.<sup>93</sup> Nevertheless, removal of zinc, copper, and mercury improved after planting in one study of field bioretention cells,<sup>64</sup> and vegetation type can be a significant factor in iron, aluminum, and chromium removal from stormwater in bioretention mesocosms.<sup>108</sup>

391 Although the majority of metal removal in bioretention is attributed to non-vegetative 392 mechanisms such as filtration and adsorption, plants can facilitate enhanced removal through 393 direct plant uptake including hyperaccumulation, rhizosphere impacts, and metal 394 sorption/desorption and complexation with the organic matter used to support plant growth. 395 Plants can directly take up metals such as zinc, copper, manganese, and nickel for micronutrients.<sup>109</sup> Other metals taken up by plants have unclear direct biological functions, such 396 as cadmium, lead and mercury.<sup>109</sup> In bioretention studies, direct uptake into plant tissue has been 397 documented for zinc,<sup>106</sup> copper, lead,<sup>93,107,110</sup> and cadmium.<sup>110</sup> Measured plant tissue metal 398 concentrations in one study ranged from 0.5–3.3%.<sup>110</sup> In another study, plant uptake of Cu, Zn, 399 and Pb accounted for 2–7% of the influent concentrations.<sup>106</sup> Plant uptake of metals provides a 400 401 route for permanent metal removal via plant harvesting.

Effective vegetation metal removal performance in bioretention has been attributed to root architecture, plant age, and leaf area. *Melaleuca ericifolia* was significantly less effective than other plant species in iron, aluminum, and chromium removal, which is hypothesized to be from preferential flow paths created by thick *Melaleuca* roots.<sup>108</sup> Metal uptake varied with time for all species in the *Melaleuca study*, indicating changes in conditions as plants grow and media

407 conditions evolve.<sup>108</sup> Mn removal has been correlated with greater root soil depth and leaf 408 area.<sup>111</sup> The tested plant species from the existing literature that facilitate metals removal are 409 listed in Table 2. Additional plant species should be tested for their metal uptake capabilities in 410 bioretention.

411 Metal hyperaccumulating plants provide the possibility of high metal uptake, but are relatively untested in bioretention.<sup>10,50</sup> Hyperaccumulators can assimilate an extremely high 412 413 concentration of metals (more than 100 times those found in non-hyperaccumulating plants) into 414 their tissues without the phytotoxic effects experienced by non-hyperaccumulators under the same conditions.<sup>112</sup> Hyperaccumulators have been identified for As, Cd, Co, Cu, Mn, Ni, Pb, Sb, 415 Se, Tl, and Zn.<sup>112</sup> Hyperaccumulators could be beneficial when designing a bioretention system 416 417 for an area with known high concentrations of heavy metals. A Thlaspi species, a known zinc 418 hyperaccumultor, was planted in bioretention in Maryland but none survived more than a few weeks after planting.<sup>10</sup> We are not aware of any other documented uses of hyperaccumulators in 419 420 bioretention. Hyperaccumulating plants often have small biomass that accumulates slowly with shallow roots.<sup>113,114</sup> Therefore, for the overall removal of the maximum mass of metals, the use 421 422 of plants that accumulate metals at less than hyperaccumulating levels but that have substantially 423 more biomass may be more effective. Further work is needed on both hyperaccumulating and 424 metal-accumulating plants with high biomass that can survive in bioretention and contribute to 425 metal removal. With both non-hyperaccumulating and (especially with) hyperaccumulating 426 plants, the presence of metals in the plant biomass can be a concern for animal consumption as 427 well as for eventual return to the media if no biomass harvesting occurs. A bioretention pot study<sup>110</sup> determined that Zn, Cu, and Pb levels in non-hyperaccumulating bioretention plants did 428 429 not exceed the toxic levels recommended for livestock forage, but Cd concentration did. Wildlife

exposure from bioretention metal ingestion warrants further investigation. Disposal of the plants
can also become a financial burden if the plant shoots qualify as hazardous waste.<sup>109</sup>

432 Vegetation also alters the microbial and chemical composition of the rhizosphere whereby metals are mobilized for plant uptake or adsorption onto the media.<sup>115</sup> Organic acids in plant root 433 434 exudates can affect the retention and mineralization of metals in the rhizosphere, e.g. increasing the available Zn fraction.<sup>115</sup> Additionally, acidification occurs when the plant or microbes take 435 436 up ammonium and release H<sup>+</sup>, and can influence metal speciation by altering the surface charge of soil particles or facilitating metal redox reactions.<sup>115,116</sup> A decrease in pH causes a decrease in 437 metal adsorption.<sup>89</sup> The stimulation or suppression of certain microbes in the rhizosphere by 438 439 plant influence can also affect metal behavior. Metals adsorb to microbes, and secreted microbial metabolites can complex metals.<sup>116</sup> 440

Finally, vegetation can indirectly impact metal removal in bioretention via organic matter (typically compost) added to the media for plant growth. Compost can leach copper, lasting for several years of simulated rainfall in one study.<sup>95</sup> Nevertheless, the presence of organic matter in general in the media can also provide a benefit to metals removal by increasing the sorption of metals to the media via complexation.<sup>9</sup> For example, increased copper retention was found with the addition of wood chips and pea straw to the media.<sup>117</sup>

447 2.2.2.5 Hydrocarbons. Although hydrocarbon removal rates are generally high in 448 bioretention, vegetated systems remove more total petroleum hydrocarbons and polycyclic 449 aromatic hydrocarbons (PAH) than soil alone.<sup>118</sup> In bioretention specifically, both column and 450 field studies have found consistent oil and grease removal of greater than 96%.<sup>20</sup> In a Maryland 451 field bioretention study, PAH event mean concentration reductions of 31–99% were 452 documented.<sup>119</sup> In Minnesota, planted columns removed 93% of the naphthalene versus 78% for

the unplanted columns, suggesting that vegetation played an important role in removal.<sup>120</sup>
Furthermore, the two plant species tested had different masses of naphthalene taken up into their
plant tissue. Beyond uptake, both plant species generated lower naphthalene export (7% for
vegetated columns) than the unplanted column (22%).

457 Hydrocarbons in stormwater are predominantly removed via sorption to and filtration by 458 bioretention mulch and media, but plant removal mechanisms also impact hydrocarbon fate, 459 especially for lower molecular weight PAHs. Abiotic filtration is an important process because 74–90% of hydrocarbons are associated with particles.<sup>121</sup> Therefore, a simple layer of mulch was 460 461 able to sorb and filter 80-95% of input toluene, naphthalene, and used motor oil in a bench-scale bioretention study.<sup>121</sup> Approximately 90% of the motor oil was biodegraded within eight days. In 462 463 a different study of planted bioretention columns, labeled naphthalene tracing demonstrated sorption to the media was the dominant fate, removing 56–73% of the added naphthalene.<sup>120</sup> 464 465 Hydrocarbons on the top of the mulch are also exposed to solar radiation, which can facilitate photodegradation.<sup>122</sup> Finally, biochar has also shown promise for PAH removal from water in 466 non-bioretention settings,<sup>123</sup> and may be a useful amendment in bioretention. 467

Plant removal processes of hydrocarbons in bioretention include direct plant uptake, influence on the rhizosphere microbial community, the introduction of additional organic matter to the media, and the prevention of photodegradation through plant shading of the mulch/media. In isotope-labeled bioretention columns, direct plant uptake accounted for 2.5% (for clover)– 23% (for grass) of naphthalene removal.<sup>120</sup> The difference in incorporation into plant biomass is likely attributable to several factors, including the extensive root structure of the grass. For both species, the majority of the naphthalene in the plant tissue was present in the shoots, indicating 475 translocation from the roots after uptake, and the possible efficacy of plant shoot harvesting for476 permanent removal.

477 Plants also influence the rhizosphere microbial community that degrades hydrocarbons. 478 Hydrocarbons that have been trapped in the media through sedimentation and filtration can be 479 degraded by indigenous microbial petroleum hydrocarbon degraders.<sup>56,120</sup> Evidence for the role 480 of vegetation in supporting these microbial communities is mixed. In a column study without 481 vegetation, microbial degradation process removed 90% of the trapped material (naphthalene, toluene, and dissolved motor oil).<sup>121</sup> In a column study with vegetation and no-vegetation 482 483 controls, complete microbial mineralization (12–18% of total removal) was not different between the treatments.<sup>120</sup> Nevertheless, the grass columns had significantly more microbial naphthalene 484 dioxygenase functional genes present than the clover or unplanted columns.<sup>120</sup> When soil 485 486 samples collected from the columns at the end of the study were used as inoculum in batch 487 biodegradation experiments, samples from vegetated columns resulted in significantly faster 488 kinetics. Similarly, in a field study, greater numbers of two bacterial genes that aid in 489 hydrocarbon breakdown were found in Minnesota bioretention field sites with deeply-rooted vegetation than those sites with grass only or mulch only (non-vegetated).<sup>56</sup> This suggests that 490 491 more complex vegetation better supports a bacterial population that can degrade hydrocarbons, 492 potentially leading to increased removal efficiencies. Root exudates can improve PAH 493 transformation by altering the bioavailability of PAHs, allowing bacteria to access and breakdown these pollutants.<sup>124</sup> 494

An additional plant mechanism related to hydrocarbon fate is the introduction of organic matter to the media for plant growth. The presence of organic matter in the media increases the sorption of hydrocarbons, especially for higher molecular weight PAHs with log K<sub>ow</sub> values of

498 >4, which are less easily biodegraded than low molecular weight PAHs.<sup>9,125</sup> Thus, the 499 contribution of organic matter from decaying bioretention cell vegetation may enhance oil and 500 grease removal in bioretention, and if present in sufficient quantity, may even make introduced 501 mulch unnessary.<sup>121</sup> Finally, plants can negatively impact the mineralization of hydrocarbons 502 filtered or sorbed to the mulch and media by blocking sunlight, thus blocking photodegradation.

503 2.2.2.6 Pathogens. As with metals and hydrocarbons, pathogens can be removed in 504 bioretention at a high level by the media alone, although vegetation can significantly influence 505 pathogen removal by altering infiltration rates. It should be noted that removal, *i.e.*, fewer 506 pathogens in effluent than influent stormwater, does not automatically constitute deactivation of 507 the pathogens. The impacts of other vegetation mechanisms on pathogen removal rates remain 508 untested. Pathogens, often measured as fecal coliform or E. coli levels but also including 509 protozoa and viruses, can be introduced from incoming stormwater, wildlife or pet waste, leaking 510 sewers, etc. In one study, unvegetated columns produced a mean removal of E. coli of 72%, which increased to 97% or greater between six and 18 months (the end of the study).<sup>126</sup> In 511 512 another study, vegetation type had a significant effect on E. coli removal through the vegetation's impact on infiltration rates.<sup>127</sup> Greater E. coli removal occurred with plants that 513 514 produced low infiltration rates. Nevertheless, another study reported *E. coli* removal of >90% in all treatments, planted and unplanted.<sup>68</sup> Fecal coliform rates varied more widely, from 56 to 515 516 99.9% removal, with media type having more of an impact on removal rate than plant presence or plant species.<sup>68</sup> 517

518 Plant-related pathogen removal mechanisms in bioretention include both documented 519 influences, such as root structure, and untested (in bioretention) influences, as explained in detail 520 herein. Root structures that facilitate slower infiltration rates are correlated with greater pathogen

removal.<sup>127,128</sup> A substantial driver of pathogen removal in bioretention cells is the presumed 521 522 result of physical filtration of the pathogens in the media. For example, in a meta-analysis, the 523 presence or absence of shrubs explained 10% of the total variance in fecal indicator bacteria (FIB) removal rates, due to the shrubs' influence on infiltration rates.<sup>128</sup> Better FIB (including *E*. 524 525 *coli*) removal occurred with plant species associated with lower infiltration rates that allow for 526 more physical filtration. Other studies noted the presence of vegetation influenced the E. coli removal rate in dry conditions,<sup>127,129</sup> including a significant correlation between vegetation type 527 and infiltration rate.<sup>127</sup> In contrast to FIB and E. coli, there was no correlation between 528 bioretention vegetation and the removal efficiency of protozoa and viruses.<sup>129</sup> These results 529 530 could be due to the decrease in soil moisture content from greater evapotranspiration in vegetated 531 sites, the macropores and preferential flow paths created by the roots, and/or the variation in size and inherent biology between FIB and *E. coli* vis-à-vis protozoa and viruses.<sup>130</sup> 532

533 Vegetation is presumed to influence pathogen presence and removal through the hosting of 534 wildlife, light screening, root exudate antimicrobial compounds, and the alteration of microbial 535 grazers, but these mechanisms are poorly illuminated for bioretention. Vegetation, through its 536 provision of habitat or food such as berries or browse, can attract wildlife and introduce pathogens through direct defecation in the bioretention cell.<sup>9</sup> Thus far, studies on animal use of 537 538 bioretention are limited to insect populations, which exhibit greater biodiversity in bioretention than lawn-type greenspace,<sup>131–133</sup> and neglect warm-blooded animals. Secondly, UV light kills 539 pathogens, as is widely used in wastewater treatment plants.<sup>134</sup> Naturally occurring sunlight 540 541 therefore has the potential to kill pathogens on the surface of bioretention cells, but dense vegetation in bioretention may hinder UV light exposure.<sup>9,135</sup> Nevertheless, no experimental data 542 543 correlating light exposure in bioretention and pathogen die off have been generated, and this

544 remains an area for future study. Additionally, plant root exudates can contain antimicrobial compounds,<sup>136</sup> which can influence rhizosphere microbes. This impact is untested in 545 546 bioretention. Lastly, the community of microbial predators of pathogens in the media is likely 547 influenced by vegetation. In unvegetated columns, indigenous protozoa in the media grew 548 logistically, with an ~10-fold increase in total number between fresh columns and  $\geq$ 13-month-old 549 columns, and may have played a role in the increase of E. coli removal over time, through predation.<sup>126</sup> The contribution of vegetation to the microbial ecology of the media, and 550 551 bioretention plant-related pathogen removal generally, is an area of research that requires further 552 study.

553 **2.2.2.7 Emerging contaminants.** Emerging contaminants are those chemicals found in 554 the aquatic environment that are not regulated, and/or those that have become of concern in recent years.<sup>137</sup> Emerging contaminants may include, but are not limited to, disinfection 555 556 byproducts, new-market pesticides/biocides, pharmaceuticals and personal care products, and endocrine-disruptors. Soluble emerging contaminants are susceptible to plant uptake,<sup>52</sup> though 557 558 knowledge of this interaction in bioretention is very limited. In one study, after the equivalent of 559 ~1.3 years of runoff applied, planted bioretention columns demonstrated >75% removal of 560 diuron, >50% removal of methylbenzotriazole, oryzalin, and tris(3-chloro-ethyl)phosphate (TCPP), and poor removal of atrazine, simazine, and prometon.<sup>138</sup> Further removal for all 561 562 contaminants occurred when the same bioretention systems were amended with biochar or 563 granular activated carbon. Biochar was the most effective of the two amendments, maintaining 564 >99% removal of all contaminants during the experiment. Additional work on the synergy 565 between vegetation and black carbon, as well as the mechanisms of vegetation's impact on 566 removal of these emerging contaminants, is warranted. Previous hydroponic plant uptake studies

567 report that the relatively polar emerging contaminant benzotriazole (anticorrosive) and 568 mercaptobenzothiazole (tire rubber vulcanizer) are rapidly assimilated by Arabidopsis plants and metabolized, in some cases with the metabolite being released from the plant.<sup>52,139</sup> These 569 metabolites were also documented in food crops,<sup>53</sup> but have not yet been documented in 570 bioretention plants. Another class of emerging contaminants of particular interest in bioretention 571 572 is polar neonicotinoid pesticides. Neonicotinoids are of concern because of their ubiquity as the most widely used insecticides in the world<sup>140</sup> including in urban applications, their harmful 573 574 impacts on non-target insect species, and their translocation within plants.

575

#### 2.3 Ancillary Benefits of Vegetation in Bioretention

576 2.3.1 Aesthetics. Plants can increase the aesthetics of bioretention, especially compared to traditional "grey" infrastructure, translating to increased property values. The Maryland 577 Stormwater Design Manual<sup>141</sup> states that, "Aesthetics and visual characteristics should be a 578 579 prime consideration" for stormwater best management practices. The 2007 Prince George's 580 County Bioretention Manual describes how designers can increase "real estate values up to 20 percent by using aesthetically pleasing landscaping,"<sup>142</sup> suggesting diverse, visually pleasing 581 582 bioretention vegetation rather than only turf grass. In addition to inherent plant aesthetics, vegetation may also cover visually unappealing sediment deposits,<sup>106</sup> and/or provide a 'green 583 screen' between pedestrian and car traffic.<sup>116</sup> 584

A critical attribute of aesthetics is plant survivorship. Plants must be able to tolerate the extremes in moisture that result from occasional inundation during / immediately following storms coupled with extended dry periods due to media with high hydraulic conductivity. For example, the measured infiltration rate in a Maryland, USA, bioretention cell<sup>32</sup> results in water moving through the root zone in 21 minutes. Vegetation must be able to take up water during

590 this short window and then survive during the antecedent dry period before the next precipitation 591 event. Additionally, plants must be able to withstand any other geographic-specific stressors on 592 plant survivorship, such as salt runoff from winter deicing operations. Vegetation must also 593 match the desired aesthetic of the bioretention cell and surrounding area under the planned maintenance regime to maximize aesthetic value.<sup>143</sup> Especially in arid regions, dead/dormant 594 vegetation can still provide aesthetic appeal that may be acceptable to the general public. 595 596 Nevertheless, green plants and flowers are typically desired, especially in regions where this is the norm.<sup>144</sup> 597

598 2.3.2 Lessened irrigation and fertilization demands. Bioretention can decrease the 599 need for supplemental irrigation and fertilization compared to 'traditional' landscaping choices. 600 Because the drainage area is typically many times the area of the bioretention cell (approximately 20 times,<sup>19</sup> although a hydraulic loading ratio of up to 49 times has been 601 suggested as a maximum<sup>73</sup>), bioretention receives a much greater quantity of stormwater and 602 603 thus more stormwater nutrients than landscaping receiving only areal rainfall. Therefore, plants 604 may be able to grow in bioretention that would not survive outside of bioretention. Nevertheless, 605 the selected bioretention must be able to withstand the other contaminants that become 606 concentrated in bioretention, such as metals and salt, and the rapid infiltration of water followed 607 by dry conditions. If plants are selected that can withstand those challenges, then the influx of 608 nutrients and water into bioretention is presumed to lessen the need for traditional fertilization 609 and irrigation compared to a non-bioretention landscape.

610 2.3.3 Provision of urban 'micro' habitats. Bioretention vegetation can provide small
 611 animal habitat in urban areas. For example, a significant difference in invertebrate biodiversity
 612 between bioretention and lawn-type greenspace has been measured, with an average of 22

invertebrate species in bioretention compared to five species in lawn-type greenspace.<sup>131,133</sup> In 613 614 this study, the highest biodiversity occurred in sites with a greater depth of leaf/plant litter, the 615 highest number of plant taxa, and a greater quantity of mid-stratum (i.e., not trees or 616 groundcover) vegetation. Thus, bioretention cells with complex and varied vegetation have the 617 potential to provide more invertebrate habitat than bioretention cells with only one low-growing 618 plant species. Habitat provision, including for pollinators, is expected to be maximized when native plants are used.<sup>145,146</sup> Additionally, soil invertebrates and earthworms have been found in 619 media, especially near the media surface.<sup>147</sup> Their presence is expected to contribute to soil 620 621 development as the bioretention cell ages, especially with the contribution of root exudates and 622 plant biomass (if the biomass is not removed after its senesce as part of bioretention cell 623 maintenance). Further study of wildlife usage of bioretention would help quantify the ecosystem 624 services that bioretention provides.

A possible concern for the provision of animal habitat is the use (on plants purchased for bioretention) of chemicals that maybe harmful to wildlife. For example, neonicotinoid pesticides are the mostly widely used insecticides worldwide,<sup>140</sup> and their inadvertent negative impacts on honeybees have received considerable attention. Neonicotinoids are used in nursery plants sold to the general public<sup>148</sup> (although in decreasing amounts due to negative publicity), and thus plants purchased for use in bioretention may contain neonicotinoids, providing an exposure route for pollinators in bioretention cells.

632 2.3.4 Food and/or Biomass Production. Plants in bioretention vegetation could be used
 633 as food crops. Global agricultural fertilizer use is projected to exceed 200 million metric tons in
 634 2018, a 25 percent increase from 2008.<sup>149</sup> Fertilizer production often requires energy intensive
 635 processes, such as mining or the Haber-Bosch process for ammonia fixation.<sup>150</sup> In contrast,

nutrient collection from stormwater is integral to bioretention without requiring additional energy input. Vegetables (beet, onion, spinach, tomato, broad bean) were grown in Australian bioretention, with yields generally similar to traditional vegetable gardens.<sup>151</sup> Sub-irrigation was used to reduce vegetable contact with potential stormwater contaminants, but further work is needed to examine the uptake of contaminants, including metals, into food crops grown in bioretention. This work could be informed by previous studies on the use of reclaimed water in agriculture, *e.g.*, references <sup>53,152</sup>.

643 Bioretention could be used to grow crops for electricity production through biomass combustion. Switchgrass (*Panicum virgatum*) has been successfully grown in bioretention.<sup>110,153</sup> 644 645 The energy for transportation to and from a bioretention cell is often expended as part of 646 bioretention maintenance, which may include plant harvesting. Assuming such maintenance 647 would occur regardless of plant type, then the net energy production of switchgrass grown in a bioretention cell could be approximately  $1.9 \times 10^6$  kJ (527 kW-h) per year (calculations shown in 648 ESI). This is 59% of the average 2016 monthly energy consumption of a U.S. home.<sup>154</sup> A case-649 650 by-case analysis will be needed, including consideration of any air pollution generated, but 651 switchgrass growth and harvesting could be energy-generating if a biomass power plant is 652 nearby. Biomass harvesting must not compact the media with heavy equipment that would 653 negatively impact hydraulic conductivity.

654

### 2.3.5 Additional Benefits.

*Thermal Attenuation.* Vegetation shades the bioretention surface, which can contribute to
 the thermal attenuation of the stormwater. Such thermal attenuation of stormwater in
 bioretention has been documented,<sup>155</sup> and is important for temperature-sensitive aquatic
 species such as trout, which may live in the receiving natural waters (lakes, streams, etc.)

659of unattenuated stormwater and/or bioretention effluent. For this reason, vegetation that660produces a near 100% canopy cover has been recommended for bioretention.9 A tradeoff661is that shading can increase pathogen survival at the surface of the media by blocking UV662light.

*Public Education.* Bioretention can provide important public education of water quantity
 and quality if signage or other communication is used (*e.g.*, Figure 3). Vegetation can
 provide an entry-point for this education, by drawing more positive attention to the
 facility that an unvegetated bioretention cell.



 **Fig. 3:** Example of onsite educational signage at a bioretention facility at the University of Maryland. (Photo: Muerdter.)

• *Climate Change Adaptation*. Bioretention has been proposed and is being implemented as a tool to help offset the hydrologic effects of climate change in urban areas.<sup>156,157</sup> Vegetation can increase the hydrologic resilience of stormwater infrastructure, as described herein. Bioretention plant selection should also consider possible climate change impacts on plant health.

*Air Quality Improvement*. Vegetation has the potential to improve air quality. For
 example, one study demonstrated that planted biofilters can remove gaseous toluene
 at a significantly higher rate than unplanted biofilters.<sup>158</sup> Additional studies of other
 gas phase pollutants in bioretention conditions are needed in order to understand the
 contribution of vegetation to improving urban air quality via vegetation in
 bioretention.

## 682 **3. DESIRABLE PLANTS AND PLANT TRAITS FOR BIORETENTION DESIGN**

Plant type and species should be chosen with prioritized pollutant/hydrology goals in mind. Due to environmental and geographic restrictions, not all plants can be used in every location. Plant traits (Table 1) are characteristics that are more widely applicable than recommendations for specific species. Table 2 presents specific plants that are effective for a given pollutant/hydrology goal in bioretention. Additionally, we aggregate multiple bioretention design resources that provide region-specific plant recommendations (ESI).

Plant Trait	Effect on Bioretention Performance <sup>(Reference)</sup>
Plant mass	Higher plant biomass decreases nutrient effluent concentration and increases transpiration. <sup>68,75</sup>
Growth rate	A rapid growth rate ( <i>e.g.</i> , $> 10 \text{ mg g}^{-1}$ day <sup>-1</sup> relative growth rate) decreases nutrient effluent concentrations, especially when coupled with the root characteristics listed below. <sup>74,111</sup>
Root lipid content	High root lipid content ( <i>e.g.</i> , >~0.6%) increases PAH uptake. <sup>159</sup> (Not yet tested in bioretention)
Root length	Long roots and a large total root length of a root system $(e.g., \sim 1,000 \text{ m})^{111}$ decreases nutrient effluent

689 **Table 1** Plant traits that benefit pollutant removal and hydrologic performance.

	concentration, although roots that reach the bottom of the media may increase nutrient effluent concentration. <sup>31</sup>	
Root mass/thickness <sup>31,111</sup>	Large total root mass and dense fine root patterns ( <i>e.g.</i> , >40% dense roots) <sup>111</sup> decreases nutrient effluent concentration (although note caveat about root length above). <sup>31</sup> Thicker roots increase hydraulic conductivity.	
High-nutrient tolerance <sup>75</sup>	Plants that are adapted to high-nutrient conditions will be more likely to increase nutrient removal. <sup>75</sup>	
High water-use efficiency	Plants with efficient water use $[e.g., >$ 78 mg N (L H <sub>2</sub> O) <sup>-1</sup> for tropical trees] will decrease nutrient effluent concentration. <sup>87</sup> (Not yet tested in bioretention)	
Adaptation to bioretention microenvironment (bowl, slope, etc.) conditions <sup>160</sup>	Plants should be matched to water and media conditions in the different areas of the cell. This will increase plant survival, and therefore increase the potential for increased pollutant removal. <sup>160</sup>	
Salt tolerance	For areas with road deicing salt use during winter, or other sources of salt, salt tolerance should be high. <sup>161</sup>	
High pollutant uptake per monetary investment in plant material <sup>162</sup>	The cost efficiency of bioretention pollutant removal can be maximized by choosing plants that have high pollutant uptake but low purchase cost. <sup>162</sup>	

690	Table 2 Summary of recommended plant traits or species to maximize pollutant removal and hydrologic performance in bioretention cells
691	

Pollutant/ Hydrologic Behavior	<b>Recommended Plant Trait or Plant Species</b> <sup>Reference</sup>	Proposed Mechanisms	Comments
Aluminum	Carex appressa <sup>108</sup>	Not specified	
Cadmium	High biomass <sup>110</sup>	Direct uptake	
Chromium	Carex appressa <sup>108</sup>	Not specified	
Clogging	Thicker roots, <sup>26,68</sup> vigorous vegetation growth <sup>27</sup> Melaleuca ericifolia, <sup>26</sup> Muhlenbergia lindheimeri <sup>68</sup>	Macropores from thicker roots, roots shrink and expand due to weather conditions, coarse roots have slower turnover rate and grow to deeper soil depths.	Fine roots did not maintain permeability, caused clumps.
Copper	<i>Carex microptera</i> <sup>107</sup> , <i>Carex praegracilis</i> , <sup>107</sup> <i>Correa</i> <i>alba</i> , <sup>75</sup> Creeping Juniper, <sup>93</sup> <i>Ficinia nodosa</i> , <sup>75</sup> <i>Kentucky</i> - <i>31</i> , <sup>110</sup> <i>Panicum virgatum</i> , <sup>110</sup> <i>Phragmites australis</i> <sup>107</sup>	Direct plant uptake	
E. coli	Plants that create low infiltration rates <sup>127</sup> Leptospermum continentale, <sup>127</sup> Melaleuca incana <sup>127</sup> , Palmetto buffalo <sup>127</sup>	Low infiltration rate, perhaps direct uptake or rhizosphere processes	
Transpiration	High biomass <sup>68</sup>	Direct plant uptake	
Hydrocarbons:	Carex hystricina, <sup>124</sup> Dalea purpurea, <sup>124</sup> Spartina	Plant root exudates can abiotically	
PAH: naphthalene	pectinate <sup>124</sup>	enhance desorption of naphthalene	
Hydrocarbons: PAHs:			
phenanthrene and pyrene	Helianthus annuus, <sup>163</sup> Zea mays <sup>163</sup>	Direct plant uptake	Not yet tested in bioretention.
Iron	Carex appressa <sup>108</sup>	Not specified	
Lead	<i>Carex microptera</i> , <sup>107</sup> <i>Carex praegracilis</i> , <sup>107</sup> Creeping Juniper <sup>93</sup>	Direct plant uptake	
Manganese	Large leaf area, <sup>111</sup> maximized root soil depth, <sup>111</sup> <i>Carex</i> appressa, <sup>75</sup> <i>Melaleuca ericifolia</i> <sup>75</sup>	Not specified	
	High plant mass, long roots, high root mass, large root		
----	---	--	--
	soil depth, extensive root systems, dense fine root		
	architecture, high number of microscopic root hairs,		
	arbuscular mycorrhizal fungi, rapid growth		
	Prairie vegetation community, <sup>37</sup> <i>Agapanthus praecox</i> , <sup>74</sup>		
	Amelanchier utahensis, <sup>71</sup> Artemisia cana, <sup>71</sup> Banksia		
	integrifolia, <sup>69</sup> Betula nigra, <sup>162</sup> Betula nigra Dura-		
	Heat, <sup>162</sup> Bouteloua gracilis, <sup>71</sup> Buchloe dactyloides, <sup>68</sup>		
	<i>Callistemon pachyphyllus</i> , <sup>69</sup> <i>Carex appressa</i> , <sup>73,75</sup> <i>Carex</i>		
	microptera, <sup>77</sup> Carex praegracilis, <sup>77</sup> Carpobrotus		
	edulis, <sup>74</sup> Carpobrotus glaucenses, <sup>69</sup> Cercocarpus		
	<i>ledifolius</i> , <sup>71</sup> <i>Cercocarpus montanus</i> , <sup>71</sup> <i>Dactylis</i>		
	glomerata, <sup>71</sup> Dianella brevipedunculata, <sup>69</sup> Elegia		
TN	<i>tectorum</i> , <sup>74</sup> E. purpureum subsp. maculatum Gateway, <sup>162</sup> Ficinia nodosa, <sup>74,75</sup> Goodenia ovata, <sup>75</sup>		
	Gateway, <sup>102</sup> Ficinia nodosa, <sup>14,15</sup> Goodenia ovata, <sup>15</sup>		
	Helianthus angustfolius, <sup>162</sup> Juncus amabilis, <sup>75</sup> Juncus		
	effusus, <sup>31,71</sup> Juncus flavidus, <sup>75</sup> Medicago sativa, <sup>71</sup>		
	Melaleuca ericifolia, <sup>73</sup> Muhlenbergia lindheimeri, <sup>68</sup>		
	Panicum virgatum Shenandoah, <sup>162</sup> Pennisetum		
	alopecuriodes, <sup>69</sup> Pennisetum clandestinum, <sup>74</sup>		
	<i>Phragmites</i> sp., <sup>71</sup> <i>Phragmites australis</i> , <sup>77</sup> <i>Poaceae</i> family, <sup>160</sup> <i>Rhododendron indicum</i> L., <sup>164</sup> <i>Salix exigua</i> , <sup>71</sup>		
	Schizachyrium scoparium, <sup>71</sup> Sorghastrum nutans, <sup>71</sup>		
	Stenotaphrum secundatum, <sup>74</sup> Typha sp., <sup>71</sup> Typha		
	capensis, <sup>74</sup> Zantedeschia aethiopica <sup>74</sup>		
	εαρεπείες, Σαπιεαειςεπια αειπισμικά		
	Avoid:		
	<i>Carex praegracilis</i> , <sup>77</sup> <i>Poa pratensis</i> , <sup>37</sup> <i>Scirpus acutus</i> , <sup>77</sup>	Direct plant uptake, microbial uptake,	
	<i>Scirpus validus</i> , <sup>77</sup> specified shrub community <sup>37</sup>	fungal uptake, increased infiltration	

ТР	Large root mass, long roots, extensive root systems, many root hairs <i>Agapanthus praecox</i> , <sup>74</sup> <i>Banksia integrifolia</i> , <sup>69</sup> <i>Betula</i> <i>nigra</i> , <sup>162</sup> <i>Betula nigra</i> Dura-Heat, <sup>162</sup> <i>Buchloe</i> <i>dactyloides</i> , <sup>68</sup> <i>Callistemon pachyphyllus</i> , <sup>69</sup> <i>Carex</i> <i>appressa</i> , <sup>75</sup> <i>Carex microptera</i> , <sup>77</sup> <i>Carex praegracilis</i> , <sup>77</sup> <i>Carpobrotus edulis</i> , <sup>74</sup> <i>Carpobrotus glaucenses</i> , <sup>69</sup> <i>Dianella brevipedunculata</i> , <sup>69</sup> <i>Eutrochium purpureum</i> subsp. <i>maculatum</i> Á. Löve & D. Löve Gateway, <sup>162</sup> <i>Helianthus angustfolius</i> , <sup>162</sup> <i>Muhlenbergia lindheimeri</i> , <sup>68</sup> <i>Panicum virgatum</i> Shenandoah, <sup>162</sup> <i>Pennisetum</i> <i>alopecuroides</i> , <sup>69</sup> <i>Pennisetum clandestinum</i> , <sup>74</sup> <i>Phragmites australis</i> , <sup>74,77</sup> <i>Rhododendron indicum</i> , <sup>164</sup> <i>Stenotaphrum secundatum</i> , <sup>74</sup> <i>Typha capensis</i> , <sup>74</sup>	Direct plant uptake, microbial immobilization (increased by plant presence), increased infiltration	Media pH should also be considered, for its effect on the sorption of P onto media
PCBs	Helianthus annuus, <sup>163</sup> Zea mays <sup>163</sup>	Direct plant uptake	Not yet tested in bioretention. Highest concentrations were in plant roots, not shoots.
Zinc	Bromus ciliates, <sup>110</sup> Carex microptera, <sup>107</sup> Carex praegracilis, <sup>107</sup> Creeping Juniper, <sup>93</sup> Kentucky-31, <sup>110</sup> Panicum virgatum, <sup>110</sup> Vinca minor <sup>106</sup>	Direct plant uptake	

693 Generally, plants with high above-ground biomass and thick, extensive roots are 694 recommended to improve pollutant removal, increase transpiration, and prevent media clogging. High-biomass plants generally (but not always) maximize the mass of contaminants assimilated 695 696 into plant biomass. Even if uptake rates are less than small plants, the overall greater biomass 697 may result in greater removal. Roots that are thick and penetrate a large proportion of the media 698 but do not reach the bottom of the bioretention cell are recommended to improve pollutant 699 removal, increase stormwater-media contact, increase transpiration, and prevent clogging. Roots 700 that do not penetrate to the bottom of the media are recommended to avoid preferential flow paths to the bottom of the bioretention cell, which may lessen pollutant removal performance.<sup>31</sup> 701 Thick roots improve hydraulic conductivity.<sup>26</sup> Bulbous roots may lead to preferential flow paths 702 and erosion, but research to confirm this assertion in bioretention is needed.<sup>160</sup> Root depths and 703 704 shapes vary widely between species: for example, roots of native North American prairie plants are typically orders of magnitude deeper than turfgrass such as Kentucky bluegrass.<sup>165</sup> In one 705 study, prairie plants were the only treatment to produce positive nitrogen removal efficiency.<sup>37</sup> 706 707 Turfgrass, shrub, and bare soil treatments had negative nitrogen removal efficiencies. Different 708 plants also alter media hydraulic performance, with prairie plants producing less total drainage out of bioretention than other plants or bare soil,<sup>37</sup> and shrub and prairie treatments having less 709 710 soil moisture between storms at their rooting depth than the turfgrass treatment and no-plant control.<sup>28</sup> 711

Bioretention plants should have high nutrient uptake capacity to maximize pollution control benefits. Nutrient uptake may be achieved through high N and P fraction in biomass and/or high total biomass. Many native plants do not exert a high nutrient demand; for example, some plants have evolved in low-nutrient soils rather than the higher-nutrient conditions in bioretention.<sup>75</sup>

Those plants may struggle with growth in bioretention and contribute less efficiently to nutrientuptake than plants adapted to high-nutrient conditions.

Vegetation maintenance is an important consideration for maximizing biomass and therefore nutrient removal. For example, experimental cutting regimes of *Juncus effusus* (recommended for bioretention) in non-bioretention conditions in Norway found that cutting back to 1 cm of remaining stubble resulted in significantly less regrowth than leaving 5 cm of stubble.<sup>166</sup> Regrowth also varied with the time of cutting.

Bioretention plants should also be suited to the microenvironment of the particular section of the bioretention cell. For example, the bottom surface of the bioretention cell, rather than the sloping sides of the bioretention cell ponding area, will receive the most stormwater. Additionally, locations close to the inlet will receive the fastest-moving stormwater. Therefore, the plants at the bottom of the bioretention cell that are closest to the inlet need to be the most tolerant of high flows and frequent inundation. Finally, local conditions should be taken into account, e.g., salt-tolerant plants in cold weather climates where deicing salt is employed.

730

## 731 4. CONCLUSIONS

732 **4.1 Bioretention Vegetation Role in Bioretention** 

The role of bioretention in vegetation is significant and complex. Plant processes in stormwater management green infrastructure have received considerably more research attention in recent years than previously, but important research gaps remain. From a hydrologic perspective, vegetation can decrease erosion of the bioretention cell surface, enhance infiltration of water into the media, prevent media clogging over time, and transpire water out of the bioretention cell. Thick roots and vigorous vegetation growth are recommended for clogging

739 prevention. Rooting depth and planting density are important parameters, with hydrologic 740 impact, that require further study. Vegetation impacts stormwater quality through a variety of 741 mechanisms, including phytoextraction, phytotransformation, and rhizosphere processes. In 742 terms of specific pollutants, vegetation does not have a large impact on TSS removal. Vegetation 743 typically has a significant impact on nitrogen removal, with important variations between plant 744 species. Phosphorus removal appears less impacted by plant selection than nitrogen, but plants 745 with high P uptake/media influence capacity can significantly affect P removal. The majority of 746 metal and hydrocarbon removal is attributed to non-plant mechanisms, though both pollutants 747 have been found in bioretention plant tissue biomass, and plants can alter the abiotic and 748 microbial removal mechanisms in the rhizosphere through root exudates. Pathogen removal is 749 similar, with influence on infiltration rate as the main documented plant-related influence. The 750 removal of some emerging contaminants has been documented in bioretention, but further work 751 on the role of vegetation in this removal is needed.

752 Bioretention vegetation has benefits beyond hydraulic and pollutant removal processes. 753 Plants make important contributions to bioretention aesthetics, can lessen irrigation and 754 fertilization demands, provide animal habitat, produce food and/or biomass, create thermal 755 attenuation of stormwater, enable public education, and contribute to climate change adaptation. 756 Plants should be chosen with specific pollutant priorities in mind based on of specific plants / 757 plant traits that have demonstrated improvement to bioretention (Tables 1 and 2). Most 758 generally, plants with high above-ground biomass and thick, extensive roots are recommended to 759 improve pollutant removal, increase transpiration, and prevent media clogging. Bioretention 760 plants should have high nutrient uptake capacity to maximize pollution control benefits, and be 761 suited to the part of the bioretention cell in which they are planted.

Page 41 of 56

762 **4.2 Future Research Areas** 

Based on the above findings, the authors propose several research needs for future work, asdescribed below.

765 A greater emphasis on the transferable basis for research findings. Focus research on 766 transferable processes that provide a mechanistic understanding of pollutant removal 767 processes and hydrology, not just "black box, in-out" findings. A deeper understanding of 768 plant traits that can transcend regional boundaries/plant ranges, e.g., those listed in Table 1, 769 to allow for the wider application of research results. Additionally, the impact of bioretention 770 age on vegetation performance, especially on bioretention of >2 years, requires study. 771 Mesocosm studies are generally conducted in less than two years; field conditions after two 772 years are expected to deviate from these results.

Better understanding of below-ground, plant-facilitated pollutant removal mechanisms.
 Specifically:

## Oreater elucidation of the interaction of plant roots, and particularly root exudates, with the media and microbial community. For example, root exudates may provide a sustainable carbon source for denitrification.

- Further work on how plant density and root depth impact contaminant removal.
   Experiments should examine differential pollutant removal in systems of varied rooting depths (*i.e.*, those that reach the bottom of the media and those that do not) and plant densities.
- The role of mycorrhizae in facilitating pollutant removal. Mycorrhizal inoculations
   have the potential to greatly improve bioretention function, especially for nutrients
   and organic contaminants, and have been understudied.

Plant shoot harvesting: quantification of the permanent removal of plant-assimilated
 pollutants from bioretention and the effect on post-harvest plant growth. If harvesting will
 occur, the feasibility of biomass crops should be investigated.

- In addition to continuing work on nutrients and other more well-studied pollutants, the
   impact of bioretention vegetation on other stormwater pollutants:
- *Emerging contaminants*, particularly polar pollutants that can be assimilated
   into plant tissues and present the greatest risk to groundwater during
   infiltration. Given the potential for recycled water use in bioretention,<sup>97,167</sup> and
   the increasing quantities of trace organic contaminants in treated and
   environmental waters, plant interactions with emerging contaminants demands
   investigation. The potential synergy between vegetation and black carbon or
   other novel geomedia in this area should be studied.
- Metals. Additional tests of metal hyperaccumulators and high-biomass metal
   accumulating plants in bioretention conditions to find plant species that can
   maximize metal removal. Also, further study is warranted on the ultimate fate
   and impacts to wildlife that consume the plant tissue.

801 Vegetation plays an important role in bioretention functioning. Studies thus far have 802 developed the understanding of many of these roles, but continued work on vegetation function 803 will further illuminate plant processes to fully maximize bioretention hydrologic and pollutant 804 removal performance.

805

806 ELECTRONIC SUPPLEMENTARY INFORMATION (ESI). Table of representative

807 vegetation bioretention design resources, biomass energy production calculation assumptions.

~	~	~
v	()	v.
о	υ	σ

- 809 **NOTES.** The authors declare no competing financial interest.
- 810
- 811 ACKNOWLEDGEMENTS. Carol Wong was funded by the U.S. National Science Foundation
- 812 Graduate Research Program Fellowship, under grant number DGE 1147470, during her work
- 813 at Stanford University. The authors acknowledge Rai Tokuhisa for the initial idea to use biomass
- 814 crops in bioretention.

## 815 **REFERENCES**

- 817 (1) Palmer, E. T.; Poor, C. J.; Hinman, C.; Stark, J. D. Nitrate and Phosphate Removal
  818 Through Enhanced Bioretention Media: Mesocosm Study. *Water Environ. Res.* 2013, 85
  819 (9), 823–832.
- Marsalek, J.; Rochfort, Q. Urban wet-weather flows: sources of fecal contamination impacting on recreational waters and threatening drinking-water sources. *J. Toxicol. Environ. Heal. Part A* 2004, 67 (20–22), 1765–1777.
- Roy-Poirier, A.; Champagne, P.; Filion, Y. Review of Bioretention System Research and
  Design: Past, Present, and Future. *J. Environ. Eng.* 2010, *136* (9), 878–889.
- 825 (4) Schiff, K.; Kinney, P. Tracking Sources of Bacterial Contamination in Stormwater
  826 Discharges to Mission Bay, California. *Water Environment Research*. Water Environment
  827 Federation pp 534–542.
- United States Environmental Protection Agency Office of Water. *Environmental Impacts of Stormwater Discharges*; Washington, 1992.
- Bavis, A. P. Green Engineering Principles Promote Low-impact Development. *Environ. Sci. Technol.* 2005, *39* (16), 338A–344A.
- Fletcher, T. D.; Shuster, W.; Hunt, W. F.; Ashley, R.; Butler, D.; Arthur, S.; Trowsdale,
  S.; Barraud, S.; Semadeni-Davies, A.; Bertrand-Krajewski, J.-L.; et al. SUDS, LID,
  BMPs, WSUD and more The evolution and application of terminology surrounding
  urban drainage. *Urban Water J.* 2015, *12* (7), 525–542.
- 836 (8) Jia, H.; Wang, Z.; Zhen, X.; Clar, M.; Yu, S. L. China's sponge city construction: A
  837 discussion on technical approaches. *Front. Environ. Sci. Eng.* 2017, *11* (4), 18.
- Hunt, W. F.; Davis, A. P.; Traver, R. G. Meeting Hydrologic and Water Quality Goals
  through Targeted Bioretention Design. *J. Environ. Eng.* 2012, *138* (6), 698–707.
- 840 (10) Davis, A. P. Field Performance of Bioretention: Hydrology Impacts. J. Hydrol. Eng. 2008,
  841 13 (2), 90–95.
- 842 (11) Trowsdale, S. A.; Simcock, R. Urban stormwater treatment using bioretention. J. Hydrol.
  843 2011, 397 (3), 167–174.
- 844 (12) Shuster, W. D.; Darner, R. A.; Schifman, L. A.; Herrmann, D. L. Factors Contributing to
  845 the Hydrologic Effectiveness of a Rain Garden Network (Cincinnati OH USA).
  846 *Infrastructures* 2017, 2 (3), 11.
- 847 (13) McCuen, R. H. *Hydrologic Analysis and Design*, 3rd ed.; Pearson Prentice Hall: Upper
  848 Saddle River, New Jersey, 2005.
- 849 (14) Berland, A.; Shiflett, S. A.; Shuster, W. D.; Garmestani, A. S.; Goddard, H. C.; Herrmann,

850 851		D. L.; Hopton, M. E. The role of trees in urban stormwater management. <i>Landsc. Urban Plan.</i> <b>2017</b> , <i>162</i> , 167–177.
852 853	(15)	Asadian, Y.; Weiler, M. A New Approach in Measuring Rainfall Interception by Urban Trees in Coastal British Columbia. <i>Water Qual. Res. J. Canada</i> <b>2009</b> , <i>44</i> (1), 16–25.
854 855	(16)	Xiao, Q.; McPherson, E. G. Surface Water Storage Capacity of Twenty Tree Species in Davis, California. <i>J. Environ. Qual.</i> <b>2016</b> , <i>45</i> (1), 188.
856 857	(17)	Muzylo, A.; Llorens, P.; Valente, F.; Keizer, J. J.; Domingo, F.; Gash, J. H. C. A review of rainfall interception modelling. <i>J. Hydrol.</i> <b>2009</b> , <i>370</i> (1–4), 191–206.
858 859 860	(18)	Hao, L. Based on the Technology of Sponge City in Urban Design Study. In 2016 International Conference on Smart City and Systems Engineering (ICSCSE); IEEE, 2016; pp 27–29.
861 862	(19)	Davis, A. P.; McCuen, R. M. Stormwater management for smart growth; Springer US: New York, 2005.
863 864	(20)	Hsieh, C.; Davis, A. P. Evaluation and Optimization of Bioretention Media for Treatment of Urban Storm Water Runoff. <i>J. Environ. Eng.</i> <b>2005</b> , <i>131</i> (11), 1521–1531.
865 866	(21)	Davis, A. P.; Hunt, W. F.; Traver, R. G.; Clar, M. Bioretention Technology: Overview of Current Practice and Future Needs. <i>J. Environ. Eng.</i> <b>2009</b> , <i>135</i> (3), 109–117.
867 868	(22)	Kandra, H. S.; Deletic, A.; McCarthy, D. Assessment of Impact of Filter Design Variables on Clogging in Stormwater Filters. <i>Water Resour. Manag.</i> <b>2014</b> , <i>28</i> (7), 1873–1885.
869 870 871	(23)	Le Coustumer, S.; Fletcher, T. D.; Deletic, A.; Barraud, S.; Lewis, J. F. Hydraulic performance of biofilter systems for stormwater management: Influences of design and operation. <i>J. Hydrol.</i> <b>2009</b> , <i>376</i> (1–2), 16–23.
872 873	(24)	Minnesota Pollution Control Agency. Minnesota Stormwater Manual https://stormwater.pca.state.mn.us/index.php?title=Main_Page.
874 875 876	(25)	Xia, J.; Wang, H.; Stanford, R. L.; Pan, G.; Yu, S. L. Hydrologic and water quality performance of a laboratory scale bioretention unit. <i>Front. Environ. Sci. Eng.</i> <b>2018</b> , <i>12</i> (1), 14.
877 878 879	(26)	Le Coustumer, S.; Fletcher, T. D.; Deletic, A.; Barraud, S.; Poelsma, P. The influence of design parameters on clogging of stormwater biofilters: A large-scale column study. <i>Water Res.</i> <b>2012</b> , <i>46</i> (20), 6743–6752.
880 881	(27)	Hatt, B. E.; Fletcher, T. D.; Deletic, A. Hydrologic and pollutant removal performance of stormwater biofiltration systems at the field scale. <i>J. Hydrol.</i> <b>2009</b> , <i>365</i> (3–4), 310–321.
882 883	(28)	Hart, T. D. Root-Enhanced Infiltration in Stormwater Bioretention Facilities in Portland, Oregon, Portland State University, 2017.

- (29) Gonzalez-Merchan, C.; Barraud, S.; Bedell, J.-P. Influence of spontaneous vegetation in stormwater infiltration system clogging. *Environ. Sci. Pollut. Res.* 2014, 21 (8), 5419– 5426.
- (30) Fan, Y.; Miguez-Macho, G.; Jobbágy, E. G.; Jackson, R. B.; Otero-Casal, C. Hydrologic
  regulation of plant rooting depth. *Proc. Natl. Acad. Sci. U. S. A.* 2017, *114* (40), 10572–
  10577.
- Muerdter, C.; Smith, D.; Davis, A. P. Vegetation Selection Impacts Nitrogen but not
   Phosphorus Processing in Bioretention Mesocosms. *Submitted*.
- 892 (32) Muerdter, C.; Özkök, E.; Li, L.; Davis, A. P. Vegetation and Media Characteristics of an
  893 Effective Bioretention Cell. J. Sustain. Water Built Environ. 2016, 2 (1), 4015008.
- (33) Abdel-Lah, A. K.; Watanabe, K.; Kurokawa, U. Seasonal Change of Evapotranspiration and Influence of Plant Density on Evapotranspiration. *J. Japan Soc. Eng. Geol.* 1997, 37
  (6), 446–451.
- 897 (34) Denich, C.; Bradford, A. Estimation of Evapotranspiration from Bioretention Areas Using
  898 Weighing Lysimeters. J. Hydrol. Eng. 2010, 15 (6), 522–530.
- (35) Wadzuk, B. M.; Asce, A. M.; Hickman, J. M.; Traver, R. G.; Asce, F. Understanding the
  Role of Evapotranspiration in Bioretention: Mesocosm Study. J. Sustain. Water Built *Environ.* 2015, 1 (2).
- 902 (36) Hess, A.; Asce, M.; Wadzuk, B.; Welker, A. Evapotranspiration in Rain Gardens Using
  903 Weighing Lysimeters. J. Irrig. Drain Eng 2017, 143 (6), 4017004.
- 904 (37) Nocco, M. A.; Rouse, S. E.; Balster, N. J. Vegetation type alters water and nitrogen
  905 budgets in a controlled, replicated experiment on residential-sized rain gardens planted
  906 with prairie, shrub, and turfgrass. *Urban Ecosyst.* 2016, *19* (4), 1665–1691.
- 907 (38) Orr, A. M. Transpiration performance in bioretention systems designed for semiarid
   908 climates, The University of Utah, 2013.
- McCarthy, H. R.; Pataki, D. E.; Jenerette, G. D. Plant water-use efficiency as a metric of urban ecosystem services. *Ecol. Appl.* 2011, 21 (8), 3115–3127.
- 911 (40) Johnston, M. R. Vegetation type alters rain garden hydrology through changes to soil
   912 porosity and evapotranspiration, University of Wisconsin Madison, 2011.
- (41) Hagishima, A.; Narita, K.; Tanimoto, J. Field experiment on transpiration from isolated urban plants. *Hydrol. Process.* 2007, *21* (9), 1217–1222.
- 915 (42) Schifman, L. A.; Tryby, M. E.; Berner, J.; Shuster, W. D. Managing Uncertainty in
  916 Runoff Estimation with the U.S. Environmental Protection Agency National Stormwater
  917 Calculator. JAWRA J. Am. Water Resour. Assoc. 2018, 54 (1), 148–159.

9	918 919 920	(43)	Kang, S.; Gu, B.; Du, T.; Zhang, J. Crop coefficient and ratio of transpiration to evapotranspiration of winter wheat and maize in a semi-humid region. <i>Agric. Water Manag.</i> <b>2003</b> , <i>59</i> (3), 239–254.
9	921 922 923	(44)	Daly, E.; Deletic, A.; Hatt, B. E.; Fletcher, T. D. Modelling of stormwater biofilters under random hydrologic variability: a case study of a car park at Monash University, Victoria (Australia). <i>Hydrol. Process.</i> <b>2012</b> , <i>26</i> (22), 3416–3424.
9	924 925 926	(45)	Schifman, L. A.; Kasaraneni, V. K.; Sullivan, R. K.; Oyanedel-Craver, V.; Boving, T. B. Bacteria Removal from Stormwater Runoff Using Tree Filters: A Comparison of a Conventional and an Innovative System. <i>Water</i> <b>2016</b> , <i>8</i> (3), 76.
	927 928	(46)	Pitt, R. National Stormwater Quality Database http://unix.eng.ua.edu/~rpitt/Research/ms4/mainms4.shtml (accessed Feb 6, 2018).
9	929 930 931	(47)	Kayhanian, M.; Fruchtman, B. D.; Gulliver, J. S.; Montanaro, C.; Ranieri, E.; Wuertz, S. Review of highway runoff characteristics: Comparative analysis and universal implications. <i>Water Res.</i> <b>2012</b> , <i>46</i> (20), 6609–6624.
9	932 933 934	(48)	Kumar, P. B. A. N.; Dushenkov, V.; Motto, H.; Raskin, I. Phytoextraction: The Use of Plants To Remove Heavy Metals from Soils. <i>Environ. Sci. Technol.</i> <b>1995</b> , <i>29</i> (5), 1232–1238.
9	935 936 937	(49)	Burken, J. G. Uptake and Metabolism of Organic Compounds: Green-Liver Model. In <i>Phytoremediation: Transformation and Control of Contaminants</i> ; McCutcheon, S. C., Schnoor, J. L., Eds.; John Wiley & Sons, Inc.: Hoboken, 2003; pp 59–84.
9	938 939 940	(50)	Kratky, H.; Li, Z.; Chen, Y.; Wang, C.; Li, X.; Yu, T. A critical literature review of bioretention research for stormwater management in cold climate and future research recommendations. <i>Front. Environ. Sci. Eng.</i> <b>2017</b> , <i>11</i> (4), 16.
9	941 942 943	(51)	Fu, Q.; Zhang, J.; Borchardt, D.; Schlenk, D.; Gan, J. Direct Conjugation of Emerging Contaminants in <i>Arabidopsis</i> : Indication for an Overlooked Risk in Plants? <i>Environ. Sci. Technol.</i> <b>2017</b> , <i>51</i> (11), 6071–6081.
9	944 945 946	(52)	LeFevre, G. H.; Müller, C. E.; Li, R. J.; Luthy, R. G.; Sattely, E. S. Rapid Phytotransformation of Benzotriazole Generates Synthetic Tryptophan and Auxin Analogs in Arabidopsis. <i>Environ. Sci. Technol.</i> <b>2015</b> , <i>49</i> (18), 10959–10968.
9	947 948 949	(53)	LeFevre, G. H.; Lipsky, A.; Hyland, K. C.; Blaine, A. C.; Higgins, C. P.; Luthy, R. G. Benzotriazole (BT) and BT plant metabolites in crops irrigated with recycled water. <i>Environ. Sci. Water Res. Technol.</i> <b>2017</b> , <i>3</i> (2), 213–223.
	950 951	(54)	Fu, Q.; Ye, Q.; Zhang, J.; Richards, J.; Borchardt, D.; Gan, J. Diclofenac in Arabidopsis cells: Rapid formation of conjugates. <i>Environ. Pollut.</i> <b>2017</b> , <i>222</i> , 383–392.
	952 953	(55)	<i>Phytoremediation: Transformation and Control of Contaminants</i> ; McCutcheon, S. C., Schnoor, J. L., Eds.; Wiley-Interscience: Hoboken, 2003.

- (56) LeFevre, G. H.; Hozalski, R. M.; Novak, P. J. The role of biodegradation in limiting the
  accumulation of petroleum hydrocarbons in raingarden soils. *Water Res.* 2012, 46 (20),
  6753–6762.
- (57) Chen, X.; Peltier, E.; Sturm, B. S. M.; Young, C. B. Nitrogen removal and nitrifying and denitrifying bacteria quantification in a stormwater bioretention system. *Water Res.* 2013, 47 (4), 1691–1700.
- Minett, D. A.; Cook, P. L. M.; Kessler, A. J.; Cavagnaro, T. R. Root effects on the spatial and temporal dynamics of oxygen in sand-based laboratory-scale constructed biofilters.
   *Ecol. Eng.* 2013, 58, 414–422.
- 963 (59) Lucas, W. C.; Greenway, M. Nutrient Retention in Vegetated and Nonvegetated
  964 Bioretention Mesocosms. J. Irrig. Drain. Eng. 2008, 134 (5), 613–623.
- (60) Bodelier, P. L. E.; Libochant, J. A.; Blom, C. W. P. M.; Laanbroek, H. J. Dynamics of
  Nitrification and Denitrification in Root- Oxygenated Sediments and Adaptation of
  Ammonia-Oxidizing Bacteria to Low-Oxygen or Anoxic Habitats. *Appl. Environ. Microbiol.* 1996, *62* (11), 4100–4107.
- 969 (61) Lynch, J. M.; Whipps, J. M. Substrate flow in the rhizosphere. *Plant Soil* 1990, *129* (1),
   970 1–10.
- 971 (62) Ulrich, B. A.; Vignola, M.; Edgehouse, K.; Werner, D.; Higgins, C. P. Organic Carbon
  972 Amendments for Enhanced Biological Attenuation of Trace Organic Contaminants in
  973 Biochar-Amended Stormwater Biofilters. *Environ. Sci. Technol.* 2017, *51* (16), 9184–
  974 9193.
- (63) Hsieh, C.-H.; Davis, A. P. Multiple-event study of bioretention for treatment of urban storm water runoff. *Water Sci. Technol.* 2005, *51* (3–4), 177–181.
- 977 (64) Alyaseri, I.; Zhou, J.; Morgan, S. M.; Bartlett, A. Initial impacts of rain gardens'
  978 application on water quality and quantity in combined sewer: field-scale experiment.
  979 *Front. Environ. Sci. Eng.* 2017, 11 (4), 19.
- (65) Sansalone, J. J.; Buchberger, S. G. Partitioning and first flush of metals in urban roadway storm water: EBSCOhost. *J. Environ. Eng.* 1997, *123* (2), 134–143.
- 982 (66) Fritioff, A.; Greger, M. Aquatic and Terrestrial Plant Species with Potential to Remove
  983 Heavy Metals from Stormwate. *Int. J. Phytoremediation* 2003, 5 (3), 211–224.
- 984 (67) David, N.; Leatherbarrow, J. E.; Yee, D.; McKee, L. J. Removal Efficiencies of a
  985 Bioretention System for Trace Metals, PCBs, PAHs, and Dioxins in a Semiarid
  986 Environment. J. Environ. Eng. 2015, 141 (6), 4014092.
- 987 (68) Barrett, M. E.; Limouzin, M.; Lawler, D. F. Effects of Media and Plant Selection on
  988 Biofiltration Performance. *J. Environ. Eng.* 2013, *139* (4), 462–470.

989 990 991	(69)	Henderson, C.; Greenway, M.; Phillips, I. Removal of dissolved nitrogen, phosphorus and carbon from stormwater by biofiltration mesocosms. <i>Water Sci. Technol.</i> <b>2007</b> , <i>55</i> (4), 183–191.
992 993 994	(70)	Zhang, Z.; Rengel, Z.; Liaghati, T.; Antoniette, T.; Meney, K. Influence of plant species and submerged zone with carbon addition on nutrient removal in stormwater biofilter. <i>Ecol. Eng.</i> <b>2011</b> , <i>37</i> (11), 1833–1841.
995 996 997	(71)	Houdeshel, C. D.; Hultine, K. R.; Johnson, N. C.; Pomeroy, C. A. Evaluation of three vegetation treatments in bioretention gardens in a semi-arid climate. <i>Landsc. Urban Plan.</i> <b>2015</b> , <i>135</i> , 62–72.
998 999 1000	(72)	Goh, H. W.; Zakaria, N. A.; Lau, T. L.; Foo, K. Y.; Chang, C. K.; Leow, C. S. Mesocosm study of enhanced bioretention media in treating nutrient rich stormwater for mixed development area. <i>Urban Water J.</i> <b>2017</b> , <i>14</i> (2), 134–142.
1001 1002 1003	(73)	Bratieres, K.; Fletcher, T. D.; Deletic, A.; Zinger, Y. Nutrient and sediment removal by stormwater biofilters: A large-scale design optimisation study. <i>Water Res.</i> <b>2008</b> , <i>42</i> (14), 3930–3940.
1004 1005 1006	(74)	Milandri, S. G.; Winter, K. J.; Chimphango, S. B. M.; Armitage, N. P.; Mbui, D. N.; Jackson, G. E.; Liebau, V. The performance of plant species in removing nutrients from stormwater in biofiltration systems in Cape Town. <i>Water SA</i> <b>2012</b> , <i>38</i> (5), 655–662.
1007 1008	(75)	Read, J.; Wevill, T.; Fletcher, T.; Deletic, A. Variation among plant species in pollutant removal from stormwater in biofiltration systems. <i>Water Res.</i> <b>2008</b> , <i>42</i> (4–5), 893–902.
1009 1010 1011	(76)	Szota, C.; Farrell, C.; Livesley, S. J.; Fletcher, T. D. Salt tolerant plants increase nitrogen removal from biofiltration systems affected by saline stormwater. <i>Water Res.</i> <b>2015</b> , <i>83</i> , 195–204.
1012 1013 1014	(77)	Rycewicz-Borecki, M.; McLean, J. E.; Dupont, R. R. Nitrogen and phosphorus mass balance, retention and uptake in six plant species grown in stormwater bioretention microcosms. <i>Ecol. Eng.</i> <b>2017</b> , <i>99</i> , 409–416.
1015 1016 1017	(78)	Passeport, E.; Hunt, W. F.; Line, D. E.; Smith, R. A.; Brown, R. A. Field Study of the Ability of Two Grassed Bioretention Cells to Reduce Storm-Water Runoff Pollution. <i>J. Irrig. Drain. Eng.</i> <b>2009</b> , <i>135</i> (4), 505–510.
1018 1019 1020	(79)	Locicero, R. C. R. Mainstreaming Green Infrastructure: The Nexus of Infrastructure and Education Using the Green Space Based Learning (GSBL) Approach for Bioretention Plant Selection, University of South Florida, 2015.
1021 1022	(80)	Fageria, N. K.; Baligar, V. C.; Clark, R. B. <i>Physiology of Crop Production</i> , 1st ed.; Food Products Press: Binghamton, New York, 2006.
1023 1024	(81)	Shuman, L. M. Mineral nutrition. In <i>Plant-environment interactions</i> ; Wilkinson, R. E., Ed.; New York and Basel, 2000; pp 65–109.

1025 1026 1027	(82)	Pilbeam, D. J.; Kirkby, E. A. Some aspects of the utilization of nitrate and ammonium by plants. In <i>Nitrogen metabolism of plants</i> ; Mengel, K., Pilbeam, D. J., Eds.; Clarendon Press and Oxford University Press: Oxford and New York, 1992; pp 55–70.
1028 1029	(83)	Morot-Gaudry, JF.; Lea, P. J. Introduction. In <i>Plant nitrogen</i> ; Morot-Gaudry, JF., Lea, P. J., Eds.; Springer: Berlin and New York, 2001; pp vi–xi.
1030 1031	(84)	Chapin, F. S.; Moilanen, L.; Kielland, K. Preferential use of organic nitrogen for growth by a non-mycorrhizal arctic sedge. <i>Nature</i> <b>1993</b> , <i>361</i> (6408), 150–153.
1032 1033 1034	(85)	Jones, D. L.; Darrah, P. R. Influx and efflux of amino acids from Zea mays L. roots and their implications for N nutrition and the rhizosphere. <i>Plant Soil</i> <b>1993</b> , <i>155–156</i> (1), 87–90.
1035 1036	(86)	Raab, T. K.; Lipson, D. A.; Monson, R. K. Soil amino acid utilization among species of the cyperaceae: plant and soil processes. <i>Ecology</i> <b>1999</b> , <i>80</i> (7), 2408–2419.
1037 1038	(87)	Cernusak, L. A.; Aranda, J.; Marshall, J. D.; Winter, K. Large variation in whole-plant water-use efficiency among tropical tree species. <i>New Phytol.</i> <b>2007</b> , <i>173</i> (2), 294–305.
1039 1040 1041	(88)	Knapp, P. A.; Soulé, P. T. Increasing water-use efficiency and age-specific growth responses of old-growth ponderosa pine trees in the Northern Rockies. <i>Glob. Chang. Biol.</i> <b>2011</b> , <i>17</i> (1), 631–641.
1042 1043	(89)	Davis, A. P.; Traver, R. G.; Hunt, W. F. Improving Urban Stormwater Quality: Applying Fundamental Principles. <i>J. Contemp. Water Res. Educ.</i> <b>2010</b> , <i>146</i> (1), 3–10.
1044 1045	(90)	Mangangka, I. R.; Liu, A.; Egodawatta, P.; Goonetilleke, A. Performance characterisation of a stormwater treatment bioretention basin. <i>J. Environ. Manage.</i> <b>2015</b> , <i>150</i> , 173–178.
1046 1047 1048	(91)	Hsieh, C.; Davis, A.; Needelman, B. A. Nitrogen Removal from Urban Stormwater Runoff Through Layered Bioretention Columns - ProQuest. <i>Water Environ. Res.</i> 2007, No. 79.12, 2404–2411.
1049 1050	(92)	Li, L.; Davis, A. P. Urban Stormwater Runoff Nitrogen Composition and Fate in Bioretention Systems. <i>Environ. Sci. Technol.</i> <b>2014</b> , <i>48</i> (6), 3403–3410.
1051 1052	(93)	Davis, A. P.; Shokouhian, M.; Sharma, H.; Minami, C. Laboratory Study of Biological Retention for Urban Stormwater Management. <i>Water Environ. Res.</i> <b>2001</b> , <i>73</i> (1), 5–14.
1053 1054 1055	(94)	Payne, E. G. I.; Pham, T.; Cook, P. L. M.; Deletic, A.; Hatt, B. E.; Fletcher, T. D. Inside Story of Gas Processes within Stormwater Biofilters: Does Greenhouse Gas Production Tarnish the Benefits of Nitrogen Removal? <i>Environ. Sci. Technol.</i> <b>2017</b> , acs.est.6b05653.
1056 1057 1058	(95)	Mullane, J. M.; Flury, M.; Iqbal, H.; Freeze, P. M.; Hinman, C.; Cogger, C. G.; Shi, Z. Intermittent rainstorms cause pulses of nitrogen, phosphorus, and copper in leachate from compost in bioretention systems. <i>Sci. Total Environ.</i> <b>2015</b> , <i>537</i> , 294–303.

1059 (96) 1060	Li, J.; Davis, A. P. A unified look at phosphorus treatment using bioretention. <i>Water Res.</i> <b>2016</b> , <i>90</i> , 141–155.
1061 (97) 1062 1063	Fowdar, H. S.; Hatt, B. E.; Cresswell, T.; Harrison, J. J.; Cook, P. L; Deletic, A. Phosphorus Fate and Dynamics in Greywater Biofiltration Systems. <i>Environ. Sci. Technol.</i> <b>2017</b> , <i>51</i> (4), 2280–2287.
1064 (98) 1065	Schachtman, D. P.; Reid, R. J.; Ayling, S. M. Phosphorus Uptake by Plants: From Soil to Cell. <i>Plant Physiology</i> . American Society of Plant Biologists (ASPB) pp 447–453.
1066 (99) 1067	Fitter, A. H.; Hay, R. K. M. <i>Environmental Physiology of Plants</i> , 3rd ed.; Academic Press: London and San Diego, 2002.
1068 (100) 1069	Van Tichelen, K. K.; Colpaert, J. V. Kinetics of phosphate absorption by mycorrhizal and non-mycorrhizal Scots pine seedlings. <i>Physiol. Plant.</i> <b>2000</b> , <i>110</i> (1), 96–103.
1070 (101) 1071	Bolan, N. S. A critical review on the role of mycorrhizal fungi in the uptake of phosphorus by plants. <i>Plant Soil</i> <b>1991</b> , <i>134</i> (2), 189–207.
1072 (102) 1073	Winfrey, B. K.; Hatt, B. E.; Ambrose, R. F. Arbuscular mycorrhizal fungi in Australian stormwater biofilters. <i>Ecol. Eng.</i> <b>2017</b> , <i>102</i> , 483–489.
1074 (103) 1075 1076	Hsieh, C.; Davis, A. P.; Needelman, B. A. Bioretention Column Studies of Phosphorus Removal from Urban Stormwater Runoff. <i>Water Environment Research</i> . Water Environment Federation 2007, pp 177–184.
1077 (104) 1078	Janke, B. D.; Finlay, J. C.; Hobbie, S. E. Trees and Streets as Drivers of Urban Stormwater Nutrient Pollution. <i>Environ. Sci. Technol.</i> <b>2017</b> , <i>51</i> (17), 9569–9579.
1079 (105) 1080	Quinn, R.; Dussaillant J., A. Modeling Heavy Metal Behavior in Sustainable Drainage Systems: A Case Study. <i>CLEAN - Soil, Air, Water</i> <b>2014</b> , <i>42</i> (2), 160–168.
1081 (106) 1082	Muthanna, T. M.; Viklander, M.; Gjesdahl, N.; Thorolfsson, S. T. Heavy Metal Removal in Cold Climate Bioretention. <i>Water. Air. Soil Pollut.</i> <b>2007</b> , <i>183</i> (1–4), 391–402.
1083 (107) 1084 1085	Rycewicz-Borecki, M.; McLean, J. E.; Dupont, R. R. Bioaccumulation of copper, lead, and zinc in six macrophyte species grown in simulated stormwater bioretention systems. <i>J. Environ. Manage.</i> <b>2016</b> , <i>166</i> , 267–275.
1086 (108) 1087 1088	Feng, W.; Hatt, B. E.; McCarthy, D. T.; Fletcher, T. D.; Deletic, A. Biofilters for Stormwater Harvesting: Understanding the Treatment Performance of Key Metals That Pose a Risk for Water Use. <i>Environ. Sci. Technol.</i> <b>2012</b> , <i>46</i> (9), 5100–5108.
1089 (109) 1090 1091	Tangahu, B. V.; Sheikh Abdullah, S. R.; Basri, H.; Idris, M.; Anuar, N.; Mukhlisin, M. A Review on Heavy Metals (As, Pb, and Hg) Uptake by Plants through Phytoremediation. <i>Int. J. Chem. Eng.</i> <b>2011</b> , <i>2011</i> , 1–31.

1092 (110) Sun, X.; Davis, A. P. Heavy metal fates in laboratory bioretention systems. *Chemosphere* 

- **2007**, *66* (9), 1601–1609.
- 1094 (111) Read, J.; Fletcher, T. D.; Wevill, T.; Deletic, A. Plant Traits That Enhance Pollutant
   1095 Removal From Stormwater in Biofiltration Systems. *Int. J. Phytoremediation* 2010, *12*,
   1096 34–53.
- 1097 (112) Rascio, N.; Navari-Izzo, F. Heavy metal hyperaccumulating plants: How and why do they
   1098 do it? And what makes them so interesting? *Plant Sci.* 2011, *180* (2), 169–181.
- (113) Cunningham, S. D.; Berti, W. R.; Huang, J. W. Phytoremediation of contaminated soils.
   *Trends Biotechnol.* 1995, *13* (9), 393–397.
- (114) Ebbs, S. D.; Lasat, M. M.; Brady, D. J.; Cornish, J.; Gordon, R.; Kochian, L. V.
  Phytoextraction of Cadmium and Zinc from a Contaminated Soil. *J. Environ. Qual.* 1997, 26 (5), 1424–1430.
- (115) Bolan, N. S.; Park, J. H.; Robinson, B.; Naidu, R.; Huh, K. Y. Chapter four Phytostabilization: A Green Approach to Contaminant Containment. In *Advances in Agronomy*; 2011; Vol. 112, pp 145–204.
- (116) Laurenson, G.; Laurenson, S.; Bolan, N.; Beecham, S.; Clark, I. Chapter Four The Role
  of Bioretention Systems in the Treatment of Stormwater. In *Advances in Agronomy*; 2013;
  Vol. 120, pp 223–274.
- (117) Blecken, G.-T.; Zinger, Y.; Deletić, A.; Fletcher, T. D.; Viklander, M. Impact of a
  submerged zone and a carbon source on heavy metal removal in stormwater biofilters. *Ecol. Eng.* 2009, *35* (5), 769–778.
- (118) United States Environmental Protection Agency Office of Research and Development.
   *Introduction to Phytoremediation*; Washington, 2000.
- (119) DiBlasi, C. J.; Li, H.; Davis, A. P.; Ghosh, U. Removal and Fate of Polycyclic Aromatic
   Hydrocarbon Pollutants in an Urban Stormwater Bioretention Facility. *Environ. Sci. Technol.* 2009, 43 (2), 494–502.
- (120) LeFevre, G. H.; Novak, P. J.; Hozalski, R. M. Fate of Naphthalene in Laboratory-Scale
   Bioretention Cells: Implications for Sustainable Stormwater Management. *Environ. Sci. Technol.* 2012, 46 (2), 995–1002.
- (121) Hong, E.; Seagren, E. A.; Davis, A. P. Sustainable Oil and Grease Removal from
   Synthetic Stormwater Runoff Using Bench-Scale Bioretention Studies. *Water Environ. Res.* 2006, 78 (2), 141–155.
- (122) Marquès, M.; Mari, M.; Sierra, J.; Nadal, M.; Domingo, J. L. Solar radiation as a swift
  pathway for PAH photodegradation: A field study. *Sci. Total Environ.* 2017, *581–582*,
  530–540.
- 1127 (123) de Jesus, J. H. F.; da C. Cunha, G.; Cardoso, E. M. C.; Mangrich, A. S.; Romão, L. P. C.

1128 1129		Evaluation of waste biomasses and their biochars for removal of polycyclic aromatic hydrocarbons. <i>J. Environ. Manage.</i> <b>2017</b> , <i>200</i> , 186–195.
1130 1131 1132	(124)	LeFevre, G. H.; Hozalski, R. M.; Novak, P. J. Root Exudate Enhanced Contaminant Desorption: An Abiotic Contribution to the Rhizosphere Effect. <i>Environ. Sci. Technol.</i> <b>2013</b> , <i>47</i> (20), 11545–11553.
1133 1134 1135	(125)	Duan, L.; Naidu, R.; Thavamani, P.; Meaklim, J.; Megharaj, M. Managing long-term polycyclic aromatic hydrocarbon contaminated soils: a risk-based approach. <i>Environ. Sci. Pollut. Res. Int.</i> <b>2015</b> , <i>22</i> (12), 8927–8941.
1136 1137 1138	(126)	Zhang, L.; Seagren, E. A.; Davis, A. P.; Karns, J. S. Long-Term Sustainability of <i>Escherichia Coli</i> Removal in Conventional Bioretention Media. <i>J. Environ. Eng.</i> <b>2011</b> , <i>137</i> (8), 669–677.
1139 1140 1141	(127)	Chandrasena, G. I.; Pham, T.; Payne, E. G.; Deletic, A.; McCarthy, D. T. E. coli removal in laboratory scale stormwater biofilters: Influence of vegetation and submerged zone. <i>J. Hydrol.</i> <b>2014</b> , <i>519</i> , 814–822.
1142 1143 1144	(128)	Parker, E.; Rippy, M. A.; Mehring, A.; Winfrey, B.; Ambrose, R. F.; Levin, L. A.; Grant, S. B. The Predictive Power of Clean Bed Filtration Theory for Fecal Indicator Bacteria Removal in Stormwater Biofilters. <i>Environ. Sci. Technol.</i> <b>2017</b> , <i>51</i> (10), 5703–5712.
1145 1146	(129)	Li, H.; Davis, A. P. Water Quality Improvement through Reductions of Pollutant Loads Using Bioretention. <i>J. Environ. Eng.</i> <b>2009</b> , <i>135</i> (8), 567–576.
1147 1148 1149	(130)	Li, Y. L.; Deletic, A.; Alcazar, L.; Bratieres, K.; Fletcher, T. D.; McCarthy, D. T. Removal of Clostridium perfringens, Escherichia coli and F-RNA coliphages by stormwater biofilters. <i>Ecol. Eng.</i> <b>2012</b> , <i>49</i> , 137–145.
1150 1151 1152	(131)	Kazemi, F.; Beecham, S.; Gibbs, J.; Clay, R. Factors affecting terrestrial invertebrate diversity in bioretention basins in an Australian urban environment. <i>Landsc. Urban Plan.</i> <b>2009</b> , <i>92</i> (3–4), 304–313.
1153 1154	(132)	Kazemi, F.; Beecham, S.; Gibbs, J. Streetscape biodiversity and the role of bioretention swales in an Australian urban environment. <i>Landsc. Urban Plan.</i> <b>2011</b> , <i>101</i> (2), 139–148.
1155 1156	(133)	Kazemi, F.; Beecham, S.; Gibbs, J. Streetscale bioretention basins in Melbourne and their effect on local biodiversity. <i>Ecol. Eng.</i> <b>2009</b> , <i>35</i> (10), 1454–1465.
1157 1158	(134)	Kalisvaart, B. F. Re-use of wastewater: Preventing the recovery of pathogens by using medium-pressure UV lamp technology. <i>Water Sci. Technol.</i> <b>2004</b> , <i>50</i> (6), 337–344.
1159 1160 1161	(135)	Hathaway, J. M.; Hunt, W. F.; Jadlocki, S. Indicator Bacteria Removal in Storm-Water Best Management Practices in Charlotte, North Carolina. <i>J. Environ. Eng.</i> <b>2009</b> , <i>135</i> (12), 1275–1285.
1162	(136)	Haichar, F. el Z.; Santaella, C.; Heulin, T.; Achouak, W. Root exudates mediated

1163		interactions belowground. Soil Biol. Biochem. 2014, 77, 69-80.
1164 1165 1166	(137)	Ibáñez, M.; Gracia-Lor, E.; Bijlsma, L.; Morales, E.; Pastor, L.; Hernández, F. Removal of emerging contaminants in sewage water subjected to advanced oxidation with ozone. <i>J. Hazard. Mater.</i> <b>2013</b> , <i>260</i> , 389–398.
1167 1168 1169	(138)	Ulrich, B. A.; Loehnert, M.; Higgins, C. P. Improved contaminant removal in vegetated stormwater biofilters amended with biochar. <i>Environ. Sci. Water Res. Technol</i> <b>2017</b> , <i>3</i> , 726–734.
1170 1171 1172	(139)	LeFevre, G. H.; Portmann, A. C.; Müller, C. E.; Sattely, E. S.; Luthy, R. G. Plant Assimilation Kinetics and Metabolism of 2-Mercaptobenzothiazole Tire Rubber Vulcanizers by Arabidopsis. <i>Environ. Sci. Technol.</i> <b>2016</b> , <i>50</i> (13), 6762–6771.
1173 1174 1175 1176	(140)	Simon-Delso, N.; Amaral-Rogers, V.; Belzunces, L. P.; Bonmatin, J. M.; Chagnon, M.; Downs, C.; Furlan, L.; Gibbons, D. W.; Giorio, C.; Girolami, V.; et al. Systemic insecticides (neonicotinoids and fipronil): trends, uses, mode of action and metabolites. <i>Environ. Sci. Pollut. Res.</i> <b>2015</b> , <i>22</i> (1), 5–34.
1177 1178 1179 1180	(141)	Center for Watershed Protection; Maryland Department of the Environment. Maryland Stormwater Design Manual http://mde.maryland.gov/programs/water/StormwaterManagementProgram/Pages/stormw ater_design.aspx.
1181 1182	(142)	Department of Environmental Resources of Prince George's County Maryland. Prince George's County, Maryland, Bioretention Manual; 2007.
1183 1184 1185	(143)	Li, MH.; Sung, C.; Kim, M.; Chu, KH. Assessing Performance of Bioretention Boxes in Hot and Semiarid Regions. <i>Transp. Res. Rec. J. Transp. Res. Board</i> <b>2011</b> , <i>2262</i> , 155–163.
1186 1187 1188	(144)	Dobbie, M. F. Designing raingardens for community acceptance https://watersensitivecities.org.au/wp-content/uploads/2016/06/TMR_A4- 1_2016_Designing_raingardens_web.pdf (accessed Jun 14, 2017).
1189 1190	(145)	Ries, L.; Debinski, D. M.; Wieland, M. L. Conservation Value of Roadside Prairie Restoration to Butterfly Communities. <i>Conserv. Biol.</i> <b>2001</b> , <i>15</i> (2), 401–411.
1191 1192	(146)	Fiedler, A. K.; Landis, D. A. Attractiveness of Michigan Native Plants to Arthropod Natural Enemies and Herbivores. <i>Environ. Entomol.</i> <b>2007</b> , <i>36</i> (4), 751–765.
1193 1194 1195	(147)	Ayers, E. M.; Kangas, P. Topsoil Development in Bioretention Cells: What Are the Implications? In <i>Low Impact Development Technology: Design Methods and Case Studies</i> ; American Society of Civil Engineers: Reston, VA, 2015; pp 16–25.
1196 1197	(148)	Mohan, G. Bee-harming pesticides are declining at plant nurseries, report shows. <i>Los Angeles Times</i> . August 17, 2016.

1198 1199	(149)	Food and Agriculture Organization of the United Nations. World Fertilizer Trends and Outlook to 2018 http://www.fao.org/3/a-i4324e.pdf (accessed Aug 30, 2017).
1200 1201 1202	(150)	Kugler, K.; Ohs, B.; Scholz, M.; Wessling, M. Towards a carbon independent and CO2- free electrochemical membrane process for NH3 synthesis. <i>Phys. Chem. Chem. Phys.</i> <b>2014</b> , <i>16</i> (13), 6129–6138.
1203 1204 1205	(151)	Richards, P. J.; Farrell, C.; Tom, M.; Williams, N. S. G.; Fletcher, T. D. Vegetable raingardens can produce food and reduce stormwater runoff. <i>Urban For. Urban Green.</i> <b>2015</b> .
1206 1207 1208	(152)	Calderón-Preciado, D.; Matamoros, V.; Bayona, J. M. Occurrence and potential crop uptake of emerging contaminants and related compounds in an agricultural irrigation network. <i>Sci. Total Environ.</i> <b>2011</b> , <i>412–413</i> , 14–19.
1209 1210 1211	(153)	T. L. Culbertson; S.L. Hutchinson. Assessing Bioretention Cell Function in a Midwest Continental Climate. In <i>2004, Ottawa, Canada August 1 - 4, 2004</i> ; American Society of Agricultural and Biological Engineers: St. Joseph, MI, 2004.
1212 1213 1214	(154)	U.S. Energy Information Administration. How much electricity does an American home use? - FAQ - U.S. Energy Information Administration (EIA) https://www.eia.gov/tools/faqs/faq.php?id=97&t=3 (accessed Nov 8, 2017).
1215 1216	(155)	Jones, M. P.; Hunt, W. F. Bioretention Impact on Runoff Temperature in Trout Sensitive Waters. <i>J. Environ. Eng.</i> <b>2009</b> , <i>135</i> (8), 577–585.
1217 1218 1219	(156)	Zahmatkesh, Z.; Burian, S. J.; Karamouz, M.; Tavakol-Davani, H.; Goharian, E. Low- Impact Development Practices to Mitigate Climate Change Effects on Urban Stormwater Runoff: Case Study of New York City. <i>J. Irrig. Drain. Eng.</i> <b>2015</b> , <i>141</i> (1), 4014043.
1220 1221 1222	(157)	Philadelphia Water Department. Regional and Global Trends   Philadelphia Water Department http://phillywatersheds.org/watershed_issues/regional_and_global_trends (accessed Oct 23, 2017).
1223 1224	(158)	Xu, Z.; Wu, M.; He, Y. Toluene Biofiltration Enhanced by Ryegrass. <i>Bull. Environ. Contam. Toxicol.</i> <b>2013</b> , <i>90</i> (6), 646–649.
1225 1226	(159)	Gao, Y.; Zhu, L. Plant uptake, accumulation and translocation of phenanthrene and pyrene in soils. <i>Chemosphere</i> <b>2004</b> , <i>55</i> (9), 1169–1178.
1227 1228 1229	(160)	Hunt, W. F.; Lord, B.; Loh, B.; Sia, A. <i>Plant Selection for Bioretention Systems and Stormwater Treatment Practices</i> ; SpringerBriefs in Water Science and Technology; Springer Singapore: Singapore, 2015.
1230 1231	(161)	Shaw, D.; Schmidt, R. <i>Plants for stormwater design</i>   <i>Minnesota Pollution Control Agency</i> ; Minnesota Pollution Control Agency: St. Paul, MN, 2003.
1232	(162)	Turk, R. P.; Kraus, H. T.; Hunt, W. F.; Carmen, N. B.; Bilderback, T. E. Nutrient

1233 1234		Sequestration by Vegetation in Bioretention Cells Receiving High Nutrient Loads. J. Environ. Eng. 2017, 143 (2), 6016009.
1235 1236	(163)	Kacálková, L.; Tlustoš, P. The uptake of persistent organic pollutants by plants. <i>Open Life Sci.</i> <b>2011</b> , <i>6</i> (2), 223–235.
1237 1238	(164)	Geronimo, F. K. F.; Maniquiz-Redillas, M. C.; Kim, L. H. Fate and removal of nutrients in bioretention systems. <i>Desalin. Water Treat.</i> <b>2015</b> , <i>53</i> (11), 3072–3079.
1239 1240 1241	(165)	Iowa Department of Transportation. Roadside Vegetation Q&A https://www.iowadot.gov/pdf_files/roadside_vegetation_q_and_a.pdf (accessed Aug 31, 2017).
1242 1243 1244	(166)	Kaczmarek-Derda, W.; Folkestad, J.; Helgheim, M.; Netland, J.; Solhaug, K. A.; Brandsaeter, L. O. Influence of cutting time and stubble height on regrowth capacity of Juncus effusus and Juncus conglomeratus. <i>Weed Res.</i> <b>2014</b> , <i>54</i> , 603–613.
1245 1246	(167)	Fowdar, H. S.; Hatt, B. E.; Breen, P.; Cook, P. L. M.; Deletic, A. Designing living walls for greywater treatment. <i>Water Res.</i> <b>2017</b> , <i>110</i> , 218–232.
1247		